

1 Frequency-dependent hybridization contributes to habitat  
2 segregation in monkeyflowers

3 Katherine Toll<sup>1,\*</sup>

4 David B. Lowry<sup>1,2</sup>

5 1. Department of Plant Biology, Michigan State University, East Lansing, Michigan 48824

6 2. Program in Ecology, Evolution, and Behavior, Michigan State University, East Lansing,  
7 Michigan 48824

8 \* Corresponding author; email: [tollkath@gmail.com](mailto:tollkath@gmail.com)

9

10 *Manuscript elements:* Figures (6), Supplemental Tables (6)

11 *Keywords:* fecundity, frequency-dependence, hybridization, reproductive interference, habitat  
12 segregation, zonation, *Mimulus*, monkeyflower

13 *Manuscript type:* Article.

14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31

## Abstract

Spatial segregation of closely related species is usually attributed to differences in stress tolerance and competitive ability. For both animals and plants, reproductive interactions between close relatives can impose a fitness cost that is more detrimental to the rarer species. Frequency-dependent mating interactions may thus prevent the establishment of immigrants within heterospecific populations, maintaining spatial segregation of species. Despite strong spatial segregation in natural populations, two sympatric California monkeyflowers (*Mimulus nudatus* and *M. guttatus*) survive and reproduce in the other's habitat when transplanted reciprocally. We hypothesized that a frequency-dependent mating disadvantage maintains spatial segregation of these monkeyflowers during natural immigration. To evaluate this hypothesis, we performed two field experiments. First, we experimentally added immigrants in varying numbers to sites dominated by heterospecifics. Second, we reciprocally transplanted arrays of varying resident and immigrant frequency. Immigrant seed viability decreased with conspecific rarity for *M. guttatus*, but not *M. nudatus*. We observed immigrant minority disadvantage for both species, but driven by different factors— frequency-dependent hybridization for *M. guttatus*, and competition for resources and/or pollinators for *M. nudatus*. Overall, our results suggest a potential major role for reproductive interference in spatial segregation that should be evaluated along with stress tolerance and competitive ability.

32

## Introduction

33           Patterns of habitat segregation are common in closely related species and have been long  
34 mechanistically attributed to trade-offs between stress tolerance and competitive ability (Baker  
35 1909, Connell 1961, Grace and Wetzel 1981, Yost et al. 2012, Peterson et al. 2013, Chen and  
36 Schemske 2015, Ferris and Willis 2018). Competition is well established to be an important  
37 mechanism driving zonation among closely related species with a high potential for niche  
38 overlap (Burns and Strauss 2011). However, mating between closely related species may also  
39 contribute to their spatial segregation (Anacker and Strauss 2014). The fitness costs of  
40 interspecific mating (reproductive interference) can drive patterns of segregation or coexistence  
41 through demographic displacement (Groning and Hochkirch 2008) or the evolution of increased  
42 pre-zygotic isolation in sympatry (Servedio and Noor 2003). The ecological and evolutionary  
43 outcomes of costly interspecific mating – spatial segregation (Groning and Hochkirch 2008) or  
44 character displacement (Brown and Wilson 1956) – depend on the relative rate of evolutionary  
45 change versus the rate of demographic decline in sympatry (Kyogoku and Wheatcroft 2020).  
46 While the initial establishment of habitat segregation could result from exclusion in sympatry or  
47 evolve directly in response to reproductive interference (Kyogoku and Kokko 2020), the long-  
48 term maintenance of zonation patterns requires mechanisms that result in the demographic  
49 decline of rare immigrants.

50           Plant species that co-occur and overlap in flowering often compete for pollinators and  
51 often suffer consequent fitness reductions (Mitchell et al. 2009). Competition for pollination can  
52 reduce fitness through reductions in visitation or through interspecific pollen transfer (Waser  
53 1978 a, b). Both mechanisms of competition for pollination can affect fitness in a frequency-  
54 dependent manner. Rarer species may receive fewer visits and are expected to receive less

55 conspecific pollen and more heterospecific pollen than common species. Models of reproductive  
56 interference thus predict fitness declines with increasing rarity (Levin and Anderson 1970, Kuno  
57 1992). Rare-species disadvantage caused by reproductive interference is major factor limiting  
58 persistence of polyploid species that arise in populations of their diploid progenitors (minority  
59 cytotype exclusion, Levin 1975). Thus, considerable focus has been placed on these dynamics in  
60 co-occurring diploid and polyploid species (Husband 2000, Baack 2005, Buggs and Pannell  
61 2006). However, the conditions experienced by rare polyploids are similar to that of parapatric  
62 species that meet in secondary contact (Lewis 1961, Ribeiro and Spielman 1986, Thum 2007),  
63 immigrants of species that are spatially segregated in regional sympatry (Singer 1990, Friberg et  
64 al 2013, Christie and Strauss 2020), or nonnative species interacting with native congeners  
65 (Takakura et al. 2009, Takakura and Fujii 2010, Takakura 2013, Takakura and Fujii 2015).

66 An extreme cost to reproductive interactions is the production of inviable seeds from  
67 interspecific pollen transfer. This hybrid seed inviability is a common reproductive isolating  
68 barrier in plants and particularly among species in the *Mimulus guttatus* species complex  
69 (Coughlan et al. 2020, Garner et al. 2016, Oneal et al. 2016, Sandstedt et al. 2021, Vickery  
70 1978). Inviabile hybrid seeds produced between species in the *M. guttatus* species complex have  
71 a characteristic flat or shriveled shape, which is caused by arrested endosperm development.  
72 These inviable seeds can be distinguished by eye from round viable seeds with fully formed  
73 endosperm produced by conspecific pollinations.

74 Populations of the widespread *M. guttatus* have repeatedly evolved tolerance to harsh  
75 serpentine soils (Selby and Willis 2018) and co-occur with the geographically restricted  
76 serpentine soil endemic *M. nudatus* in the Northern coast range of California. *M. guttatus*  
77 typically grows in streams, seeps, or meadows, whereas *M. nudatus* typically grows in washes,

78 on rock outcroppings, or drier rocky areas adjacent to the streams inhabited by *M. guttatus*. *M.*  
79 *nudatus* typically grows in barer habitats with fewer co-occurring con- and heterospecifics than  
80 *M. guttatus*, a common pattern in serpentine soil endemics compared to species that occur both  
81 on and off serpentine soils (Sianta and Kay 2019). Since habitats occupied by *M. guttatus* and *M.*  
82 *nudatus* are often in close proximity (within meters) and immigrant individuals are regularly  
83 found, dispersal limitation does not explain spatial segregation in these species. When  
84 transplanted reciprocally at equal frequencies, *M. guttatus* has higher survival and produces more  
85 seed than *M. nudatus* in both habitats, demonstrating that stress tolerance or competition are not  
86 sufficient to explain spatial segregation patterns either (Toll and Willis 2018). In these reciprocal  
87 transplants, both species produced many inviable F1 hybrid seeds as immigrants, demonstrating  
88 the occurrence reproductive interference (Toll and Willis 2018). While a combination of factors  
89 limits the persistence of *M. nudatus* immigrants in habitats occupied by *M. guttatus*, including a  
90 lack of tolerance to flooding in *M. nudatus* (Toll and Willis 2018) and inbreeding depression  
91 (Toll et al. 2021), these factors do not sufficiently explain the absence of *M. guttatus* in the drier  
92 habitats occupied by *M. nudatus*. We therefore hypothesize that frequency-dependent  
93 reproductive interference may maintain this spatial segregation pattern by imposing a strong  
94 fitness cost on natural immigrants, which usually occur alone or in very small numbers.

95       Using observational and experimental data collected across two field seasons, we asked  
96 three questions in this study. First, does hybridization reduce fecundity when *M. guttatus* and *M.*  
97 *nudatus* naturally occur in close proximity? If hybridization contributes to the maintenance of  
98 spatial segregation, we expect that seed viability will increase with increasing conspecific  
99 frequency. Second, when species immigrate to sites occupied by heterospecifics, do components  
100 of fitness in immigrants or residents (flowers per plant, seeds per flower, and seed viability)

101 increase with their relative frequency? If resource competition contributes to the maintenance of  
102 habitat segregation, we expect that flowers per plant will increase with increasing conspecific  
103 frequency. If a lack of conspecific mates or pollinator competition contributes to the maintenance  
104 of habitat segregation, we expect that seeds per flower will increase with increasing conspecific  
105 frequency. Finally, if hybridization contributes to the maintenance of habitat segregation, we  
106 expect that seed viability will increase with increasing conspecific frequency. Third, does  
107 lifetime fecundity (measured as viable seeds per plant) in immigrants or residents increase with  
108 increasing conspecific frequency in experimental sympatry? If minority disadvantage contributes  
109 to the maintenance habitat segregation, we would expect that viable seeds per immigrant or  
110 resident plant to increase with their increasing frequency, changing the relative performance of  
111 immigrants relative to residents.

112

113

## Materials and methods

114

### *Reproductive interference in natural populations: an observational transect*

115

116

117

118

119

120

121

122

123

To characterize the existence and possible magnitude of reproductive interference in co-occurring populations of *M. guttatus* and *M. nudatus*, we surveyed plants along a transect spanning an exposed serpentine outcrop that was inhabited by a pure stand of *M. nudatus*, a transition zone where both species were present at lower density, terminating in a serpentine seep with a pure stand of *M. guttatus* (38°51.528' N, 122°24.614' W). The entire length of the transect, spanning the two pure stands through a small zone of overlap, was six meters. We counted all the fruit produced by each species within eight 42 x 66 cm gridded quadrats centered on the transect line, with the long side parallel to the transect line and evenly spaced along the length of the transect. We counted all fruit present within the quadrat at 7cm intervals along the

124 long side of the quadrat grid. At approximately 21cm intervals along the long side of the quadrat,  
125 we collected one plant if only one species was present (n=13 *M. nudatus* and n=9 *M. guttatus*) or  
126 two plants if both species were present (n=10 *M. nudatus* and n=10 *M. guttatus*) at the mid-point  
127 of the short side of the quadrat. We counted fruit numbers instead of species abundances because  
128 flower number, not plant number, is what pollinators in dense overlapping stands (Gardner and  
129 MacNair 2000). We counted all seeds produced by the plants collected at 21 cm intervals and  
130 categorized them as round viable (conspecific) or flat inviable F1 hybrid seeds based on  
131 morphology (Oneal et al. 2016). We calculated seed viability by dividing the number of round  
132 viable conspecific seeds by the total number of seeds produced by each plant. We estimated  
133 Pearson correlation coefficients for the relationship between the proportion of conspecific fruit  
134 and seed viability per sampled plant in R v 4.0.3 (R Core Team 2020).

