1 2 3	Divergent epigenetic profiles from two differentially impacted wild populations of estuarine cordgrass (<i>Sporobolus alterniflorus</i>)
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- 36 VM, LP and WW designed and executed methylation studies. LD, FM, MN, JB, LC, OC, RC,
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- 41
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45 Highlights for manuscript submission:

- estuarine grasses native to the Bronx River, NY face stresses associated with low
 dissolved oxygen and urbanization
- 48 differentially impacted populations of estuarine grasses exhibit inverse global DNA
 49 methylation profiles in response to acute heat stress
- DNA methylation may represent a mechanism by which plants transiently respond to
 environmental stressors, and this may represent a form of rapid adaptive evolution
- stress priming by transgenerational epigenetic modification may enhance fitness in
 grasses native to the heavily impacted Bronx River estuary
- 54

55 Abstract

56 The effects of urbanization on watershed ecosystems present critical challenges to 57 modern survival. Organisms in urbanized areas experience high rates of evolutionary change, but 58 genetic adaptation alone cannot mitigate the rapid and severe effects of urbanization on biodiversity. Highly resilient, foundation species are key to maintaining an ecosystem's integrity 59 60 in the face of urban stressors. However, the rapid collapse and disappearance of watershed 61 ecosystems calls into question the extent to which we can rely on such species for their services. 62 Our research investigates the molecular mechanisms by which the foundation ecosystems 63 provider, Sporobolus alterniflorus, adapts to life in an urbanized environment. To elucidate these 64 mechanisms, we quantified changes in global DNA methylation (% 5-mC) as a result of acute 65 heat stress. Specimens from two differentially impacted populations across an urban to suburban 66 geographical transect formed the basis of this study. These two populations of Sporobolus alterniflora exhibit inverse global DNA methylation patterns when exposed to the same acute 67 68 heat stress. Our findings suggest that epigenetic mechanisms, such as DNA methylation, control 69 rapid and transient adaptation, in the form of differential stress responses, to distinct environment 70 challenges.

71

72 1- Introduction

Urban estuaries are highly productive ecosystems that provide critical services including
storm surge protection, water filtration and wildlife habitat (Barbier et al., 2011; Franca et al.,
2012). Coastal salt marshes comprise approximately 25% of the global soil carbon sink through

76 plant production and high carbon burial rates (Chmura et al., 2003). Human activity,

industrialization, and climate change continue to negatively impact estuarine ecosystems,

especially in urbanized areas (Limburg et al., 2005; Astaraie-Imani et al., 2012; Chin et al.,

79 2013). Estuarine species face stresses associated with adverse environmental conditions such as

80 low dissolved oxygen, combined sewage overflow, toxin contamination, bank destabilization,

81 habitat degradation, and extreme temperature fluctuation (Van Dolah, et al., 2008; Courrat et al.,

82 2009; Halem et al., 2014; Ravaschiere et al., 2017). Knowledge of the molecular mechanisms

underlying stress adaptation is essential if we hope to rehabilitate estuarine ecosystems adverselyaffected by urbanization.

Urbanization is often associated with rapidly changing anthropogenic stressors (Alberti,
2015; Donihue and Lambert, 2015). As sessile organisms, plants must respond to these
environmental challenges by rapidly regulating gene expression, and this is often accomplished
via epigenetic alterations such as DNA methylation and histone modification (Arikan et al.,
2018; Wang et al., 2010). These transient and rapid modifications provide essential "on demand"
phenotypic variation (Rey et al., 2016), that may represent critical adaptive mechanisms for
species such as marsh grasses that provide essential ecological services to urban communities.

