

# Relative effects of edaphic conditions and climate on palm communities in the Central Andes

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**Running title:** Environmental effects on Andean palm communities

Received: \_\_\_\_\_; Revised: \_\_\_\_\_(optional); Accepted: \_\_\_\_\_.

## ABSTRACT

24 Palms (family Arecaceae) are conspicuous and structural elements in forests ecosystems of  
tropical regions and mountain forests in South America. Additionally, many species of palms are  
26 culturally and economically important to human populations. Because of their ecological and  
ethnobotanical significance, understanding the drivers of palm distribution and diversity is  
28 critical. However, most past research has focused in tropical lowland palm communities, while  
our understanding of montane tropical palm ecology and biogeography is comparatively lacking.  
30 We investigate the environmental factors that influence patterns of richness, composition, and  
abundance of palms in the Central Andes. In particular, we are interested in the relative effects  
32 that soil edaphic conditions and climate have on palm community structure. For these analyses,  
we used a network of 88 forest plots distributed along a broad elevational gradient (1,000 – 3,200  
34 meters), which are part of the Madidi Project in north-western Bolivia. We carried out palm  
community-level analysis, as well as species-specific analyses for each of the 16 most common  
36 species. We found that soils and climate contribute differentially to shaping Andean palm  
diversity and distributions. We found that soils explain more variation in species composition  
38 (14.4%) than climate (3.45%), but that climate explains more variation in species richness (13%)  
than soils (6.1%). Moreover, species-specific analyses reveal that there is great variation in how  
40 different common species respond to their abiotic environment. Our results contribute to  
understanding the drivers of biodiversity for a highly important group of plants in one of the  
42 most important hotspots for biodiversity.

44 *Key words:* Elevational gradient; Andean forest; Arecaceae; climate; Madidi region; redundancy  
analysis; soils

46 **RESUMEN**

En el neotrópico, las palmeras (Arecaceae) son un grupo diverso y abundante de plantas que  
48 constituyen elementos estructurales en bosques tropicales tanto de tierras bajas como de  
montaña. Además, muchas especies de palmeras son culturalmente y económicamente  
50 importantes para muchas poblaciones humanas. Debido a su importancia ecológica y  
etnobotánica, entender los mecanismos que controlan la diversidad y la distribución de las  
52 palmeras es extremadamente importante. Sin embargo, la mayoría de la investigación hasta el  
momento se ha enfocado en comunidades de palmeras de tierras bajas, mientras que la ecología y  
54 biogeografía de las palmeras de montañas es relativamente poco entendida. En este estudio,  
nosotros investigamos los factores ambientales que influyen la riqueza, composición y  
56 abundancia de palmeras en los Andes Centrales. En particular, estamos interesados en entender  
los efectos relativos de las condiciones edáficas del suelo y el clima en la estructura de  
58 comunidades de palmeras. Para nuestros análisis, usamos una red de 88 parcelas de árboles  
distribuidas a lo largo de un gradiente elevacional (1,000 – 3,200 metros), la misma que es parte  
60 del Proyecto Madidi en Bolivia. Encontramos que el suelo y el clima tienen efectos diferentes.  
El suelo explica más variación en la composición de especies (14.4%) que el clima (3.45%), pero  
62 el clima explica más variación en la riqueza de especies (13%) que los suelos (6.1%). Además,  
análisis independientes para las 16 especies más comunes demuestran gran variación en como  
64 cada especie responde a las condiciones ambientales. Nuestros resultados contribuyen a entender  
los factores que controlan la diversidad de un grupo importante de plantas en una de las regiones  
66 más diversas del planeta.

68

## 70 1. INTRODUCTION

Palms (family Arecaceae) are a charismatic and diverse group of plants, which play critical  
72 ecological roles in tropical ecosystems, and provide many important natural services to human  
populations. Many palms are important elements of tropical and montane forests across South  
74 America, contributing significantly to the structure and diversity of uppermost canopy, as well as  
of the mid- and understorey (Svenning 2001, Henderson 2002, Garibaldi & Turner 2004,  
76 Eiserhardt *et al.* 2011, Balslev *et al.* 2012). In these ecosystems, palm species provide food and  
shelter, and their local abundance can make them a keystone resource for populations of  
78 vertebrate and invertebrates animals (palm fruits are edible, either for their soft mesocarp or for  
their endosperm (Zona & Henderson 1989, Muñoz *et al.* 2019). Palms also have many important  
80 ethnobotanical uses for human societies across the tropics. For example, people use stems as  
poles in house construction and leaves for thatching roofs. Leaves also provide fibers for  
82 weaving baskets, fishing nets and other tools (Paniagua-Zambrana *et al.* n.d., Garibaldi & Turner  
2004). Other palms are used in ceremonial or religious activities and other activities of cultural  
84 importance (Paniagua-Zambrana *et al.* n.d.). Indeed, palms species provide fuel, fabrics and  
medicine for many rural and urban communities (Paniagua-Zambrana *et al.* n.d., Henderson  
86 2002, Eiserhardt *et al.* 2011, Montoya & Moraes R. 2014, Moraes R. 2015). For all these  
reasons, natural populations of palms are exploited, and their environments are threatened. Thus,  
88 understanding the drivers of the amazing geographic variation in palm richness, composition and  
abundance is fundamental to conservation strategies focused on their preservation.

90 Multiple evolutionary and ecological forces have been proposed to explain variation in  
the diversity and composition of natural communities. These processes range in scale from  
92 geological changes such as mountain uplift that creates new habitats and physical barriers for  
dispersal (Pintaud *et al.* 2008) to local interactions such as apparent competition mediated by

94 shared enemies (Pintaud *et al.* 2008, Eiserhardt *et al.* 2013). However, one of the most important  
factors that shape the distribution and abundance of species, and that often mediate the effects of  
96 biogeographic history and species interactions, is the abiotic environment (Condit *et al.* 2013). In  
this way, understanding the effects abiotic conditions on community diversity and composition is  
98 often a fundamental first step needed before deeper insights can be obtained. This is the main  
objective of our study.

100 Past research has suggested that climate and soils are important drivers of plant  
communities. For example, (Eiserhardt *et al.* 2013) showed that habitat types, characterized by  
102 inundation regime, play an important role in the spatial turnover of species in Amazonian  
rainforests (Sesnie *et al.* 2009, Eiserhardt *et al.* 2013). In particular palms are well known as a  
104 tropical clade of species that strongly prefer wet and warm conditions (Eiserhardt *et al.* 2011).  
Thus, we expect that precipitation and temperature will be important drivers of palm abundance  
106 across elevational gradients. Indeed, Sesnie *et al.* (2009) found that precipitation affect the palm  
abundance in north-eastern Costa Rica, while Costa *et al.* (2009) suggested that water drainage  
108 patterns seem to be a major factor controlling patterns of palms distribution along topographic  
gradients. Precipitation might be particularly important in montane regions because high  
110 topographic heterogeneity can cause dramatic variation in precipitation from place to place  
(Eiserhardt *et al.* 2011).

112 Plant distributions also often track soil conditions, particularly soil nutrient availability.  
For example, Figueiredo *et al.* (2018) found that soils are a more important predictor of plant  
114 species ranges in Amazonia than climate. Their study suggests that species distribution are  
limited by edaphic factors that reduce species' abilities to track suitable climate conditions. In a  
116 different study, (Condit *et al.* 2013) found that soil phosphorus was the strongest factor affecting  
the distribution of more than half of the species of tropical vegetation in central Panama. The

118 finding that many species have associations with either high or low phosphorus reveals an  
important role for this nutrient in limiting tropical tree distributions. The distribution of palm  
120 species has also been associated with soil conditions like texture and nutrients. For example,  
Phillips *et al.* (2003) and Tuomisto *et al.* (2003) found an important role of soil properties in  
122 turnover of species composition within *terra firme* forests.

Climate and edaphic conditions can sometimes interact. Indeed, changes in temperature,  
124 precipitation or topography (such as slope and aspect) can play a big role in shaping variation in  
soil conditions and nutrient availability. For example, temperature controls microbial activity,  
126 which in turn reduce the decomposition rate (Lloyd & Taylor 1994), affecting the total amount of  
nutrients available and organic carbon stocks (Garibaldi & Turner 2004, Girardin *et al.* 2010,  
128 Figueiredo *et al.* 2018). For all these reasons, to understand the effects of soils we need to  
simultaneously understand the effects of climate on plant distributions.

130 Many studies have investigated the effect of climate and edaphic conditions on palm  
distribution in lowlands (e.g., the Amazon) or premontane forest (Jones *et al.* 2008, 2013,  
132 Pintaud *et al.* 2008, Eiserhardt *et al.* 2011, 2013, Kristiansen *et al.* 2012, Condit *et al.* 2013,  
Prada *et al.* 2017, Schlindwein *et al.* 2017). In contrast, there is a paucity of studies that  
134 simultaneously consider the effects of edaphic and climate conditions on palms communities in  
the Andes. This is so, despite a high diversity of palms in montane regions. Indeed, the tropical  
136 Andes above 1,000 m host 121 species of palms and 24 genera (Pintaud *et al.* 2008, Arias &  
Stauffer 2012). Just as in lowland forests, palms can constitute an important component of the  
138 structure and ecology of montane forest (Moraes *et al.* 1995, Borchsenius & Moraes 2006).

