1 Carbon limitation leads to thermodynamic regulation of aerobic metabolism

- 2 Vanessa A. Garayburu-Caruso¹, James C. Stegen¹, Hyun-Seob Song^{1,2}, Lupita Renteria¹,
- 3 Jaqueline Wells^{1,3}, Whitney Garcia¹, Charles T. Resch¹, Amy Goldman¹, Rosalie Chu⁴, Jason
- 4 Toyoda⁴, Emily B. Graham^{1*}
- 5
- ⁶ ¹Pacific Northwest National Laboratory, Richland WA 99352, USA
- 7 ²University of Nebraska-Lincoln, Lincoln NE 68588, USA
- ³Oregon State University, Corvallis, OR 97331, USA
- 9 ⁴ Environmental Molecular Sciences Laboratory, Richland WA 99352, USA
- 10 *Correspondence: emily.graham@pnnl.gov; 509-372-6049
- 11

13 Abstract

14 Organic matter (OM) metabolism in freshwater ecosystems is a critical source of uncertainty in 15 global biogeochemical cycles, yet aquatic OM cycling remains poorly understood. Here, we 16 present the first work to explicitly test OM thermodynamics as a key regulator of aerobic 17 respiration, challenging long-held beliefs that organic carbon and oxygen concentrations are the 18 primary determinants of respiration rates. We pair controlled microcosm experiments with 19 ultrahigh-resolution OM characterization to demonstrate a clear relationship between OM 20 thermodynamic favorability and aerobic respiration under carbon limitation. We also 21 demonstrate a shift in the regulation of aerobic respiration from OM thermodynamics to nitrogen 22 content when carbon is in excess, highlighting a central role for OM thermodynamics in aquatic 23 biogeochemical cycling particularly in carbon-limited ecosystems. Our work therefore 24 illuminates a structural gap in aquatic biogeochemical models and presents a new paradigm in 25 which OM thermodynamics and nitrogen content interactively govern aerobic respiration.

26

28	Metabolism of organic matter (OM) in freshwater ecosystems plays a large role in global
29	biogeochemical cycles ¹⁻³ , as freshwater ecosystems emit more than 2 Pg C yr ⁻¹ into the
30	atmosphere ^{4,5} . These emissions are largely dominated by contributions from river corridors ^{1,5,6} ,
31	and within the river corridor, areas of groundwater-surface water mixing (hyporheic zones) have
32	a disproportionate impact on aerobic respiration ⁷⁻⁹ . Recent field observations have suggested that
33	OM chemistry, and in particular OM thermodynamics, are key to predicting aerobic respiration
34	in hyporheic zones ¹⁰⁻¹² . If supported, these observations challenge a widespread paradigm that
35	organic carbon and oxygen concentrations are the primary determinants of aerobic respiration
36	rates and highlight a key source of model uncertainty. Yet, no work has provided direct evidence
37	for OM thermodynamics as a regulator of aerobic respiration in a controlled laboratory
38	environment. Demonstrating this behavior would identify mechanisms that drive field-based
39	phenomena and would enable key properties of OM to be represented in predictive models,
40	thereby contributing to reducing the uncertainty in modeling river corridor biogeochemical
41	cycling ^{13,14} .
42	We use highly controlled aerobic microcosms, non-invasive dissolved oxygen consumption
43	rates, and ultrahigh-resolution OM characterization to investigate the role of OM chemistry in
44	determining aerobic respiration in hyporheic zone sediments. Based on field observations ¹⁰⁻¹² ,
45	we hypothesized that OM chemistry, including thermodynamic favorability and nitrogen (N)
46	content, would regulate aerobic respiration. Historically, investigations of thermodynamic

47 constraints on microbial metabolism have primarily focused on oxidation-reduction reactions

48 that are controlled by the availability of various terminal electron acceptors (e.g., oxygen, nitrate,

49 sulfate)¹⁵⁻¹⁷. Theory indicates that respiration under aerobic conditions is governed by rate

50 kinetics, while thermodynamic regulation has no influence. This expectation is based on the

51 premise that using oxygen as the terminal electron acceptor provides sufficient energy for ATP generation regardless of thermodynamic properties of the electron donor^{18,19}. Consequently, the 52 53 role of OM thermodynamics in microbial metabolism has been mainly explored under anaerobic conditions²⁰⁻²³. However, OM chemistry has recently emerged as a possible regulator of aerobic 54 55 metabolism based on correlative field-based observation. These studies have suggested that OM 56 thermodynamics interact with N content and carbon concentration to influence aerobic respiration¹⁰⁻¹². Here, we present the first work to explicitly test OM thermodynamics as a key 57 58 regulator of aerobic respiration and demonstrate a clear relationship between OM 59 thermodynamic favorability and aerobic respiration under carbon limitation, thus challenging long-held beliefs that respiration rates are govern solely by kinetics. 60

