1 Polyvinyl Chloride Biodegradation Fuels Survival of Invasive Insect

2 Larva and Intestinal Degrading Strain of Klebsiella

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21 Abstract

22 Microbial degradation of polyvinyl chloride (PVC) is eco-friendly and economically 23 attractive, but extremely challenging due to the lack of mechanistic understanding on the 24 degrading strains and enzymes. Motivated by an accidental discovery that the larva of an 25 agricultural invasive pest, Spodoptera frugiperda, effectively survived solely on PVC film, 26 we profiled the intestinal microbiota of S. frugiperda and screened for PVC-degrading strains. 27 The results showed PVC film feeding significantly changed the larvae intestinal microbiota 28 through selective enrichment of Enterococcus, Ochrobactrum and Klebsiella. From the larva 29 intestines, we isolated and named a biofilm-forming strain EMBL-1, and experimentally 30 verified it as the first Klebsiella bacterium that can actively degrade and utilize PVC based on 31 various classic physicochemical and morphological analyses. We further used multi-omics 32 analyses that complementarily integrate whole genomic, transcriptomic, proteomic, and 33 metabolic insights to identify enzyme-coding genes responsible for PVC degradation and 34 proposed a putative biodegradation pathway by the bacterial strain. All in all, both S. 35 frugiperda and its intestinal strain EMBL-1 are discovered to effectively survive on PVC film 36 by exploiting its polymer as a sole energy source. Moreover, this work exemplifying PVC 37 biodegradation provides reference for discovering more degrading microbes and enzymatic 38 resources of other recalcitrant plastics.

Keywords: Spodoptera frugiperda; intestinal microorganisms; polyvinyl chloride (PVC);
 Klebsiella; degrading enzymes; multi-omics

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42 Introduction

43 The globally increasing accumulation and pollution of wasted plastics is a serious eco-44 environmental and socio-economic problem¹. Polyvinyl chloride (PVC) is one of the six 45 widely used plastic polymers, such as polyethylene (PE), polystyrene (PS), polypropylene 46 (PP), polyurethane (PUR) and polyethylene terephthalate (PET). Its market share (10.0%) is second only to PE (29.7%) and PP (19.3%) based on the European polymer demand², 47 48 producing tremendous plastic wastes. Landfill and incineration processes are commonly used 49 for the treatment and final disposal of plastic wastes. However, such energy-intensive 50 technologies and industries, unfortunately, treat plastics solely as useless wastes rather than 51 recyclable resources. Worse still, numerous harmful secondary pollutants (e.g., chloride and 52 dioxins) and green-house gases are largely and continuously released into our water, soil and 53 air environment³, making these economically unattractive technologies eco-environmentally 54 risky and unsustainable. Unfortunately, no green or sustainable approach is practically 55 available for cost-effective and environmentally friendly disposal of PVC wastes, calling for 56 methodological and technological innovations in the treatment and recycling of such man-57 made plastics.

58 Biological treatment and recycling of organic wastes is a promising approach for the future development of circular green economy ^{4,5}. PVC is considered very difficult to be 59 biodegraded because of its recalcitrant nature 6 . It has the same linear structure as PE except 60 61 for one of its hydrogens replaced with a chloride atom. However, current research progress on the biodegradation of PVC polymers far lags behind that of PE⁷⁻¹² and other plastics (e.g., 62 PET¹³⁻¹⁵). Until now, although limited research reported the biodegradation of PVC materials 63 including plasticizers by microbial consortia ^{3,16,17}, little is, however, known about the 64 65 responsible PVC-degrading microbes or biodegradation pathways (e.g., enzymes and 66 products). Other research empirically speculated PVC biodegradation by several fungal (i.e., 67 Basidiomycotina, Semi-bacterial, Ascomycota) or bacterial (i.e., *Pseudomonas*, 68 Mycobacterium, Bacillus and Acinetobacter) groups based on solely morphological and 69 physicochemical changes (e.g., surface damage and molecular weight loss) that visually

signify plastic degradation ¹⁸⁻²². While these prior studies may generally provide preliminary 70 71 evidence for microbial PVC degradation, the biodegradation mechanisms remain unexplored. 72 In particular, there is no report on microbial or bacterial genes and enzymes in the degradation of PVC ²³, excluding Nazia Khatoon's documentation of PVC-degrading activity of an 73 74 extracellular lignin peroxidase of fungal Phanerochaete chrysosporium²⁴. Therefore, global 75 efforts are urgently needed to discover PVC-degrading microbes and enzymatic resources, 76 and untangle the underlying biodegradation mechanisms to develop oriented engineering 77 tools and biotechnology for prospering future industry of PVC wastes treatment and recovery.

78 During laboratory cultivation of an agriculturally invasive insect Spodoptera frugiperda, we 79 accidentally discovered that the insect larva actively bit and fed on PVC film (Figure 1a). 80 Motivated by curiosity and recent reports that some insect species (particular the larvae of wax moths and meal moths²⁵) can consume plastic polymers of PE or PS ²⁶⁻²⁹, we specifically 81 82 designed triplicate cultivation experiments to check whether the larva can survive solely on 83 PVC film and whether their intestinal microbiota may play a role in the film digestion, 84 leading to our first discovery of intestinal microbiota-dependent ability of S. frugiperda larva 85 to degrade PVC for effective survival. Further, using PVC film as the sole energy and organic 86 carbon source, we successfully isolated the first PVC-degrading Klebsiella strain (named 87 EMBL-1) from the intestinal microbiota of the larvae. We explored the biodegradation 88 enzymes and genes, and metabolic products of PVC film and proposed the first metabolic 89 pathway for bacterial PVC biodegradation based on multi-omics approaches (e.g., genomics, 90 transcriptomics and proteomics). This work systematically exemplifies an integrated use of 91 multi-omics approaches universal for comprehensively mining biodegradation strains, genes 92 and enzymes of plastics other than, providing a theoretical basis for future biodegradationbased enzymatic recycling of plastic wastes, as has been recently demonstrated for PET¹³⁻¹⁵. 93

94 **Results**

95 Discovery and verification of *S. frugiperda* larva survival fueled by PVC biodegradation

96 To verify our discovery that larva of S. frugiperda can consume PVC film for effective

97 survival, laboratory cultivation experiments were specifically designed and conducted in 98 triplicates to compare the key physiological indexes (i.e., survival rate, body weight and body 99 length) and intestinal microbiota between the larva under starvation (Starvation group, n =100 15), feeding solely on PVC film (PVC group, n = 50), and feeding normally on corn leaves 101 (Corn group, n = 30), respectively (Figure 1a). Overall, the survival rate of the larvae in the 102 PVC film group (70%) after 5-day cultivation was significantly higher than that of the 103 Starvation group (25%), although lower than that of the Corn group (100%, P < 0.0001) 104 (Figure 1b). This result consisted with the significantly increase (P < 0.0001) in the body 105 weight (Figure 1c) and length (Figure S1a) of the larva groups in the following order: 106 Starvation < PVC < Corn, indicating that the PVC film can provide energy for and maintain 107 the survival of the larva, although the growth efficiency of this specialized feeding with PVC 108 is lower than that with the normal corn leaves, as also morphologically manifested by the 109 contrasting excreted feces (Figure S1b). Moreover, scanning electron microscopy (SEM) 110 analysis of the PVC fragments recovered from excreted feces showed strong surface damage 111 (Figure S1c). These results together verify our discovery that the gut PVC biodegradation 112 fuels the survival of S. frugiperda larva.