92 Here, we take an *in vivo* approach to the study of epigenetic adaptation in wild 93 populations of smooth cordgrass Sporobolus alterniflorus (formerly Spartina alterniflora), a 94 critical ecosystems provider and foundation species that is native to estuaries throughout the 95 North American east coast (Gedan and Bertness, 2010; Peterson, 2014). This halophyte species 96 is particularly resilient in the face of urbanization (Gedan and Bertness, 2010) and thus 97 represents a particularly suitable model for our study. In research findings presented here, we 98 track two robust mechanisms for transient stress response in plants: genomic DNA cytosine 99 methylation (% 5-mC) and heat shock protein 70 expression (HSP70). The methylation of 5-100 cytosine (% 5-mC) occurs when a methyl group is enzymatically attached to the 5' carbon of 101 cytosine's pyrimidine ring, resulting in a decrease in gene expression at the methylation site. 102 This reversible chemical modification elicits transient phenotypic changes required for stress 103 response, especially in plants (Meyer, 2008; Boyko and Kovalchuk, 2011; Arikan et al., 2018). 104 HSP70 chaperone protein is a well-characterized and universal stress responder that facilitates 105 refolding of denatured proteins, and thus restoration of vital metabolic activities (Walter and 106 Ron, 2011).

107 We compared global DNA methylation (% 5-mC) in response to acute heat stress in two 108 differentially impacted wild populations: the Bronx River estuary, New York and the Greenwich 109 Cove estuary, Connecticut, using HSP70 as an indicator of acute stress response. These estuaries 110 are located about 25 miles from each other and share similar biota. However, their diametric 111 histories make these estuaries uniquely suited for a comparative analysis of urban stress 112 response. The Bronx River estuary has a history of factory waste and sewage dumping 113 (Crimmens and Larson, 2006). Construction of the Bronx River Parkway led to degradation of 114 water quality and a decrease in species diversity (Rachlin, 2007). Ongoing industrialization 115 eventually led to benzo-a pyrene (Litten et al., 2007) and sewage contamination (Rachlin, 2007; 116 Wang and Pant, 2010) of the Bronx River estuary. Whereas, the suburban Greenwich Cove 117 estuary lies at a greater geographic distance from urbanized centers and thus experiences 118 comparatively less impact from urbanization and industrialization (Halem et al., 2014; 119 Ravaschiere et al., 2017). Previous studies of the Bronx River estuary document endocrine 120 disruption (Halem et al., 2014) and altered heat shock response (Ravaschiere et al., 2017) in the 121 native ecosystem provider (Galimay et. al., 2017) Atlantic ribbed mussel (Geukensia demissa). 122 In addition, our nine-year longitudinal study, demonstrates that dissolved oxygen (DO) 123 concentrations in the Bronx River estuary remain consistently lower than those recorded from 124 the Greenwich Cove estuary (Table 1). This is likely a result of ongoing urbanization proximal to 125 the Bronx River estuary (Slattery, 2018). Consistent and long-lasting differences in dissolved 126 oxygen levels from these two estuaries lend quantitative support to our claim that salt marsh 127 grasses native to the Bronx River estuary are coping with challenges that rarely arise for their 128 Greenwich Cove conspecifics.

Ultimately, the health of urban wetlands will depend on phenotypic adaptation of native species to the pressures introduced by anthropogenic processes (Alberti, 2015). Results of this study provide insight into the molecular mechanisms deployed by a dominant and foundational ecosystem provider that is facing consistent and long-term environmental challenge. Our results will help to inform future conservation and management decisions regarding critical urban plant species that must persist regardless of increasing urbanization and environmental degradation.

135

136 2- Materials and Methods

137

138 2.1 Water Collection and analyses

139 Concentrations of dissolved oxygen (DO) were obtained using the Winkler titration method.

- 140 Oxygen was fixed on site using manganous sulfate, alkaline potassium iodide azide, and
- 141 sulfamic acid. Sodium thio- sulfate was used to titrate the water sample with starch indicator to a
- 142 clear endpoint.
- 143

144 *2.2 Collection of plant material and heat shock*

145 *Sporobolus alterniflorus* (formerly *Spartina alterniflora*) plants were removed from their native

- 146 environments in Harding Lagoon near Soundview Park, Bronx River estuary, NY (40°48'35.6"N,
- 147 73°34'18.0"W), and at Todd's Point estuary in Greenwich Cove, CT (41°0' 30.8088"N,
- 148 73°34'14.1709"W). Plants were taken along with their roots, soil, and native water at low tide.