61 Thermodynamic regulation of aerobic respiration

62 To test the role OM chemistry as a key regulator of aerobic metabolism, we incubated sediments 63 with thermodynamically distinct N-bearing or N-free OM treatments at concentrations commonly observed in freshwater systems (from 0.3 to 9 mg C L⁻¹, Supplementary Table 64 1)^{12,24,25}. We inferred carbon limitation at low treatment concentrations ($< 3 \text{ mg C L}^{-1}$) because 65 sediments were collected from a low carbon area (%C < 0.6 as previously discussed in Graham 66 et al.¹¹) with observed C:N $< 5^{11}$ which is typically associated with carbon limitation^{26,27}. In 67 68 addition, sediments were pre-processed prior to incubation until dissolved organic carbon 69 concentrations were below instrument detection (see Methods and Supplementary Fig. 1). Moreover, aerobic respiration increased with treatment concentration from 0.3 to 3 mg C L⁻¹ and 70 stabilized between 3 and 9 mg C L⁻¹, indicating that additional carbon only stimulated respiration 71 when 3 mg C L⁻¹ or less was amended (Fig. 1a, p = 0.005 & p = 0.43). 72

74 We provide direct evidence for OM thermodynamics in controlling aerobic respiration by 75 demonstrating that thermodynamically-favorable OM supported enhanced respiration in 76 microcosms under carbon limitation. We use the mean Gibbs free energy of the half reaction of organic carbon oxidation under standard conditions (ΔG°_{Cox}) as a proxy for thermodynamic 77 favorability throughout this paper, as per LaRowe and Van Cappellen²⁸, Graham et al.¹¹, and 78 Stegen et al.¹². Aerobic respiration was negatively correlated with ΔG°_{Cox} when OM was 79 amended at low concentrations (Fig. 1b-c 0.3 mg C $L^{-1} R^2 = 0.23 p = 0.03$, 3 mg C $L^{-1} R^2 = 0.34$ 80 81 p = 0.007).

82

While our results support previous field observations that emphasize the role of OM thermodynamics in predicting aerobic respiration 11,12 , we uniquely underscore the central role for 83 84 OM thermodynamics in the metabolism of carbon-limited ecosystems using highly controlled laboratory experiments. A field study by Graham et al.¹¹ showed that thermodynamically 85 86 favorable OM was preferentially metabolized in sediments across a vegetation gradient. Similarly, Stegen et al.¹² highlighted the importance of OM thermodynamics in regulating 87 88 aerobic respiration within hyporheic zones. While the previous works are based on correlative 89 observations, our controlled laboratory microcosms allowed us to clearly demonstrate 90 thermodynamic regulation of aerobic metabolism and define the conditions under which OM 91 thermodynamics provide an avenue for improving aquatic biogeochemical models. 92 Specifically, we reveal a dependency of thermodynamic OM regulation on organic carbon 93 concentration, as sediments putatively not experiencing carbon limitation had no evidence of thermodynamic constraints on aerobic respiration (Fig. 1d, 9 mg C $L^{-1} p = 0.71$). Low molecular 94

- 95 weight carbon compounds, such as the amino and organic acids used for treatments in this
- 96 experiment (see Methods), are highly bioavailable to microorganisms which have direct uptake

pathways for these molecules²⁹⁻³¹ in contrast to polymeric OM in sediments that requires the
production of extracellular enzymes^{32,33}. Thus, we hypothesize that excess low molecular weight
and highly bioavailable OM diminishes the benefits of thermodynamic preference in metabolism,
because the low microbial cost of direct uptake outweighs the energy benefit gained from
selective carbon metabolism.

102 Pathways of OM metabolism vary with thermodynamic control of aerobic respiration

103 We underscore a need for improved model structures for predicting aerobic respiration, as we 104 observed different metabolic processes between carbon-limited and carbon-replete environments. 105 To investigate metabolic processes involved in aerobic respiration, we calculated inferred 106 biochemical transformations in ultrahigh-resolution OM profiles following procedures described previously^{10-12,34-36}. This method relies on the mass accuracy of Fourier transform ion cyclotron 107 108 resonance mass spectrometry (FTICR-MS) and produces a count of the number of times a 109 specific molecule (e.g., glucose, valine, glutamine, etc.) is putatively gained or lost in reactions 110 (see Methods).