113 Interconnections between intestinal microbiota and PVC biodegradation

114 The lack of evidence to date for plastic degradation by germ-free invertebrate larvae generally 115 supports that intestinal microbiota are key drivers of plastic degradation²⁵. Supporting our 116 hypothesis that gut microbiota is essential for PVC degradation by S. frugiperda larva, the 117 survival rate of the larva after 5 days was significantly reduced because of the gentamicin 118 inhibition of intestinal microbiota. Accordingly, the body weight (Figure 1c) and body length 119 (Figure S1a) of PVC-fed larva group treated by gentamicin were significantly lower than that 120 without gentamicin treatment. The microbial biomass in the larvae intestinal microbiota of the Corn group kept stable during the experiment (from 3.90 ± 0.58 to $3.70\pm0.63\times10^{6}$ CFUs/piece), 121 122 while the antibiotic-treated group showed an over 99% reduction in microbial biomass. These 123 results suggest the dependence of PVC biodegradation on the larva intestinal microbiota.

124 The PVC film degradation by intestinal microbiota can release transformation products, 125 which we hypothesize should create new ecological niches for microbiome selection through 126 cross feeding. Consistent with the assumption, 16S rRNA gene amplicon sequencing analysis 127 of intestinal microbiota showed that PVC film degradation triggered a dramatic compositional 128 shift from Proteobacteria-dominated $(87.5\pm8.0\% \text{ to } 49.5\pm16.0\%)$ to Firmicutes-abundant 129 (11.9±7.2% to 44.2±17.0%,) microbiota (Figure 1d). Further cross-group comparisons down 130 to the levels of genus and amplicon sequence variants (ASVs) showed that compared with the 131 normal feeding with corn leaves, PVC feeding and biodegradation dramatically increased the 132 alpha diversity of ASVs in the larvae gut microbiota (i.e., Shannon's H index increased from 133 ~ 0.7 to 2.0 and observed species from 30.6 to 70.0) and largely favored the selective 134 enrichment of an unclassified Enterococcus (4.7±3.7% to 37.0±19.8%), Ochrobactrum (0.1± 135 0.2% to $3.4\pm0.7\%$) and *Klebsiella* (1.4±0.6% to 1.7±0.4%) (Figure 1e), revealing close 136 interconnection between intestinal microbiota and PVC biodegradation.

137 Identification of strain EBML-1 as the first PVC-degrading Klebsiella bacterium

138 Because the larvae intestinal microbiota of S. frugiperda are found to associate with PVC film 139 degradation, we then assumed that the larvae intestine should represent an important reservoir 140 of PVC-degrading strains and promiscuous enzymatic resources. During laboratory screening 141 (Figure S2a), a gram-negative strain (Figure S2b-2c), named as EMBL-1, formed a visible 142 biofilm on the surface of the PVC film after 10-day incubation, causing cracks on the surface 143 of the PVC film (Figure 2a), accompanying a dramatic increase in biomass concentration, i.e., 144 OD600 from 0.20 to 0.94 (Figure 2b). The cracks formed during initial film degradation could facilitate further plastic degradation ³⁰. The strain was further identified as a new *Klebsiella* 145 146 bacterium most closely related to Klebsiella variicola and Klebsiella pneumoniae based on 147 PCR cloning, sequencing, and phylogenetic analysis of 16S rRNA gene (Figure S2d-2e).

148 The surface of the PVC film already formed a compact biofilm after 90-day incubation, 149 and after removing the biofilm it showed more pits and cracks. The results of contact angle 150 (Figure S3a) and tensile strength tests of PVC films (Figure 2c) indicated the surface 151 hydrophobicity and tensile strength of PVC film cultured with strain EMBL-1 had changed 152 significantly. These results together suggested that EMBL-1 did damage the physical integrity 153 of the PVC film. During the experiment, the weight loss of the PVC film inoculated with 154 strain EMBL-1 continued to significantly increase over 90 days, and the final average weight 155 loss of the PVC film reached 19.57% (Figure 2d). The results of Advanced Polymer 156 Chromatography (APC, Waters, China) showed that compared with the control group, the 157 molecular weight measures, i.e., Mn and Mw, of the PVC film in EMBL-1 strain group were 158 decreased by 12.4% and 15.0%, respectively (Figure 2e), indicating that the long-chain 159 structure of PVC was depolymerized, producing lower molecular weight fragments. Moreover, 160 Thermogravimetric Analysis (TGA/DSC 3+/1600 HT, Mettler-Toledo, Switzerland) results 161 showed that the T_{max} and T_{onset} of the PVC film in the EMBL-1 group markedly dropped from 162 316°C to 279°C, and 273°C to 253°C (Figure 2f), respectively, while those metrics showed 163 limited change (T_{max}: 316°C to 310°C and T_{onset}: 273°C to 265°C) in the control group (Figure 164 S3b-3c), suggesting that the strain EMBL-1 had attacked the PVC polymer chain and reduced 165 the chemical stability of the PVC film.

166 Strain EBML-1 degrades PVC polymer but not plasticizers in the film

167 Once PVC is demonstrated to be depolymerized by EMBL-1, we further questioned on the 168 diversity of transformation products by the strain. Semiquantitative FTIR MICRO 169 SPECTROMETER (ThermoFisher, Nicolet iS50, China) analysis showed that the infrared 170 spectrum of the surface of the PVC film inoculated with strain EMBL-1 gradually differed 171 from that of the control group over time (Figure 2f), forming new functional groups such as 172 hydroxyl (3500-3300 cm⁻¹) and carbonyl (1550-1650 cm⁻¹), indicating microbial oxidation of 173 the PVC film. Moreover, the peaks intensity of C-Cl stretch³ (690 cm⁻¹) showed a decreasing 174 trend over time, indicating that the PVC film had a dechlorination reaction under the action of 175 the EMBL-1 strain (Figure 2g). Further, we used GC-MS to quantitatively profile the 176 degradation products of PVC films within 90 days. By comparative inspection of the peaks 177 with significant differences between the EMBL-1 group and control group, we identified six

degradation products (Compounds 1 to 6) (Figure 2h), which were sequentially identified as "2-ethylhexanol" (1), "2-nonanol adipic acid" (2), "adipic acid, methy octyl ester" (3), "octyl myristate" (4), "dodecanoic acid, isooctyl ester" (5) and "hexadecenoic acid, 1-methylheptyl ester (6)" according to high match score (> 800) of each compound in the NIST library (Figure S4). In contrast, the GC-MS results of the liquid culture did not show inspectable soluble degradation products of PVC (data not shown).