135

### 136 *Experimental Transplants*

137 To test whether frequency-dependence contributes to the maintenance of habitat  
138 segregation, we performed two reciprocal transplants over two years. In the first field season  
139 (2018), we transplanted immigrants in varying abundances into natural populations of resident  
140 heterospecifics. This allowed us to simulate a situation where immigrants arrive at low  
141 abundances to sites already established by heterospecifics. However, since we did not transplant  
142 the naturally occurring resident species with the immigrants, we lacked an appropriate  
143 comparison for immigrant performance relative to residents. In the second field season (2019),  
144 we transplanted immigrants and residents at varying frequencies reciprocally, allowing us to  
145 compare the overall fecundities of experimental immigrants and residents.

146

147 *M. guttatus* immigration experiment

148           In February 2018, we transplanted variable numbers of immigrants into naturally dense  
149 patches of native seedlings (>100 per half meter). While we performed this experiment  
150 reciprocally, all but nine immigrant *M. nudatus* transplants died before flowering. We lacked  
151 sufficient seedlings to replace them, and they were not analyzed further. Therefore, this  
152 experiment only allowed us to examine the immigration dynamics of *M. guttatus* into *M. nudatus*  
153 habitat (the direction less-well explained by our previous studies).

154           We collected seeds from a local population of *M. guttatus* at the University of California  
155 McLaughlin reserve (38°51.528' N, 122°24.614' W) in 2013. We grew seeds from field-collected  
156 maternal families in a greenhouse and crossed a single plant from each family to a single plant  
157 from a different family to produce outbred maternal families in 2013 (described in more detail in  
158 Toll et al. 2021). To minimize effects of inbreeding on fecundity (see Toll et al 2021), we  
159 germinated 6 outbred maternal families of *M. guttatus* to use as focal plants. These plants were  
160 initially grown in the shade house at the UC McLaughlin reserve in 2018. We did not have  
161 sufficient seed to use these same outbred lines as neighbors, so we pooled equal numbers of seed  
162 from twenty field-collected maternal families of *M. guttatus* to germinate in the shade house to  
163 use as neighbors. We planted 72 blocks of focal *M. guttatus* plants in random order by maternal  
164 family, then randomly assigned each family to one of the 6 levels of the initial conspecific  
165 neighbor treatments (0, 8, 24, 48, 80, and 120 neighbors, Figure 1A). We transplanted a range of  
166 seedlings that were ecologically relevant if a *M. guttatus* fruit dispersed into *M. nudatus* habitat,  
167 given that each fruit produces an average of 200 seeds (Toll and Willis 2018). *M. guttatus*  
168 seedlings were planted within naturally dense patches of *M. nudatus* (>100 native *M. nudatus*  
169 within a half meter of each focal *M. guttatus* plant, Figure 1B). We planted *M. guttatus* neighbors



170 within a half-meter of each focal plant in a regularly spaced array and we planted these blocks  
171 over three meters away from one another.

172 We replaced *M. guttatus* focal transplants as they died of transplant shock, resulting in 62  
173 focal plants (of the original 72 plus 28 replacements) that survived to flower by the end of the  
174 season. Neighbor transplant mortality rates were also high; 78% of *M. guttatus* transplants did  
175 not survive to flowering, resulting in a range of flowering neighbors from 0 to 103, with a  
176 median of 1 neighbor surviving to flowering. In subsequent analyses of the effect of conspecific  
177 neighbors on components of fitness, we only included neighbors that successfully survived to  
178 flower, as only they could potentially reduce reproductive interference. Mortality rates from this  
179 transplant experiment are consistent with previous estimates at this site: 73%, 33%, and 82% of  
180 *M. guttatus* seedlings died at this site in 2015, 2016, and 2017 respectively (site “QV” in Toll  
181 and Willis 2018). Thus, we are confident that our transplants represent an ecologically realistic  
182 range of potential surviving immigrant *M. guttatus* seedlings into *M. nudatus* habitat. We  
183 counted all seeds produced by the 62 focal plants that produced fruit and categorized them as  
184 viable round conspecific-pollinated or inviable flat heterospecific-pollinated (hybrid) seeds based  
185 on morphology.

186 *Statistical analysis.* To examine how the number of conspecific neighbors affected focal  
187 plant fitness components (flowers per plant, total seed per flower, and seed viability per focal *M.*  
188 *guttatus* transplant) we used generalized linear mixed models in the R package *glmmTMB*  
189 (Brooks et al. 2017). We fit models with each fitness component as the dependent variable, the  
190 number of surviving conspecific neighbors as a fixed effect, and maternal family (n=6) of the  
191 focal plant as a random effect. Models of total seeds per flower had an additional nested random  
192 effect term for individual plant (n=62) since each focal plant produced multiple flowers.

193 We fit each count model first with a Poisson error distribution and identified the best  
194 fitting error distribution by evaluating model diagnostics with the R package DHARMA (Hartig  
195 2019). The effect of conspecific neighbors on flowers per focal transplant was estimated using a  
196 negative binomial model because Poisson models were significantly overdispersed. The effect of  
197 conspecific neighbors on seeds per flower was estimated using zero-inflated negative binomial  
198 models with a constant intercept because Poisson and negative binomial models were  
199 significantly overdispersed and zero-inflated. The effect of conspecific neighbors on seed  
200 viability was estimated using a beta-binomial model because the binomial model was  
201 significantly overdispersed.

202 We assessed the significance of the fixed effect (the number conspecific neighbors) with  
203 the best fitting model for each fitness component using Wald Type II Chi-squared tests (R  
204 package *car*, Fox and Weisberg 2019). We predicted marginal effects and bootstrapped (n=500  
205 iterations) 95% confidence intervals with the R package *ggeffects* (Lüdtke 2018). We plotted  
206 raw data and predictions with *ggplot2* (Wickham 2016) and combined plots with *patchwork*  
207 (Pederson 2020).

208

### 209 *Frequency-manipulation reciprocal transplant*

210 To test whether frequency-dependence contributes to the maintenance of habitat  
211 segregation, we reciprocally transplanted *M. guttatus* and *M. nudatus* at different frequencies but  
212 with the same number of individuals in experimental patches. These mixed species patches were  
213 transplanted into a seep dominated by *M. guttatus* and a wash dominated by *M. nudatus*. The  
214 plants used in this experiment were all collected directly from natural populations at each  
215 transplant site in the year (2019) of the experiment. In April 2019, we dug up 380 seedlings per

216 species from each site and repotted them in yellow 49-mL (2.54 cm diameter) Cone-tainers™  
217 filled with potting soil (Stuewe and Sons, Inc., Tangent, OR, USA). We maintained seedlings in  
218 a shadehouse at the UC McLaughlin Reserve for a week prior to redistributing seedlings in  
219 Cone-tainers™ at both sites. We sunk Cone-tainers™ into the ground, so the top was flush with  
220 the soil surface at each site in 19 plant hexagonal arrays (Figure 1C). Arrays were planted within  
221 natural populations of each species, but we cleared native *Mimulus* plants within 0.25 m of each  
222 array. We designed this experiment following the frequency manipulation arrays from Nagy  
223 (1997): 0% immigrants (0 immigrants, 19 residents); 5% immigrants (1 immigrant, 18 residents);  
224 ~25% immigrants (4 immigrants, 15 residents); ~50% immigrants (9 immigrants, 10 residents)  
225 (Figure 1D). Higher levels of immigrants than natives would have been unrealistic in this system.  
226 For clarity, we abbreviate frequency treatments as the ratio of *M. guttatus* (G) and *M. nudatus*  
227 (N) transplants. We assigned plants to four frequency treatments in the seep site (19G, 18G:1N,  
228 15G:4N, and 10G:9N) which is naturally occupied by *M. guttatus*, and four frequency treatments  
229 in the wash site (18N, 18N:1G, 15N:4G, and 10N:9G) which is naturally occupied by *M.*  
230 *nudatus*. We planted 5 replicates per frequency treatment per site, for a total of 20 blocks per site  
231 and 40 blocks overall. Blocks were planted at least three meters away from each other. We added  
232 supplemental water to blocks for a month after transplanting and replaced seedlings that died  
233 prior to flowering to ensure that the same number of plants flowered in each frequency treatment.  
234 We left containers in the field until all plants terminated flowering and ripened their fruit. We  
235 collected fruit from each transplant in the field prior to digging up containers. We counted and  
236 categorized all seeds produced by experimental plants and estimated the hybridization rate using  
237 the same methods as the observational transect. We estimated lifetime fecundity (viable seeds  
238 per plant) from three fitness components: the total number of flowers produced per plant, the

239 total number of seeds produced per flower, and seed viability (=viable seeds/total seeds  
240 produced).