111 While previous work has suggested both thermodynamic and N-related regulation of aerobic respiration¹⁰⁻¹², we postulate a carbon limitation threshold beyond which thermodynamic 112 113 controls do not persist. Our data suggest that beyond this threshold (i.e., in the absence of carbon 114 limitation), aerobic respiration is coupled to organic N metabolism. We found no differences in 115 OM chemistry (Supplementary Fig. 2a-b, all p > 0.05), respiration rates (p = 0.11 at 0.3 mg C 116 L^{-1} and p = 0.24 at 3 mg C L^{-1}) or biochemical transformations (Fig. 2a-b, 0.3 mg C L^{-1} p =117 0.67, 3 mg C L⁻¹ p = 0.91) across microcosms with N-bearing vs. N-free OM amended when 118 respiration was thermodynamically-controlled (i.e., in carbon limited conditions). However, 119 respiration in carbon-replete microcosms (9 mg C L^{-1}) increased with the addition of organic N

120	relative to N-free OM, and pathways of OM metabolism varied between treatments with or
121	without added organic N (Fig. 2c-d, $p = 0.04$, $p = 0.04$). Biochemical transformations in
122	microcosms receiving 9 mg C L^{-1} more frequently involved N when organic N was added to
123	microcosms (Supplementary Fig. 3, $p = 0.02$). Additionally, we provide evidence that N-
124	enriched molecules are preferably consumed in natural environments with surplus carbon.
125	Bioavailable organic N addition increased the relative abundance of more complex protein-like
126	OM, suggesting the preservation of sediment-bound OM containing N in the presence of more
127	accessible N sources (Supplementary Fig. 2c, $p < 0.01$). Together, these results are consistent
128	with N-mining observed previously in this system, whereby OM is oxidized for microbial
129	acquisition of N ^{10,37} .
130	More broadly, we highlight that when organic carbon is in excess, there is a dependency of
131	aerobic respiration on specific nutrient limitations rather than on OM thermodynamics.
132	Therefore, OM thermodynamics appear to be most informative of aquatic biogeochemistry
133	within carbon-limited ecosystems, while N availability governs aerobic respiration at higher
134	carbon to nitrogen ratios (C:N). While N-dependent aerobic respiration at high C:N is consistent
135	with nutrient limitations observed in variety of systems ^{38,39} , thermodynamic regulation at low
136	C:N ratios challenges the widespread notion that organic carbon and oxygen concentrations are
137	the main variables driving respiration.

138 Enhancing model predictions with a new paradigm of aquatic biogeochemical cycling

While many processed-based models represent biogeochemical cycles, most model structures contain only a few lumped carbon pools that do not fully represent the complexity of natural OM sources^{20,21}. Organic matter cycling is typically modeled following Michaelis-Menten kinetics, with OM separated into particulate or dissolved pools⁴⁰. In some cases, these pools are further 143 categorized by environmental properties or bioavailability, and each subpool is assigned a fixed mineralization rate^{20,40-42}. These traditional approaches do not address OM chemistry because 144 145 commonly used bulk characterization techniques do not provide sufficient molecular detail, and 146 because representing individual OM molecules in a given ecosystem is computationally 147 unfeasible. Our work demonstrates that processes associated with OM thermodynamics and N 148 content strongly influence aerobic respiration—and, in turn, biogeochemical reaction networks— 149 thus providing a link between OM chemistry and biogeochemical rates that will help inform and 150 parameterize new and existing models.

151 We therefore propose a new conceptual model with direct avenues for model incorporation in 152 which OM thermodynamics regulates aerobic respiration until organic carbon concentrations are 153 sufficient to induce nutrient limitations (Fig. 3). We suggest that a combination of organic carbon 154 concentration and thermodynamic limitation governs aerobic respiration in sediments with low 155 carbon to nutrient ratios. In contrast, nutrient limitation regulates respiration when bioavailable 156 carbon is in excess, and in this scenario, organic N may be a key constraint on respiration. This is 157 consistent with previous reports of a strong role from organic N cycling in hyporheic zones and freshwater systems^{10,12,43,44}. While our work represents a single system, our conceptual model is 158 159 meant to provide proof-of-concept for more spatially extensive studies that will allow broader 160 transferability.

To our knowledge, this is the first work to provide direct evidence for OM thermodynamics as a regulator of aerobic respiration in a controlled laboratory environment and highlights a key gap in current mechanistic understanding of OM cycling. In order to improve model accuracy, we reveal a need to explicitly represent the interactions between OM thermodynamics, nutrient limitations, and organic carbon concentration in process-based models of aquatic carbon cycling.

Our results constrain the environments under which specific chemical attributes of OM are valuable for rate predictions and provide guidance on the data types that are needed to accurately represent hyporheic zone biogeochemistry. These processes cannot be represented by lumped OM pools informed by decades of coarse measurements, and thus we highlight the utility of high-resolution technologies in models of aquatic biogeochemistry and present a new paradigm in which aerobic respiration is governed by a combination of carbon concentration, OM thermodynamics, and nutrient limitations.