184 PVC film is composed of both polymer and plasticizers, the latter usually accounting for 40% to 65% in amounts ³¹. As expected, three main plastic additives, i.e., dioctyl adipate 185 186 (DOA), dioctyl terephthalate (DOTP) and erucylamide (Figure S5a-5d and Table S1), were 187 detected in the PVC film by Pyrolysis-GC-MS. Recently, PVC film degradation by enriched 188 anaerobic marine consortia is implicated to start from the degradation of plasticizers and then extend to the degradation and destruction of the PVC polymer¹⁷. However, whether strain 189 190 EMBL-1 effectively degrade the main plasticizers detected in the PVC film is unknown. Our 191 triplicate degradation experiments showed no or negligible growth of strain EMBL-1 over 30-192 day incubation using each plasticizer as a sole energy source (Figure S5f), revealing that the 193 strain could not degrade examined plasticizers under tested conditions.

194 Genome-level taxonomy and functional profiles of strain EMBL-1

195 To explore the biodegradation mechanisms of PVC film, the complete genome of strain 196 EMBL-1 was constructed based on co-assembly of short reads (150 bps \times 2) and long reads 197 (average 27628 bps) derived from Illumina next-generation sequencing and Nanopore third-198 generation sequencing, respectively. The result showed that the genome contains a 5,662,860 199 bps circular chromosome with a G-C content of 57.31% and 5646 open reading frames (ORF) 200 predicted as protein-coding genes. Interactive Tree of Life (iTOL) analysis with whole-201 genome strain information of *Klebsiella* extracted from the GTDB database (Figure 3a) 202 showed that the genome sequence of strain EMBL-1 showed 99.01% average nucleotide 203 identity (ANI) to it closest relative *Klebsiella variicola* (RS GCF 000828055.2).

204 Functional annotation of strain EMBL-1 genome (Figure 3b) was performed by 205 homology-based search against four databases, i.e., NR, GO, KEGG and CAZy. Through 206 KEGG metabolic pathways and network analysis, strain EMBL-1 genome was found to 207 encode 87 genes involving in the metabolism and biodegradation of xenobiotics such as 208 phenylacetic acids and 4-hydroxyphenylacetate. Carbohydrate-Active Enzyme (CAZy) 209 analysis showed the genome encoded 74 glycoside hydrolases (GHs), which should 210 contribute to the strong digestion of lignocellulose-containing biomass (e.g., corn leaf) and 211 absorption capacity of carbohydrates by the larvae of S. frugiperda. To facilitate identification 212 of PVC-degrading enzymes, we systematically summarized reported genera and enzymes 213 related to PE, PVC, and PET degradation, and found that strain EMBL-1 had a total of 11 214 candidate plastic-degrading genes annotated as laccase, alkane monooxygenase, lipase, 215 esterase, peroxidase and carboxylesterase in the KEGG database (Table 1 and Dataset S1), 216 which provide firsthand genomic evidence for our finding that the EMBL-1 strain has the 217 ability to degrade PVC polymer.

218 Proteomic analysis of PVC degradation by strain EMBL-1

219 To further verify PVC-degrading enzymatic activities, strain EMBL-1 was regrown with PVC 220 film for 30 days before harvesting cells for intracellular (IN) or extracellular (OTU) protein 221 extraction and expression activity tests. The control group supplied with 1% glucose (glu) was 222 set up to better differentiate metabolic activities of PVC degradation from those of common 223 carbohydrates. The result showed no significant difference in the weight loss of PVC film 224 between the experimental and control groups (Figure 4a), revealing that additional organic 225 carbon source did not improve the PVC degradation efficiency of the strain. Instead, glucose 226 addition only increased the glucose metabolism of the EMBL-1 strain (as evidenced by the 227 following proteomic analysis), which in turn increased the protein (a) and biomass (b) 228 contents of the strain (Figure S6). Moreover, the *in vitro* activity measurement showed PVC-229 degrading activity of the four protein extracts in the following decreasing order: OUT 230 (13.5%) > OUTglu (10.4%) > IN (5.2%) > INglu (5.0%) (Figure 4b), suggesting that the231 strain exhibited stronger extracellular than intracellular activities for PVC degradation.

232 To comprehensively explore metabolic functions underlying PVC degradation by strain 233 EMBL-1, we used LC-MS/MS quantitative proteomics to identify a total of 29 proteins 234 jointly expressed in all four experimental and control groups (Figure 4c), plus 10 proteins 235 jointly expressed in only the two PVC-degrading experimental groups (Dataset S2). By 236 inspecting the differential expression activities of the 39 key proteins between intracellular 237 (IN) and extracellular (OUT) proteomes, we nominated two main categories of proteins 238 associated with PVC degradation. Firstly, we focused on the five most extracellularly (OTU) 239 expressed proteins (Figure 4d), including i) catalase-peroxidase, enolase, and aldehyde 240 dehydrogenase which are most likely responsible for the biodegradation of PVC or 241 depolymerized byproducts, and ii) highly-conserved and universal elongation factor Tu and 242 chaperone protein closely related to microbial translation and protective cell responses to 243 nutrient starvation ⁴⁶ and heat shock (or other co-defensed harmful conditions such as 244 alcohols, inhibitors of energy metabolism, and heavy metal)⁴⁷, respectively. Among them, 245 catalase-peroxidase has strong redox capacity and polymer depolymerization ability, which 246 has also been reported to degrade lignin ⁴⁸, while enolase with lyase activity and aldehyde 247 dehydrogenase with redox activity on aldehyde groups are considered the putative PVC-248 degradation proteins. Moreover, we identified another five proteins strongly upregulated 249 (defined here as $Log_2(OUT/IN) \ge 3$) extracellularly (OTU) than intracellularly (IN) (Figure 250 4d), such as i) dihydroxy-acid dehydratase which can degrade depolymerized products 251 through cleavage of carbon-oxygen bonds (Figure 5), ii) entericidin EcnA/B family protein 252 that can manifest the strain's stress responses to toxic substances (e.g., PVC plasticizers), iii) 253 porin OmpC and outer membrane proteins that can transport some small molecule metabolites, 254 and iv) glutamate synthase large subunit known to involve in ammonia assimilation ⁴⁹.