149 Plants were transported to the laboratory, where each plant was transplanted into native soil in a

150 separate plastic cup with holes punched in the sides and base to allow for adequate aeration. All

151 plants were equilibrated at approximately 23°C in a large basin of native water with aeration, for

152 at least 24 hours. For the initial heat shock, four individual plants from each were incubated in a

153 stand-up incubator at $42 \pm 2^{\circ}$ C for 30 minutes in pre-heated native water, and the incubator was

154 left dark. Water was not aerated during treatment. Concurrently, four individual plants from each

site (8 total plants), were incubated in their respective water at room temperature $(21 \pm 1^{\circ}C)$ for

156 30 minutes in the dark without aeration to serve as controls. After treatments, all plants were

157 moved to room temperature (both air and native water), ambient light, with no aeration to rest for

158 30 minutes. After resting, plants underwent a second treatment, heat shock or control, as

159 described above. Post second treatment, all plants were moved to room temperature air and

160 native water and allowed to rest overnight in ambient light with aeration. Grass samples were

161 collected over multiple years, at the same time of year, and on the same tides, and as such

- 162 represent appropriate biological replicates.
- 163

164 *2.3 DNA extraction*

DNA was isolated using the Qiagen DNeasy Plant Mini Kit (Qiagen, #69104). Leaves were cut with clean dissecting scissors from near the stem of each individual plant and wiped clean of any dirt or debris. Plant matter (0.5 g) was taken for each treatment from each of four plants. Plant matter was pooled and cut into small pieces with clean dissecting scissors and crushed with a

169	mortar and pestle and 400 μ L AP1 (lysis buffer) + and 4 μ L RNase. All contents were transferred
170	to a micro centrifuge tube. Pipetting up and down further agitated samples. Samples were
171	vortexed and incubated in a heat block at 65°C for 10 minutes, vortexing 2-3 times during
172	incubation. The remaining steps were performed according to the Qiagen DNeasy Plant Mini
173	Kit. DNA concentrations and quality were assessed using Nano Drop Microvolume
174	Spectrophotometer (ThermoFisher Scientific). DNA samples were stored at -20°C in AE buffer
175	from Qiagen DNeasy Plant Mini Kit.
176	
177	2.4 Global DNA methylation analysis
178	Global genomic methylation levels were determined using 100 ng of DNA for each sample using
179	the MethylFlash Methylated DNA Quantification Kit Colorimetric (Epigentek, #P-1034). The

absolute methylation level was determined using a standard curve. The percentage of globally

- 181 methylated DNA was calculated using absolute quantification from the Epigentek Protocol.
- 182

183 2.5 Protein extraction and concentration determination

184 Leaves were cut with clean dissecting scissors from near the stem of each individual plant and 185 wiped clean of any dirt or debris. Plant tissue (500 mg) was pooled from four individuals from 186 each site/condition. Tissue was immediately placed into lysis bags provided in the P-PER Plant 187 Protein Extraction Kit (Thermo Scientific, #89803, Rockford, IL) and lysed according to 188 provided instructions with the following modification: the volume of working solution was 189 reduced to half of the recommended volume per sample. Protein solutions were stored at -20 °C 190 for no longer than 4 weeks. Concentrations of proteins in plant leaves were determined according 191 to the Pierce BCA Protein Assay Kit-Reducing Agent Compatible (Thermo Scientific, #23250, 192 Rockford, IL).

193

194 2.6 Quantification of heat shock protein

- 195 Specific heat shock protein (HSP70) protein levels were quantified by Western blot (SDS-
- 196 PAGE) analysis. Plant leaf protein samples (30-100 μg) were run on a 10% Mini-PROTEAN[®]
- 197 TGXTM Precast Protein Gel (Bio-Rad) and transferred to PVDF membranes, Ponceau stained to
- 198 visualize protein transfer and evenly loaded across samples. Membranes were blocked for 1 hour
- 199 at room temperature overnight at 4 °C in Superblock (Thermo-scientific, #37535, Rockford, IL).

