1 Life-history predicts global population responses to the

2 weather in the terrestrial mammals

- 3 Jackson, John^{1,2} john.jackson@zoo.ox.ac.uk
- 4 Le Coeur, Christie³ <u>c.l.coeur@ibv.uio.no</u>
- 5 Jones, Owen R¹ jones@biology.sdu.dk
- 6

7 1. Interdisciplinary centre for population dynamics (CPop), Department of Biology, University of8 Southern Denmark, Odense, Denmark

9 2. Department of Zoology, University of Oxford, Oxford, UK

3. Centre for Ecological and Evolutionary Synthesis (CEES), Department of Biosciences, University of
 Oslo, Oslo, Norway

12

13 Corresponding Author: John Jackson, Department of Zoology, University of Oxford, UK

14

15 Abstract

16 With the looming threat of abrupt ecological disruption due to a changing climate, predicting which 17 species are most vulnerable to environmental change is critical. The life-history of a species is a 18 promising candidate for explaining differences in climate-change responses, but we now need data 19 linking population change, weather and life-history to explore these predictions. Here, we use long-20 term abundance records from 157 species of terrestrial mammals to investigate the link between weather 21 and annual population growth rates. Overall, we found no consistent effect of temperature or 22 precipitation anomalies on annual population growth rates, but there was variability in weather 23 responses for populations within a species. Crucially, however, long-lived mammals with smaller litter 24 sizes had responses with a reduced absolute magnitude compared to their shorter-living counterparts 25 with larger litters. These results highlight the role of species-level life-history in driving responses to 26 the environment.

27

28 Introduction

Climate change is one of the greatest challenges we face in the twenty first century¹. Although habitat 29 loss and direct exploitation are currently the greatest drivers of extinction in the natural world^{1,2}, changes 30 31 to the climate are predicted to cause widespread declines to global biodiversity in the coming decades³⁻ 32 ⁵. For mammals and birds, temperature increases are already associated with declining population 33 trends⁶ and many endangered species have already been negatively impacted by climate change in at 34 least part of their range⁷. Perhaps more worryingly, abrupt ecological disruption due to runaway climate 35 change has been predicted to have large negative impacts on biodiversity, with tropical ecosystems 36 being affected as early as 2030⁸. Furthermore, these future impacts will likely be exacerbated by 37 synergism between the climate and other drivers of extinction such as habitat loss^{9,10}. Research 38 highlighting the species and ecosystems that are most vulnerable to climate change impacts will 39 therefore provide crucial knowledge to prevent future losses to global biodiversity.

40 Not all species are equally vulnerable to climate change. Species vary in their climatic niches 41 and in their behavioural, physiological, and demographic responses to environmental change and we therefore expect there to be both climate 'winners' and 'losers'¹¹⁻¹³. At a macro-scale, species 42 43 occupancy data highlight that geographic range shifts are the key response associated with climate 44 change across taxa, resulting in changes to community composition, but not necessarily population 45 decline^{14,15}. Recent broad assessments of biodiversity change metrics from species assemblage and 46 abundance records mirror this paradigm with both occupancy and abundance trend patterns suggesting a balanced frequency of winners and losers^{15,16}. In the Marine realm, species richness increases were 47 48 associated with warming temperatures, consistent with the expectation that warming marine ecosystems 49 will receive an influx of species tracking suitable temperatures¹². However, there was no consistent 50 temperature-related biodiversity change effect on land¹². Changes in species richness and other 51 biodiversity metrics however do not necessarily equate to population declines. For birds and mammals 52 average abundance trends were negatively associated with rates of climate warming⁶. Studies unpicking how changes in weather patterns cause population change are therefore vital¹⁷, and a growing body of 53 54 literature is exploring the relationship between the climate and the demographic processes driving 55 population decline¹⁸⁻²². Applying these concepts at a comparative scale and assessing finer-scale 56 population changes with respect to changes in the weather, and their relationship to species traits, will 57 aid in illuminating consistent or disparate climate change responses across the tree of life^{21,23}.

58 Variation in the demographic responses of organisms climate change suggests that life-history is 59 a promising target for explaining responses to environmental change. The timing of demographic events 60 relating to the key demographic rates of survival and recruitment are evolved responses to the 61 environment, and characteristics relating to both 'slow' and 'fast' life-histories are therefore adaptive 62 in different environmental contexts²⁴. Indeed, Pacifici et al.⁷ concluded that intrinsic traits, including 63 habitat specialisation and aspects of life-history, were associated with negative climate-mediated 64 population effects reported in mammals and birds. Life-history differences between three amphibian 65 species in Western Europe drove differences in survival and reproduction in response to the North Atlantic Oscillation²⁵. Generally, we expect that organisms with slower life-histories are better-adapted 66 to cope with fluctuations in the environment. Longer-lived organisms have a reduced relative effect of 67 68 variability in vital rates, variability which is expected during environmental change, on population growth rates²⁶ and long-lived plants have weaker absolute demographic responses to weather²³. 69 70 Furthermore, the recently developed concept of demographic resilience uses demographic rates 71 characterising the life cycle of an organism to quantify their resilience to perturbations²⁷. Comparative 72 approaches linking life-history traits to climate change responses may therefore provide a crucial 73 predictive link to improve our understanding of climate vulnerability.

74 In this study, we investigated annual population responses to temperature and precipitation in 75 populations of terrestrial mammals across the world's ecosystems. Importantly, we tested whether life-76 history predicts population responses to the weather, and therefore its utility in assessing vulnerability

77 to climate change. We addressed these questions using 486 long-term (≥ 10 years) abundance records 78 from 157 species of terrestrial mammal obtained from the Living Planet Database⁴, by implementing a 79 two-step meta-regression framework. First, for each abundance record, we assessed how observed 80 annual population growth rates were influenced by weather anomalies (annual deviation from long-81 term average weather patterns) using autoregressive additive models that accounted for temporal 82 autocorrelation in abundance records and overall abundance trends. Then, we used a phylogenetically 83 controlled Bayesian meta-regression with weather effect coefficients as the response variable to address 84 three key questions: 1) Are there consistent temperature and precipitation effects on abundance change 85 across the terrestrial mammals? 2) How are these patterns influenced by covariance both within and 86 between species, and are there vulnerable biomes? 3) Can species-level life-history traits predict the 87 magnitude of population responses to the weather? The terrestrial mammals are an ideal study system 88 to explore the predictors of population responses to climate change because they are a well-studied 89 group with a combination of intensive abundance monitoring across the globe⁴, detailed life-history information for hundreds of species^{28,29} and a highly-resolved phylogeny to facilitate phylogenetic 90 91 comparative analyses³⁰. Furthermore, there is growing evidence from the mammals of the mechanistic links between the climate, demography, and population dynamics^{17,20–22}. 92