255 Transcriptomic analysis of PVC degradation by strain EMBL-1

To identify enzyme-coding genes associated with PVC film degradation, strain EMBL-1 was first grown on the PVC film in triplicate for 10 days. The weight loss of the film reached about 7% (Figure 4a), accounting for ~ 3 times increase in the strain biomass (OD₆₀₀ increase 259 from 0.20 to 0.61, Figure 4e). We then performed whole transcriptomic analysis of strain 260 EMBL-1 to screen out 77 out of 96 differentially expressed genes that were significantly 261 (FDR-adjusted $P \le 0.05$) up-regulated ($\log_2(FC) > 0.5$) or downregulated ($\log_2(FC) < -0.5$) in 262 the PVC film group (see red spheres, Figure 4f), compared with the control group (Dataset 263 S3). Most of the gene expression activities were ascribed to cell growth and death (e.g., 264 elongation factor G, 50S ribosomal protein, and DNA-directed RNA polymerase), followed 265 by transport and catabolism (e.g., MFS transporter, amino acid ABC transporter, and TonB-266 dependent siderophore receptor) (Figure 4g). Among them, three transcriptionally active 267 genes were also highly represented in the extracellular (OUT) and/or intracellular (IN) 268 proteomes of the strain (see proteins marked in red, Figure 4d). Notably, biodegradation-269 related genes, such as MBL fold metallo-hydrolase, phenylacetic acid degradation protein, 270 and alkyl hydroperoxide reductase, showed active upregulated expression during the strain's 271 PVC-dependent growth (Figure 4g), revealing their involvement in the PVC degradation.

272 Multi-omics analysis of PVC degradation pathway by strain EMBL-1

Based on the above multi-omics approaches that incorporates complementary results of genome, transcriptome, proteome and metabolites analyses, strain EMBL-1 was found to encode peroxidase, alkane monooxygenase, laccase, lipase, esterase, and carboxylesterase known to degrade PE and/or PET (Table 1 and Dataset S1) ^{48,50-53}. These enzymes probably participate in the biodegradation of PVC and its byproducts through a putative pathway which includes the abiotic effect, extracellular enzymatic depolymerization, and the intracellular metabolism of degradation by products (Figure 5).

First, abiotic factors including light and oxygen are widely considered to induce plastic degradation reactions which are initiated via C-C and C-H scission^{12,50,52,54}. These factors are most likely to attack and modify PVC polymer via hydroxyl dichlorination and carbonylation, as supported by FTIR diagram (Figure 2f). Then, catalase-peroxidase, known to degrade polymers (e.g., lignin ⁴⁸) and found by our study as the 4th most extracellularly expressed protein during the PVC-dependent growth of strain EMBL-1 (Figure 4d), depolymerizes the 286 modified polymer and converts it into C24~C5 byproducts, as detected by GC-MS/MS 287 analysis (Figure 2g). The multi-omics profiles of other metabolites and degradation enzymes 288 support further stepwise transformation of the long-chain products eventually into shorter 289 ones (importable by strain EMBL-1) through a series of enzymatic reactions catalyzed by 290 laccase³⁴, monooxygenase⁵⁵, dioxygenase, aldehyde dehydrogenase, esterase and dihydroxy-291 acid dehydratase (Figure 5), among which dioxygenase has been reported to modify and 292 degrade plastic polymer via oxidation of C=C functional group⁵⁶. Although laccase, alkane 293 monooxygenase, dioxygenase, esterase, and lipase are below detection by proteomic analysis 294 of EMBL-1, the strain possesses genes encoding these enzymes (Table 1). It is likely that the 295 binding substrate intermediates of these enzymes are promptly consumed by the strain during 296 PVC-dependent growth. Last but not the least, many genes encoding transport and catabolic 297 proteins are found to be highly represented in the proteomes (d) and transcriptomes (g) of 298 strain EMBL-1 (Figure 4). They are responsible for the transportation of small organic 299 molecules and fatty acids to support intracellular catabolism to support the strain's growth 300 using PVC as sole energy source.

301 Discussion

302 Microbial and enzymatic degradation of PVC is a grand global challenge. The limited number 303 of reported PVC-degrading strains are originated from natural and non-host environments such as soil²², landfills¹⁶, and marine environments^{17,21,57}, and the underlying biodegradation 304 305 mechanisms remain unexplored. This study is the first report on PVC-degrading microbiota, 306 strain, enzymes, and mechanism in the larvae intestines of an agriculturally invasive insect. 307 We discover that the larva of S. frugiperda can survive solely on the energy derived from 308 PVC degradation by intestinal microbiota, and further successfully isolate and experimentally 309 verify strain EMBL-1 as the first PVC-degrading Klebsiella bacterium. The isolate is taxonomically identified to belong to Klebsiella variicola (ANI > 99%). This species is 310 regarded as an emerging pathogen of human and other animals⁵⁸. It also harbors plant-311 associated isolates⁵⁹ that can fix nitrogen (to enable its co-existence with plants)⁶⁰, and 312 degrade xenobiotic pollutants (e.g., atrazine⁶¹) and natural polymers (e.g., lignin and 313

cellulose^{62,63}). Most importantly, our discovery of strain EMBL-1 and multi-omics exploration
of its PVC biodegradation mechanisms (as further discussed below) provide promising
methodology framework and research direction to guide future mechanistic investigations on
other PVC-degrading microbes as well as microbial degradation of other plastics.

318 Besides the discovery of PVC-degrading strain EMBL-1, we also demonstrate an 319 innovative use of multi-omics approaches that complementarily integrate DNA, mRNA, 320 protein, and metabolite analyses for mining and elucidating microbial PVC biodegradation. 321 Our multi-omics results show that PVC polymers and the monomers are effectively utilized 322 by the strain to derive energy for growth. The Klebsiella strain is found to utilize combined 323 actions of extracellular and intracellular enzymes to achieve a powerful function in 324 depolymerizing and decomposing monomers of PVC (Figure 5). It also exhibits strong 325 adhesion ability and easily forms a biofilm on the PVC film (Figure 2a), which should largely 326 facilitate its initial destruction of the hydrophobic surface structure of the PVC film by 327 secreting extracellular proteins (as shown by proteome analysis, Figure 4b-4d), thus paving 328 the road for its alternative lifestyle that utilizes PVC for growth⁶⁴. During PVC degradation, 329 39 proteins are differentially co-expressed by the strain, both extracellularly and 330 intracellularly. They are responsible for either direct degradation of PVC and its 331 depolymerized byproducts, or fundamental metabolism related with genetic or environmental 332 information processing and cellular processes that supports PVC-dependent growth of the 333 strain (Figure 4d). In particular, comparative transcriptomics showed significant upregulation 334 of functional genes involved in the xenobiotics biodegradation and metabolism, cell growth 335 and death, and transport and catabolism during cell growth (Figure 4g), revealing the strain's 336 strong growth ability and molecule transport ability closely linked with extracellular PVC 337 polymerization, further conversion, and eventual intracellular utilization.