173 Acknowledgements

- 174 This research was supported by the U.S. Department of Energy (DOE), Office of Biological and
- 175 Environmental Research (BER), as part of Subsurface Biogeochemical Research Program's
- 176 Scientific Focus Area (SFA) at the Pacific Northwest National Laboratory (PNNL). Data were

177 generated under EMSL user proposal 51180. A portion of the research was performed at

- 178 Environmental Molecular Science Laboratory User Facility. PNNL is operated for DOE by
- 179 Battelle under contract DE-AC06-76RLO 1830.

180 Author Contributions

- 181 V.G.C., E.B.G, H.S.S. and J.C.S., conceptualized the study; V.G.C., L.R., J.W., W.G, and A.G.,
- 182 carried out the study; C.T.R., R.C. and J.T. conducted instrumental analyses; V.G.C. and E.B.G
- 183 drafted the manuscript and all authors contributed to the writing.

184 **Competing interests**

185 The authors declare no competing financial interests.

186 Materials and methods

187 Study site and sediment collection

- 188 This study was conducted using sediments from the Columbia River hyporheic zone within the
- 189 Hanford Site 300 Area (approximately 46° 22' 15.80"N, 119° 85 16' 31.52"W) in eastern
- 190 Washington, USA^{10,11,45}. Hyporheic zone sediments were collected in April 2018 at five
- locations, separated by $\sim 2 \text{ m}$ (depth: $\sim 30 \text{ cm}$). Sediments were sieved in the field to < 2 mm,
- 192 homogenized, and kept on ice until same-day laboratory processing. We performed sequential
- 193 organic carbon extractions with synthetic river water (see Supplementary Methods for
- 194 composition) prior to incubations to minimize the influence of carbon mobilized from sediments
- during field sampling. Details regarding pre-processing step are provided in the Supplementary
- 196 Methods.

197 Laboratory Microcosms

198 We used a total of 85 microcosms in a full factorial design (a) four chemically distinct OM

amendments at three concentrations, (b) four autoclaved controls (heat kills), and (c) one

200 synthetic water control, each treatment with five replicates. Incubations were performed over the

201 course of 5 days, where each day we incubated 17 bioreactors (i.e., 1 replicate treatment per day,

see design in Supplementary Table 1). On the day prior to the experiment, 10 g of pre-processed

sediments were removed from 4°C storage and subsampled into 20 mL borosilicate glass vials.

204 Vials were left in the dark at ambient laboratory temperature for 8 h before incubation to

- 205 acclimate to room temperature.
- 206 To initiate microcosms, 18 mL of treatment solution was added to vials containing sediment,
- 207 leaving <1 mL headspace. Treatment solution consisted of synthetic river water and the specific

208 OM compound at the desired concentration (Supplementary Table 1). Nitrate and phosphate 209 were added to the synthetic river water to provide sufficient nutrients for the duration of the 210 experiment - nitrate concentration matched ambient groundwater while phosphate concentration 211 matched the Redfield ratio relative to groundwater N (16N:P). We added the following 4 types of 212 OM because their thermodynamic properties encompassed the extremes experienced by the surface and the groundwater in situ¹² and were either N-bearing or N-free: Lysine (Gibbs free 213 214 energy of the half reaction of organic carbon under standard conditions, $\Delta G^{\circ}_{Cox} = 79.40$ kJ (mol C)⁻¹, N-bearing), Serine ($\Delta G^{\circ}_{Cox} = 41.21 \text{ kJ} \text{ (mol C)}^{-1}$, N-bearing), Propionate ($\Delta G^{\circ}_{Cox} = 79.40 \text{ kJ}$ 215 $(mol C)^{-1}$, N-free), and Ascorbate $(\Delta G^{\circ}_{Cox} = 41.21 \text{ kJ} (mol C)^{-1}$, N-free). The vials were placed 216 217 horizontally on a shaker at 250 rpm in the dark at 21±1 °C for the duration of the experiment, 218 except during dissolved oxygen measurements.

219 **Respiration rates**

220 Dissolved oxygen (DO) concentration (μ mol L⁻¹) was measured in each microcosm every hour 221 for 6 h using 0.5 cm diameter factory-calibrated oxygen sensors and an oxygen optical meter 222 (Fibox 3; PreSens GmbH, Regensburg, Germany). The DO measurements were automatically 223 corrected for temperature and the data were recorded using PST3v602 software (PreSens 224 GmbH). Respiration rates were calculated as the slope of the linear regression between DO 225 concentration and incubation time for each microcosm (Supplementary Fig. 4-6). We infer that 226 changes in DO were driven by aerobic respiration, as DO in heat kills did not change during the 227 incubation (Supplementary Fig. 7). pH measurements were also collected using an optical meter 228 and factory calibrated pH sensor spots (pH-1 mini; PreSens GmbH). pH values did not change 229 during the incubation (Supplementary Fig. 8). After 6 hours, microcosm contents were 230 transferred to 50 mL sterile polypropylene centrifuge tubes and centrifuged for 5 min at 3200 rcf

- and 20° C. The supernatant was filtered through a 0.22 µm polyethersulfone membrane filter
- 232 (Millipore Sterivex, USA) frozen at -20°C until further analysis.