- 200 After blocking, blots were incubated in mouse HSP70 monoclonal antibody (1:1000, Enzo,
- 201 #AZI-SPA-820, Plymouth Meeting, PA). After washing 3x 10min in TBS + Tween20,
- 202 membranes were incubated in goat anti-mouse secondary antibody (1:10,000, Thermo-scientific,
- 203 #31430, Rockford, IL) for 1 hour at room temperature. Membranes were washed 3 x 10min in
- TBS + Tween 20 before proteins were detected by Immuno-Star luminol-peroxide (Bio-Rad,
- #170-5070). Bio-Rad Chemi-doc was used to visualize and quantify proteins with densitometry.
- 206

207 2.7 DNA barcoding

208 DNA isolated from 12 organisms from each site (described above) was used to PCR amplify the 209 universal plant gene *rbcL* using the following primers:

210 *rbcL*aF 5'- TGTAAAACGACGGCCAGTATGTCACCACAAACAGAGACTAAAGC-3'

211 *rbcLa* rev 5'CAGGAAACAGCTATGACGTAAAATCAAGTCCACCRCG-3'

- 212 Primers were a generous gift from the DNA Learning Center, Harlem, NY. PCR reactions were
- 213 performed using GE illustra PURETaq Ready-To-Go PCR beads, 25µL of PCR reaction, 2µL of
- template DNA, and 11.5µL of primers. Reactions were run for 50 cycles: 30 seconds
- 215 denaturation at 94°C, 45 seconds annealing at 54°C, and 45 seconds extending 72°C using
- 216 Techne Genius Thermal Cycler. PCR amplicons were visualized on a 2% agarose gel using
- 217 pBR322/BstNI molecular weight standards to ensure that products were the predicted size. PCR
- 218 products were sent to Genewiz Inc. for sequencing. Sequences were aligned using Nucleotide
- 219 BLAST. Percent similarity was ascertained using the CLUSTAL W (Thompson et al., 1994; Ni
- et al., 2012).
- 221

222 2.8 Statistical analysis

- 223 All water analysis data are presented as mean \pm SD. p < 0.0001 for dissolved oxygen
- 224 concentrations, p=0.0063 for pH values. Error bars and error ranges display \pm S.E.M. as in Table
- 1 and Figure 2 respectively. P-values were determined through unpaired t-tests for statistical
- comparisons of two discrete populations as in Table 1. ANOVA tests for statistical comparisons
- of more than two experimental groups and/or discrete populations were performed as in Figure 2.
- 228 P-values less than or equal to 0.05 are assumed to display statistical significance.
- 229
- **3- Results**

231

232 *3.1 Water analyses*

233	Over a 9-year survey of the waters in the Bronx River estuary versus its less impacted
234	counterpart in Greenwich Cove estuary, we collected physicochemical data regarding pH,
235	temperature and dissolved oxygen levels. As shown in Table 1, the temperature and pH of the
236	two sites are comparable. Temperature and pH for Bronx River estuary are 21.93 ± 1.05 °C and
237	7.28 ± 0.05 , while Greenwich Cove estuary are 24.62 ± 0.807 °C and 7.89 ± 0.024 . However,
238	dissolved oxygen concentrations at the two sites consistently vary widely, with the Bronx River
239	estuary at 5.89 ± 0.30 mg/L and Greenwich Cove estuary at 9.69 ± 2.2 mg/L. This difference is
240	statistically significant ($p < 0.0001$). The National Oceanic and Atmospheric Administration
241	(NOAA) National Estuarine Eutrophication Survey classifies water quality falling between 2 and
242	5 mg/L dissolved oxygen as stressed (Bricker et al., 1999). These data indicate that organisms
243	residing in the Bronx River estuary are exposed to chronic, low oxygen levels, a condition that
244	induces stress and metabolic dysfunction (Halem, et al., 2014; Ravaschiere et al., 2017).