241 *Statistical analysis.* To examine how the immigrant frequency treatments affected plant  
242 fitness components (flowers per plant, total seed per flower, and seed viability) and lifetime  
243 fecundity (viable seeds per plant), we used generalized linear mixed models in the R package  
244 *glmmTMB* (Brooks et al. 2017). We tested whether fitness components differed among frequency  
245 treatments for each species in each site separately because no immigrants were transplanted in  
246 the 0% immigrant frequency treatment (Figure 1D). We fit models with each fitness component  
247 as the dependent variable, immigrant frequency as a fixed effect and block as a random effect  
248 (block n=15 for immigrants, n=20 for residents). We also tested whether lifetime fecundity  
249 differed between species, but only within the treatments where both immigrants and residents  
250 were present (5%, 25%, and 50% immigrant frequency treatments) in each site separately. We fit  
251 models with lifetime fecundity as the dependent variable, immigrant frequency, species and their  
252 interaction as fixed effects and block as a random effect (block n=15).

253 Poisson models of count fitness components and lifetime fecundity were significantly  
254 overdispersed, and models of seeds per flower and viable seeds per plant were significantly zero-  
255 inflated. Thus, the effect of immigrant frequency on flowers per plant was estimated using a  
256 negative binomial distribution, while total seeds per flower and viable seeds per plant were  
257 estimated using a zero-inflated negative binomial distribution with a constant zero-inflation  
258 intercept applied to all observations. The effect of immigrant frequency on seed viability was  
259 analyzed using a binomial model for immigrants, and a beta-binomial model for residents  
260 because binomial models for residents were significantly overdispersed.

261 We assessed the significance of the fixed effect (immigrant frequency treatment) with the  
262 best fitting model for each fitness component using Wald Type II Chi-squared tests and assessed  
263 the significance of the fixed effects (immigrant frequency treatment, species, and their  
264 interaction) for lifetime fecundity using Wald Type III Chi-squared tests (R package *car*, Fox  
265 and Weisberg 2019). We tested whether fitness components differed among treatments within  
266 species, and whether species differed in lifetime fecundity within each frequency treatment with  
267 Tukey post-hoc contrasts (R package *emmeans*, Lenth 2020). We estimated marginal means and  
268 bootstrapped (n=500 iterations) 95% confidence intervals with the R package *ggeffects* (Lüdtke  
269 2018). We plotted marginal means with *ggplot2* (Wickham 2016) and combined plots with  
270 *patchwork* (Pederson 2020).

271

272

## Results

273

### *Reproductive interference in natural populations: an observational transect*

274

275

276

277

278

279

280

281

282

283

Individuals collected along a transect produced a high proportion of viable seeds in zones where only conspecifics were present (average 0.92 for *M. guttatus*; 0.84 for *M. nudatus*), even though heterospecifics were in very close proximity (always within 3m). In contrast, individuals collected in the 3m overlap zone of the two species produced far lower proportions of viable seeds (average 0.53 in *M. guttatus*; 0.66 in *M. nudatus*, Figure 2). Seed viability was positively correlated with the local frequency of conspecific fruit for *M. guttatus* plant sampled along the transect (Pearson correlation coefficient= 0.66; 95% CI= 0.29, 0.86;  $t = 3.61$ ,  $df=17$ ,  $p = 0.002$ ), meaning that greater numbers of adjacent conspecific flowers and fruits were associated with greater proportions of viable seeds. However, seed viability was not significantly correlated with the local frequency of conspecific fruit *M. nudatus* plants sampled along the transect (Pearson

284 correlation coefficient= 0.18; 95% CI= -0.28, 0.57;  $t = 0.79$ ;  $df = 19$ ;  $p = 0.44$ ). These data from  
285 a natural transition zone suggest that reproductive interference is more likely to be frequency-  
286 dependent in *M. guttatus* immigrants than *M. nudatus* immigrants.

287

### 288 *Experimental Transplants*

289 Seed viability was positively associated with immigrant frequency for immigrant *M.*  
290 *guttatus* transplants, but not immigrant *M. nudatus* transplants, consistent with the findings from  
291 the observational transect. Resident transplants of both species also suffered from decreased seed  
292 viability as immigrant transplants increased in frequency. Unexpectedly, fitness components  
293 declined for immigrant *M. nudatus* transplants when they were most common, resulting in  
294 immigrant disadvantage in the highest frequency treatment. Immigrant transplants had lower  
295 lifetime fecundity, measured as viable seeds per plant, than resident transplants when immigrants  
296 were rarest. For immigrant *M. guttatus*, this rarity disadvantage could not be explained by fitness  
297 component differences prior to hybridization. In contrast, rarity disadvantage for immigrant *M.*  
298 *nudatus* was observed for fitness components prior to hybridization.

299

### 300 *M. guttatus* immigration experiment

301 As expected from the observational transect, *M. guttatus* immigrant transplants  
302 hybridized less and produced a greater proportion of viable conspecific seeds when surrounded  
303 by more flowering conspecific neighbor transplants (Figure 3C), as would occur if an intact fruit  
304 or multiple seeds colonized a habitat patch occupied by heterospecifics. Seed viability increased  
305 with increasing numbers of conspecific neighbors (Wald type II Chi-square test:  $\chi^2 = 7.49$ ,  $df =$   
306  $1$ ,  $p = 0.006$ ; coefficient estimate = 0.023, SE = 0.009, z-value = -2.74,  $p = 0.006$ ). The model-

307 predicted seed viability for rare immigrants increased from 0.24 for immigrants without  
308 conspecific neighbors, to 0.78 in the plant with 103 neighbors (Figure 3C).

309 Flowers per plant and total seeds per flower for focal immigrant *M. guttatus* transplants  
310 were not significantly associated with the number of flowering conspecific neighbor transplants  
311 (Figure 3A, 3B; Wald type II Chi-square tests: flowers per plant  $\chi^2 = 0.11$ ,  $df = 1$ ,  $p = 0.73$ ;  
312 seeds per flower  $\chi^2 = 0.52$ ,  $df = 1$ ,  $p = 0.47$ ).

313

314 *Frequency-manipulation reciprocal transplant*

315 *Fitness components in wash.* For immigrant *M. guttatus* transplants in the wash, seed  
316 viability was significantly associated with the frequency treatment (Fig 4E, Wald type II Chi-  
317 square test:  $\chi^2 = 10.602$ ,  $df = 2$ ,  $p = 0.005$ ). In contrast, flowers per plant (Fig 4A, Wald type II  
318 Chi-square test:  $\chi^2 = 1.75$ ,  $df = 2$ ,  $p = 0.42$ ) and seeds per flower (Fig 4C, Wald type II Chi-  
319 square test:  $\chi^2 = 0.44$ ,  $df = 2$ ,  $p = 0.80$ ) were not associated with the frequency treatment. Seed  
320 viability in immigrant *M. guttatus* transplants increased from 0.14 to 0.48 between the 5%  
321 (1G:18N) and 50% (9G:10N) immigrant frequency treatments (Tukey Post-hoc contrast:  $p =$   
322 0.006, Table S1).

323 For resident *M. nudatus* transplants in the wash, seed viability was significantly  
324 associated with the frequency treatment (Fig 4F, Wald type II Chi-square test:  $\chi^2 = 11.40$ ,  $df = 3$ ,  
325  $p = 0.01$ ), but flowers per plant (Fig 4B, Wald type II Chi-square test:  $\chi^2 = 3.20$ ,  $df = 3$ ,  $p = 0.36$ )  
326 and seeds per flower were not (Fig 4D, Wald type II Chi-square test:  $\chi^2 = 6.47$ ,  $df = 3$ ,  $p = 0.09$ ).  
327 Seed viability in resident *M. nudatus* transplants decreased from 0.83 to 0.68 between the 0%  
328 (1G:18N) and 25% (9G:10N) immigrant frequency treatments (Tukey Post-hoc contrast:  $p =$   
329 0.005, Table S2).

330

331 *Fitness components in seep.* For immigrant *M. nudatus* transplants in the seep, flowers  
332 per plant (Fig 5A, Wald type II Chi-square test:  $\chi^2 = 11.74$ ,  $df = 2$ ,  $p = 0.003$ ) and seeds per  
333 flower (Fig 5C, Wald type II Chi-square test:  $\chi^2 = 12.21$ ,  $df = 2$ ,  $p = 0.002$ ) were significantly  
334 associated with frequency treatment, but seed viability was not (Fig 5E, Wald type II Chi-square  
335 test:  $\chi^2 = 1.52$ ,  $df = 2$ ,  $p = 0.47$ ). Flowers per plant in immigrant *M. nudatus* transplants  
336 decreased from ~6 to ~2 flowers per plant between the 25% (4N:15G) and 50% (9N:10G)  
337 immigrant frequency treatments (Tukey post-hoc contrast:  $p = 0.005$ , Table S3). This decline  
338 resulted in a lower proportion of *M. nudatus* flowers than expected in the highest immigrant  
339 frequency treatment (Proportion immigrant *M. nudatus* flowers by immigrant frequency  
340 treatment: 5% treatment (1N:18G) = 0.06, 25% (4N:15G) treatment = 0.33, 50% (9N:10G) =  
341 0.29). Seeds per flower in immigrant *M. nudatus* transplants increased from 27 to 64 seeds per  
342 flower between the 5% (1N:18G) and 25% (4N:15G) immigrant frequency treatments and  
343 decreased from 64 to 29 between the 25% (4N:15G) and 50% (9N:10G) immigrant frequency  
344 treatments (Tukey post-hoc contrasts:  $p < 0.05$ , Table S3).