233 Fourier transform ion cyclotron resonance mass spectrometry (FTICR-MS)

- Fourier transform-ion cyclotron resonance mass spectrometer (FTICR-MS) (12 Tesla (12T)
- 235 Bruker SolariX, Billerica, MA) located at the Environmental Molecular Sciences Laboratory in
- 236 Richland, WA, was used to collect high-resolution mass spectra of the OM. Resolution was
- 237 220K at 481.185 m/z. The FTICR-MS was outfitted with a standard electrospray ionization (ESI)
- source, and data was acquired in negative mode with the voltage set to +4.4kV. Data were
- collected with an ion accumulation time of 0.3 sec from 98 900 m/z at 4M. One hundred
- 240 fourty-four scans were co-added. BrukerDaltonik (version 4.2) was used to convert raw spectra
- to a list of m/z values by applying FTMS peak picker module. Chemical formulas were then
- assigned using in-house software following the Compound Identification Algorithm⁴⁶⁻⁴⁹, using
- the criteria previously described by Graham et al. 10,11 . The chemical character of the compounds
- identified in the FTICR-MS spectrum and their biochemical classes were evaluated using Van
- 245 Krevelen diagrams.
- 246 We calculated the ΔG°_{Cox} to evaluate relationships between aerobic respiration and OM
- 247 thermodynamics, as per LaRowe and Van Cappellen²⁸. An expanded description of sample
- 248 preparation, instrument and FTCR-MS data processing for estimating Van Krevelen diagrams
- and ΔG°_{Cox} is presented in the Supplementary Methods.

250 Identification of biochemical transformations using FTICR-MS

- 251 Biochemical transformations were inferred by calculating all possible pairwise mass differences
- 252 within a sample's spectrum and matching differences (within $1 \Box ppm$) to a list of common

253	biochemical transformations ⁵⁰ . Biochemical transformations were identified following the
254	procedures described by Breitling et al. ⁵⁰ and previously employed by Bailey et al. ³⁴ , Graham et
255	al. ^{10,11} , Moritz et al. ³⁶ , Kaling et al. ³⁵ , and Stegen et al. ¹² . Briefly, pairwise mass differences
256	between all m/z peaks in a sample were compared with a reference list of 1298 commonly
257	observed biochemical reactions of organic matter (Supplementary Table 2). For mass differences
258	matching to compounds in the reference list, we inferred the gain or loss of that compound via a
259	biochemical transformation.
260	Statistical analyses
260 261	Statistical analyses All statistical analyses were completed using R (version 3.4.1). Linear regressions were used to
260 261 262	Statistical analyses All statistical analyses were completed using R (version 3.4.1). Linear regressions were used to assess the relationship between respiration rates and $\overline{\Delta G^{\circ}_{Cox}}$. Differences across groups were
260 261 262 263	Statistical analyses All statistical analyses were completed using R (version 3.4.1). Linear regressions were used to assess the relationship between respiration rates and ΔG°_{Cox} . Differences across groups were evaluated with one-sided Mann-Whitney U test. Permutational multivariate analysis of variance
260 261 262 263 264	Statistical analyses All statistical analyses were completed using R (version 3.4.1). Linear regressions were used to assess the relationship between respiration rates and ΔG° _{Cox} . Differences across groups were evaluated with one-sided Mann-Whitney U test. Permutational multivariate analysis of variance (PERMANOVA) of Bray–Curtis distances was used to assess dissimilarities among

266 ΔG_{Cox} to account for any differences due to the thermodynamic properties of the treatment

267 solution. Biochemical transformations were visualized with non-metric multidimensional scaling

268 (NMDS).