Site	Dissolved O ₂ (mg/L)	Temperature (°C)	рН
Bronx River	5.59 ± .302	21.93 ± 1.05	7.28 ± 0.05
Greenwich Cove	9.10 ± .492	24.62 ± .807	7.89 ± .024

245

Table 1. Estuarian water metrics. Average levels of dissolved oxygen (DO) over a 9-year survey, water
temperature over a 7-year survey, and pH over a 5-year survey of the Bronx River estuary, NY and
Greenwich Cove, CT. All water collections occurred in June and July, around low tide. Data are presented
as mean ± SD. p < 0.0001 for dissolved oxygen concentrations, p=0.0063 for pH values.

251 3.2 Global methylation in response to stress for grasses from the differentially impacts sites 252 As shown in Figure 2, acute heat stress, as monitored by heat shock protein levels, results in 253 inverse global DNA methylation profiles when comparing grasses collected from sites with 254 distinctly different levels of environmental challenge. Percent global DNA methylation is 255 increased by heat shock for grasses from the Bronx River estuary but is decreased by heat shock 256 for Greenwich Cove organisms. These results are statistically significant (p values < 0.05). 257 Interestingly, Bronx River grasses consistently demonstrate lower baseline global methylation 258 levels as compared with their Greenwich Cove conspecifics (in Figure 1, compare Bronx and 259 Greenwich controls).



260

Figure 1: Quantification of global DNA methylation over two years. Average percent 5-methlycytosine (5-mC) of *S. alterniflorus* from the Bronx River and Greenwich Cove estuaries (n=4 for each group). The

two populations launch inverse methylation responses in acute heat stress. Bronx individuals increase

- 264 global methylation (p < 0.05), while Greenwich Cove individuals decrease global methylation, (p < 0.05). 265 Statistical results are based on ANOVA tests.
- 266
- 267 *3.3 HSP70 protein levels indicate acute stress*
- 268 Western blots analyses of heat shock immuno-reactive proteins bands (HSP70) across two years
- 269 (Figure 2) show that heat shock proteins were successfully elicited in controlled laboratory
- 270 settings in grasses from both sites.



271

Figure 2: HSP70 Western blot analyses, *S. alterniflorus* (2013 and 2015), collected from Bronx River
 Estuary (top) and Greenwich Cove (bottom). Arrows point to 70 kDa HSP70 immunoreactive band. C=
 control and HS = heat shocked.

275

276 *3.4 Barcoding for species identification*

277 Among *rbcL* chloroplast DNA gene sequences derived from individuals native to both the Bronx 278 River and Greenwich Cove, CLUSTAL W alignment revealed negligible heterogeneity (0 - 1%). 279 In addition, CLUSTAL W alignment also showed insignificant (0 - 1%) divergence in *rbcL* 280 sequence within both populations. These results indicate that all specimens used in this study 281 from both sites are members of the same species, either Sporobolus alterniflorus or Sporobolus 282 maritimus. Based on geographical data Sporobolus maritimus is not native to the Bronx or 283 Connecticut (USDA, plants database). Thus, we determined that Sporobolus alterniflorus is the 284 species we analyzed at both sites, Greenwich Cove and Bronx River estuaries. 285

286 4- Discussion

287 Rapid global urbanization and its impact on the environment are fundamentally changing
288 the course of evolution in organisms that represent our urban co-inhabitants (Hendry and

Kinnison, 1999; Alberti, 2015; Donihue and Lambert, 2015; Johnson and Munshi-South, 2017).
Plant life is critical to sustainable development of urban ecosystems (Barbier et al., 2011; Mexia et al., 2018). Thus, an important challenge in ecology is to gain a fuller understanding of the
mechanisms by which plants cope with current unprecedented rates of environmental change.
Information presented here contributes to a growing field of research that documents key
molecular mechanisms by which plants adapt to the challenges of urbanization.