345 For resident *M. guttatus* transplants in the seep, seed viability was significantly associated  
346 with frequency treatment (Fig 5F, Wald type II Chi-square test:  $\chi^2 = 21.56$ ,  $df = 3$ ,  $p < 0.001$ ),  
347 but flowers per plant (Fig 5B, Wald type II Chi-square test:  $\chi^2 = 3.14$ ,  $df = 3$ ,  $p = 0.37$ ) and seeds  
348 per flower were not (Fig 5D, Wald type II Chi-square test:  $\chi^2 = 5.80$ ,  $df = 3$ ,  $p = 0.122$ ). Seed  
349 viability decreased from 0.84 to 0.60 between the 0% (0N:19G) and 50% (9N:10G) immigrant  
350 frequency treatments (Tukey post-hoc contrast:  $p < 0.001$ , Table S4) and decreased from 0.81 to  
351 0.60 between the 5% (1N:18G) and 50% (9N:10G) immigrant frequency treatments (Tukey post-  
352 hoc contrast:  $p = 0.001$ , Table S4)



353

354 *Lifetime fecundity in wash.* Viable seeds per plant in the wash was significantly  
355 associated with immigrant frequency treatment, species, and an immigrant frequency by species  
356 interaction (Fig 6A, Wald type III Chi-square test: frequency  $\chi^2 = 6.83$ ,  $df = 2$ ,  $p = 0.03$ ; species  
357  $\chi^2 = 12.14$ ,  $df = 1$ ,  $p < 0.001$ ; frequency x species  $\chi^2 = 14.12$ ,  $df = 2$ ,  $p < 0.001$ ). Immigrant *M.*  
358 *guttatus* transplants produced one-sixth as many viable seeds per plant as resident *M. nudatus*  
359 transplants in the 5% (1N:18G) immigrant frequency treatment (60 vs 347; Tukey post-hoc  
360 contrast  $p = 0.008$ , Table S5).

361

362 *Lifetime fecundity in seep.* Viable seeds per plant in the seep was significantly associated  
363 with species and an immigrant frequency treatment by species interaction (Fig 6B, Wald type III  
364 Chi-square test: frequency  $\chi^2 = 2.14$ ,  $df = 2$ ,  $p = 0.34$ ; species  $\chi^2 = 18.85$ ,  $df = 1$ ,  $p < 0.001$ ;  
365 frequency x species  $\chi^2 = 15.21$ ,  $df = 2$ ,  $p < 0.001$ ). Immigrant *M. nudatus* transplants produced  
366 one-twelfth as many viable seeds per plant as resident *M. guttatus* transplants in the 5%  
367 (1N:18G) immigrant frequency treatment (15 vs 187, Tukey post-hoc contrast  $p < 0.001$ ) and  
368 one-sixth as many in the 50% (9N:10G) treatment (19 vs 111, Tukey post-hoc contrast  $p = 0.001$ ,  
369 Table S6).

370

371

## Discussion

372

373

374

375

In this study, we found that frequency-dependent hybridization likely contributes to habitat segregation between two closely related monkeyflower species. Each species suffered from hybridization when in close contact with heterospecifics in both field experiments as well as in natural populations. Immigrants of both species had lower lifetime fecundity than residents

376 when rare. However, this minority disadvantage was caused by different factors for each species.  
377 Partitioning of the fitness components demonstrated that rarity disadvantage for *M. guttatus*  
378 immigrants was most consistent with reproductive interference, whereas aboveground resource  
379 and/or pollinator competition together with reproductive interference all likely contributed to  
380 rarity disadvantage for *M. nudatus* immigrants. *M. nudatus* immigrants also suffered fitness  
381 reductions consistent with intra-specific competition when they were at their highest frequency.  
382 This negative frequency dependence is predicted to facilitate immigrant invasion and break down  
383 spatial segregation (Grainger et al. 2019). However, since it was identified only for *M. nudatus*,  
384 it is unlikely to lead to breakdown. Our previous studies in this system identified flooding and  
385 inbreeding depression as major factors limiting the long-term establishment of *M. nudatus* in  
386 habitats occupied by *M. guttatus* (Toll and Willis 2018, Toll et al. 2021). The strong rarity  
387 disadvantage for *M. guttatus* immigrants is further evidence that multiple asymmetric processes  
388 contribute to the maintenance of strong spatial segregation between these closely related species.

389 In foundational studies on intertidal algae and barnacles, stress tolerance and competition  
390 differed between species, determining zonation patterns (Baker 1909, Connell 1961). The factors  
391 limiting each species are mechanistically different in our study, but the asymmetric nature of the  
392 limitation is fundamentally similar to the mechanisms of zonation found in those classic studies.  
393 For example, we found that a strong barrier limiting *M. nudatus* persistence in wetter sites,  
394 occupied by *M. guttatus*, was its inability to survive seasonal flooding (Toll and Willis 2018).  
395 Persistent flooding occurs during the winter rainy season in the seeps and meadows inhabited by  
396 *M. guttatus*. In contrast, flooding is uncommon in quickly draining washes and rocky outcrops  
397 inhabited by *M. nudatus*. An additional asymmetrical barrier limiting immigrant persistence in  
398 this system is inbreeding depression (Toll et al. 2021). Since immigrants have fewer

399 opportunities to mate with conspecifics, they likely produce most viable seeds through selfing.  
400 While recurrent selfing reduces fitness for both species, the reduction is greater for *M. nudatus*  
401 especially in foreign habitats. Selfing reduces immigrant *M. nudatus*' fitness to zero in fewer  
402 generations than it does for immigrant *M. guttatus* (Toll et al. 2021). Combined, these prior  
403 studies provided strong mechanistic rationale for why *M. nudatus* does not invade *M. guttatus*  
404 streams. However, why *M. guttatus* does not invade *M. nudatus* washes was still largely  
405 unexplained, and the experiments presented in this study suggest frequency-dependent  
406 hybridization is a contributing, and probably dominant, factor.

407

#### 408 *Frequency dependence of hybridization*

409 This study identifies frequency-dependent hybridization as a strong barrier limiting the  
410 persistence of *M. guttatus* in habitats occupied by *M. nudatus* (Figure 4E). This result is similar  
411 to that of reciprocal transplant experiments of *Limnanthus* species and *Gilia* subspecies, where  
412 frequency-dependent reductions in fecundity for one species (or subspecies) but not the other  
413 was attributed to asymmetric impacts of heterospecific pollen transfer on seed set (Nagy 1997,  
414 Runquist 2012, Runquist and Stanton 2012). In contrast, Christie and Strauss (2020) observed  
415 symmetric declines in seed viability of reciprocally transplanted *Streptanthus* species, even  
416 though experimental crosses reduced seed set asymmetrically (Christie and Strauss 2019). Since  
417 heterospecific pollen transfer between *M. nudatus* and *M. guttatus* reduces viable seed set  
418 symmetrically (Oneal et al. 2016), asymmetry of the frequency-dependence of seed viability is  
419 surprising. However, this asymmetry can be explained when we consider how fecundity  
420 components prior to hybridization changed across immigrant frequency treatments for *M.*  
421 *nudatus* immigrants. *M. nudatus* immigrants produced the fewest flowers per plant in the highest

422 immigrant frequency treatment, resulting in a lower proportion of *M. nudatus* gametes than  
423 suggested by the plant frequency treatment (~50% plants vs ~29% flowers). *M. nudatus* has  
424 fewer ovules and pollen grains per flower than *M. guttatus* (Oneal et al. 2016, Ritland and  
425 Ritland 1989). This difference in pollen production further amplified the difference in effective  
426 gamete frequency than would be expected solely by the manipulation of the frequency of *M.*  
427 *nudatus*. Thus, while other studies were able to predict asymmetries in frequency-dependence  
428 based on lab crossing data (Runquist and Stanton 2012, Nagy 1997), the asymmetry we observe  
429 is driven by ecological interactions (the reduction in flower number with increasing conspecific  
430 frequency) and evolved life history differences (pollen production).

431         While naturally occurring *M. nudatus* also hybridizes with *M. guttatus*, seed viability was  
432 also not associated with conspecific fruit frequency in the observational transect (Figure 2). Our  
433 result differs from that of Gardner and MacNair (2000), who found a significant negative  
434 correlation between fruit frequency and seed viability in *M. nudatus* co-occurring with *M.*  
435 *guttatus*. Our studies occurred in the same region but in qualitatively different sites. Gardner and  
436 MacNair (2000) only sampled *M. nudatus* plants with heterospecifics in close proximity,  
437 equivalent to the middle portion of our transect where both species co-occurred but did not  
438 extend the transect into pure populations. Compared to *M. guttatus*, seed viability in the  
439 transition zone for *M. nudatus* was more variable (Figure 2), but since we did not record  
440 phenology or pollinator behavior over time, the factors contributing to this pattern remain open  
441 questions.

442         Our observations from the transect study suggests that fitness costs of hybridization are  
443 highly localized and decay within meters from the nearest heterospecific (Figure 2), which  
444 makes sense given the narrow boundaries that we see in most populations. In the reciprocal

445 transplant experiment, resident transplants of both species suffered decreased seed viability when  
446 immigrant transplants were at high frequencies (Figure 4F, 5F), despite the high abundance of  
447 native conspecifics for resident transplants at each site. These results suggest that low pollinator  
448 constancy leads to high rates of interspecific pollen transfer between species at the local scale,  
449 and that most pollination is occurring at a very small scale. These localized and frequency-  
450 dependent fitness costs to residents are consistent with the observed stability of spatial  
451 segregation patterns. Immigration rates would have to be extremely high, and immigrants would  
452 have to be widespread to negatively affect resident populations.

453

454 *Competition may contribute asymmetrically to habitat segregation*

455 The *M. guttatus*/*M. nudatus* study system presented a unique opportunity to separate  
456 effects of hybridization on fitness from effects of competition and abiotic factors. While the  
457 combined effects of reproductive and ecological niche overlap could amplify positive frequency-  
458 dependence and hasten exclusion (Schreiber et al. 2019), we instead observed declines in flowers  
459 per plant and seeds per flower for *M. nudatus* immigrants in their highest frequency treatment.  
460 The decline at high densities suggests that some form of competition was more important than  
461 release from reproductive interference for *M. nudatus* immigrants at high immigrant frequencies.  
462 As the transplants occurred in plastic containers, pre-empting below-ground competition,  
463 above-ground competition for space, light, and/or pollinator visits may have suppressed flower  
464 and seed production in these potted immigrants (Figure 1D). Due to the experimental design of  
465 our frequency arrays (Figure 1C), we were unable to estimate the strength of intra- vs  
466 interspecific competition (Inouye 2001).