In this study, we used 88 forest plots distributed along a 2,200 m elevational gradient to  
140 assess the effects of environmental conditions (climate and soils) on Andean palm species  
communities. Specifically, we address three interrelated question: (1) Is there an elevational

142 gradient in the diversity, abundance and species composition of Andean palm communities? We  
expect that palm communities will change across the elevation gradient in response to diverging  
144 environmental conditions. (2) Do environmental conditions, particularly climate and soils,  
explain variation in the diversity, abundance and composition of Andean palm communities?  
146 This question explores the relative contribution of the environmental conditions on the structure  
of communities. (3) Do climatic and soil conditions explain spatial variation in abundance of the  
148 most common species of Andean palms? Here, we study how individual palm species respond to  
their abiotic environment to understand the factors that have species-specific effects on  
150 abundance patterns. We found that soils explain more variation in species composition (14.4%)  
than climate (3.45%), and climate explains more variation species richness (13%) than soils  
152 (6.1%). Species-specific analyses reveal that there is great variation in how different common  
species respond to their abiotic environment. Our results contribute to understanding the drivers  
154 of biodiversity for a highly important group of plants in one of the most important hotspots for  
biodiversity.

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## 2. METHODS

158 **1. Species composition and abundance data.** — In this study, we used data from the Madidi  
Project (<https://madidiproject.weebly.com/>), which includes a large network of nearly 500 forest  
160 plots distributed along an elevational gradient in the Amazon and Andes of northwestern Bolivia.  
Because we were interested in studying Andean palm communities, our analyses used only 88  
162 plots located above 1,000 m in elevation that contained at least one palm individual. These plots  
are 0.1-ha in area (50 x 20 m), are typically located in mature forest at least 500 m apart from  
164 one another and cover a broad range of environmental conditions ranging in elevation from  
1,000 to 3,200 m (Figure 1; (Tello *et al.* 2015)). Mean annual temperature varies from 14.6 to

166 21.6°C, while total annual precipitation ranges from 652.8 to 2,882 mm (climate data from  
WorldClim 2.0; (Fick & Hijmans 2017)). The soils in the region can vary from moderately deep  
168 to very deep, and can be alkaline, silty-clay-loam, silty-clay, loamy-clayey and clayey, in  
addition to frequent gravel and deep stones (Fuentes 2005). Within our forest plots, all palm  
170 individuals with a diameter at breast height (DBH) equal or greater than 2.5 cm were measured  
and taxonomically identified. Herbarium specimens to document each species at each site are  
172 deposited mainly at the Herbario Nacional de Bolivia and the Missouri Botanical Garden. All  
palms were identified to species level. In total, we recorded 3,148 individuals belonging to 16  
174 species. The data used in our analyses correspond to version 2.2 of the Madidi Project Dataset,  
which is deposited in Zenodo (DOI 10.5281/zenodo.4280178) (Tello *et al.* 2018).

176 From these data, we constructed a community matrix that contains 16 palm species  
(columns) distributed across 88 plots (rows). This matrix is filled with the abundance of each  
178 species at each site. For each plot, we calculated the number of palm species (richness) and the  
total number of individuals (total abundance). We then transformed the community matrix using  
180 the Hellinger method described by Legendre & Gallagher (2001) and implemented by the  
function “*decostand*” in R package “*vegan*” (Oksanen *et al.* 2016). This transformation reduces  
182 the weight of rare species on the estimation of dissimilarities between plots and is recommended  
to diminish bias in ordination methods (Legendre & Gallagher 2001). Moreover, when used in  
184 combination with methods based on Euclidean distances, such as principal component analysis  
or redundancy analysis, this transformation avoids the double-zero problem (Legendre &  
186 Gallagher 2001).

**2. Climate predictors.** — We used the geographic coordinates of each plot to extract values for  
188 19 bioclimatic variables from WorldClim Version 2.0 (Fick & Hijmans 2017). These variables  
reflect the annual average and temporal variability in temperature and precipitation. The



190 WorldClim dataset is generated by an interpolation of long-term monthly temperature and  
precipitation data (mostly for the 1970-2000 period) from weather stations in a large number of  
192 global, regional, national and local sources. The interpolation uses satellite-derived weather data  
and other covariables to create global climate surfaces (Fick & Hijmans 2017).

194 To reduce the dimensionality of the climate dataset, we employed a principal component  
analysis (PCA) on centered and standardized variables. The first three PC axes accounted for  
196 93% of variation in climate (Figure S1A), and only these were retained for further analyses. The  
first PC axis (hereafter Clim1) is dominated by positive loadings for mean and extreme  
198 temperature and by negative loadings of temperature variability (Figure S1B). Thus, Clim1  
reflects variation from colder sites with large temperature variation to warm sites that change  
200 little in temperature through the year. The second PC axis (Clim2) reflects a variation in  
precipitation from dry to wet sites (Figure S1C). Finally, the third PC axis (Clim3) represents a  
202 gradient from sites with low precipitation seasonality and high precipitation in the driest part of  
the year to sites with high seasonality in precipitation and less rain during the driest part of the  
204 year (Figure S1D, Table S1).

**3. Soil predictors.** — Soil characteristics were measured in samples collected during the  
206 vegetation census of each plot. Within each plot, three sub-samples were taken at randomly  
selected locations between trees. The samples were taken from a 50 × 50 cm square, free of leaf  
208 litter and stones. Each sub-sample was taken from a depth of 0 to 30 cm below the litter layer,  
and then combined and mixed into a full sample of 1 kg. All soil samples were analyzed by the  
210 Laboratorio de Calidad Ambiental of the Instituto de Ecología at the Universidad Mayor de San  
Andrés in La Paz, Bolivia. Soil conditions were characterized by 11 variables (including 6  
212 nutrients). Total nitrogen (N) was measured using the semi-micro Kjeldahl method, while  
phosphorus (P), sodium (Na), potassium (K), calcium (Ca) and magnesium (Mg) using the 1M

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214 ammonium acetate solution method. We also measured pH and electro-conductivity (EC) with a  
potentiometer, cationic exchangeable capacity (CEC) was measured using the summation of total  
216 bases. Finally, silt and sand percentages were estimated with the hydrometer method. Clay  
percentage was excluded because of its perfect multicollinearity with silt and sand (Text S1).

218 To reduce dimensionality, we log-transformed all soil variables and conducted a PCA on  
centered and standardized variables. For all subsequent analyses, we used the first seven  
220 principal components, which accounted for 92% of the variation in edaphic conditions (Figure  
S2A). We determined that the first PC axis (Soils1) represents a gradient from low to high  
222 concentrations in most micronutrients (P, Mg, Ca and K; Figure S2B). The second PC axis  
(Soils2) represents a gradient from silty soils to sandy soils with high Na concentrations. Soils3  
224 is a gradient from acidic soils to neutral soils with high concentration in N and high electric-  
conductivity (Figure S2D). Soils4 characterizes a gradient from high to low Na concentrations  
226 and cationic exchangeable capacities (Figure S2E). Soils5 represents a gradient from high  
percentages of silt and sand (i.e. low clay) and high concentration of K to soils with high electro-  
228 conductivity and high percentages of clay (Figure S2F). Soils6 characterizes sites with high P  
concentration to sandy sites with high cationic exchangeable capacities (Figure S2G). Finally,  
230 Soils7 represents a gradient from low to high concentrations of P and low to high cationic  
exchangeable capacity (Figure S2H). For more details, see Tables S2 and S3.

232 **4. Statistical analyses.** — In order to address our first question, we regressed palm species  
richness and total abundance separately against elevation. For these analyses, we used linear and  
234 quadratic ordinary least square models (OLS) and compared them using an ANOVA. If the  
quadratic model was significantly better, it was retained and used for interpretation (Table 1).

236 In addition, we conducted a redundancy analysis (RDA) where the Hellinger-transformed matrix  
of composition was the response, and the linear and quadratic terms of elevation were predictors.

238 For these analyses, we used the function “*rda*” in the R Package “*vegan*” (Oksanen *et al.* 2016).

For our second question, we conducted separate analyses for (1) species richness, (2) total  
240 abundance and (3) the Hellinger-transformed matrix of species composition. First, we conducted  
a series of variation partitioning analyses (function “*varpart*” in R package “*vegan*”). These  
242 analyses partition the variation in the response into fractions explained by different sets of  
predictors. In our case, we partitioned the variation (in richness, abundance or composition) into  
244 fractions explained (1) only by soils, (2) only by climate, (3) simultaneously by both soils and  
climate (i.e shared variation) and (4) the variation that remains unexplained. In each analysis, we  
246 used all three climate principal components and all seven soil principal components as  
predictors.