295 Studies that focus specifically on wild populations native to urban landscapes are 296 especially well suited for predicting and conserving life in urban ecosystems. To date, few 297 studies employ in vivo techniques to analyze the mechanisms of stress tolerance in wild 298 populations as they respond to urbanization in their native habitats (Thiebau et al., 2019). For 299 this study, two distinct wild populations of estuarine grasses were chosen as the biological 300 models. Sporobolus alterniflorus was collected from two sites: the Bronx River, NY and 301 Greenwich Cove, CT. These estuaries are located along an urban to suburban gradient, that 302 features differential degrees of urbanization both quantitatively and historically (Rachlin et al., 303 2007; Halem et al., 2014; Ravaschiere et al., 2017). Physio-chemical analyses of estuary water, 304 presented in Table 1, demonstrate that the Bronx River estuary consistently experiences long 305 term, comparatively low dissolved oxygen, and these results are statistically significant. Hypoxic 306 conditions in the Bronx River are associated with chronic stress response and metabolic 307 disruptions in a foundation molluscan species (Halem, et al., 2014; Ravaschiere et al., 2017). 308 Low dissolved oxygen primarily results from phosphorus and nitrogen run-off associated with 309 wastewater and urban run-off. In addition, gas exchange in Sporobolus alterniflorus relies on 310 direct access to dissolved oxygen from the aquatic environment via stems and roots that are often 311 tidally submerged. These submerged parts of the plant, exposed to chronically low dissolved 312 oxygen, must fulfill respiratory needs for the entire plant (Teal and Kanwisher, 1966). Thus, a 313 low dissolved oxygen concentration directly inhibits vital metabolic processes and serves as a 314 predictable indicator for urbanization and contamination.

Using estuarine grasses, from two distinct populations as described above, we simulated an abiotic stressor (heat stress) that is commonly associated with urbanization and climate change (Schlesinger et al., 2008; Fossog et al., 2013; Simon et al., 2016; Duarte et al., 2017). We did this by experimentally delivering a controlled heat shock, and monitoring stress response via immunochemical detection of the reliable stress indicator, heat shock protein 70 (HSP70). After 320 heat shock was applied and induction of heat shock protein was confirmed by Western Blot 321 analyses, we quantified changing levels of % 5-mC DNA methylation. This allowed us to 322 monitor epigenetic modification as a response to acute stress in two differentially impacted plant 323 populations. Figure 1 demonstrates that the Bronx River and Greenwich Cove grasses exhibited 324 statistically different inverse DNA methylation profiles in response to the controlled application 325 of acute heat stress. Western Blot analyses (Figure 2) confirm heat stress as indicated by the 326 predicted and significant increases in heat shock protein (molecular weight 70 kD), and this 327 response was documented across two years.

328 Results of our global DNA methylation analyses suggest that the two distinct populations of grasses counter the same heat stress by eliciting DNA methylation profiles that differ 329 330 fundamentally and in biologically relevant ways. Our results show that grasses from the Bronx 331 River estuary counter an acute stress via DNA hyper-methylation. Further studies are necessary 332 to confirm our speculation that hyper-methylation may be an essential metabolic trade-off that 333 globally shuts down gene expression, sparing only the basic housekeeping functions that are 334 necessary for survival during acute stress. Whereas, for grasses native to the less impacted 335 Greenwich Cove, application of acute stress elicits a reverse pattern, a hypo-methylation of the 336 genome that may facilitate expression of those loci required for a successful acute stress 337 response. In support of our findings, multiple recent studies demonstrate that individual plant 338 populations counter similar environmental stressors in different ways, and this may be controlled 339 at the epigenetic level (Saez- Laguna et al., 2014; Thiebaut et al., 2019; Rehman and Tanti, 340 2020).

341 The inverse epigenetic profiling that we report here between Bronx and Greenwich Cove 342 grasses do not result from speciation. Our barcoding results demonstrate that grasses collected 343 from these two geographically distinct populations have negligible differences in chloroplast 344 *rbcL* gene sequence (Kress and Erikson, 2007). Very low rate of sequence divergence (1% or 345 less) between individuals from the Bronx River and Greenwich Cove estuaries indicates that 346 these individuals belong to the same species and are thus genetically identical. Differential DNA 347 methylation within a species is now recognized as a critical indicator of short-term ecological experience (Rey et al., 2020). This notion of "ecological populations" will become essential in 348 349 conservation biology with the recognition that epigenetics offers a key link between 350 environmental change and phenotypic plasticity in wild populations (Rey et al., 2020).