93 In light of previous findings from studies exploring biodiversity change metrics and life-history 94 theory in the context of environmental change, we make the following predictions related to our three 95 key questions. First, reflecting the complexity of overall biodiversity change and the paradigm of climate winners and losers^{12,15,16}, we predict that across the mammals there will be no consistent 96 97 responses to temperature and precipitation. Second, recent evidence has highlighted phylogenetic covariance between demographic traits³¹, which may also scale to population responses. However, 98 99 spatial variation between study sites, heterogeneity across a species range and the complex interaction 100 of demographic traits to drive population dynamics may reduce phylogenetic covariance. Finally, we 101 predict that organisms with slow life-histories will be more resilient to weather effects, reflecting an 102 evolutionary adaptation to buffer against fluctuating environments²⁶. We therefore expect responses to 103 weather to have a higher magnitude in short-living organisms with larger reproductive output compared 104 to their longer-lived counterparts.

105

106 Results

We assessed population responses to weather in 486 long-term abundance time-series records of 157 species of terrestrial mammals from across the world's ecosystems (Fig. S1). The time-series records ranged in duration from 10 years to 35 years, with mean and median record lengths across records of 15.7 and 14 years, respectively (Fig. S1). The records were distributed across 13 terrestrial biomes, including both tropical and temperate regions, but were generally biased towards north western Europe and North America. We had records from 12 of 27 mammalian orders recognised by the IUCN Red List for threatened species³², but most densely in the Artiodactyla (n = 172), Carnivora (n = 127) and

Rodentia (n = 82) (Fig. S1). The number of records for each species ranged from 1-17, with a mean of
3.1 and median of 2 records per species.

116

117 No consistent population response to weather

118 Overall, we found no consistent effect of either temperature or precipitation anomalies on annual 119 population growth rates in the terrestrial mammals (Fig.1). The raw weather effects on population 120 growth rates, ω , varied across species and records but were centred close to 0, with a mean temperature 121 coefficient of -0.08 (± 0.73 SD; Fig. 1a) and mean precipitation coefficient of 0.02 (± 0.36 SD; Fig. 1b). 122 Furthermore, 95% of records had temperature and precipitation coefficients between -1.87-1.38 and -0.59-0.83, respectively. Nevertheless, approximately 10% (n = 51) of temperature effects and 3% of 123 124 precipitation were greater than 1 or less than -1, indicating that small clusters of populations experienced 125 more extreme annual responses to the weather (Fig.1). Our Bayesian meta-regression, controlling for 126 both within species variance, phylogenetic covariance and differences in sample size (number of years) 127 between records, mirrored the lack of consistent weather effects on population growth. The posterior 128 mean global intercept, $\bar{\alpha}$, for temperature effects was 0.07 [-0.17-0.33] (95% credible intervals) and for 129 precipitation effects was -0.06 [-0.28-0.14] (Fig.1a and Fig.1b). There was, however, a small positive 130 effect of the number of years of population data for a record and the response to temperature, with a 131 linear slope, β_N of 0.09 [0.0-0.18]. Together with the results of the global intercept $\bar{\alpha}$, this suggests that 132 shorter records were associated with more negative temperature effects. Overall, these results highlight 133 the paradigm of the existence of both winners and losers in weather responses, but no clear effect across

- the Mammalia.
- 135

136 Spatial effects and variation between species

137 We tested whether there were differences in weather responses among ecological biomes because biome 138 effects may be indicative of more extreme responses to weather in some habitats. Using leave-one-out 139 cross-validation, we compared the predictive performance of the model including the effect of biome 140 relative to the base model, and we found no evidence for an influence of biome on either temperature 141 $(\Delta elpd = -0.67 relative to base model)$ or precipitation ($\Delta elpd = -0.73$) effects (see Fig. S15-16 for more 142 information). Furthermore, we explored the role of spatial autocorrelation at driving differences in 143 weather coefficients across records using Morans I tests and spatially explicit meta-regressions but did 144 not find evidence for spatial autocorrelation in weather effects (Figs. S18-S20). We also incorporated 145 both phylogenetic covariance (σ_{PHYLO}^2) and species-level variance ($\sigma_{SPECIES}$) to capture both among-146 and within-species variation. Interestingly, we found far greater levels of within-species variation in 147 temperature responses compared to among-species variance (Fig. 1c). The posterior mean for species-148 level variance in temperature effects was 0.2 [0.01-0.4] which was 20 times greater than the posterior 149 estimate of 0.01 [0.0-0.03] for phylogenetic covariance (Fig. 1c). Similarly, for precipitation the

posterior mean for species-level variance was five times greater than for phylogenetic covariance, with a value 0.05 [0.0-0.15] compared to 0.01 [0.0-0.02] (Fig. 1c). These patterns are reflected in the temperature and precipitation coefficients, for which large variation can be seen among records of the same species. For example, *Myodes glareolus* (bank vole) in the Rodentia had 9 population records, and a range of temperature/precipitation effects of -1.64-3.59 and -1.27-1.44 respectively, compared to -3.32-5.67 and -1.27-1.44 across Rodentia as a whole (Fig.1). This result highlights the potential importance of within-species variability in population responses to environmental change.

157

158 *Life-history predicts population responses to weather*

159 Across the terrestrial mammals, we found that both maximum longevity and mean litter size at the 160 species-level predicted the magnitude of population responses to weather. We tested a set of Gamma 161 models incorporating univariate, multivariate and 2-way interactions between maximum longevity, 162 litter size, and adult body mass and their influence on the absolute magnitude of 163 temperature/precipitation effects using model selection and leave-one-out cross-validation (Table S1 & 164 S2). As with our Gaussian models of overall weather effects, we found that the sample size of a record 165 had strong negative impact on the absolute temperature and precipitation responses, with posterior 166 estimates on the linear predictor scale of $\beta_N = -0.26$ [-0.35 - -0.16] and $\beta_N = -0.35$ [-0.46 - -0.24], 167 respectively (Fig. S17). Namely, shorter records were associated with temperature and precipitation 168 responses of a larger magnitude. We found no association between adult body mass and either temperature ($\beta_{BODYMASS} = -0.01$ [-0.19-0.16]) or precipitation responses ($\beta_{BODYMASS} = 0.03$ [-0.20-169 170 0.25]). Furthermore, we found no strong evidence for any two-way interactions between life-history 171 variables (Table S1 & S2). For both temperature and precipitation effects, the most competitive model 172 was the univariate model including mean litter size ($\Delta elpd = 3.03$ and $\Delta elpd = 3.00$, compared to the 173 base model for temperature and precipitation, respectively; Table S1 & S2). However, univariate 174 models including maximum longevity alone also had a higher predictive performance than the base 175 model ($\Delta elpd = 0.94$ and $\Delta elpd = 1.67$ for temperature and precipitation, respectively). In both cases 176 the second-best predictive model was the one that included univariate effects for longevity, bodymass 177 and litter size (Δ elpd = 2.88 and Δ elpd = 2.94; Table S1 & S2). Therefore, we selected the models 178 including all univariate life-history effects.