467 Immigrant *M. nudatus* transplants produced the greatest number of seeds per flower in  
468 the intermediate immigrant frequency treatment (Figure 5C). We do not have a clear  
469 interpretation for this phenomenon. However, we hypothesize that pollinator behavior might  
470 have driven to this U-shaped pattern in fecundity. Benadi and Pauw (2018) observed that South  
471 African Fynbos species received the highest visitation rates at intermediate flower abundances,  
472 relative to rare and common species. Future studies on pollinator visitation rates are needed to  
473 test this hypothesis.

474 In greenhouse studies, Brigham (2003) failed to find deleterious effects of interspecific  
475 competition to either *M. guttatus* or *M. nudatus*. In the field, Brigham (2001) found that effects of  
476 intra-specific competition for *M. nudatus* varied by year, site and the timing of germination.  
477 Neighbor removals increased, decreased, or had no effect on *M. nudatus* survival (Brigham  
478 2001). In our study, we did not observe reductions in *M. guttatus* fecundity prior to hybridization  
479 (i.e., in flower or fruit set). These findings accord well with expected competitive ability of *M.*  
480 *guttatus*, which occur in streams that are full of both conspecifics as well as a high density of  
481 many other plant species. The rocky serpentine washes inhabited by *M. nudatus* are mostly  
482 sparsely vegetated, with only occasional areas of high density of monkeyflowers and very few  
483 other small forbs. This is consistent with the broader pattern of serpentine endemics occurring in  
484 barer (low competition) habitats compared to closely related species (Sianta and Kay 2019).  
485 Therefore, it is perhaps not surprising that in our arrays, which intentionally created a density  
486 somewhat intermediate and appropriate for both habitats, *M. nudatus* would suffer from  
487 competition.

488

489 *Reciprocal transplant experiments and the asymmetry of barriers*

490 Reciprocal transplant experiments provide strong evidence of local adaptation, whether  
491 between populations, ecotypes, or closely related species (Clausen et al. 1940, Kawecki and  
492 Ebert 2004, Hereford 2009). In annual plants, the response variable in these studies is usually  
493 seed set. Disentangling the relative effects of interspecific mating from environmental  
494 differences between habitats can be achieved by varying the frequencies of each species  
495 reciprocally in native and foreign sites. This allows comparisons across frequencies of fitness  
496 components that are not affected by interspecific mating interactions (e.g., survival, flower  
497 number) to those that could be (e.g., seed abortion, hybrid seed production, Nagy 1997, Runquist  
498 and Stanton 2013, Christie and Strauss 2020).

499 In reciprocal transplants, the fitness of local residents and foreign immigrants are  
500 compared at native and foreign sites and changes in the rank of fitness, where local residents  
501 outperform foreign immigrants at each site, is taken as evidence of adaptation to local conditions  
502 (Kawecki and Ebert 2004). Beyond differential habitat-based selection, transplanting individuals  
503 of closely related species reciprocally reduces the distance between heterospecific individuals.  
504 This in turn increases the chances of reproductive interference. While reciprocal transplant  
505 experiments are often used to explain habitat segregation between closely related species (e.g.,  
506 Yost et al. 2012, Peterson et al. 2013, Chen and Schemske 2015, Ferris and Willis 2018,  
507 DiVittorio et al. 2020), reproductive interactions are rarely considered as important drivers of  
508 these patterns (but see Toll and Willis 2018, Christie and Strauss 2020). Since heterospecific  
509 pollen transfer can reduce seed set in a pattern indistinguishable from divergent adaptation,  
510 deleterious effects of interspecific mating on fitness need to be accounted for in assessments of  
511 local adaptation.

512 As demonstrated here, both divergent habitat adaptation (i.e., differences in flooding  
513 tolerance) and reproductive interference contribute the spatial separation of the two  
514 monkeyflower species. The underlying mechanisms of resident species advantage in reciprocal  
515 transplants commonly operate asymmetrically (Popovic and Lowry 2020), and thus a pattern  
516 consistent with divergent adaptation may arise as an emergent property of several asymmetric  
517 barriers to foreign habitat persistence. Therefore, interpretations of a single fitness metrics (i.e.,  
518 flower production, seed set) across species are difficult to interpret without knowledge of  
519 frequency- and density-dependent processes and the underlying mechanisms. In a previous  
520 study, *M. guttatus* and *M. nudatus* were transplanted reciprocally at equal frequencies (Toll and  
521 Willis 2018) and thus, could not test for of minority disadvantage as we did here.

522

## 523 Conclusion

524 Patterns of spatial segregation are ubiquitous across animal and plant taxa and diverse  
525 environments (Wisheu 1998). In closely related species, negative reproductive interactions can  
526 contribute to these patterns if they occur in a frequency-dependent manner and are stronger than  
527 intraspecific competition (Schreiber et al 2019, Weber and Strauss 2016). Using both  
528 experimental and observational approaches, we found that hybridization operated in a frequency-  
529 dependent manner for one species of monkeyflower when invading habitats dominated by its  
530 congener but was not frequency-dependent when the other species invaded. Despite this  
531 asymmetry, hybridization contributed to rare immigrant disadvantage for both species. Overall,  
532 this result, combined with other results from this system (Toll and Willis 2018, Toll et al. 2021),  
533 suggests that spatial segregation is an emergent property of many asymmetric barriers operating  
534 in complementary and contrasting ways. Therefore, it seems likely that evidence of divergent



535 adaptation from reciprocal transplant experiments of closely related species have the potential to  
536 be driven by reproductive interference, in addition to the more commonly invoked differences in  
537 stress tolerance and competitive ability. Parsing out the specific effects of each mechanism  
538 across many systems will be necessary for a fuller understanding of how closely related species  
539 segregate into different habitats.

540

541

## Acknowledgments

542 The authors wish to thank K. Christie, J.M. Coughlan, and E.F. LoPresti whose comments  
543 helped improve and clarify this manuscript. This work was performed at the University of  
544 California Natural Reserve System McLaughlin Natural Reserve, Reserve DOI:  
545 10.21973/N3W08D. We thank C. Koehler and P. Aigner for facilitating field work. This work  
546 was supported by Michigan State University and a grant from the National Science Foundation  
547 (IOS-1855927).

548

## References

- Anacker, B. L., and S. Y. Strauss. 2014. The geography and ecology of plant speciation: range overlap and niche divergence in sister species. *Proceedings of the Royal Society B: Biological Sciences* 281: 20132980.
- Baack, E. J. 2005. Ecological factors influencing tetraploid establishment in snow buttercups (*Ranunculus adoneus*, Ranunculaceae): minority cytotype exclusion and barriers to triploid formation. *American Journal of Botany* 92: 1827–1835.
- Baker, S.M. 1909. On the causes of the zoning of brown seaweeds on the seashore. *New Phytologist* 8: 196-202.
- Benadi, G., and A. Pauw. 2018. Frequency dependence of pollinator visitation rates suggests that pollination niches can allow plant species coexistence. *Journal of Ecology* 106: 1892–1901.
- Buggs, R. J. A., and J. R. Pannell. 2006. Rapid displacement of a monoecious plant lineage is due pollen swamping by a dioecious relative. *Current Biology* 16: 996–1000.
- Burns, J. H. and S. Y. Strauss. 2011. More closely related species are more ecologically similar in an experimental test. *Proceedings of the National Academy of Sciences* 108: 5302-5307.
- Brigham, Christy A. 2001. Maintenance of Diversity: Mechanisms of Persistence in Rare Plants. University of California, Davis, PhD Dissertation.
- Brigham, C. A. 2003. Factors Affecting Persistence in Formerly Common and Historically Rare Plants. Pages 59–97 in C. A. Brigham and M. W. Schwartz, editors. *Population viability in plants: conservation, management, and modeling of rare plants*. Springer Science and Business Media, New York, New York, USA.
- Brown, W. L., and E. O. Wilson. 1956. Character displacement. *Systematic Zoology* 7: 49–64.
- Christie, K., and S. Y. Strauss, S.Y. 2019. Reproductive isolation and the maintenance of species boundaries in two serpentine endemic Jewelflowers. *Evolution* 73: 1375-1391.
- Christie, K., and S. Y. Strauss. 2020. Frequency-dependent fitness and reproductive dynamics contribute to habitat segregation in sympatric jewelflowers. *Proceedings of the Royal Society B: Biological Sciences* 287: 20200559.
- Chen, G. F., and D. W. Schemske. 2015. Ecological differentiation and local adaptation in two sister species of Neotropical *Costus* (Costaceae). *Ecology* 96: 440-449.
- Clausen, J., D. D. Keck, and H. M. Hiesey. 1940. Experimental studies on the nature of species. I. Effect of varied environment on Western North American plants. Washington, DC, USA: Carnegie Institution of Washington. Publications No. 520.