248 Additionally, we constructed minimum adequate models (MAMs) for each response. For  
species richness and total abundance, we used spatial generalized least square models (GLS;  
250 details below), while for community composition we used a redundancy analysis. In both cases,  
models started with a full complement of predictors (all climate and soil principal components)  
252 and the model selection was based on a stepwise procedure using the Akaike information  
criterion (AIC). For model selection, we used the function “*stepAIC*” in the R package “*MASS*”  
254 (Venables & Ripley 2002). The final model (the MAM) contains the combination of predictors  
that most reduce the AIC value of the model, and that are likely the most important factors  
256 responsible for variation in the response variable. All predictors were centered and standardized  
to make regression coefficients comparable.

258 As mentioned above, for richness and total abundance, we used generalized least-squared  
(GLS) models. These models used geographic coordinates of each plot to account for spatial

260 autocorrelation among sites. We used a Gaussian correlation structure in all analyses because it  
was the best fit for our models (functions “*gls*” and “*corGaus*” in the R package “*nlme*”;  
262 (Pinheiro *et al.* 2020)). After producing the minimum adequate model (MAM) by AIC selection,  
we compared this model against a null model. This allowed us to obtain a model-wide p-value  
264 for the MAM via ANOVA. The null model was another GLS regression that only contained an  
intercept term and the same spatial autocorrelation structure estimated for the MAM. As  
266 estimates of relative fit, we calculated the difference in AIC between the MAM and the null  
model, as well as a pseudo- $R^2$  using the function “*rsquared*” in the R package “*piecewiseSEM*”  
268 (Lefcheck 2016).

Unfortunately, there is no GLS equivalent when using a multivariate response. Thus, to  
270 study the climatic and edaphic determinants of species composition we used a redundancy  
analysis (RDA). Here, the Hellinger-transformed matrix of community composition was the  
272 response and climatic and edaphic PC axes were predictors. For the redundancy analysis, an AIC  
cannot be computed, so the variable selection is based on a metric that mimics AIC but is based  
274 on the residual sums of squares (function “*deviance.cca*” in R package “*vegan*”). After variable  
selection, an adjusted  $R^2$  and  $P$ -values were obtained directly from model output.

276 For our third question, we related the abundance of each species separately with climate  
and soil predictors. Not all palm species had sufficient data for these analyses; thus, we only used  
278 the seven species that occur in at least 15 plots: *Aiphanes horrida* (26 plots), *Chamaedorea*  
*angustisecta* (16), *Dictyocaryum lamarckianum* (41), *Euterpe precatoria* (52), *Geonoma*  
280 *orbignyana* (29), *Iriartea deltoidea* (19) and *Socratea exorrhiza* (28). For all analyses, we  
retained only the plots where the focal species was present (i.e., eliminated all zeroes).

282 Additionally, because of the reduced sample size, we only considered the first three principal  
components of climate and the first four principal components of soils. Including more predictors

284 would likely lead to overfitting each model. Just like for the analyses on richness and total  
abundance, analyses for each species involved (1) a variation partitioning procedure using all  
286 predictors and (2) the construction of a minimum adequate model using AIC variable selection  
and spatial GLS models. The details of these methods are the same as those described for  
288 question two.

### 290 3. RESULTS

Is there an elevational gradient in the diversity, abundance and species composition of palm  
292 communities? Species richness showed a humped-shaped relationship with elevation (i.e.  
quadratic regression better than lineal regression; Figure 2A;  $F = 7.04$ ;  $P = 0.002$ ). Richness  
294 reached its maximum between 1,200 – 1,600 m and decreased to only one species above 2,200m.  
In total, elevation explained 11.4% of variation in species richness. In contrast, abundance was  
296 not significantly related to elevation (neither the quadratic nor linear regressions were  
significant; Figure 2B;  $P > 0.05$ ;  $R^2 = 0.014$ ). Finally, the RDA showed that elevation (linear and  
298 quadratic terms) explains a small proportion of variation in palm species composition among  
plots (6.3%; Figure 2C).

300 Do environmental conditions, particularly climate and soils, explain variation in the  
diversity, abundance and composition of palm communities? Regarding the effects of edaphic  
302 and climatic factors on species richness (Figure 3), we found that partitioning this variation into  
components (Figure 3A) shows that the shared effect was the largest (19.1%). Among the unique  
304 effects, the variation associated solely with climate was larger than that associated solely with  
soils (13% vs. 6.1%). However, both sets of predictors combined explained about 38.5% of the  
306 variation. In our minimum adequate regression model, we found also evidence that both soils and  
climate contribute to explaining variation in species richness (Figure 3C). Indeed, the best GLS

308 model retained Clim1, Clim2, Clim3, Soils1, Soils3 and Soils 4 as significant predictors ( $R^2_{\text{pseudo}} = 0.447$ ;  $P < 0.05$ ). The effects of climate indicate that richness increases with temperature and  
310 precipitation, but decreases with seasonality (Figure S1B, S1C and S1D). The effects of soils suggest that richness changes with concentrations of soil nutrients (Soils1), increasing toward  
312 soils with high nitrogen concentrations (Soils3). Richness also increases with cation exchange capacity (Soils4) and decreases towards acidic soils (Soils3) (Figure S2B, S2D and S2E).

314 Soils and climate explained a smaller fraction of variation in total palm abundance (Figure 3): all predictors combined accounted for only 18.1% of the variation across plots  
316 (Figure 3B). As with richness, variation partitioning showed that the shared effect of soils and climate was the largest component of variation (9.1%) while the unique effects were very small  
318 (<6%). The best GLS regression on total palm abundance retained both climatic and soil variables: Clim1, Soils1 and Soils4 (Figure 3D). The sign of these coefficients is the same as in  
320 the richness model. This suggests that total palm abundance increases with temperature (Clim1), is reduced with overall nutrient availability (Soils1), but increases with cation exchange capacity  
322 (Soils4).

When we analyzed species composition, we found that soils and climate explain about  
324 one quarter of the variation across plots (Figure 4). In contrast to the results for richness, the largest fraction of explained variation in composition is associated solely with soils (14.4%).  
326 This fraction is larger than that associated solely with climate (3.5%), or than the variation explained simultaneously by both sets of predictors (7.9%) (Figure 4A). The minimum adequate  
328 model based on redundancy analyses (RDA) retained both climatic and soil predictors (Figure 4B). Specifically, it retained all three climate principal components (Clim1 to Clim3), as well as  
330 four participant components from the soil data: Soils1, Soils2, Soils3 and Soils6.

Do climatic and soil conditions explain spatial variation in abundance of common species  
of Andean palms? We found a high degree of variability in analyses for each of the seven  
common species (Figure 5). Total amount of variation explained varied from nearly 60% (*C.*  
*angustisecta*) to around 10% (*G. orbignyana*). However, for two species, *A. horrida* and *G.*  
*orbignyana*, we could not find a minimum adequate model that was significantly better than a  
null model with only an intercept (Figure 6). Thus, we conclude that climatic and soil predictors  
do not account for variation in abundance of these two species. Among the remaining five  
species, climate seems to be more important than soils for three species (*C. angustisecta*, *D.*  
*lamarckianum*, and *E. precatorea*), while soils seem to be more important than climate for *I.*  
*deltoidea* and *S. exorrhiza* (Figure 5).

For *Chamaedorea angustisecta*, the best model (MAM) includes both two climatic and  
soils variables (Table 2, Figure 6B). The abundance of this species increases towards drier and  
more seasonal sites with soils rich in nutrients like P, Mg, Ca, K, but low in Na. The best model  
for *Dictyocaryum lamarckianum* includes Clim1, Clim2, Clim3 and Soils1. This species is more  
abundant in colder and drier sites with higher variation in temperature, and is also more abundant  
in sites with nutrient rich soils. The best model for *Euterpe precatorea* just contains two climate  
variables: Clim1 and Clim3, which means that abundance of this species is high in cooler sites  
with higher variation in precipitation across the year (Figure 6D). *Iriarteia deltoidea* shows a  
significant relation with Soils1 and Soils3 (Table 2); its abundance increases towards sites with  
pH neutral soils and high nutrient concentration, particularly N (Figure 6F). Finally, for *Socratea*  
*exorrhiza* the best model includes Clim3, Soils1 and Soils2. This means that the abundance of  
this species increases in sites with high seasonality in precipitation and towards less sandy, more  
silty soils with low concentrations of nutrients like P, Mg, Ca and K (Figure 6G).