338 Until now, except for the only report on the PVC-degrading activities of fungal lignin 339 peroxidase in *Phanerocheate chrysosporium*²⁴, little is known about PVC-degrading pathway, 340 genes and enzymes in bacteria. PVC and PE have similar stable chain structures. Oxidase is a 341 class of enzymes with better activity in the PE-degradation enzymes, which can add oxygen 342 to the long carbon chain⁶⁵, then it generates free radicals to form carboxyl groups, alcohols, 343 ketones and aldehydes³¹, such as monooxygenase and dioxygenase, etc.⁶⁶ The oxidation and 344 cleavage of PE make the polymer more hydrophilic, thus facilitate its contact with other 345 extracellular enzymes (e.g., lipase and esterase) after carboxyl group formation, or endopeptidase of amide group³¹. In addition, Eyheraguibel et al. reported that the Facilitor 346 347 superfamily (MFS) or a vector containing an ATP binding cassette (ABC) can integrate small 348 molecular weight PE oligomers into cells to achieve PE degradation⁶⁷. The results of genomic, 349 transcriptomic, and proteomic of the EMBL-1 strain together lead to our proposal of its 350 microbial PVC biodegradation pathway (Figure 5), advancing current mechanistic 351 understanding of bacterial PVC degradation.

352 In summary, we discover and fully verify EMBL-1 as the first PVC-degrading bacterium 353 belong to *Klebsiella*, and further demonstrate the use of multi-omics approaches to achieve 354 first in-depth insights into and systematic knowledge on its degradation mechanism of PVC. 355 In addition, our discovery of the S. frugiperda larva survival fueled by microbial 356 biodegradation of polyvinyl chloride lays the foundation for a follow-up study to elucidate the 357 biodegradation mechanisms of PVC polymers in the larvae intestinal microbiota of this 358 agricultural invasive insect and to mine other hidden plastic-degrading strains and enzymatic 359 resources therein.

360 Conclusion

361 Microbial biodegradation-based recycling of plastic wastes is promising but extremely 362 challenging, especially when biodegradation mechanism is unknown. In this study, we started 363 from the discovery of S. frugiperda larvae feeding on PVC film, explored the intestinal 364 microbiota changes underlying effective larvae survival, and finally isolated PVC-degrading 365 Klebsiella strain EMBL-1 which exhibited PVC-dependent growth. Through multi-omics 366 analyses integrating first genome, transcriptome, protein and metabolite-level insights into in 367 the strain, we identified a group of functional genes, enzymes, and metabolic pathways 368 closely related to PVC degradation and proposed a hypothetical biodegradation pathway via 369 dechlorination, oxidation, depolymerization, and further degradation and mineralization. This 370 study not only opens the gate to PVC-degrading microbiota, microbe, and enzymatic 371 resources in the larvae intestine of an invasive agricultural insect, but also provide a multi-372 omics framework and an intriguing scenario that inspire future studies to exploit microbial 373 biodegradation of other recalcitrant plastics or xenobiotic contaminants (e.g., pesticides). 374 More importantly, when today microbial and enzymatic biodegradation is demonstrated as an 375 eco-environmentally sustainable and commercially promising biotechnology for the recycling 376 of wasted PET ¹³⁻¹⁵, our study striving to systematically decipher the mechanisms underlying 377 bacterial PVC biodegradation should lay a foundation for a follow-up study to explore and 378 eventually realize sustainable treatment and recycling of PVC plastics enabled by the iterative 379 cycle of "Design-Build-Test-Learn (DBTL)" for DNA, strain and enzymatic engineering.

380 Methods

381 Field sampling and laboratory cultivation experiments

382 The larva of S. frugiperda were collected from corn fields in Jiangcheng County, Yunnan 383 Province, China. They were then cultured indoors by feeding corn leaves (artificial insect 384 breeding room with temperature of 25°C and humidity of 50%-60%). To validate the larva's 385 ability to live on PVC film, 120 pieces of 4_{th} -instar larva with the same growth status were 386 divided into three groups: 1) Control group (starvation, 15 pcs), 2) Corn group (fed with corn 387 leaves, 40 pcs), and 3) PVC group (fed with PVC film, 60 pcs). Other experimental 388 conditions, such as specific breeding conditions and PVC film cleaning, were kept the same, 389 as detailly described in Method S1. The weight and body length of all numbered larva were 390 measured after 24-h starvation and after the 5-day experiment. By the end of the experiment, 391 the excreted feces of each experimental group were collected and the number of survivals was 392 counted. Subsequently, the surviving larva in the Corn group and PVC group were dissected 393 under aseptic conditions to obtain intestinal samples (see details in Method S2), which were 394 labeled and temporarily stored at $4 \Box$ until further operation. The intestinal feces (IF) samples 395 collected from the PVC and Corn groups were labeled as PVC IF and Corn IF, respectively.

In addition, residual PVC fragments were also recovered from the excreted feces of the PVC
group to characterize their surface morphological changes by Hitachi Field Emission
Scanning Electron Microscope (Regulus 8230, Japan).

We hypothesized that intestinal microbiota of *S. frugiperda* larva should play an essential role in digesting PVC film to enable its observed and experimentally verified survival on the PVC film. To test the hypothesis, cultivation experiments were conducted to compare the key physiological indexes (i.e., body length, weight and survival rate) of PVC-fed larva with and without gentamicin pretreatment of their intestinal microbiota. The number of culturable cells in the intestinal microbiota were counted after ending the experiment.

405 **16S rRNA gene amplicon sequencing analysis of intestinal microbiota**

406 Molecular experiments and sequencing Total DNA was extracted from intestinal feces 407 samples collected from the experimental groups using QIAamp Fast DNA Stool Mini Kit 408 following the manufacturer's recommendations (QIAGEN GmbH, Germany). Then, 409 hypervariable V4-V5 regions of prokaryotic 16S rRNA gene were amplified using 515F (5'-410 GTGYCAGCMGCCGCGGTAA-3') and 926R (5'-CCGYCAATTYMTTTRAGTTT-3'). The 411 amplicon products of each sample were evenly mixed and sequenced using a paired-end 412 sequencing strategy (PE250) on the Illumina HiSeq2500. More details on the PCR conditions 413 and experimental procedures were available in the Method S3.

414 **Bioinformatics and statistical analyses** For the 16S rRNA gene amplicon data analysis, 415 FastQC (v0.11.9) and cutadapt (v1.18) were first used to check the quality of the raw data and 416 excise double-ended primers (fastaq files). Dada2 (v1.14) was then used to cluster the input 417 sequence with default parameter settings and further denoise after importing double-ended 418 data through Quantitative Insight into Microbial Ecology (QIIME2-2020.6) and input-format 419 setting parameters. The next step was to select high-quality areas based on FastQC's report 420 results. The taxonomic classification was conducted using the gime2 built-in package, the 421 feature-classifier classify-sklearn machine learning method was used for taxonomic

422 annotation using the SILVA 138 SSU as the reference database. The generated files were 423 imported into R studio Version 1.1.414 (R version 4.0.3), and phyloseq (v1.32.0) was used for 424 data statistical analysis and visualization. In addition, survival (v3.2.7) and survminer 425 packages were used to calculate and draw the survival curve, while ggplot2 (v3.3.3) was used 426 to draw box plots of weight and body length. The ANOVA test was used to check the 427 significance of differences between experimental groups.