270 References

- 271 1 Battin, T. J. *et al.* Biophysical controls on organic carbon fluxes in fluvial networks.
- 272 *Nature geoscience* **1**, 95 (2008).
- 273 2 Cole, J. J. *et al.* Plumbing the global carbon cycle: integrating inland waters into the
- terrestrial carbon budget. *Ecosystems* **10**, 172-185 (2007).
- 275 3 Tranvik, L. J. *et al.* Lakes and reservoirs as regulators of carbon cycling and climate.
- 276 *Limnology and Oceanography* **54**, 2298-2314 (2009).
- 277 4 Sawakuchi, H. O. *et al.* Carbon dioxide emissions along the lower Amazon River.
- 278 Frontiers in Marine Science 4, 76 (2017).
- 279 5 Raymond, P. A. *et al.* Global carbon dioxide emissions from inland waters. *Nature* 503,
 280 355 (2013).
- 281 6 Ruhala, S. S. & Zarnetske, J. P. Using in-situ optical sensors to study dissolved organic
- carbon dynamics of streams and watersheds: A review. *Sci Total Environ* **575**, 713-723,
- 283 doi:10.1016/j.scitotenv.2016.09.113 (2017).
- 284 7 Naegeli, M. W. & Uehlinger, U. Contribution of the hyporheic zone to ecosystem
- 285 metabolism in a prealpine gravel-bed-river. *Journal of the North American Benthological*286 *Society* 16, 794-804 (1997).
- 8 Battin, T. J., Kaplan, L. A., Newbold, J. D. & Hendricks, S. P. A mixing model analysis
- 288 of stream solute dynamics and the contribution of a hyporheic zone to ecosystem
- 289 function. *Freshwater Biology* **48**, 995-1014 (2003).
- 290 9 Kaplan, L. A., Wiegner, T. N., Newbold, J., Ostrom, P. H. & Gandhi, H. Untangling the
- 291 complex issue of dissolved organic carbon uptake: a stable isotope approach. *Freshwater*
- 292 *Biology* **53**, 855-864 (2008).

- 293 10 Graham, E. B. et al. Multi 'omics comparison reveals metabolome biochemistry, not
- 294 microbiome composition or gene expression, corresponds to elevated biogeochemical
- function in the hyporheic zone. *Sci Total Environ* **642**, 742-753,
- 296 doi:10.1016/j.scitotenv.2018.05.256 (2018).
- 297 11 Graham, E. B. et al. Carbon Inputs From Riparian Vegetation Limit Oxidation of
- 298 Physically Bound Organic Carbon Via Biochemical and Thermodynamic Processes.
- *Journal of Geophysical Research: Biogeosciences* **122**, 3188-3205,
- 300 doi:10.1002/2017jg003967 (2017).
- 301 12 Stegen, J. C. et al. Influences of organic carbon speciation on hyporheic corridor
- biogeochemistry and microbial ecology. *Nat Commun* **9**, 585, doi:10.1038/s41467-018-
- 303 02922-9 (2018).
- Regnier, P. *et al.* Anthropogenic perturbation of the carbon fluxes from land to ocean.
 Nature geoscience 6, 597 (2013).
- 306 14 Butman, D. & Raymond, P. A. Significant efflux of carbon dioxide from streams and
- 307 rivers in the United States. *Nature Geoscience* **4**, 839 (2011).
- 308 15 Hedin, L. O. et al. Thermodynamic constraints on nitrogen transformations and other
- 309 biogeochemical processes at soil-stream interfaces *Ecology* **79**, 684-703,
- 310 doi:10.1890/0012-9658(1998)079[0684:tconao]2.0.co;2 (1998).
- 311 16 McClain, M. E. *et al.* Biogeochemical hot spots and hot moments at the interface of
- 312 terrestrial and aquatic ecosystems. *Ecosystems* **6**, 301-312 (2003).
- 313 17 Craig, L., Bahr, J. M. & Roden, E. E. Localized zones of denitrification in a floodplain
- 314 aquifer in southern Wisconsin, USA. *Hydrogeology Journal* **18**, 1867-1879,
- 315 doi:10.1007/s10040-010-0665-2 (2010).

- Jin, Q. & Bethke, C. M. A new rate law describing microbial respiration. *Appl. Environ. Microbiol.* 69, 2340-2348 (2003).
- 318 19 Keiluweit, M., Nico, P. S., Kleber, M. & Fendorf, S. Are oxygen limitations under
- 319 recognized regulators of organic carbon turnover in upland soils? *Biogeochemistry* **127**,
- 320 157-171, doi:10.1007/s10533-015-0180-6 (2016).
- Burd, A. B. *et al.* Terrestrial and marine perspectives on modeling organic matter
 degradation pathways. *Global change biology* 22, 121-136 (2016).
- 323 21 Li, L. *et al.* Expanding the role of reactive transport models in critical zone processes.
- 324 *Earth-science reviews* **165**, 280-301 (2017).
- 325 22 Boye, K. et al. Thermodynamically controlled preservation of organic carbon in

326 floodplains. *Nature Geoscience* **10**, 415-419, doi:10.1038/ngeo2940 (2017).