354

#### 4. DISCUSSION

##### 356 **1. Richness and composition show clear elevational gradients, but abundance does not.** —

Our results indicate that species richness decreases non-linearly with elevation (Figure 2A), but  
358 that total palm abundance does not have a significant elevational gradient (Figure 2B). The  
relationship between richness and elevation is widely documented in many groups of organisms  
360 (Rahbek 1995, Lomolino 2008), including palms (Svenning 2001, Eiserhardt *et al.* 2011) and in  
montane forests of the Bolivian Yungas (Gerold *et al.* 2008). Despite the generality of the  
362 relationship, the shape of elevational gradients in richness is less well understood. In the  
literature, three types of gradients have been described: linearly decreasing, hump-shaped with  
364 an intermediate peak, or a low-elevational plateau followed by a linear decrease (Colwell & Lees  
2000, McCain & Grytnes 2010). We found that the gradient is non-linear and that resembles  
366 more a low-elevation plateau. Indeed, richness seems to be somewhat constant until around  
1,500 m in elevation, and then decreases rapidly to zero at about 2,500 m. We found that species  
368 composition also changes with elevation, but elevation was able to explain only 6.3% of the  
variation in composition. This suggests that although communities at different elevations are  
370 different, other factors beyond elevation must be more important. For example, climate, soils and  
other abiotic or biotic forces might be more important than simply elevation to account for  
372 changes in community composition.

We were surprised to find no clear change in total palm abundance with elevation.

374 Neither our linear or quadratic models were significant (Figure 2B). Despite this lack of  
statistical support, the distribution of abundance seems nonrandom. Indeed, total palm abundance  
376 seems to peak at intermediate elevations, between 1,400 and 1,500 m. This peak of abundance at  
intermediate elevations may result from the dispersal of species from lower and higher  
378 elevations (Fischer *et al.* 2011). According to (Fischer *et al.* 2011), mountain forest compared to



flat and lowlands the higher difference between site types per area is due to different inclinations,  
380 exposures, and geological substrates; many different micro-sites. Further studies with increased  
sample size and larger plots might be needed to clarify this relationship between abundance and  
382 elevation in montane palm communities (Nettesheim *et al.* 2018).

**2. Climate explains more variation in richness, but soils explain more variation in species  
384 composition.** — According to our results, 13% of the variation in richness is explained by  
temperature (principal component Clim1) and precipitation (Clim3). This is in agreement with  
386 several studies that have shown that climate conditions have major effects on tropical palms  
(Baldeck *et al.* 2013, Jones *et al.* 2013, Nettesheim *et al.* 2018), and also agree more broadly  
388 with the literature that shows strong correlations between species richness and climate for many  
groups of organisms (Francis & Currie 2003, Currie *et al.* 2004, Tello & Stevens 2010). In our  
390 study system, climate, particularly temperature, changes dramatically between lowlands and  
montane regions (Jones *et al.* 2013, Nettesheim *et al.* 2018). Thus, climate might be the main  
392 driver of the elevational gradient in species richness that we observe in our data (Figure 2A).  
Indeed, Jones *et al.* (2011) has proposed an important role for temperature in determining the  
394 distribution of species and patterns of diversity in montane habitats. How climate shapes richness  
patterns is highly contentious, and many hypotheses have been proposed to explain this  
396 relationship (Hawkins *et al.* 2003, Currie *et al.* 2004, Evans *et al.* 2005). Our analyses do not  
attempt to disentangle these possibilities, but shows that richness patterns of an iconic group of  
398 tropical organisms respond strongly to climate across a montane region of mainly temperate  
environmental conditions.

400 While climate is the best predictor of richness patterns, edaphic conditions seem to be  
more strongly associated to community composition. Indeed, the fraction of variation in  
402 composition associated solely to soils is four times larger than that associated solely to climate

(Figure 4A). Three principal components of soils were selected for our final model (Soils1, 404 Soils2 and Soils3), which represent variation in a broad range of soils nutrients (mainly Soils1) and soil physical conditions (texture in Soils2 and pH in Soils3). Previous research has shown 406 that variation in abundance of palm species depends on nutrient requirements (Svenning 2001). For example, in tropical forests across Panamá, (Turner *et al.* 2018) showed that within the same 408 community there are species adapted to high or low phosphorous (P) concentrations. In this way, different palm species can be located at same elevation with the same climate, but in sites with 410 contrasting soil properties (Poulsen *et al.* 2006, Blach-Overgaard *et al.* 2010). This micro-habitat differentiation could represent an ecological strategy to reduce competition among co- 412 occurring palm species, as has been proposed for Amazonian palm communities (Eiserhardt *et al.* 2013). Our analyses for individual common species also support this view, but suggest that 414 while community composition as a whole is mainly driven by soil properties, some individual species might respond more to climate variation.

416

**3. Different common species show varying responses to climate and soil conditions.** — In 418 our analyses, we found that the abundances of *I. deltoidea* and *S. exorrhiza* responded to different soil conditions. *S. exorrhiza* decreases in abundance with increasing amounts of P, Mg, 420 Ca and K (Soils1), while it increases with sand concentration (Soils2; Figure 6). These results are in agreement with Duivenvoorden *et al.* (2005), Carlos-Copete *et al.*, (2019) and Henderson *et al.* (2019), who showed that the presence of this species is within the parameters characteristic of 422 well-drained forest soils. Our results also support findings by Cámara-Leret *et al.* (2017) who suggest that this species can grow in both rich and poor soils. *I. deltoidea* shows the opposite 424 pattern of variation increasing in abundance with the soil characteristics represented by Soils1 (P, Mg, Ca and K), which contradict the findings of Copete *et al.* (2019). *C. angustisecta* also has 426

some important associations with soil conditions, despite responding primarily to climate. Like *I. deltoidea*, *C. angustisecta* responded positively to increasing P, Mg, Ca and K, but its abundance declined with higher Na concentrations and higher cation exchange capacities.

On the other hand, we found three species, *C. angustisecta*, *D. lamarckianum* and *E. precatoria*, which respond mostly to climatic conditions. One pair of these species respond positively to average temperature (Clim1): *D. lamarckianum* and *E. precatoria*. Both become more abundant in warmer sites. A different pair of species, *C. angustisecta* and *D. lamarckianum* increase in abundance with increasing total precipitation (Clim2). Surprisingly, all three species seem to increase in abundance with increasing seasonality in precipitation (Clim3). This effect is relatively small for *C. angustisecta* and *D. lamarckianum*, but might be important for *E. precatoria*.

Overall, our analyses on the most common palm species demonstrate a high level of idiosyncrasy in how species respond to environmental conditions. Among the seven species analyzed, two had no significant associations with climate or soils, two responded more strongly to soil conditions, and three responded mainly to climate. It is important to note that nine species were too rare to be included in these analyses, and that their environmental preferences need to be further studied.

**4. Limitations of this study.** — Although we found that climate and soils explain significant fractions of variation in richness, abundance and composition of palm communities in our study region, much of the variation remains unexplained. There is a multitude of methodological and biological reasons for this. First, the resolution of the environmental data might not be the best reflection of the scales at which the environment varies and species respond to it. For example, our climate data comes from Worldclim (Fick & Hijmans 2017), which uses weather stations and satellite-derived data to create global climate surfaces. To do so, however, the data needs to be

interpolated, which is difficult to do in topographically heterogeneous regions like the Andes.

452 Similarly, we have a single measurement of soil properties per forest plot. Although our plots are  
small, soil conditions can change quickly at the scale of a few meters (John *et al.* 2007). Thus,  
454 our measure of soil properties might also be coarse related to how species experience these  
environmental conditions. Despite the deficiencies in the data, there are no better data sources  
456 available. Our study is similar to many previous studies that have tried to measure the effects of  
environment on species composition, so it is comparable in methodology (Myers *et al.* 2013,  
458 Arellano *et al.* 2016). Without doubt, there is need to develop better datasets to further our  
understanding of how species and communities respond to environmental change in remote  
460 tropical regions.

Some of the unexplained variation could also be the result of other biological processes  
462 that drive the distribution of species and assembly of communities, but that were not part of the  
focus of our analyses. Processes such as glaciation history, human and natural disturbance,  
464 species-species interactions, dispersal and chance could all be important. Indeed, many previous  
studies have highlighted one or more of these mechanisms as drivers of plant community  
466 assembly (Tuomisto, Ruokolainen, *et al.* 2003, Kraft *et al.* 2008, Antonelli *et al.* 2009, Myers *et al.*  
*et al.* 2013). Finally, we believe that much of the unexplained variation results from simple  
468 sampling effects. This means that variation from one community to another in terms of species  
composition and richness is the result of the small sample sizes given by our 0.1-ha plots. These  
470 size plots are widely used because it is sometimes better to spread data-collection effort in small  
grain sizes (plot size) distributed across broad extents (environmental or geographic gradients).  
472 That is exactly the case in our study that covered a broad elevational gradient. This small grain  
size, however, introduced measurement error in describing the species composition of a site,  
474 which in turn translates into unexplained variation in statistical analyses. Despite the small size

of our plots, we were able to capture clear environmental signal in the diversity and composition  
476 of palm communities in our region, which was the main objective of our study. Future work  
would be needed to clarify whether more intensive sampling in each community improves our  
478 ability to predict community structure using climatic and soil predictors.