For both temperature and precipitation, our results highlight that shorter-living mammals with greater litter sizes experienced weather effects of a significantly greater magnitude than longer-living, slowly reproducing mammals (Fig. 2). The magnitude of weather responses was negatively associated with longevity, with posterior means on the linear predictor scale of $\beta_{LONGEVITY} = -0.19$ [-0.44-0.07] and $\beta_{LONGEVITY} = -0.24$ [-0.57-0.07] for temperature and precipitation, respectively (Fig. 2a & 2c). Thus, a maximum longevity change from 10 months (*Akodon azarae*) to 80 years (*Loxodonta africana*) was associated with a 2.3-fold and 2.9-fold decrease in the predicted magnitude of responses to

186 temperature and precipitation. So, for every additional 5 years of life, there was a 14% decrease in the 187 magnitude of responses to temperature and 18% decrease in the magnitude of responses to precipitation. 188 An organism's longevity is strongly correlated to their body mass, but the effect of longevity held 189 irrespective of whether adult body mass was also included in the model. In contrast, but also following 190 key predictions from life-history theory, the magnitude of weather responses had a strong positive 191 association with litter size, with posterior means of $\beta_{LITTER} = 0.33$ [0.13-0.52] and $\beta_{LITTER} = 0.30$ [0.06-192 0.54] for temperature and precipitation, respectively (Fig. 2b & 2d). In other words, mammals bearing 193 more offspring in a single litter had greater responses to temperature and precipitation. A change in 194 litter size from 1 (monotocous species, various) to 17 (*Thylamys elegans*) was associated with a 4.17-195 fold and 3.71-fold increase in the predicted magnitude of temperature and precipitation responses. For 196 every additional offspring invested into at the litter stage, there is a 24% increase in the magnitude of

- temperature responses and 22% increase in the magnitude of precipitation responses.
- 198

199 Discussion

200 Our results provide an important empirical link between a species' life-history and its population 201 responses to environmental change. While we found no consistent patterns of responses to temperature 202 and precipitation anomalies across the mammals, life-history traits relating to the pace of life were 203 associated responses to weather. Namely, shorter-living species with increased litters sizes, or species 204 characterised with 'fast' life-history traits, responded with a greater magnitude compared to those with 205 'slow' life-history traits. While it has long been theorised that an organism's life-history traits evolve 206 in response to, and as an adaptation to, environmental conditions²⁴, rarely has this theory been tested at 207 a global scale. We find strong support for the hypothesis that longevity, and 'slow' life-history characteristics more generally, buffer organisms against variability in the environment²⁶ and add to a 208 209 small number of studies linking population demography and the climate^{21,23}. Critically, our results 210 highlight the potential utility of life-history traits for predicting species vulnerability to climate change. 211 Based on our findings, abrupt ecological disruption from climate change will disproportionately impact 212 shorter-lived species with higher reproductive output.

213 Demography has a vital role to play in predicting population declines in the Anthropocene and in highlighting targets for conservation management 28,33 . Our study emphasises this role, demonstrating 214 215 the predictive power of demographic life-history traits when investigating responses to environmental 216 change. However, there are limitations and barriers to the utility of demography in conservation. Only 217 1.3% of tetrapods globally have sufficient demographic information with which to estimate population 218 dynamics²⁸. Here, we used summary traits that are available for many species (maximum longevity and 219 mean litter size), but in particular maximum recorded longevity, while sufficient as a broad indicator, 220 is strongly influenced by sampling variance and a flawed measure of longevity differences between 221 $taxa^{34}$. Ideally, lifetables with mortality and reproduction trajectories across the lifecycle can be 222 combined data on external drivers to investigate detailed patterns in population dynamics^{35,36}. The

223 recent development of the demographic resilience framework, which uses demographic data across the 224 lifecycle to simulate how a population may respond to perturbations²⁷, has excellent potential in 225 extending these findings to explore demographic relationships with climate responses in detail. 226 Unfortunately, however, detailed (st)age-specific demographic information is not currently available 227 for a majority of species, but growing in availability rapidly³⁷. Therefore, there is a need to continue to 228 increase the collection of demographic data (and other traits) for many more species than are currently 229 available²⁸, so that we may predict population changes with respect to environmental change. Achieving this target may revolutionise the way we quantify species vulnerability to climate change^{12,15,16,21}, 230 231 helping to prevent extinctions before they occur.

232 In line with recent global assessments of biodiversity in the face of climatic change²¹, we did 233 not find an overall consistent effect of weather anomalies on population growth rates. This may in part 234 reflect the fact that abundance changes are a higher-order process determined by complex interactions between demographic processes that counteract each other^{16,21}. However, our results contrast with 235 236 findings of linear associations between mammal abundance and temperature change⁶. These differences 237 may reflect our approach to investigate annual changes, rather than long-term trends. Significant 238 population trends from long time-series are detectable from smaller component time-series even when 239 sampling is incomplete³⁸, and thus responses detected in trends may reflect broader changes in response 240 to the climate that are not detected in models of annual change. Furthermore, we estimated linear, annual 241 effects of weather on population growth rates, where population responses may actually be more 242 complex non-linear patterns or lagged effects. However, the detection of climate effects on average 243 trends may also be confounded by effects of other (sometimes more dominant) drivers (e.g. habitat 244 loss)². Nevertheless, our findings can be explained in light of recent studies from the Living Planet 245 Database that have found that the large majority of records do not exhibit population declines¹⁶.