- 327 23 Pracht, L. E., Tfaily, M. M., Ardissono, R. J. & Neumann, R. B. Molecular
- 328 characterization of organic matter mobilized from Bangladeshi aquifer sediment: tracking
- 329 carbon compositional change during microbial utilization. *Biogeosciences* **15**, 1733-1747,
- doi:10.5194/bg-15-1733-2018 (2018).
- 331 24 Musolff, A. *et al.* Spatio-temporal controls of dissolved organic carbon stream water

332 concentrations. *Journal of Hydrology* **566**, 205-215, doi:10.1016/j.jhydrol.2018.09.011

- 333 (2018).
- Yang, Q., Zhang, X., Xu, X. & Asrar, G. R. An analysis of terrestrial and aquatic
 environmental controls of riverine dissolved organic carbon in the conterminous United
 States. *Water* 9, 383 (2017).
- Egli, T., Lendenmann, U. & Snozzi, M. Kinetics of microbial growth with mixtures of
 carbon sources. *Antonie van Leeuwenhoek* 63, 289-298 (1993).

339	27	Zinn, M., Witholt, B. & Egli, T. Dual nutrient limited growth: models, experimental
340		observations, and applications. Journal of biotechnology 113, 263-279 (2004).
341	28	LaRowe, D. E. & Van Cappellen, P. Degradation of natural organic matter: A
342		thermodynamic analysis. Geochimica et Cosmochimica Acta 75, 2030-2042,
343		doi:10.1016/j.gca.2011.01.020 (2011).
344	29	Berggren, M. & del Giorgio, P. A. Distinct patterns of microbial metabolism associated
345		to riverine dissolved organic carbon of different source and quality. Journal of
346		Geophysical Research: Biogeosciences 120, 989-999, doi:10.1002/2015jg002963 (2015).
347	30	Sundh, I. Biochemical composition of dissolved organic carbon derived from
348		phytoplankton and used by heterotrophic bacteria. Appl. Environ. Microbiol. 58, 2938-
349		2947 (1992).
350	31	Rosenstock, B. & Simon, M. Use of dissolved combined and free amino acids by
351		planktonic bacteria in Lake Constance. Limnology and Oceanography 38, 1521-1531
352		(1993).
353	32	Khatoon, H., Solanki, P., Narayan, M., Tewari, L. & Rai, J. Role of microbes in organic
354		carbon decomposition and maintenance of soil ecosystem. (2017).
355	33	Arnosti, C. et al. Extracellular enzymes in terrestrial, freshwater, and marine
356		environments: perspectives on system variability and common research needs.
357		Biogeochemistry 117, 5-21 (2014).
358	34	Bailey, V. L., Smith, A., Tfaily, M., Fansler, S. J. & Bond-Lamberty, B. Differences in
359		soluble organic carbon chemistry in pore waters sampled from different pore size
360		domains. Soil Biology and Biochemistry 107, 133-143 (2017).

361	35	Kaling, M. et al. Mycorrhiza-triggered transcriptomic and metabolomic networks
362		impinge on herbivore fitness. Plant physiology 176, 2639-2656 (2018).
363	36	Moritz, F., Kaling, M., Schnitzler, J. P. & Schmitt-Kopplin, P. Characterization of poplar
364		metabotypes via mass difference enrichment analysis. Plant Cell Environ 40, 1057-1073,
365		doi:10.1111/pce.12878 (2017).
366	37	Moorhead, D. L. & Sinsabaugh, R. L. A theoretical model of litter decay and microbial
367		interaction. Ecological Monographs 76, 151-174 (2006).
368	38	Treseder, K. K. Nitrogen additions and microbial biomass: A meta analysis of
369		ecosystem studies. <i>Ecology letters</i> 11 , 1111-1120 (2008).
370	39	Vitousek, P. M. & Howarth, R. W. Nitrogen limitation on land and in the sea: how can it
371		occur? Biogeochemistry 13, 87-115 (1991).
372	40	Wang, G., Post, W. M. & Mayes, M. A. Development of microbial enzyme mediated
373		decomposition model parameters through steady state and dynamic analyses. <i>Ecological</i>
374		Applications 23, 255-272 (2013).
375	41	Sulman, B. N., Phillips, R. P., Oishi, A. C., Shevliakova, E. & Pacala, S. W. Microbe-
376		driven turnover offsets mineral-mediated storage of soil carbon under elevated CO 2.
377		Nature Climate Change 4, 1099 (2014).
378	42	Vachon, D., Prairie, Y. T., Guillemette, F. & Del Giorgio, P. A. Modeling allochthonous
379		dissolved organic carbon mineralization under variable hydrologic regimes in boreal
380		lakes. <i>Ecosystems</i> 20 , 781-795 (2017).
381	43	Brookshire, E. N. J., Valett, H. M., Thomas, S. A. & Webster, J. R. Coupled cycling of
382		dissolved organic nitrogen and carbon in a forest stream. Ecology 86, 2487-2496 (2005).