480 **Data availability statement.** — The species and soil data used in this study are part of the  
Madidi Project, a long-term research collaboration aimed at studying plant diversity in the  
482 tropical Andes. The data used in our analyses correspond to version 2.2 of the Madidi Project  
Dataset, which is deposited in Zenodo (DOI 10.5281/zenodo.4280178). Additionally, raw data of  
484 the Madidi Project are stored and managed in Tropicos® (<https://tropicos.org/home>), the  
botanical data database of the Missouri Botanical Garden. These data can be viewed and  
486 accessed via <http://legacy.tropicos.org/Project/MDI>.

## 5. TABLES

488 **Table 1. Most important climate and soil predictors for richness and abundance.** The table  
 shows result of spatial GLS regressions with the predictors that were selected based on a search  
 490 for a minimum most adequate model using AIC (see methods). These analyses were conducted  
 separately for palm richness and total palm abundance. Predictors includes principal components  
 492 derived from climate and soil data (see Methods). For each predictor in each model, a coefficient  
 estimate, standard error (*SE*), *t*-value and *p*-value are reported. Bold font highlights  
 494 statistically significant predictors.

	<b>Variable</b>	<b>Estimate</b>	<b><i>SE</i></b>	<b><i>t</i>-value</b>	<b><i>p</i>-value</b>
	(Intercept)	2.621	0.125	21.000	<b>&lt;0.001</b>
	Clim1	0.503	0.134	3.762	<b>&lt;0.001</b>
	Clim2	0.220	0.138	1.592	0.115
<b>Richness</b>	Clim3	-0.469	0.172	-2.725	<b>0.008</b>
	Soils1	-0.372	0.152	-2.452	<b>0.016</b>
	Soils3	0.216	0.151	1.427	0.158
	Soils4	-0.363	0.140	-2.594	<b>0.011</b>
	(Intercept)	37.092	3.969	9.345	<b>&lt;0.001</b>
<b>Abundance</b>	Clim1	9.652	4.001	2.413	<b>0.018</b>
	Soils1	-10.689	3.838	-2.785	<b>0.007</b>
	Soils4	-11.322	3.910	-2.896	<b>0.005</b>

**Table 2. Most important climate and soil predictors for each species.** The table shows result  
 498 of spatial GLS regressions with the predictors that were selected based on a search for the  
 minimum most adequate model using AIC (see methods). The analysis was used for abundance  
 500 of each species of palms that is present in fifteen or more tree-plots. For each predictor in each  
 model, a coefficient estimate, standard error (*SE*), *t*-value and *p*-value are reported. Bold font  
 502 highlights statistically significant predictors.

Species	Variable	Estimate	<i>SE</i>	<i>t</i> -value	<i>p</i> -value
	(Intercept)	0.564	0.139	4.071	<b>&lt;0.001</b>
<i>Aiphanes horrida</i>	Soils1	-0.114	0.073	-1.559	0.133
	Soils2	-0.087	0.049	-1.773	0.090
	Soils3	0.133	0.082	1.625	0.118
	(Intercept)	0.692	0.053	12.962	<b>&lt;0.001</b>
<i>Chamaedorea angustisecta</i>	Clim2	-0.229	0.063	-3.661	<b>0.004</b>
	Clim3	0.148	0.054	2.759	<b>0.019</b>
	Soils1	0.147	0.053	2.767	<b>0.018</b>
	Soils4	0.135	0.054	2.502	<b>0.029</b>
	(Intercept)	0.510	0.037	13.815	<b>&lt;0.001</b>
<i>Dictyocaryum lamarckianum</i>	Clim1	-0.168	0.036	-4.659	<b>&lt;0.001</b>
	Clim2	-0.122	0.041	-2.970	<b>0.005</b>
	Clim3	0.066	0.045	1.484	0.147
	Soils1	0.100	0.043	2.310	<b>0.027</b>
	(Intercept)	0.548	0.043	12.790	<b>&lt;0.001</b>
<i>Euterpe precatorea</i>	Clim1	-0.094	0.034	-2.773	<b>0.008</b>
	Clim3	0.146	0.041	3.590	<b>&lt;0.001</b>
		(Intercept)	0.441	0.052	8.486
<i>Geonoma orbignyana</i>	Clim1	-0.086	0.054	-1.598	0.122
	(Intercept)	0.580	0.058	9.956	<b>&lt;0.001</b>
<i>Iriartea deltoidea</i>	Soils1	0.217	0.064	3.418	<b>0.004</b>
	Soils3	-0.127	0.064	-1.999	0.063
	(Intercept)	0.621	0.039	15.863	<b>&lt;0.001</b>
<i>Socratea exorrhiza</i>	Clim3	0.096	0.058	1.664	0.110
	Soils1	-0.221	0.058	-3.830	<b>&lt;0.001</b>
	Soils2	0.071	0.042	1.707	0.101

504

## 6. FIGURE LEGENDS

506 **FIGURE 1.** Location of the study region and forest plots used in this study. (A) The Madidi  
Region is located in the areas in and around the Madidi National Park in northwestern Bolivia  
508 (B) The 88 temporary plots (yellow dots) with palm abundance data are distributed along an  
elevational gradient from 1,000 to 2,500 m.

510

**FIGURE 2.** Variation in palm species richness, total abundance and composition in relation to  
512 elevation. (A) Palm species richness shows a unimodal relationship with elevation with a peak at  
around 1,500 m, while the most species poor communities are located above 2,000 m. (B) Total  
514 palm abundance doesn't have a statistically significant relationship with elevation. In both panels  
(A) and (B), the adjusted  $R^2$  and  $p$ -values correspond to the quadratic model (solid line), which is  
516 significantly better than the linear one (broken line). (C) A redundancy analysis (RDA) also  
shows that species compositional has a significant relationship with elevation, but the  
518 relationship is very weak. Gray dots represent tree plots used in the analysis (88 in total).

520 **FIGURE 3.** Variation in palm richness and abundance predicted by climatic and edaphic factors.  
Panels (A) and (B) show results of the variation partitioning analyses using all climate and soil  
522 principal components. Panels (C) and (D) show the principal components of climate and soils  
retained in the final models. In these panels, the pseudo- $R^2$  and  $p$ -values from spatial GLS  
524 models are presented (A) Variation partitioning shows that less than half of the variation in  
richness can be explained by abiotic predictors, and that climate explains more variation than  
526 soils. (B) Variation in total palm abundance is less well explained by abiotic conditions, but both  
soils and climate contribute with similar fractions of variation. (C) The most important predictors  
528 for richness were climatic principal components: Clim1 and Clim3. Richness was also explained



by three different soil principal components (just two are significant). (D) The most important  
530 predictors for total palm abundance included one climatic and two edaphic principal  
components: Clim1, Soils1 and Soils4. All of these predictors has similar relative importance.

532

**FIGURE 4.** Variation in palm species composition predicted by climatic and edaphic factors.

534 (A) Variation partitioning analysis using all principal components of climate and soils show that  
around one fourth of the variation can be explained. Soils explain more variation than climate.

536 (B) An RDA with the best predictors shows the effect of both climate and soils on community  
composition. Four principal components of soils and three of climate were retained during  
538 variable selection (see Methods). Adjusted  $R^2$  and  $p$ -values for this RDA are also represented.

Ir del = *Iriartea deltoidea*; So exo = *Socratea exorrhiza*; Ge orb = *Geonoma orbignyana*; Eu pre  
540 = *Euterpe precatoria*; Di lam = *Dictyocaryum lamarckianum*; Ch ang = *Chamaedorea*  
*angustisecta*.

542

**FIGURE 5.** Variation partitioning between climate and soils for individual palm species. Each

544 bar shows fractions of variation using all climate and soil predictors. Two species shows more  
variation explained by edaphic conditions than climate: *I. deltoidea* and *S. exorrhiza*. The other  
546 five species show more variation explained by climate than soils: *A. horrida*, *C. angustisecta*, *D.*  
*lamarckianum*, *E. precatoria* and *G. orbignyana*. Species also vary considerably in the total  
548 amount of variation explained. Only *C. angustisecta* have more than 50% of their variation in  
abundance explained.