428 Enrichment, isolation, and identification of PVC-degrading strain EMBL-1

429 To isolate PVC-degrading microbe, intestine materials of ten larva were suspended in 10-mL 430 PBS solution and vortexed for 5 min. Then, the intestinal mucosa was removed from the 431 mixed solution. The remaining suspension was used as a microbial inoculum and transferred 432 to a 250-mL flask containing 0.1-g PVC film and 100-mL MSM liquid medium. Then it was 433 cultivated on a shaker (150 rpm/min) at 30 \Box , transferred every 15 days. After 45 days, the 434 culture medium was first diluted and then spread onto MSM agar medium plate with PVC 435 film as the sole carbon source to cultivate and enrich for degrading strains. The enrichment 436 degrading strain was further sub-cultured until a pure colony of isolate was obtained. 437 Depending on whether the isolates were grown in the PVC film-amended liquid MSM, the 438 surface changes of the PVC film were inspected by SEM, until a PVC-degrading strain 439 (named EMBL-1) was successfully obtained. To identify the strain, near full-length 16S 440 rRNA gene sequence was PCR amplified using the universal primer set 27F (5'-441 AGAGTTTGATCCTGGCTCAG-3') and 1492R (5'- GGTTACCTTGTTACGACTT-3'). The 442 16S rRNA gene amplicon sequence obtained from Sanger sequencing were deposited in the 443 National Center for Biotechnology Information (NCBI accession no.: MZ475068) and 444 annotated using the NCBI's online Basic Local Alignment Search Tool (BLAST) in August 445 2021.

446 Biodegradation experiments of PVC film by strain EMBL-1

447 To further explore the ability of strain EMBL-1 to degrade PVC film, three experimental

448 groups were designed: 1) MSM liquid medium (20 mL) + EMBL-1 strain (OD600=0.2); 2) 449 MSM liquid medium (20 mL) + EMBL-1 strain (OD600=0.2) + PVC film (weighed); 3) 450 MSM liquid medium (20 mL) + PVC film (weighed). Each experimental group was prepared 451 in 27 replicates and immediately started upon inoculation of the strain cells. The following 452 operational procedures were repeated every 10 days: i) take out the PVC films from three 453 replicates of each group, ii) SEM microscopically inspect degradation of the PVC films and 454 weigh them, and iii) replace the half of the MSM liquid medium. The entire experiment was 455 lasted for 90 days, during which PVC films on 10 d and 90 d were collected and stored. The 456 post-treatment methods of PVC film and downstream morphological and physicochemical 457 characterization methods were described with details in the Method S4 and Method S5, 458 respectively.

To ensure the biodegradation ability of strain EMBL-1 to PVC polymer, Gas Chromatography Mass Spectrometry (GC-MS, Trace1300-ISQ7000, ThermoFisher, Singapore) method was established to analyze the composition of additives in the film (Method S6). Further, culture experiments were set up in triplicates and operated for 30 days to check whether the EMBL-1 strain is able to degrade and utilize the three major additives pre-identified from the PVC film (Method S7).

465 Characterization of PVC film damage and biodegradation products

466 **PVC film damage** To validate and follow PVC film biodegradation progress by strain 467 EMBL-1, multiple classic physicochemical methods were co-used to analyze the temporal 468 changes in the morphological, compositional, and other physiochemical properties over 90 469 days. First, the colonization of the strain was morphologically characterized by SEM after cell 470 fixation. Meanwhile, the degradation efficiency of the strain was directly measured based on 471 the weight loss of the PVC film on a 10-day basis (Method S5). Further, changes in the 472 physical properties of the PVC film were detected by contact angle and tensile strength tests 473 (Method S5). The depolymerization of the plastic materials was also recorded by the change 474 of molecular weight. FTIR MICRO SPECTROMETER (FTIR, ThermoFisher, Nicolet iS50,

475 China) was used to analyze and detect the changes in the surface chemical composition and 476 functional groups of the PVC film. Thermogravimetric Analysis (TGA/DSC 3+/1600 HT, 477 Mettler-Toledo, Switzerland) was used to compare the initial degradation temperature and the 478 maximum degradation temperature of the PVC film, and obtain the composition and heat 479 stability and thermal decomposition of the PVC film and the possible intermediate products. 480 The Advanced Polymer Chromatography (APC, Waters China) was additionally used to 481 determine the molecular weight of different groups of PVC films. The specific processing 482 steps and testing conditions of FTIR, TGA and APC analyses were recorded in Method S5.

483 **Biodegradation products**

484 To master more biodegradation evidence of the PVC film by EMBL-1 strain, we detected 485 potential biodegradation products of PVC in the PVC films and cultural mediums by using 486 GC-MS. Degraded PVC film weighing 0.3 g were cut into pieces and mixed with 10 mL 487 tetrahydrofuran, the mixture was ultrasonicated for 30 min at room temperature. The extract 488 was concentrated to 0.5 ml by nitrogen blowing treatment and it was mixed with 1 mL N-489 hexane to obtain some possible products by vortex an ultrasonic for 10 min. The samples filtered using 0.22 um PTFE syringe filter for further step⁶⁸. The liquid culture (150 mL) was 490 491 centrifuged at 10,000 rpm for 15 min, and the supernatant was extracted with the same 492 volume of n-hexane. The extract was concentrated to 1 mL by nitrogen blowing treatment. 493 The soluble daughter products in the filtrate were analyzed using GC-MS.

494 **GC-MS analysis of PVC-degradation products** The sample was injected at an initial 495 temperature of $40 \square$ (hold 4 min) which was progressively increased at 10°C per minute and 496 held at 280°C (hold 5 min). Moreover, the detector conditions such as transfer line 497 temperature, ion source temperature, ionization mode electron impact and scan time were 498 maintained at 250°C,280°C, 70 eV and 0.3 s respectively.

499 Whole-genome sequencing analysis of PVC-degrading strain EMBL-1

500 To further explore the biodegradation mechanism of PVC film by strain EMBL-1 and

501 discover the responsible PVC-degrading genes or enzymes, TIANamp Bacteria DNA Kit was 502 used to extract the genomic DNA of the strain and the genomic DNAs of the EMBL-1 strain 503 was split into two fetches and sequenced using on both Illumina next-generation sequencing 504 (PE150) and Oxford Nanopore (PromethION). The experimental procedures including sample 505 quality testing, library construction, library quality testing, and library sequencing were 506 performed in accordance with the standard protocol provided by sequencer providers. 507 Bioinformatics analysis includes five major steps: raw data quality control, genome assembly, 508 genome component analysis, functional annotation, and genome visualization. In brief, the 509 quality control of raw short reads from Illumina sequencing and raw long reads from 510 Nanopore sequencing were performed in Fastp 0.19.5 and Mecat 2, respectively. Then, the 511 clean short and long reads were co-assembled to reconstruct compete genome using Unicycle 512 (https://github.com/rrwick/Unicycler) to generate complete. Predictive coding sequence (CDS) was predicted using Glimmer version 3.0269). Databases such as KEGG, COG, GO, and 513 514 CAZy are used for function annotation. In addition, the MUMmer software (v3.23) was used 515 to compare the target genome with the reference genome to determine the collinearity 516 between the genomes.