246 Interestingly, we did not find evidence for phylogenetic covariance in weather responses 247 between species. Recent evidence from birds indicated strong phylogenetic covariance in vital rates, 248 particularly in adult survival, and the incorporation of phylogenetic information greatly improved predictive performance when imputing vital rates³¹. Therefore, as with overall patterns, our findings 249 250 may reflect the trade-offs between vital rates, which cancel one another out when scaling up to 251 population-level processes such as population growth rates in response to the weather²¹. Furthermore, 252 for long-term time-series, there may also be temporal trade-offs in vital rates, where for example 253 investing heavily into survival in one year (in response to climate) may impact subsequent reproduction 254 for several years, decreasing the magnitude of population growth rates. The extent of phylogenetic 255 covariance in vital rate responses and trade-offs remains unknown, understanding how the climate 256 impacts demographic rates across species may provide a useful tool for imputing population responses 257 to the climate across the tree of life³¹.

We highlight the importance of variation in population responses to climate within a species range. Sampling heterogeneity has recently been shown to have broad implications for metrics of 260 population dynamics, where demographic rates are poorly correlated among sampling sites for the same 261 species^{39,40}. Therefore, inferences obtained from monitoring single populations or studies may not 262 accurately portray species-level variability. This has broad implications for macroecology, particularly 263 for population viability assessments (PVA) and species-distribution modelling. First, as well as 264 suffering from data quality issues in their parameterisation⁴¹, our findings suggest that PVAs based on 265 data from a single population may not accurately reflect population viability across a species' 266 geographic range. Therefore, incorporating detailed demographic data, and investigating differences in 267 population responses across a range, could greatly improve our perspective on population viability³⁶. 268 Second, presence-only models of species distributions that do not account for the fact that responses to 269 the environment within a species range do not accurately represent species distributions⁴². Moving 270 towards trait-based monitoring and explicitly including demographic processes with mechanistic links 271 to appropriate drivers into species distribution models could greatly improve predictions of climate 272 change impacts on the biosphere⁸.

Ultimately, improving our predictions of how humans are influencing the natural world is
paramount to prevent rapid declines to global biodiversity⁴³. This however requires a large shift towards
both broad and detailed monitoring of global biodiversity. We show that linking species traits such as
life-history to changes in the environment may equip us with tools to predict and prevent future losses.

277

278 Methods

To assess the effects of weather on population growth rates we collated information on global weather and the abundance, life-history and phylogeny of the terrestrial mammals. All analyses were carried out using R version 4.0.4⁴⁴. For all data on the terrestrial mammals, taxonomies were resolved using the *taxize* package version 0.9.98⁴⁵ and matched using the Global Biodiversity Information Facility database (https://www.gbif.org/). All code used in the current study and full descriptions of the analyses are archived in the Zenodo repository (doi:10.5281/zenodo.4707232), which was created from the following GitHub repository https://github.com/ijackson-eco/mammal weather lifehistory.

286

287 *Time-series abundance data*

288 Annual time-series abundance data from across the terrestrial mammals were obtained from the Living Planet Database found at https://livingplanetindex.org/data portal. This database was developed by the 289 290 World Wildlife Fund and the Zoological Society of London as a tool to monitor global biodiversity, 291 and contains over 20,000 population records for over 4000 species of vertebrate⁴. The records measure 292 annual abundance in a variety of ways (e.g. full population counts, density, indices), and contain 293 information on the location, realm, biome and taxonomy of the species in the record. First, we included 294 only data for the terrestrial mammals that had species-level life-history information and coordinate 295 locations, which referred to either specific or more general locations for each population (accounted for 296 using weather data from a buffered radius around each location). We natural-log-transformed the raw

297 abundance data to ensure that weather coefficients were comparable across records and abundance 298 measure types. Because our analyses were focussed on estimating weather effects on annual population 299 growth rates using regression models with several variables, we included only long-term records with 300 10 or more consecutive years of abundance data, and only for years in which there was also weather 301 data (1979-2013). In one record (for *Bettongia penicillata*), there were two blocks with ≥ 10 years of 302 data, which were analysed separately. We also removed records with a high proportion (>32%) and 303 consecutive occurrences of 0 in the raw abundance time-series. Our final dataset contained 486 records from 157 terrestrial mammal species, which was used in all subsequent analyses (Fig. S1). 304

305

306 Global weather data

307 We used version 1.2.1 of the CHELSA monthly gridded temperature and precipitation timeseries 308 dataset, which is at a spatial resolution of 30 arc seconds (~ 1km²) for all months between 1979-2013 across the globe's land surface. The CHELSA datasets can be obtained from Karger et al.⁴⁶. Raster files 309 310 of the raw monthly mean temperature and total precipitation data were processed using the *raster*, *rgeos*, 311 and sf packages^{47–49}. Using the Living Planet Database record coordinate locations as a centroid, we 312 averaged the monthly weather data for a buffered radius of 5km around each record location to account 313 for the lack of specificity in record locations using the *exact extractr* package⁵⁰. Averaged weather 314 variables and weather effects for alternate buffer radii (50m and 50km) were highly correlated (Fig. S2; 315 Fig. S7). Generally, we expect that organisms will respond to deviations in the weather compared to the 316 average values, as opposed to raw weather variables. Furthermore, across the globes surface the 317 variance in weather variables changes substantially, which may influence population responses. Thus, 318 we explored population responses for the key weather variable of standardised annual anomalies, and 319 then validated our approach using annual weather variance. For the anomalies, we decomposed z-320 transformed averaged monthly weather data for each location for the full timeseries (1979-2013) using 321 a Seasonal-Trend Decomposition by Loess (STL). We used a seasonal window of 7 and trend window 322 of 1000 for the decomposition and extracted the anomaly component, which describes the remainder 323 when accounting for the trend and seasonal components of the timeseries. We then used annual mean 324 temperature and precipitation anomalies as the key weather variables in subsequent analyses. These 325 weather anomalies are the average deviation of the temperature and precipitation from expected values 326 in a given year. We also investigated population responses to weather using annual weather variance, 327 which was calculated as the annual variance of monthly mean temperature and total precipitation values.