550

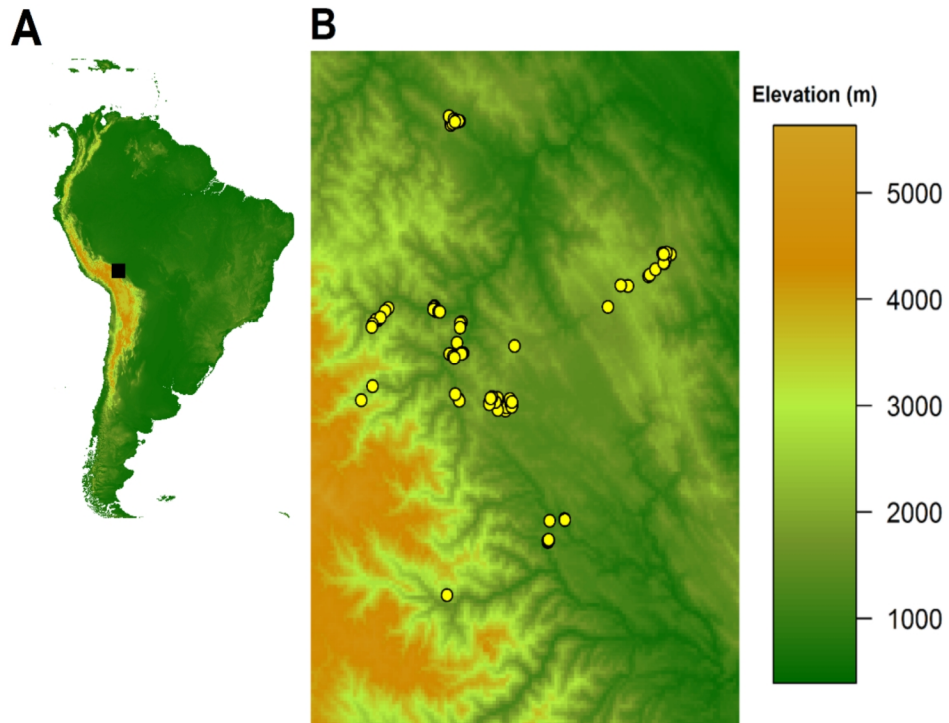
**FIGURE 6.** Abundance of individual palm species predicted by climate and soil factors. Each

552 panel shows the predictors selected in the minimum adequate model of a different common

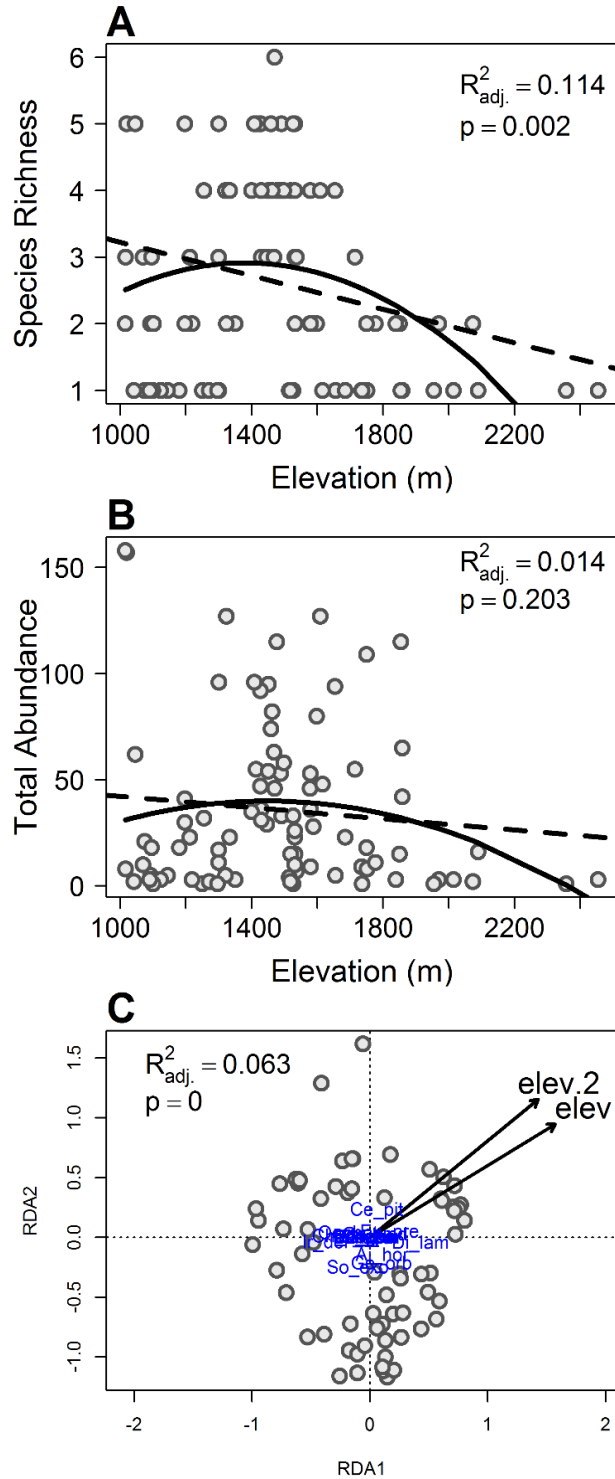
species. In each panel, the pseudo- $R^2$  and  $p$ -values from the corresponding spatial GLS model are  
554 presented. For two species, *A. horrida* and *G. orbignyana* the GLS models are not statistically  
better than null models with just an intercept.