517 Proteomic analysis of PVC film degradation by strain EMBL-1

518 To mine enzymatic activities and metabolic pathways related to PVC degradation, 519 biodegradation experiments of PVC film by the EMBL-1 strain were conducted with and 520 without additional supply of 1% (w/v) glucose (to resolve its proteomic signals from that of 521 PVC film). Then, both intracellular and extracellular proteins were separately extracted from 522 the cells harvested after 30 days based on the acetone precipitation method. The ability of the 523 protein solutions to degrade PVC film was further tested in vitro (Method S8). Meanwhile, 524 proteins were resolved with the Thermo Ultimate 3000 integrated nano-HPLC system which 525 is directly interfaced with the Thermo orbitrap fusion lumos mass spectrometer (LC-MS/MS) 526 to explore some related PVC degradation proteins. Details on the experimental setups and 527 procedures, protein extraction methods, protein activity tests, and proteomic analysis were 528 described in Method S9.

529 Transcriptomic analysis of PVC film degradation by strain EMBL-1

530 **Experimental design** To mine genes related to PVC degradation, a degradation experiment 531 of PVC film by the EMBL-1 strain were conducted for 10 d. a) MSM medium + EMBL-1 532 (OD600=0.2), b) MSM medium +EMBL-1 (OD600=0.2) + PVC film (weighed), three 533 repeats per group. All treatments were cultured in a shaker (30°C, 150 rpm). By the end of the 534 experiment, the liquid culture was centrifuged in 4°C to harvest the cells. The total RNA of 535 each treatment was extracted, and RNA integrity was assessed using the RNA Nano 6000 536 Assay Kit of the Agilent Bioanalyzer 2100 system (Agilent Technologies). Sequencing 537 libraries were generated using NEBNext Ultra Directional RNA Library Prep Kit for Illumina 538 (NEB) and sequenced on an Illumina Hiseq 2500 platform. Sequencing was performed at 539 Beijing Novogene Bioinformatics Technology Co., Ltd. The raw data of transcriptomic of 6 540 samples from two groups were uploaded in the CNGB database (Sub 022027).

541 **Bioinformatics analysis** Raw reads were first filtered using fastp to remove the reads that 542 contain 10 low-quality bases (base quality score less than 20) or length shorter than 36bp. 543 Then, the resulting high-quality (HQ) reads were aligned to *Klebsiella variicola* reference 544 genome (*Klebsiella variicola* strain FH-1) using hisat2. After alignment the read counts for 545 each gene were extracted using htseq-count. The gene expression profiles of triplicate 546 transcriptomes in two groups were compared with PCoA, which inspected one outlier dataset 547 in each group (due to unexpected experimental errors) that was discarded from downstream 548 analysis. Differential expression (DE) at a gene level in our two groups (group a) and group 549 b)) was evaluated using edgeR version 3.30.3, implemented in R 4.0.3. The p-values 550 presented are adjusted for multiple-testing with the procedure of Benjamini and Hochberg to 551 control the type I error rate, and a cut off of $p \le 0.05$ was used as a threshold to define 552 differential expression. Kraken2 was used to check the contamination in RNA-seq.

553 Multi-omics-based prediction of degradation pathway of PVC film

554 To further explore the degradation mechanism of PVC film by strain EMBL-1, multi-omics 555 results from genome, transcriptome, proteome and metabolite analyses were co-used to 556 propose a putative pathway of PVC degradation. In brief, the potential plastic-degrading 557 genes encoded in the EMBL-1 genome (Table 1 and Datasets S1) and the metabolites detected 558 by GC-MS were co-used to build a PVC degradation pathway. Further, 39 proteins jointly 559 expressed during PVC-dependent growth of strain EMBL-1 (Figure 4d and Dataset S2) were 560 aligned against the 96 differentially expressed genes revealed by transcriptomic analysis 561 (Figure 4f and Dataset S3) using NCBI's BLAST+ 2.9.0 at an e-value cut off of 0.01, 562 generating a list of gene expression and proteomic activities ascribed to the PVC-dependent 563 metabolism of the strain.

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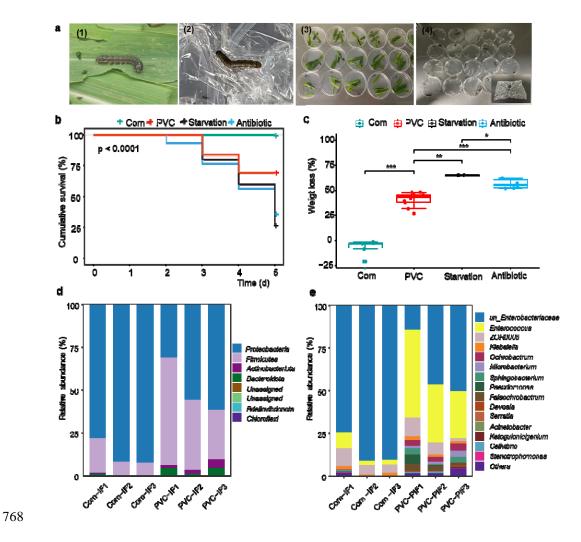
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761 Figure and Table Legends

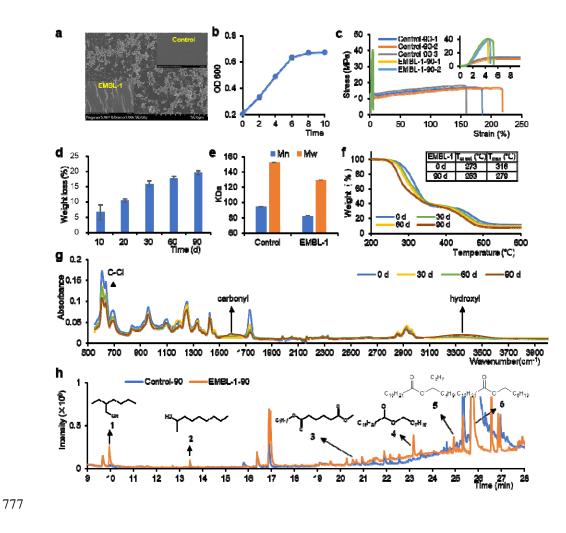
Figure 1 Laboratory cultivation and intestinal microbiota composition of *Spodoptera frugiperda* **larva feeding on corn leaf and PVC film.** a, laboratory feeding of *S. frugiperda* larva with corn leaf (Corn group) and PVC film (PVC group). b-c, cumulative survival and body weight loss of larva in the Corn group, PVC group, Starvation group (no feeding) and Antibiotic group (gentamicin pretreatment of intestinal microbiota before PVC feeding). d-e, intestinal microbiota composition at the phylum (d) and genus (e) levels.