- Connell, J. H. 1961. The influence of interspecific competition and other factors on the distribution of the barnacle *Chthamalus stellatus*. *Ecology* 42: 710-723.
- Coughlan, J. M., M. W. Brown, and J. H. Willis. 2020. Patterns of hybrid seed inviability in the *Mimulus guttatus* sp. complex reveal a potential role of parental conflict in reproductive isolation. *Current Biology* 30: 83- 93.
- DiVittorio, C. T., S. Singhal, A.B. Roddy, F. Zapata, D. D. Ackerly, B. G. Baldwin, C. R. Brodersen, A. Búrquez, P. V. A. Fine, M. P. Flores, E. Solis, J. Morales-Villavicencio, D. Morales-Arce, D. W. Kyhos. 2020. Natural selection maintains species despite frequent hybridization in the desert shrub *Encelia*. *Proceedings of the National Academy of Sciences* 117: 33373-33383.
- Ferris, K.G., and J. H. Willis. 2018. Differential adaptation to a harsh granite outcrop habitat between sympatric *Mimulus* species. *Evolution* 72: 1225-1241.
- Fox, J., and S. Weisberg. 2019. *An R Companion to Applied Regression*, Third Edition. Thousand Oaks CA: Sage.
- Friberg, M., Leimar, O., and C. Wiklund. 2013. Heterospecific courtship, minority effects and niche separation between cryptic butterfly species. *Journal of Evolutionary Biology* 26: 971–979.
- Gardner M., and M. Macnair. 2000. Factors affecting the co-existence of the serpentine endemic *Mimulus nudatus* Curran and its presumed progenitor, *Mimulus guttatus* Fischer ex DC. *Biological Journal of the Linnean Society* 69: 443-459.
- Garner, A. G., A. M. Kenney, L. Fishman, and A. L. Sweigart. 2016. Genetic loci with parent of-origin effects cause hybrid seed lethality in crosses between *Mimulus* species. *New Phytologist* 211: 319–331.
- Grace, J. B., and R. G. Wetzel. 1981. Habitat partitioning and competitive displacement in cattails (*Typha*): Experimental field studies. *The American Naturalist* 118: 463–474.
- Grainger, T.N., Levine, J.M., and B. Gilbert. 2019. The invasion criterion: a common currency for ecological research. *Trends in Ecology and Evolution* 34: 925-935.
- Groning, J., and A. Hochkirch. 2008. Reproductive interference between animal species. *Quarterly Review of Biology* 83: 257–282.
- Hereford, J. 2009. A quantitative survey of local adaptation and fitness trade-offs. *The American Naturalist* 173: 579-588.
- Husband, B. C. 2000. Constraints on polyploid evolution: a test of the minority cytotype exclusion principle. *Proceedings of the Royal Society B: Biological Sciences* 267: 217-223.

- Inouye, B.D. 2001. Response surface experimental designs for investigating interspecific competition. *Ecology* 82: 2696–2706.
- Kawecki, T. J., and D. Ebert. 2004. Conceptual issues in local adaptation. *Ecology Letters* 7: 1225–1241.
- Kuno, E. 1992. Competitive exclusion through reproductive interference. *Researches on Population Ecology* 34: 275–284.
- Kyogoku, D., and H. Kokko. 2020. Species coexist more easily if reinforcement is based on habitat preferences than on species recognition. *Journal of Animal Ecology* 89: 2605–2616.
- Kyogoku, D., and D. Wheatcroft. 2020. Heterospecific mating interactions as an interface between ecology and evolution. *Journal of Evolutionary Biology* 33: 1330–1344.
- Lenth, R. V. 2020. emmeans: Estimated Marginal Means, aka Least-Squares Means. R package version 1.5.3.
- Levin, D.A. 1975. Minority cytotype exclusion in local plant populations. *Taxon* 24: 35–43.
- Levin, D. A., and W. W. Anderson. 1970. Competition for pollinators between simultaneously flowering species. *The American Naturalist* 104: 455–467.
- Lewis, H. 1961. Experimental sympatric populations of *Clarkia*. *The American Naturalist* 95:155–168.
- Lüdtke, D. 2018. ggeffects: Tidy data frames of marginal effects from regression models. *Journal of Open Source Software* 3: 772.
- Mitchell, R. J., R. J. Flanagan, B. J. Brown, N. M. Waser, and J. D. Karron. 2009. New frontiers in competition for pollination. *Annals of Botany* 103: 1403–1413.
- Nagy, E.S. 1997. Frequency-dependent seed production and hybridization rates: Implications for gene flow between locally adapted plant populations. *Evolution* 51: 703–714.
- Oneal, E., J. H. Willis, and R. G. Franks. 2016. Disruption of endosperm development is a major cause of hybrid seed inviability between *Mimulus guttatus* and *Mimulus nudatus*. *New Phytologist* 210: 1107–1120.
- Pederson, T. L. 2020. patchwork: The Composer of Plots. R package version 1.1.1.
- Peterson, M., K.J. Rice, and J.P. Sexton. 2013. Niche partitioning between close relatives suggests trade-offs between adaptation to local environments and competition. *Ecology and Evolution* 3: 512–522.

- Popovic D., and D. B. Lowry. 2020. Contrasting environmental factors drive local adaptation at opposite ends of an environmental gradient in the yellow monkeyflower (*Mimulus guttatus*). *American Journal of Botany* 107: 298-307.
- R Core Team. 2020. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria.
- Ribeiro, J. M. C., and A. Spielman. 1986. The Satyr effect: A model predicting parapatry and species extinction. *The American Naturalist* 128: 513–528.
- Ritland, C., and K. Ritland. 1989. Variation of sex allocation among eight taxa of the *Mimulus guttatus* species complex (Scrophulariaceae). *American Journal of Botany* 1989: 1731–1739.
- Runquist, R.D.B. 2012. Pollinator-mediated competition between two congeners, *Limnanthes douglasii* subsp. *rosea* and *L. alba* (Limnanthaceae). *American Journal of Botany* 99: 1125-1132.
- Runquist R.B., and M. L. Stanton. 2013. Asymmetric and frequency-dependent pollinator-mediated interactions may influence competitive displacement in two vernal pool plants. *Ecology Letters* 16: 183-190.
- Sandstedt, G.D., Wu, C.A. and Sweigart, A.L. 2021. Evolution of multiple postzygotic barriers between species of the *Mimulus tilingii* complex. *Evolution* 75: 600-613.
- Schreiber, S. J., M. Yamamichi, and S. Y. Strauss. 2019. When rarity has costs: coexistence under positive frequency-dependence and environmental stochasticity. *Ecology* 100: e02664.
- Selby, J. P., and J. H. Willis. 2018. Major QTL controls adaptation to serpentine soils in *Mimulus guttatus*. *Molecular Ecology* 27: 5073–5087.
- Servedio, M. R., and M. A. F. Noor. 2003. The role of reinforcement in speciation: theory and data. *Annual Review of Ecology, Evolution and Systematics* 34: 339–364.
- Sianta, S.A., and K. M. Kay. 2019. Adaptation and divergence in edaphic specialists and generalists: serpentine soil endemics in the California flora occur in barer serpentine habitats with lower soil calcium levels than serpentine tolerators. *American Journal of Botany* 106: 690-703.
- Singer, F. 1990. Reproductive costs arising from incomplete habitat segregation among three species of *Leucorrhinia* dragonflies. *Behaviour* 115: 188–201.
- Takakura, K.I. 2013. Two-way but asymmetrical reproductive interference between an invasive *Veronica* species and a native congener. *American Journal of Plant Sciences* 4: 535–542

Takakura K.I., and S. Fujii. 2010. Reproductive interference and salinity tolerance differentiate habitat use between two alien cockleburs: *Xanthium occidentale* and *X. italicum* (Compositae). *Plant Ecology* 206: 309–319

Takakura, K.I., and S. Fujii. 2015. Island biogeography as a test of reproductive interference. *Population Ecology* 57: 307–319.

Takakura, K.I., T. Nishida, T. Matsumoto, S. Nishida. 2009. Alien dandelion reduces the seed-set of a native congener through frequency-dependent and one-sided effects. *Biological Invasions* 11: 973–981.

Thum R.A. 2007. Reproductive interference, priority effects and the maintenance of parapatry in *Skistodiaptomus* copepods. *Oikos* 116: 759–768

Toll, K. and J. H. Willis. 2018. Hybrid inviability and differential submergence tolerance drive habitat segregation between two congeneric monkeyflowers. *Ecology* 99: 2776-2786.

Toll, K, E. F. LoPresti, and D. B. Lowry. 2021. Inbreeding depression contributes to the maintenance of habitat segregation between closely related monkeyflower species. *Evolution* 75: 832-846.

Vickery, R. K. 1978. Case studies in the evolution of species complexes in *Mimulus*. *Evolutionary Biology* 11: 405–507.

Waser, N.M. 1978a. Competition for hummingbird pollination and sequential flowering in two Colorado wildflowers. *Ecology* 59: 934–944.

Waser N.M. 1978b. Interspecific pollen transfer and competition between co-occurring plant species. *Oecologia* 36: 223–236.

Weber, M. G., and S. Y. Strauss. 2016. Coexistence in close relatives: Beyond competition and reproductive isolation in sister taxa. *Annual Review of Ecology, Evolution, and Systematics* 47: 359–381.

Wickham, H. 2016. *ggplot2: Elegant Graphics for Data Analysis*. Springer-Verlag New York. ISBN 978-3-319-24277-4.

Wisheu, I. C. 1998. How organisms partition habitats: Different types of community organization can produce identical patterns. *Oikos* 83: 246-258.

Yost, J. M., T. Barry, K. M. Kay, and N. Rajakaruna. 2012. Edaphic adaptation maintains the coexistence of two cryptic species on serpentine soils. *American Journal of Botany* 99: 890–897.

## Figure Legends

**Figure 1.** (A) Schematic of the 2018 *M. guttatus* immigration experiment showing positions of focal immigrant *M. guttatus* plants (filled circle) and conspecific neighbors (open circles) within each block. (B) Photograph of an *M. guttatus* transplant with no conspecific neighbors, surrounded by native *M. nudatus*. (C) Schematic of the 2019 frequency-manipulation reciprocal transplant experiment showing the showing positions of immigrant (X) and resident (●) plants within each immigrant frequency treatment. Figure adapted from Nagy (1997). Five replicates per treatment were planted at each site, resulting in 20 blocks containing residents and 15 blocks containing immigrants at each site. (D) Photograph of an example 50% immigrant (9N:10G) plot in the seep site. The diameter of each pot was 2.54 cm and the largest diameter of each array was 12.7cm. Native *Mimulus* plants within 0.25m of each array were removed prior to transplanting.