## 7. FIGURES

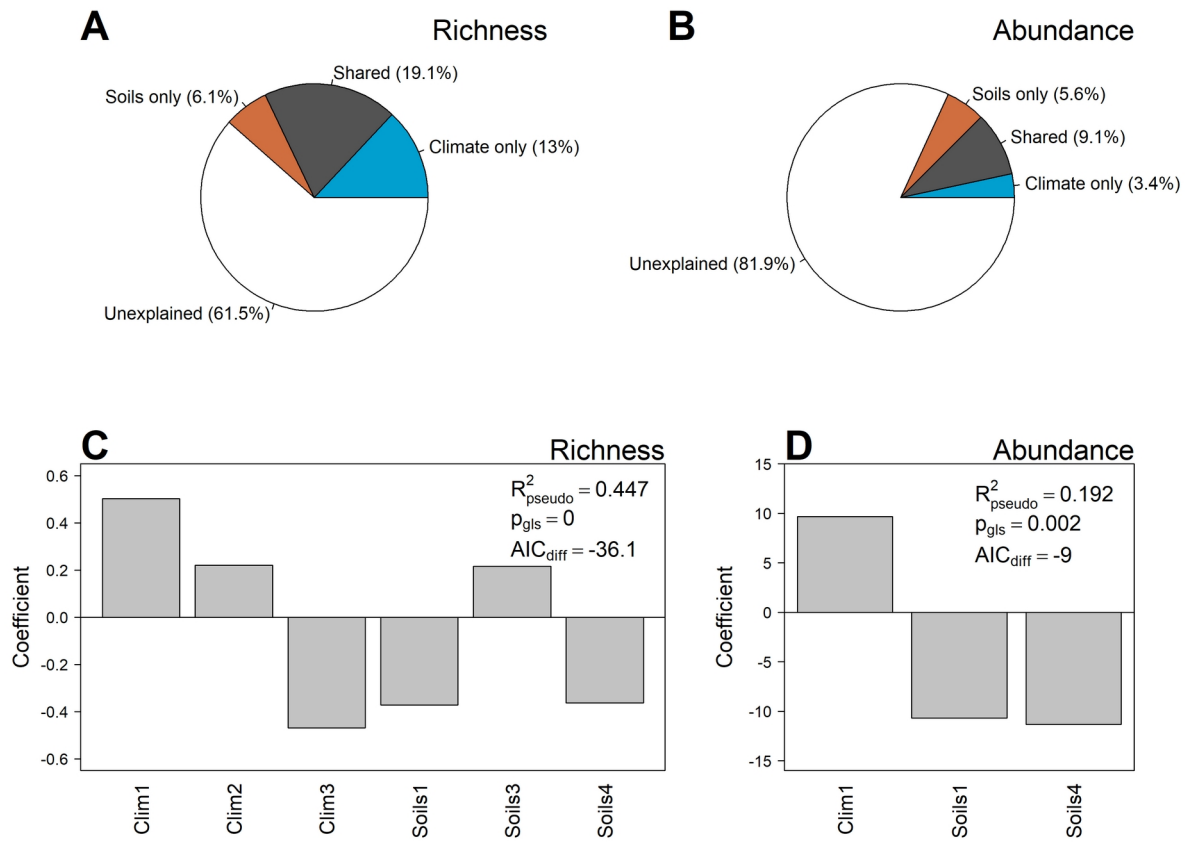
Figure 1



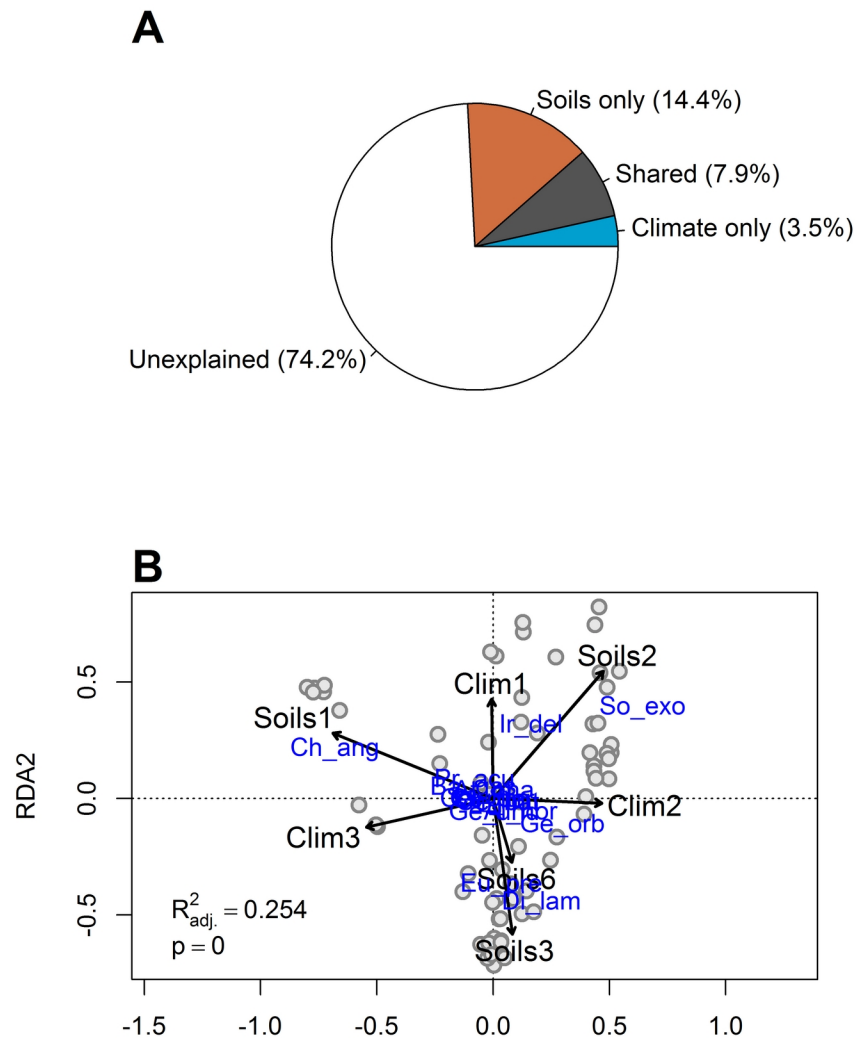
562 **Figure 2**



566 **Figure 3**



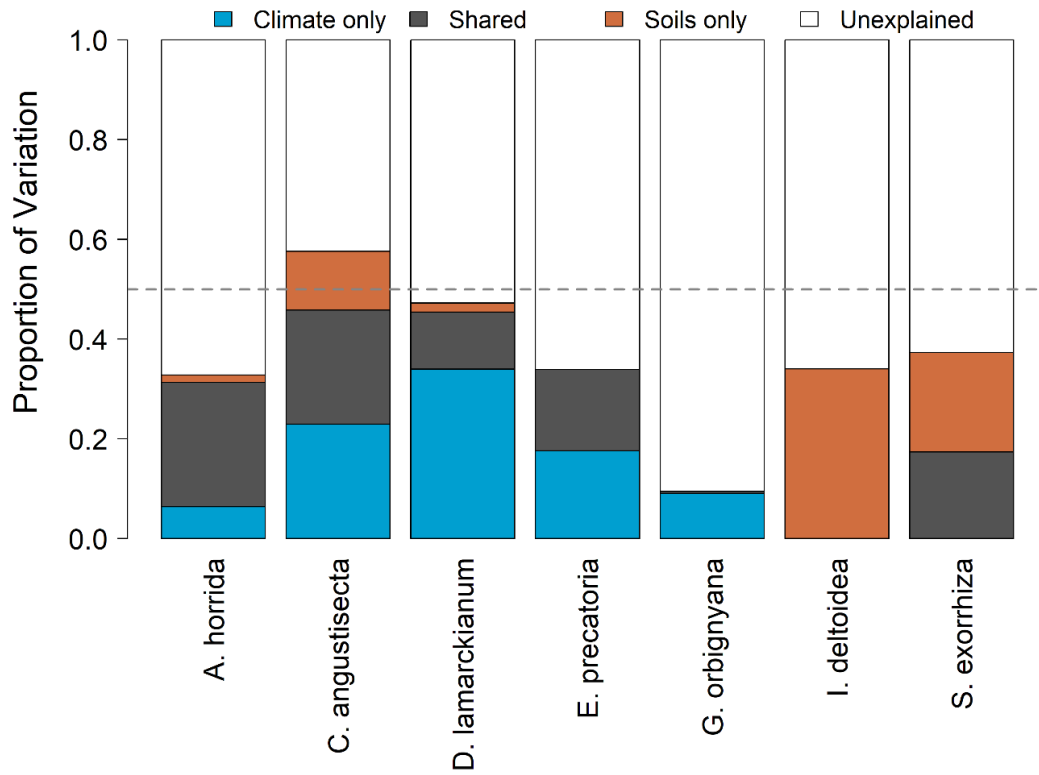
568 **Figure 4**



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574 **Figure 5**



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580

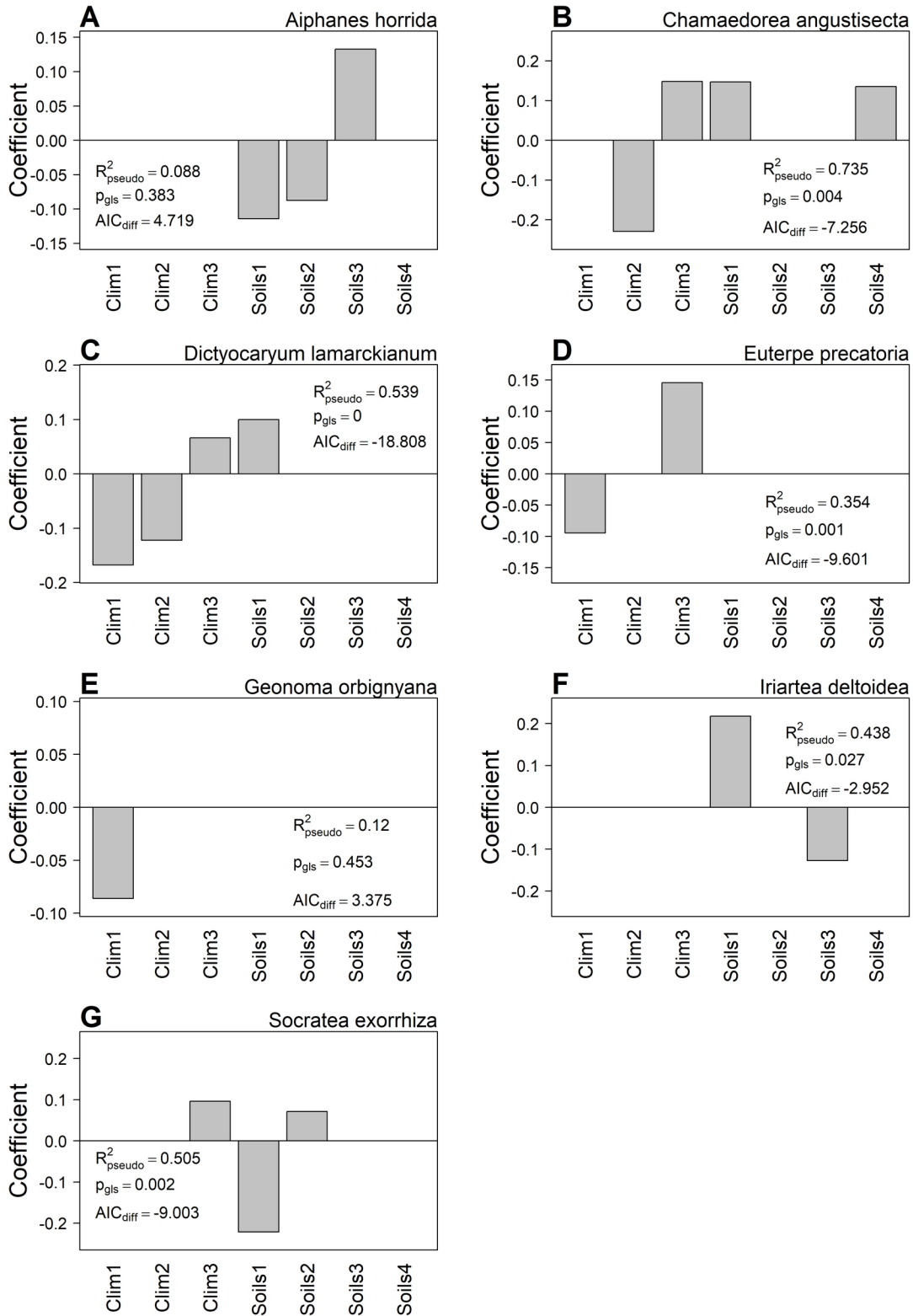
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586

Figure 6

588



590



## 8. ACKNOWLEDGMENTS

592 We thank the Herbario Nacional de Bolivia, the Dirección General de Biodiversidad, the  
Bolivian Park Service (SERNAP), the Madidi National Park and local communities for permits,  
594 access and collaboration in Bolivia. This project was supported by generous grants from the  
National Science Foundation (DEB 0101775, DEB 0743457, DEB 1836353), and the National  
596 Geographic Society (NGS 7754-04 and NGS 8047-06). Additional financial support for the  
Madidi Project has been provided by the Missouri Botanical Garden, the Comunidad de Madrid  
598 (Spain), the Universidad Autónoma de Madrid, and the Taylor and Davidson families. This work  
was completed as part of the master's thesis of the lead author and was supported by the National  
600 Science Foundation (DEB 1836353), the Missouri Botanical Garden and the Shirley Graham's  
fellowship. We are grateful to Olga Martha Montiel and Shirley Graham for the support and  
602 confidence. We thank all the researchers, students and local guides that were involved in the  
collection of the data, particularly Carla Maldonado, Maritza Cornejo, Alejandro Araujo, Javier  
604 Quisbert and Narel Paniagua. We are thankful to all the taxonomic experts that provided  
identifications for plant specimens. Finally, we would like to thank Sebastián Gonzalez-Caro,  
606 Isabel Morales-Belpaire and Ramiro P. López for their valuable comments and support.