770 Figure 2 Physicochemical characterization of PVC film degradation by strain EMBL-1.

a-b, the SEM image (a) and the growth curve (b) of strain EMBL-1 after co-culturing with
PVC film for 10 days. c & e, the tensile strength (c) and the molecular weight (e) of PVC film
in the Control group and the EMBL-1 group after 90 days. d, f & g, the temporal change in
the weight loss (d) and the thermogravimetric analysis TGA (f) and FTIR (g) diagrams of
PVC film in the EMBL-1 group over 90 days. h, the total ion chromatogram (TIC) depicting
PVC degradation products.



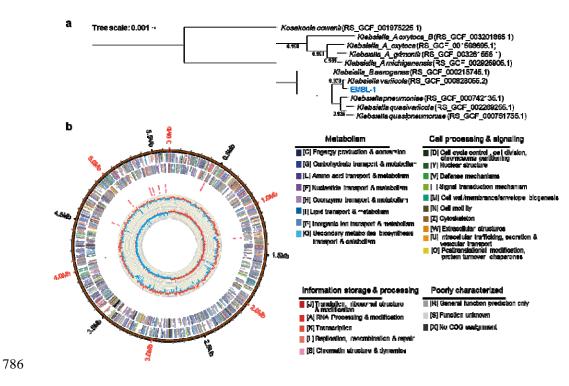
779 Figure 3 Genome-level phylogeny and functional annotation of strain EMBL-1. a, the

780 phylogenetic tree of EMBL-1 strain built with ITOL. b, the whole genome map of strain

781 EMBL-1. Rings from the outermost to the center: 1) scale marks of the genome, 2) protein-

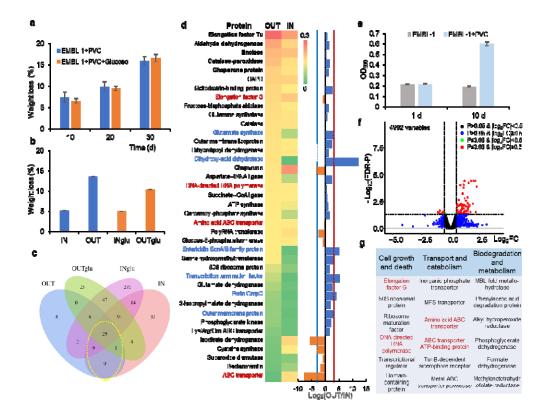
782 coding genes on the forward strand, 3) protein-coding genes on the reverse strand, 4) tRNA

- 783 (black) and rRNA (red) genes on the forward strand, 5) tRNA (black) and rRNA (red) genes
- 784 on the reverse strand, 6) GC content, 7) GC skew. Protein-coding genes are color coded
- 785 according to their COG categories.





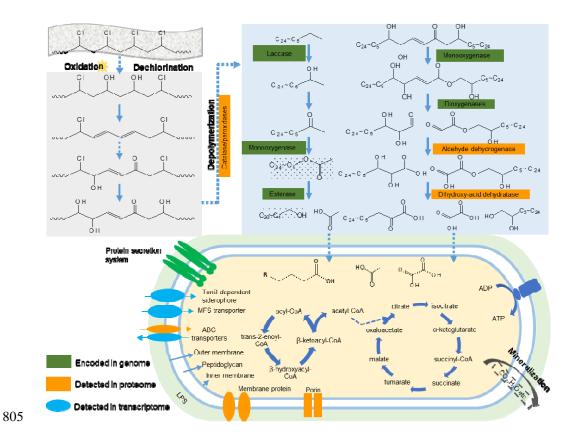
788 Figure 4 Proteomic and transcriptomic analyses of PVC-degrading metabolic functions 789 of strain EMBL-1. a-b, the weight loss of PVC film under the actions of strain EMBL-790 without and with glucose (glu) as supplementary carbon source (a), and four intracellular (IN) 791 or extracellular (OUT) protein extracts under in vitro conditions (b), respectively. c, venn 792 diagram showing shared and unique proteins in the protein extracts; d, 39 proteins involved in 793 intracellular and extracellular metabolism of PVC film. The five most abundant proteins in 794 the OUT extracts were marked in bold, while those were more upregulated in OUT extracts 795 than IN extracts (OUT/IN≥8) were marked in blue. The proteins marked in red were detected 796 in both proteomes and transcriptomes. e, strain EMBL-1 cell culture concentration (as OD_{600}) 797 using PVC film as sole carbon source; f, volcano map showing protein-coding genes (read 798 spheres) significantly upregulated in the PVC group than control group. g, functional 799 annotation of up-regulated genes with those marked in red also detected in the proteome.



801 Figure 5 Schematic diagram of the putative PVC degradation pathway by strain

802 EMBL-1. The pathway was proposed based on the multi-omics analyses that integrate

- 803 genomic, proteomic, transcriptomic and metabolic results of the strain during growth on PVC
- 804 film.



806 Table 1 Literature review showing limited understanding on degrading microbial strains and degrading

807 enzymes of PVC relative to PE and PET. Strain EMBL-1 is the first experimentally verified PVC-degrading

808 Klebsiella isolate by this study.

	Enzyme name	Source	Functional validation	Protein-coding gene ID in the genome of strain EMBL-1
	alkane hydroxylase ¹	Pseudomonas sp. E4	Recombinant strains (E coli)	
	laccases ^{2,3}	Rhodococcus ruber	Crude culture supernatant	orf01799, orf03107
	lignin peroxidases ⁴	Phanerochaete chrysosporium	Partially purified enzyme used	
PE	alkane monooxygenase ⁵	Pseudomonas aeruginosa	Recombinant strains (E coli)	orf03472
	rubredoxin reductase ⁵	Pseudomonas putida	Recombinant strains (E coli)	
	Manganese peroxidase ^{6,7}	Phanerochaete chrysosporium	Partially purified enzyme used	orf03592
	soybean peroxidase ⁸	Commercial enzyme	Purified enzyme used	
	PETase ⁹⁻¹¹	Ideonella sakaiensis	Purified enzyme used	
	lipases ¹²	Ideonella sakaiensis	Purified enzyme used	orf03181
	esterase ^{12,13}	Ideonella sakaiensis	Purified enzyme used	orf00935,orf01938,orf03042,orf06956
PET	cutinases ^{12,14}	Ideonella sakaiensis , Humilica insolens, Pseudomonas mendocina, Fusarium solani	Purified enzyme used	
	carboxylesterases ^{15,16}	Thermobifida fusca, Bacillus licheniformis, Bacillus subtilis , Thermobifida fusca	Purified enzyme used	orf00475, orf03403
PVC	lignin peroxidase ¹⁷	Phanerocheate chrysosporium	Partially purified enzyme used	