Bernal, S., Lupon, A., Catalán, N., Castelar, S. & Martí, E. Decoupling of dissolved
organic matter patterns between stream and riparian groundwater in a headwater forested
catchment. Hydrology and Earth System Sciences 22, 1897-1910, doi:10.5194/hess-22-
1897-2018 (2018).
Goldman, A. E. et al. Biogeochemical cycling at the aquatic-terrestrial interface is linked
to parafluvial hyporheic zone inundation history. Biogeosciences 14, 4229-4241,
doi:10.5194/bg-14-4229-2017 (2017).
Kujawinski, E. B. & Behn, M. D. Automated analysis of electrospray ionization Fourier
transform ion cyclotron resonance mass spectra of natural organic matter. Analytical
Chemistry 78, 4363-4373 (2006).
Minor, E. C., Steinbring, C. J., Longnecker, K. & Kujawinski, E. B. Characterization of
dissolved organic matter in Lake Superior and its watershed using ultrahigh resolution
mass spectrometry. Organic geochemistry 43, 1-11 (2012).
Tfaily, M. M. et al. Sequential extraction protocol for organic matter from soils and
sediments using high resolution mass spectrometry. Analytica chimica acta 972, 54-61
(2017).
Tolić, N. et al. Formularity: software for automated formula assignment of natural and
other organic matter from ultrahigh-resolution mass spectra. Analytical chemistry 89,
12659-12665 (2017).
Breitling, R., Ritchie, S., Goodenowe, D., Stewart, M. L. & Barrett, M. P. Ab initio
prediction of metabolic networks using Fourier transform mass spectrometry data.
Metabolomics 2, 155-164, doi:10.1007/s11306-006-0029-z (2006).



406 **Fig. 1** Aerobic respiration rates in microcosms and its relationship with $\overline{\Delta G^{\circ}_{Cox}}$. (a) Respiration 407 rates were higher in microcosms amended with 3 mg C L⁻¹ vs. 0. 3 mg C L⁻¹, while microcosms 408 receiving 3 mg C L⁻¹ and 9 C mg L⁻¹ had similar respiration rates, indicating carbon limitation at 409 low concentrations of amended OM that was alleviated with increasing OM addition. In 410 microcosms with OM added at (b) 0.3 C mg L⁻¹ and (c) 3 mg C L⁻¹, $\overline{\Delta G^{\circ}_{Cox}}$ shows a negative 411 relationship with respiration, while microcosms with (d) 9 mg C L⁻¹ amendments show no 412 relationship between OM thermodynamics and respiration rates.



414 Fig. 2 Biochemical transformations between N-bearing and N-free OM amendments and its link 415 with respiration rates. Non-metric multidimensional scaling (NMDS) plots of microcosms receiving low OM concentrations (a) 0.3 mg C L^{-1} and (b) 3 mg C L^{-1} showed no difference in 416 417 biochemical transformation profiles between N-bearing and N-free amendments. In contrast, microcosms receiving high OM concentrations (c) 9 mg C L^{-1} had significantly different 418 419 biochemical transformation profiles between N-bearing and N-free amendments. (d) Microcosms 420 amended with N-bearing OM at 9 mg C L^{-1} also showed enhanced respiration vs. those receiving 421 N-free OM at the same concentration. In contrary, respiration rates were not statistically different between N-bearing and N-free microcosms receiving low OM (0.3 and 3 mg C L⁻¹). Colors in 422 423 all panels indicate N-bearing (teal) and N-free (orange) OM amendments. P-values in (a-c) were 424 derived from PERMANOVAs, and the p-value in (d) was calculated using a one-sided Mann-425 Whitney U test.



426

427 Fig. 3 Conceptualization of thermodynamic and nutrient regulations on aerobic respiration. We 428 propose a new conceptual model in which thermodynamic and nutrient limitations dually control 429 aerobic respirations. We suggest that thermodynamic properties of OM govern aerobic 430 respiration rates in ecosystems with low carbon to nutrient ratios. When OM concentration 431 reaches a threshold, thermodynamic controls do not persist, and nutrient availability, particularly 432 N regulate respiration. This work highlights a structural gap in aquatic biogeochemical models 433 and challenges long-held beliefs about aerobic metabolism being solely governed by reaction 434 kinetics. This new paradigm provides a link between OM chemistry and biogeochemical rates 435 with direct avenues for model incorporation, where OM chemistry regulates OM oxidation 436 through its thermodynamic properties until OM concentrations are sufficient to induce nutrient 437 limitations.