351 Enhanced tolerance to stress could provide an evolutionary advantage for Bronx River 352 grasses when confronted with the unpredictable challenges inherent to their native habitat. A 353 phenomenon called defense priming conditions plants to better tolerate abiotic stressors, and this 354 constitutes an important evolutionary benefit (Matinez-Medina et al., 2011; Crisp et al., 2016). 355 This switch-like and reversible process offers sufficient genomic flexibility for sessile organisms, 356 like plants, to respond rapidly to fluctuating environmental challenges (Meyer, 2008; Boyko and 357 Kovalchuk, 2011; Arikan et al., 2018). Evidence suggests that defense priming improves fitness 358 (Conrath et al., 2006; Matinez-Medina et al., 2011), and is regulated by epigenetic modification 359 (Luna et al., 2012). In defense priming, a "stimulus" targets individual loci such that their 360 methylation states become altered. This affects future accessibility of transcriptional machinery 361 and gene expression eliciting a rapid and sustained response upon a later "triggering stimulus". 362 We hypothesize that long term hypoxia, as well as potential additional environmental challenges 363 at the urbanized Bronx River estuary, constitute the priming "stimulus" for this particular 364 population of grasses. Chromatin restructuring, specifically changes to DNA methylation 365 patterns, is a mechanism by which plants engage in defense priming (Conrath et al., 2015; 366 Savvides et al., 2016; Pastor et al., 2013), and here we demonstrate that two distinct plant 367 populations deploy inverse patterns of global DNA methylation when encountering the same 368 acute stressor. These results suggest that grasses from the Bronx River Estuary may be engaged 369 in defense priming. Our research contributes to rapidly growing body of evidence demonstrating 370 that successful strategies for coping with rapid environmental change do not rely solely on 371 random heritable changes in DNA sequence. For sessile species such as plants, critical 372 evolutionary adaptations are propelled by epigenetic change (Youngson et al., 2008; Holeski et 373 al., 2012; Robertson et al., 2017; Thiebau et al., 2019; Whittle et al., 2019; Rey et al., 2020). 374 In conclusion, current predictions regarding species tolerance to urbanization and climate 375 change are based on the assumption that all members of a species elicit the same, or similar 376 responses to an environmental stressor (Sih et al., 2011). However, data presented here 377 corroborate growing evidence suggesting that populations separated geographically or by 378 differential stressors indeed demonstrate key distinct coping mechanisms (Fossog et al., 2013; 379 Halem et al., 2014; Ravaschiere et al., 2017; De Almeida Duarte et al., 2017). In other words, 380 members of the same species may utilize diverging mechanisms in order to continue to thrive in

381 the face of rapid and challenging environmental stressors. In addition, results presented here

382 suggest that stress priming by epigenetic modification may enhance fitness in grasses native to 383 the heavily impacted Bronx River estuary. These complex molecular mechanisms of biological 384 adaptation deserve our attention so that we may better understand the ways in which particular 385 populations develop unique adaptive responses that are specifically suited to unique ecological 386 challenges.

387 As molecular ecologists and evolutionary biologists, we must accept the notion that 388 environmentally induced adaptation is a key feature to sustaining resilience and biodiversity in 389 rapidly expanding urban areas. As conservationists and urban planners, we must gain better 390 understandings of the unique epigenetic mechanisms by which foundational, ecological service 391 providers adapt and evolve to tolerate the pressures of urbanization. Future studies are required 392 in order to identify genes that are differentially methylated in response to stress in Sporobolus 393 alterniflorus, a species that is vital to urban sustainability. These studies await the full sequence 394 analysis and annotation of the Sporobolus alterniflorus genome. In addition, future studies 395 should be initiated to more fully understand the intricacies of adaptation in populations that 396 remain chronically exposed to low oxygen and other stressors associated with urbanization. 397

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