328

329 *Species-level life-history and phylogeny*

We used three key traits that broadly characterise species-level life-history that are available for a large number of species: maximum longevity, litter size and adult body mass. We collected these traits from the compendium developed by Conde et al.²⁸, combining information from three primary database

333 sources: The Amniote Life-History Database²⁹, PanTHERIA⁵¹ and AnAge⁵² databases. Adult body

mass data was obtained exclusively from the Amniote Life-History Database²⁹. Where multiple records 334 335 were available for a single species, we took the largest maximum longevity value and the mean litter 336 size/adult body mass. We removed erroneous raw litter size data for Hydrochoerus hydrochaeris and 337 Marmota broweri. For analysis, we z-transformed the natural-logarithm of raw life-history trait data, 338 and verified that the life-history variables were represented across the range of weather anomaly 339 variables in the raw data (Fig. S3). The mammal phylogeny was obtained from Upham et al.³⁰, which 340 uses a 'backbone-and-patch' Bayesian approach for a newly assembled 31-gene supermatrix and is part 341 of the Vertlife project (https://vertlife.org/). We used the maximum clade credibility tree in analysis, 342 which was processed using the *ape* package⁵³. Loxodonta cyclotis (African forest elephant) was 343 considered as Loxodonta Africana (African elephant) for analysis so that the abundance record and 344 phylogenetic data matched.

345

346 Weather effects on annual population growth rates

To assess comparative population responses to weather in the terrestrial mammals we used a two-step meta-regression approach. First, for each record we estimated the effect of annual weather anomalies (and weather variance) on population growth rates. We calculated the realised, proportional population growth rate in year *t* as

351

$$\lambda_t = \frac{X_{t+1}}{X_t}, \qquad (1)$$

where X is the natural logarithm of abundance in year t. Then, with λ_t as the response variable, we 352 353 estimated the effect of temperature and precipitation anomalies on population growth using generalised 354 additive mixed models (GAMMs) fit using the gamm function of the mgcv package⁵⁴. Changes in 355 abundance are influenced by several drivers of population dynamics including habitat loss² and 356 population processes such as density dependence⁵⁵, which may confound any influence of the weather 357 on abundance. Therefore, because we aimed to assess the isolated impact of weather anomalies, 358 accounting for these trends in abundance and temporal autocorrelation was crucial. We initially 359 explored the extent of autocorrelation in abundance patterns using timeseries analysis and found 360 evidence for lag 1 autocorrelation in abundance, but not for greater lags (Fig. S4-5). Furthermore, we 361 tested the potential impact of density dependence on estimating environmental effects using an 362 autoregressive timeseries simulation and found that environmental effects were robust to density 363 dependence even for short timeseries (Fig. S6). Thus, for each record, we model population growth rate 364 in each year as

365

$$\lambda_t = \beta^0 + \omega W_t + f(y_t), \quad (2)$$

366 where β^0 is the intercept and ωW_t is a linear parametric term with coefficient ω for the weather W367 (temperature or precipitation anomaly) in year t. Here, positive coefficients indicate that positive 368 weather anomalies i.e. hotter/wetter years, were associated with population increases, and *vice versa*. 369 Identical additive regression models were run using weather variances as the weather variable W. The

370 term $f(y_t)$ captures the effect of year y_t as a non-linear trend, where the function f is a thin plate regression spline with a basis dimension of five⁵⁶. The function f was also fitted with an order 1 371 372 autoregressive (AR(1)) correlation structure, as specified in the *nlme* package⁵⁷. Thus, the term $f(y_t)$ 373 incorporates both the non-linear trend in abundance and temporal autocorrelation. Finally, we validated 374 our additive model approach by testing other models to calculate weather effects, including linear 375 regressions both including and excluding temporal trends or density dependence, and a temporally autocorrelated model fit using the glmmTMB package⁵⁸ (Fig. S8-S9). Weather coefficients ω generated 376 377 using linear year effects were positively correlated to those from additive models (Fig. S10).

378

379 Bayesian meta-regression

380 Second, with the weather effects ω from each record as the response variable, we explored comparative 381 patterns in population responses to weather using a Bayesian meta-regression framework implemented 382 in the *brms* package⁵⁹. Separate models were fit for temperature and precipitation. Bayesian meta-383 regression was used to address three key questions: 1) Were there consistent population responses to 384 weather across the terrestrial mammals? 2) How did population responses vary within and between 385 species and were there spatial patterns across biomes? 3) Does life-history predict the magnitude of 386 population responses? To address questions 1 and 2, we used Gaussian models controlling for both 387 phylogenetic and species-level covariance. The full model for record *i* and species *j* is given by equation 388 3

 $\omega \sim \text{MVNormal}(\mu, \mathbf{S})$

 $\alpha_i \sim \text{Normal}(\bar{\alpha}, \sigma_{SPECIES})$

 $\bar{\alpha} \sim \text{Normal}(0, 0.3)$

 $\beta_{BIOME} \sim \text{Normal}(0, 0.15)$

 $\sigma_{SPECIES} \sim \text{Exponential(8)}$

 $\beta_N \sim \text{Normal}(0, 0.5)$ $\sigma_{PHYLO}^2 \sim \text{Exponential}(8)$

 $\mathbf{S} = \sigma_{PHYLO}^2 \mathbf{V}$

 $\mu_i = \alpha_{SPECIES[i]} + \beta_{BIOME[i]} + \beta_N N_i$

Linear model

Varying effects

Priors

- - -

389

390

391

392 393

394

395

396

- 397
- 398

where the weather effect ω (z-transformed for analyses), is given by a multivariate normal distribution with mean μ and phylogenetic covariance matrix **S**. The global intercept is given by $\bar{\alpha}$, which estimates overall patterns in weather effects across records, addressing question 1. We incorporated phylogenetic covariance using a Brownian motion model, with the correlation matrix given by **V** (calculated from the maximum clade-credibility tree) and variance factor σ_{PHYLO}^2 , from which between-species variance was estimated. We incorporated an intercept-only varying effect for species with the term $\alpha_{SPECIES[j]}$, from which within-species variance was estimated with $\sigma_{SPECIES}$. The term β_{BIOME} gives the spatial 406 effect of biome on weather responses. Thus, estimating within ($\sigma_{SPECIES}$) species variance, between 407 (σ_{PHYLO}^2) species variance and the spatial effect of biome (β_{BIOME}), we explored question 2. All meta-408 regression models also included the linear effect of record length N (scaled number of years in the 409 record) on weather effects, which was estimated using β_N . For all meta-regression models, we used 410 regularising priors obtained from prior predictive simulations of the slope, intercept and exponential 411 variance terms^{60,61}, to reflect the constraints in the raw data across species (see S1 and Fig. S11-14 for 412 details). Gaussian meta-regression models were also fit for weather effects calculated using the annual 413 weather variance, and the results obtained were largely identical to those obtained for weather anomalies 414 (Fig. S21).