**Figure 2.** Transect of co-occurring *M. nudatus* and *M. guttatus*. Background colors depict the frequency of *M. nudatus* fruit (blue) vs *M. guttatus* fruit (red). Points indicate the proportion of inviable F1 hybrid seeds produced by individual *M. nudatus* (triangles) and *M. guttatus* (circles) plants collected every 21cm along the transect.

**Figure 3.** Components of fitness from the 2018 *M. guttatus* immigration experiment. Fitness components: flowers per plant (A), total seeds per flower (C, D), and seed viability. Each red circle represents a single immigrant focal *M. guttatus* transplant in a wash dominated by *M. nudatus*. Conspecific neighbor abundance was significantly associated with focal plant hybridization rate.

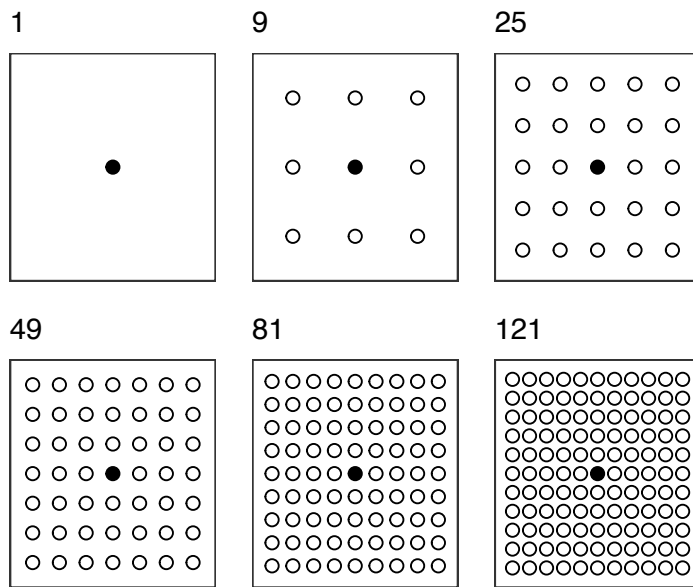
**Figure 4.** Predicted marginal means and 95% confidence intervals for components of fitness in the wash transplant site from the 2019 reciprocal transplant experiment. Models: Fitness component = immigrant frequency + (1|Block). Fitness components: flowers per plant (A, B), total seeds per flower (C, D), and seed viability (E, F). Immigrant *M. guttatus* = red circles (A, C, E), Resident *M. nudatus* = blue triangles (B, D, F). Letters indicate groupings from Tukey post-hoc contrasts.

**Figure 5.** Predicted marginal means and 95% confidence intervals for components of fitness in the seep transplant site from the 2019 reciprocal transplant experiment. Models: Fitness component = immigrant frequency + (1|Block). Fitness components: flowers per plant (A, B), total seeds per flower (C, D), and seed viability (E, F). Immigrant *M. nudatus* = blue triangles (A, C, E), resident *M. guttatus* = red circles (B, D, F). Letters indicate groupings from Tukey post-hoc contrasts.

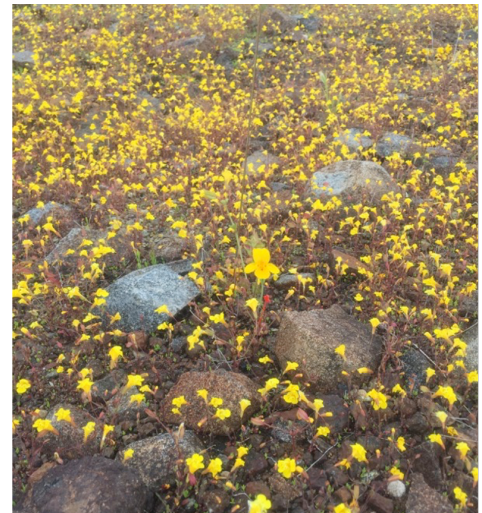
**Figure 6.** Predicted marginal means and 95% confidence intervals for lifetime fecundity in a wash (A) and seep (B) in the 2019 reciprocal transplant experiment. *M. nudatus* = blue triangles, *M. guttatus* = red circles. Marginal means were predicted from models using data from plots where both species were present at each site. Models: Lifetime fecundity = Species + Immigrant frequency + Species x Immigrant Frequency + (1|Block). Letters indicate significant groupings from Tukey post-hoc tests.



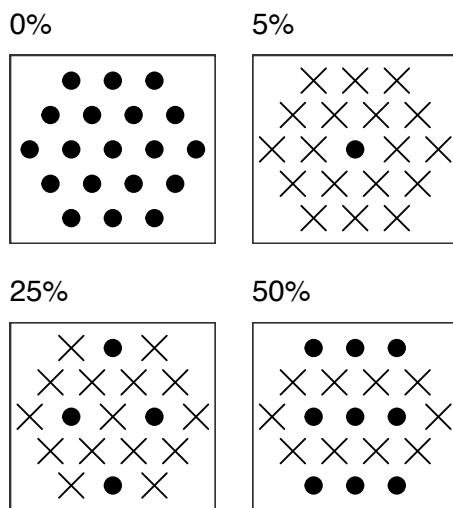
**A** *M. guttatus* immigration experiment



**B**



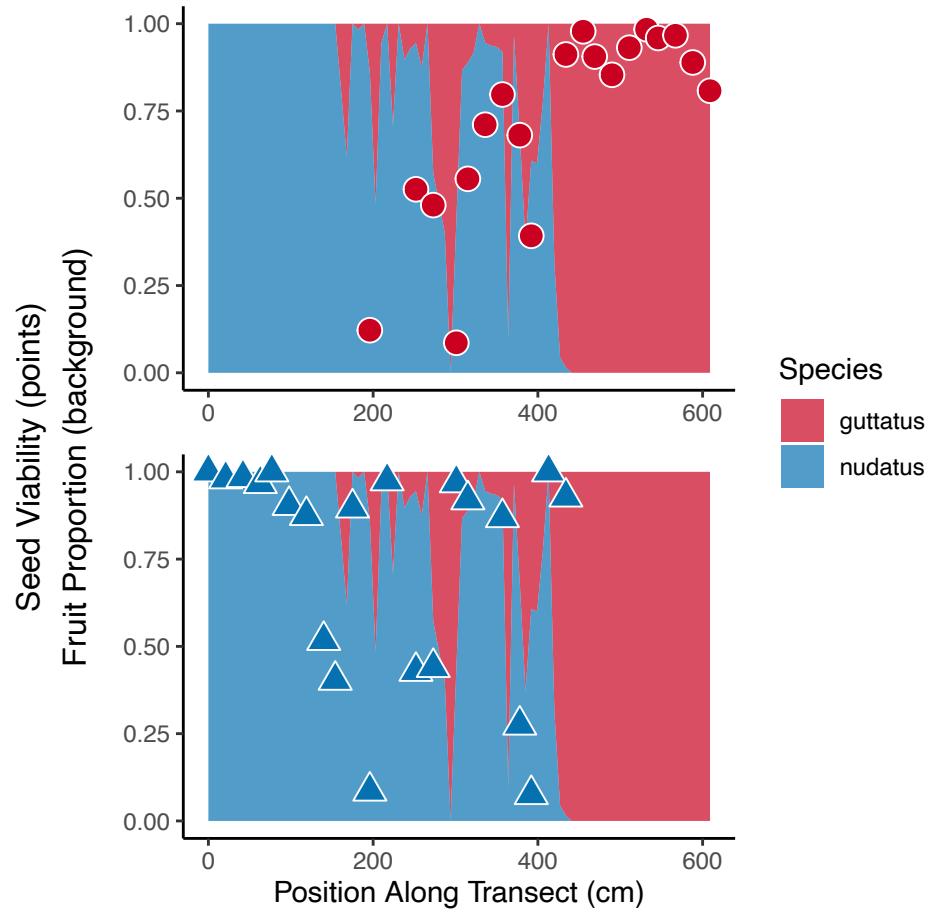
**C** Frequency-manipulation reciprocal transplant