415

For question 3, although on average we expect that species life-history influences the magnitude of responses to the environment, we have little evidence to suggest that life-history *per se* influences the directionality of responses²⁶. Thus, to address this question we explored how maximum longevity, litter size and adult body mass influenced the absolute magnitude of weather responses, $|\omega|$, using Gamma regression models with a log link. The full model for record *i* and species *j* is given by equation 4

421

$$\begin{array}{l} \underline{\text{Linear model}} \\ |\omega| \sim \text{Gamma}(\eta, \mu) \\ \log \mu_i = \alpha_{SPECIES[j]} + \gamma_{PHYLO[j]} + LH + \beta_N N_i \\ \hline \\ \underline{\text{Varying effects}} \\ \alpha_j \sim \text{Normal}(\bar{\alpha}, \sigma_{SPECIES}) \\ \gamma_j \sim \text{MVNormal}(0, \mathbf{S}) \\ \mathbf{S} = \sigma_{PHYLO}^2 \mathbf{V} \\ \hline \\ \frac{Priors}{\bar{\alpha}} \sim \text{Normal}(0, 0.3) \\ \beta_{LH} \sim \text{Normal}(0, 0.2) \\ \beta_N \sim \text{Normal}(0, 0.3) \\ \sigma_{PHYLO}^2 \sim \text{Exponential}(11) \\ \sigma_{SPECIES} \sim \text{Exponential}(2) \\ \eta \sim \text{Gamma}(2, 0.6) \end{array}$$

422

423 where η is a shape parameter that was fit with a Gamma prior, and LH refers to a set of linear life-424 history terms $(\beta_1 x_1 + \cdots + \beta_k x_k)$ that were explored using model selection. Specifically, for the three life-425 history traits, we explored a set of models incorporating univariate, multivariate and 2-way interaction 426 terms, as well as a base model excluding all life-history effects. For the full set of ten candidate models 427 please refer to the supplementary information (table S1 & S2). All life-history effects were fit with the 428 same Normal prior, with mean 0 and standard deviation 0.3 (S1; Fig. S13). We assessed the predictive 429 performance of candidate models using leave-one-out cross-validation implemented in the loo 430 package⁶². Models were compared using the Bayesian LOO estimate of out-of-sample predictive performance, or the expected log pointwise predictive density (elpd)⁶². All final meta-regression models 431

432 were run over 3 Markov chains, with 4000 total iterations and 2000 warmup iterations per chain. Model

- 433 convergence was assessed by inspecting Markov chains, and the degree of mixing between chains using
- 434

Â.

435

436 Acknowledgements

- 437 We extend a warm thank you to the Living Planet Index team from the Zoological Society of London
- 438 and the World Wildlife Foundation for the opportunity to work with this rich data source, and all the
- 439 contributors to this amazing resource. We also thank Dylan Z Childs for his support on modelling the 440 impact of density dependence on the abundance timeseries. We thank Rob Goodsell for advice on the
- 441 Bayesian modelling framework, Christopher Cooney for advice on the mammal phylogeny and the
- 442 Vertlife project, and Morgane Tidière for stimulating discussion and advice on the project. Thank you
- 443 also to Dalia Conde and Johanna Stärk for their help with the demographic data from Conde et al. [28],
- 444 and the members of the Interdisciplinary Centre for Population Dynamics (CPop), and in particular Jim
- 445 Oeppen, for useful feedback on the methodology. This work was supported by the Danish Independent
- 446 Research Fund.
- 447

448 References

- 449 1. Díaz, S. et al. IPBES: Summary for policymakers of the global assessment report on biodiversity and ecosystem 450 services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services. www.ipbes.net (2019).
- 451 452 453 2. Daskalova, G. N. et al. Landscape-scale forest loss as a catalyst of population and biodiversity change. Science 368, 1341-1347 (2020).
- 454 455 3. Soroye, P., Newbold, T. & Kerr, J. Climate change contributes to widespread declines among bumble bees across continents. Science 367, 685-688 (2020).
 - 4. Almond, R., Grooten, M. & Petersen, T. Living Planet Report 2020 : Bending the Curve of Biodiversity Loss. (2020).
- Thomas, C. D. et al. Extinction risk from climate change. Nature 427, (2004). 5.
- 455 456 457 458 459 460 6. Spooner, F. E. B., Pearson, R. G. & Freeman, R. Rapid warming is associated with population decline among terrestrial birds and mammals globally. Global Change Biology 24, 4521-4531 (2018).
 - 7. Pacifici, M. et al. Species' traits influenced their response to recent climate change. Nature Climate Change 7, 205-208 (2017).
- 461 462 463 8. Trisos, C. H., Merow, C. & Pigot, A. L. The projected timing of abrupt ecological disruption from climate change. Nature 580, 496-501 (2020).
- 464 465 466 9. Williams, J. J., Bates, A. E. & Newbold, T. Human-dominated land uses favour species affiliated with more extreme climates, especially in the tropics. *Ecography* 43, 391–405 (2020).
- 10. Brook, B. W., Sodhi, N. S. & Bradshaw, C. J. a. Synergies among extinction drivers under global change. Trends in 467 468 Ecology and Evolution 23, 453–460 (2008).
- 11. Bellard, C., Bertelsmeier, C., Leadley, P., Thuiller, W. & Courchamp, F. Impacts of climate change on the future of biodiversity. Ecology Letters vol. 15 365-377 (2012).
- 12. Antão, L. H. et al. Temperature-related biodiversity change across temperate marine and terrestrial systems. Nature Ecology and Evolution 4, 927–933 (2020).
- 408 469 470 471 472 473 474 Moritz, C. & Agudo, R. The Future of Species Under Climate Change: Resilience or Decline? Science 341, 504-508 13. (2013).
- 14. Chen, I.-C., Hill, J. K., Ohlemüller, R., Roy, D. B. & Thomas, C. D. Rapid Range Shifts of SpeciesAssociated with 475 476 477 High Levelsof Climate Warming. Science 333, 1019–1024 (2011).
- 15. Dornelas, M. et al. A balance of winners and losers in the Anthropocene. Ecology Letters vol. 22 847-854 (2019).
- 16. Leung, B. et al. Clustered versus catastrophic global vertebrate declines. Nature 588, 267-271 (2020).
- 478 17. Coulson, T. et al. Age, Sex, Density, Winter Weather, and Population Crashes in Soay Sheep. Science 292, 1528-479 480 1531 (2001).
- Cordes, L. S. et al. Contrasting effects of climate change on seasonal survival of a hibernating mammal. Proceedings 18. 481 of the National Academy of Sciences 117, undefined (2020).
- 482 19. Layton-Matthews, K. et al. Environmental change reduces body condition, but not population growth, in a high-483 arctic herbivore. Ecology Letters (2020) doi:10.1111/ele.13634.

484 20. Paniw, M., Maag, N., Cozzi, G., Clutton-Brock, T. & Ozgul, A. Life history responses of meerkats to seasonal 485 changes in extreme environments. Science 363, 631-635 (2019).

486 21. Paniw, M. et al. Global analysis reveals complex demographic responses of mammals to climate change. in review.

- 487 Woodroffe, R., Groom, R. & McNutt, J. W. Hot dogs: High ambient temperatures impact reproductive success in a 22. 488 tropical carnivore. Journal of Animal Ecology 86, 1329-1338 (2017).
- 489 23. Compagnoni, A. et al. Herbaceous perennial plants with short generation time have stronger responses to climate 490 anomalies than those with longer generation time. Nature Communications 12, (2021).
- 491 24. Stearns, S. C. The evolution of life histories. (Oxford University Press, 1992).
- 492 Cayuela, H. et al. Life history tactics shape amphibians' demographic responses to the North Atlantic Oscillation. 25. 493 Global Change Biology 23, 4620–4638 (2017).
- 494 Morris, W. F. et al. Longevity can buffer plant and animal populations against changing climatic variability. Ecology 26. 495 89, 19-25 (2008).
- 496 27. Capdevila, P., Stott, I., Berger, M. & Salgero-Gómez. Towards a Comparative Framework of Demographic 497 Resilience. Trends in Ecology & Evolution 35, (2020).
- 498 28. Conde, D. A. et al. Data gaps and opportunities for comparative and conservation biology. Proceedings of the 499 National Academy of Sciences (2019) doi:10.5061/dryad.
- 500 29. Myhrvold, N. P. et al. An amniote life-history database to perform comparative analyses with birds, mammals, and 501 502 reptiles. Ecology 96, 3109 (2015).
- 30. Upham, N. S., Esselstyn, J. A. & Jetz, W. Inferring the mammal tree: Species-level sets of phylogenies for questions 503 in ecology, evolution, and conservation. PLoS Biology 17, (2019).
- 504 James, T. D., Salguero-Gómez, R., Jones, O. R., Childs, D. Z. & Beckerman, A. P. Bridging gaps in demographic 31. 505 analysis with phylogenetic imputation. Conservation Biology (2020) doi:10.1111/cobi.13658.
- 506 32. IUCN. The IUCN Red List of Threatened Species. http://www.iucnredlist.org (2016).
- 507 508 Richards, C., Cooke, R. S. C. & Bates, A. E. Biological traits of seabirds predict extinction risk and vulnerability to 33. anthropogenic threats. Global Ecology and Biogeography geb.13279 (2021) doi:10.1111/geb.13279.
- 509 34. Moorad, J. A., Promislow, D. E. L., Flesness, N. & Miller, R. A. A comparative assessment of univariate longevity 510 measures using zoological animal records. Aging Cell 11, 940–948 (2012).
- 511 35. Jackson, J., Childs, D. Z., Mar, K. U., Htut, W. & Lummaa, V. Long-term trends in wild-capture and population 512 dynamics point to an uncertain future for captive elephants. Proceedings of the Royal Society B: Biological sciences 513 286, 20182810 (2019).
- 514 515 36. Desforges, J.-P. et al. Predicting global killer whale population collapse from PCB pollution. Science (2018).
- 37. Salguero-Gómez, R. et al. COMADRE: A global data base of animal demography. Journal of Animal Ecology 85, 516 371-384 (2016).
- 517 518 38. Wauchope, H. S., Amano, T., Sutherland, W. J. & Johnston, A. When can we trust population trends? A method for quantifying the effects of sampling interval and duration. Methods in Ecology and Evolution 10, 2067–2078 (2019).
 - 39. Engbo, S. et al. Census data aggregation decisions can affect population-level inference in heterogeneous populations. Ecology and Evolution 10, 7487–7496 (2020).
 - 40. Römer, G. et al. Drivers of large-scale spatial demographic variation in a perennial plant. Ecosphere 12, (2021).
 - 41. Chaudhary, V. & Oli, M. K. A critical appraisal of population viability analysis. Conservation Biology vol. 34 26-40 (2020)
- 519 520 521 522 523 524 525 526 527 528 529 Benito Garzón, M., Robson, T. M. & Hampe, A. ATraitSDMs: species distribution models that account for local 42. adaptation and phenotypic plasticity. New Phytologist vol. 222 1757-1765 (2019).
 - Kissling, W. D. et al. Towards global data products of Essential Biodiversity Variables on species traits. Nature 43. Ecology and Evolution vol. 2 1531-1540 (2018).
- 44. R Core Team. R: A language and environment for statistical computing. (2021).
- 45. Chamberlain, S. et al. taxize: Taxonomic information from around the web. (2020).
- 530 46. Karger, D. N. et al. Climatologies at high resolution for the earth's land surface areas. Scientific Data 4, (2017).
- 531 47. Hijmans, R. J. raster: Geographic Data Analysis and Modeling. (2020).
- 532 Bivand, R. & Rundel, C. rgeos: Interface to Geometry Engine - Open Source ('GEOS'). (2020). 48.
- 533 534 49. Pebesma, E. Simple Features for R: Standardized Support for Spatial Vector Data. The R Journal 10, 439-446 (2018).
- 535 50. Baston, D. exactextractr: Fast Extraction from Raster Datasets using Polygons. (2020).
- 536 537 51. Jones, K. E. et al. PanTHERIA: a species-level database of life history, ecology, and geography of extant and recently extinct mammals. Ecology 90, (2009).
- 538 539 Tacutu, R. et al. Human Ageing Genomic Resources: Integrated databases and tools for the biology and genetics of 52. ageing. Nucleic Acids Research 41, (2013).
- 540 53. Paradis, E. & Schliep, K. ape 5.0: an environment for modern phylogenetics and evolutionary analyses in R. 541 Bioinformatics 35, 526–528 (2019).
- 542 54. Wood, S. N. Generalized Additive Models: An Introduction with R. (Chapman and Hall/CRC, 2017).
- 543 55. Brook, B. W. & Bradshaw, C. J. A. Strength of evidence for density dependence in abundance time series of 1198 544 species. Ecology vol. 87 (2006).
- 545 56. Wood, S. N. Thin-plate regression splines. Journal of the Royal Statistical Society. Series B (Methodological) 65, 546 95-114 (2003).
- 547 57. Pinheiro, J., Bates, D., DebRoy, S. & Sarkar, D. nlme: Linear and nonlinear mixed effects models. 1-117 (2014).
- 548 58. Brooks, M. R. et al. glmmTMB Balances Speed and Flexibility Among Packages for Zero-inflated Generalized 549 Linear Mixed Modeling. The R Journal 9, (2017).