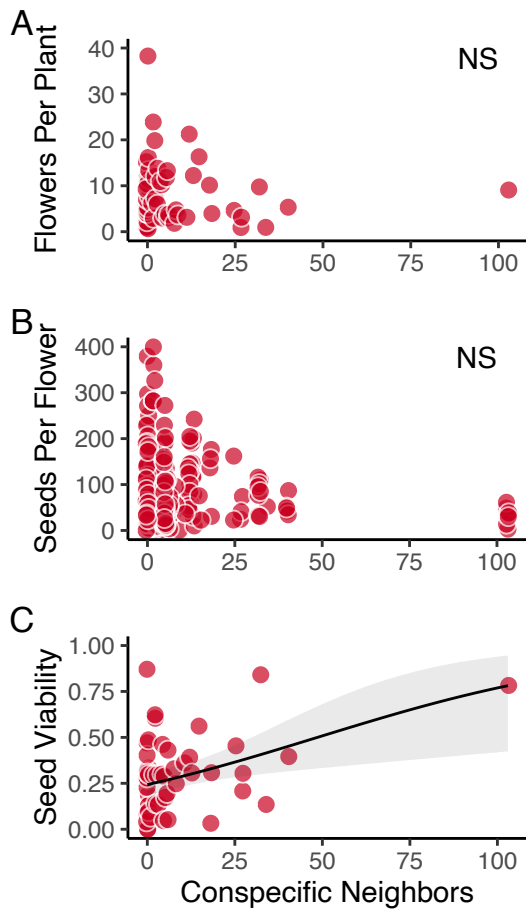
**D**







**Figure 2.**



**Figure 3.**

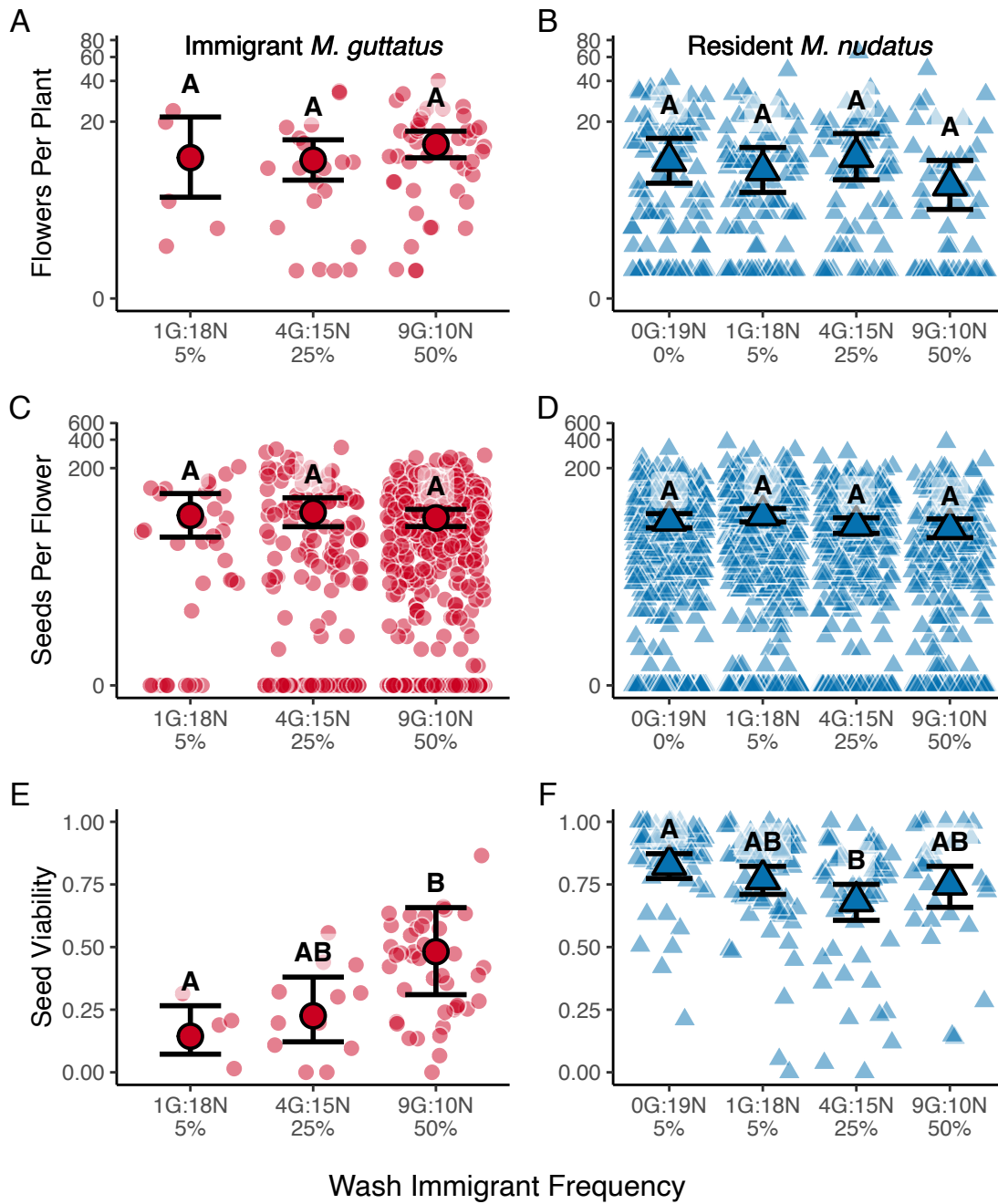


Figure 4.

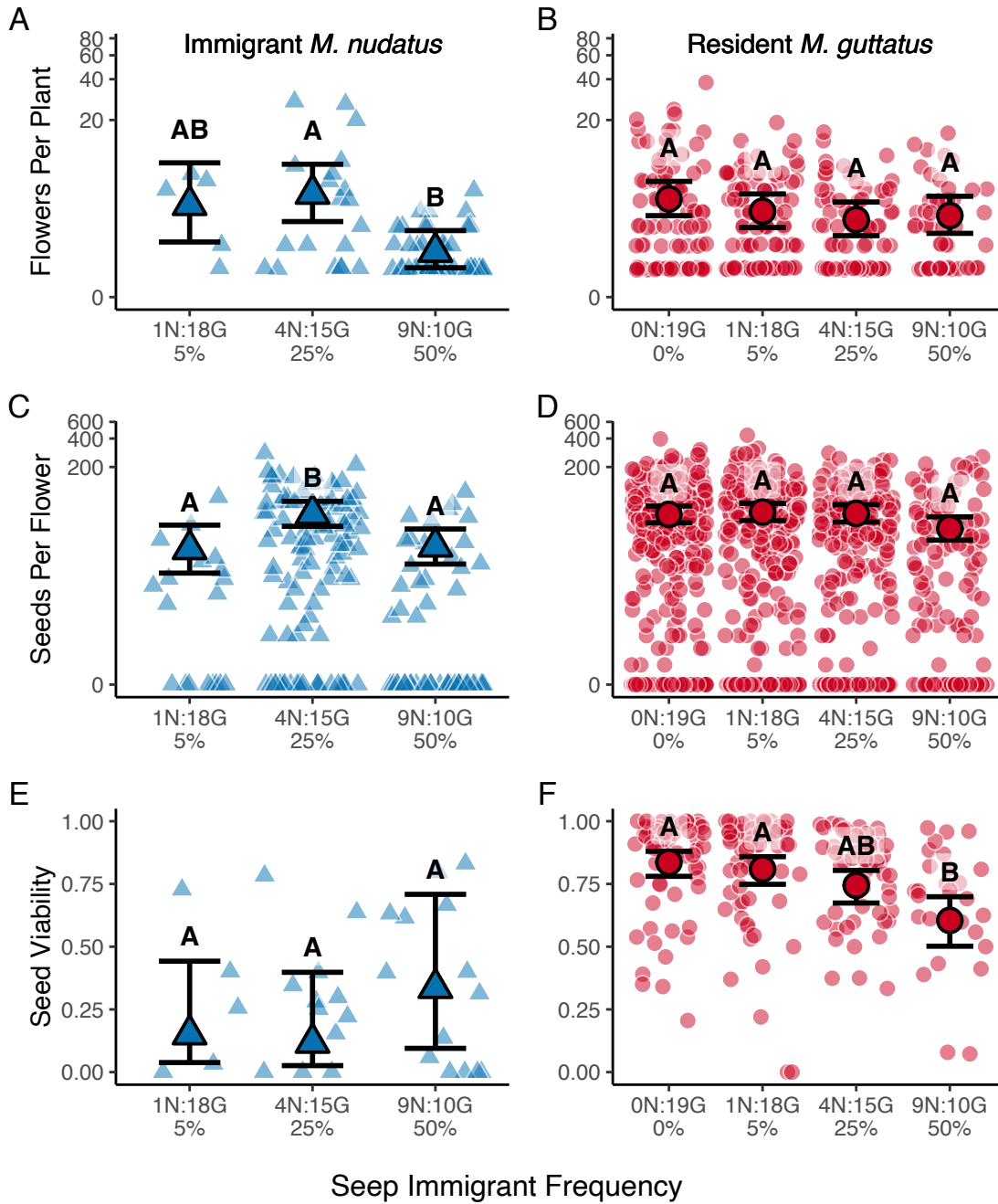
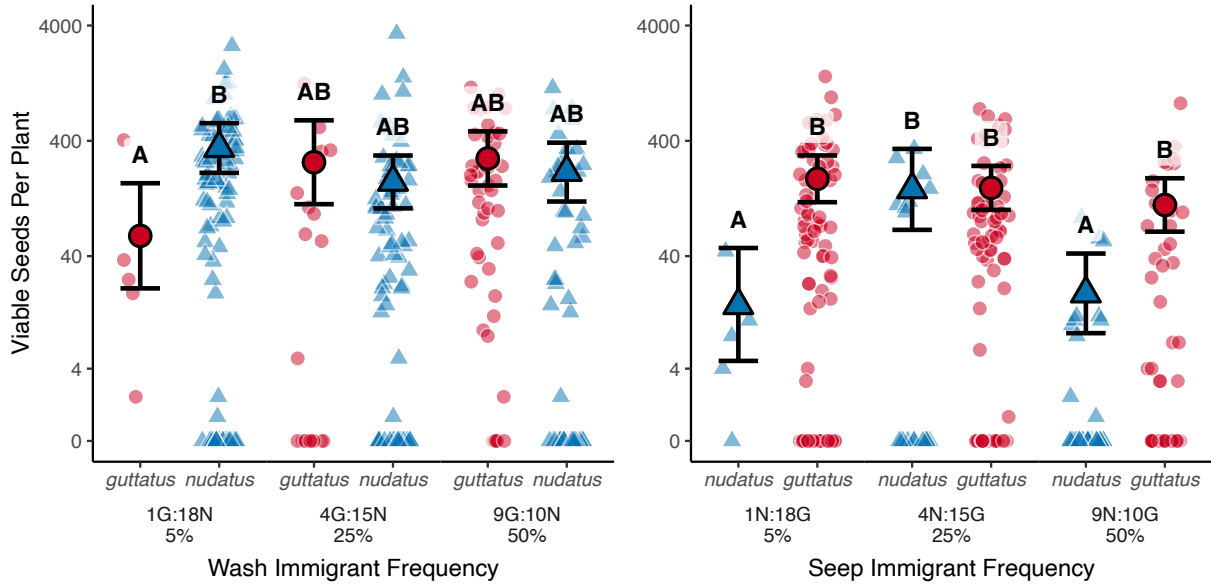


Figure 5.



**Figure 6.**