- 550 551 552 553 554 555 556 59. Bürkner, P. C. brms: An R package for Bayesian multilevel models using Stan. Journal of Statistical Software 80, (2017).
 - 60. McElreath, R. rethinking: Statistical Rethinking book package. (2020).
 - 61. McElreath, R. Statistical rethinking: A Bayesian course with examples in R and Stan. (CRC press, 2020).
 - 62. Vehtari, A., Gelman, A. & Gabry, J. Practical Bayesian model evaluation using leave-one-out cross-validation and WAIC. Statistics and Computing 27, 1413–1432 (2017).
- 557
- 558
- 559
- 560
- 561

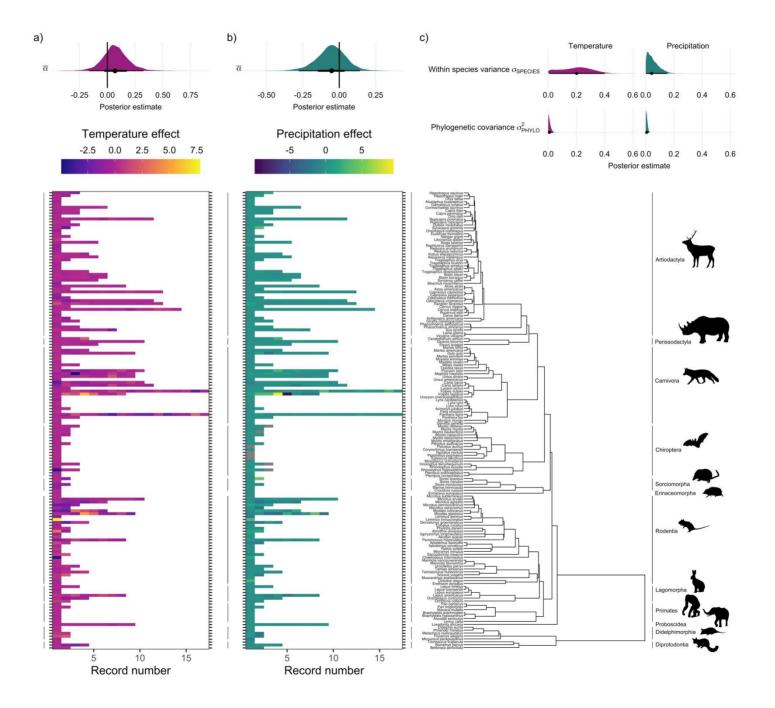
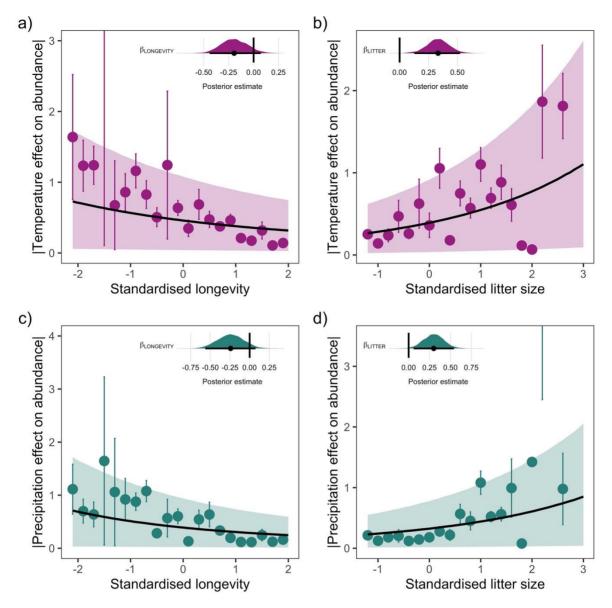


Figure 1. Global population responses to weather in the terrestrial mammals. Heatmaps for 563 564 population responses to temperature (a) and precipitation (b) for 157 species of terrestrial mammal. 565 Each row of the heatmap is a species, and coloured rectangles are the population records. The colour 566 denotes the coefficient of temperature/precipitation effects derived from autoregressive additive 567 models. Here, positive numbers indicate that positive temperature/precipitation anomalies 568 (hotter/wetter than average in a given year) were associated with increases in population size, and vice 569 versa. The distribution half-eye plots in (a) and (b) (top) are summaries of the posterior distribution for 570 the global intercept ($\bar{\alpha}$) of temperature and precipitation responses across records, fit with a Gaussian 571 Bayesian meta-regression. The points give the approximate posterior mean and the error bar is 572 calculated using a cumulative distribution function. Bayesian models were fit incorporating phylogenetic covariance using the maximum clade credibility tree from Upham et al.⁵⁹, which is plotted 573 on the right with annotations indicating the mammal order 60 . The distribution half-eye plots in (c) are 574 575 the posterior distribution summaries for phylogenetic covariance and within-species variance from the 576 Gaussian Bayesian meta-regression.





578 Figure 2. Life-history predicts population responses to weather in the terrestrial mammals. Each 579 panel presents the mean absolute effect of temperature (a & b) and precipitation on population growth 580 rates, $|\omega|$, for standardised maximum longevity (a & c) and standardised mean litter size (b & d). 581 Standardisation was performed using a z-transformation of natural-log of the raw life-history trait. The 582 values on each x-axis are split into equal bins of 0.2 from the minimum to the maximum life-history 583 value. Points are coefficient means, with standard error bars. The black lines are the mean posterior 584 predictions from the best predictive model, where predictions were calculated averaging over all other 585 covariates and varying effects in the model. The shaded intervals are the 80% prediction intervals calculated in the *rethinking* package⁶⁰. Panel insets give posterior distribution summaries for the slope 586 587 terms presented in each panel. Two points are omitted from the plotting panel due to large mean 588 coefficient values and high standard errors, which are visible on the plot.

589