**Author Contribution Statement.** — FM, JST and MMR conceptualized the study. AFF, LC,  
608 AA, TM, EMM, MIL collected the data. JST and LC coordinated the Madidi Project and the data  
collection. AFF was responsible for the curation of herbarium specimens. FM and JST conducted  
610 the analyses and wrote the original draft. All authors contributed to revisions.

## 9. DISCLOSURE STATEMENTS

612 The corresponding author confirms on behalf of all authors that there have been no involvements  
that might raise the question of bias in the work reported or in the conclusions, implications, or  
614 opinions stated.

## 10. REFERENCES

- ANTONELLI, A., J. A. A. NYLANDER, C. PERSSON, and I. SANMARTÍN. 2009. Tracing the impact of the Andean uplift on Neotropical plant evolution. *PNAS* 106: 9749–9754.
- ARELLANO, G., J. S. TELLO, P. M. JØRGENSEN, A. F. FUENTES, M. I. LOZA, V. TORREZ, and M. J. MACÍA. 2016. Disentangling environmental and spatial processes of community assembly in tropical forests from local to regional scales. *Oikos* 125: 326–335.
- ARIAS, J., and F. W. STAUFFER. 2012. Notas sobre siete especies de palmeras (Arecaceae) de bosque nublado, presentes en la vertiente noreste de la serranía de Perijá, estado Zulia, Venezuela. *Revista Biodiversidad Neotropical* 2: 93–101.
- BALDECK, C. A., K. E. HARMS, J. B. YAVITT, R. JOHN, B. L. TURNER, R. VALENCIA, H. NAVARRETE, S. J. DAVIES, G. B. CHUYONG, D. KENFACK, D. W. THOMAS, S. MADAWALA, N. GUNATILLEKE, S. GUNATILLEKE, S. BUNYAVEJCHEWIN, S. KIRATIPRAYOON, A. YAACOB, M. N. N. SUPARDI, and J. W. DALLING. 2013. Soil resources and topography shape local tree community structure in tropical forests. *Proc. R. Soc. B* 280: 20122532.
- BALSLEV, H., Z. PÉREZ DURÁN, D. PEDERSEN, W. L. EISERHARDT, A. SANJINÉS ASTURIZAGA, and N. PANIAGUA-ZAMBRANA. 2012. Comunidades de palmas de bosques subandinos y de las zonas bajas adyacentes en Bolivia. *Ecología en Bolivia* 47: 7–36.
- BLACH-OVERGAARD, A., J.-C. SVENNING, J. DRANSFIELD, M. GREVE, and H. BALSLEV. 2010. Determinants of palm species distributions across Africa: the relative roles of climate, non-climatic environmental factors, and spatial constraints. *Ecography* 33: 380–391.
- BORCHSENIUS, F., and M. MORAES. 2006. Diversidad y usos de palmeras andinas (Arecaceae). *Botánica económica de los Andes Centrales* 412–433.
- CÁMARA-LERET, R., H. TUOMISTO, K. RUOKOLAINEN, H. BALSLEV, and S. M. KRISTIANSEN. 2017. Modelling responses of western Amazonian palms to soil nutrients. *Journal of Ecology* 105: 367–381.
- CARLOS-COPETE, J., R. CÁMARA-LERET, M. SANCHEZ, H. BALSLEV, J. CARLOS-COPETE, R. CÁMARA-LERET, M. SANCHEZ, and H. BALSLEV. 2019. Relación entre la composición florística y los nutrientes del suelo en comunidades de palmas del Chocó biogeográfico en Colombia y Ecuador. *Revista de Biología Tropical* 67: 716–732.
- COLWELL, R. K., and D. C. LEES. 2000. The mid-domain effect: geometric constraints on the geography of species richness. *Trends in Ecology & Evolution* 15: 70–76.
- CONDIT, R., B. M. J. ENGELBRECHT, D. PINO, R. PÉREZ, and B. L. TURNER. 2013. Species distributions in response to individual soil nutrients and seasonal drought across a community of tropical trees. *Proceedings of the National Academy of Sciences* 110. Available at: <https://www.readcube.com/articles/10.1073%2Fpnas.1218042110> [Accessed November 5, 2020].

- COSTA, F. R. C., J.-L. GUILLAUMET, A. P. LIMA, and O. S. PEREIRA. 2009. Gradients within gradients: The mesoscale distribution patterns of palms in a central Amazonian forest. *Journal of Vegetation Science* 20: 69–78.
- CURRIE, D. J., G. G. MITTELBACH, H. V. CORNELL, R. FIELD, J. F. GUEGAN, B. A. HAWKINS, D. M. KAUFMAN, J. T. KERR, T. OBERDORFF, E. O'BRIEN, and J. R. G. TURNER. 2004. Predictions and tests of climate-based hypotheses of broad-scale variation in taxonomic richness. *Ecology Letters* 7: 1121–1134.
- DUIVENVOORDEN, J., A. DUQUE, J. CAVELIER, A. GARCÍA, C. GRÁNDEZ, M. MACÍA, H. ROMERO-SALTOS, M. SANCHEZ, and R. VALENCIA. 2005. Density and diversity of plants in relation to soil nutrient reserves in well-drained upland forests in the north-western Amazon basin. *Biologiske Skrifter ISSN 0366-3612* 55: 25–35.
- EISERHARDT, W. L., J.-C. SVENNING, F. BORCHSENIUS, T. KRISTIANSEN, and H. BALSLEV. 2013. Separating environmental and geographical determinants of phylogenetic community structure in Amazonian palms (Arecaceae). *Botanical Journal of the Linnean Society* 171: 244–259.
- EISERHARDT, W. L., J.-C. SVENNING, W. D. KISSLING, and H. BALSLEV. 2011. Geographical ecology of the palms (Arecaceae): determinants of diversity and distributions across spatial scales. *Ann. Bot.* 108: 1391–1416.
- EVANS, K. L., J. J. D. GREENWOOD, and K. J. GASTON. 2005. Dissecting the species-energy relationship. *Proceedings of the Royal Society B-Biological Sciences* 272: 2155–2163.
- FICK, S. E., and R. J. HIJMANS. 2017. WorldClim 2: new 1-km spatial resolution climate surfaces for global land areas. *Int. J. Climatol* 37: 4302–4315.
- FIGUEIREDO, F. O. G., G. ZUQUIM, H. TUOMISTO, G. M. MOULATLET, H. BALSLEV, and F. R. C. COSTA. 2018. Beyond climate control on species range: The importance of soil data to predict distribution of Amazonian plant species. *Journal of Biogeography* 45: 190–200.
- FISCHER, A., M. BLASCHKE, and C. BÄSSLER. 2011. Altitudinal gradients in biodiversity research: The state of the art and future perspectives under climate change aspects. *Waldökologie Online* 11: 35–47.
- FRANCIS, A. P., and D. J. CURRIE. 2003. A globally consistent richness-climate relationship for angiosperms. *American Naturalist* 161: 523–536.
- FUENTES, A. 2005. Una introducción a la vegetación de la región de Madidi. *Ecología en Bolivia* 40: 1–31.
- GARIBALDI, A., and N. TURNER. 2004. Cultural Keystone Species: Implications for Ecological Conservation and Restoration. *Ecology and Society* 9. Available at: <https://www.ecologyandsociety.org/vol9/iss3/art1/> [Accessed November 5, 2020].

- GEROLD, G., M. SCHAWÉ, and K. BACH. 2008. Hydrometeorologic, Pedologic and Vegetation Patterns along an Elevational Transect in the Montane Forest of the Bolivian Yungas. *Die Erde* 139.
- GIRARDIN, C. A. J., Y. MALHI, L. E. O. C. ARAGAO, M. MAMANI, W. H. HUASCO, L. DURAND, K. J. FEELEY, J. RAPP, J. E. SILVA-ESPEJO, M. SILMAN, N. SALINAS, and R. J. WHITTAKER. 2010. Net primary productivity allocation and cycling of carbon along a tropical forest elevational transect in the Peruvian Andes RID A-7631-2009 RID H-1280-2011. *Glob. Change Biol.* 16: 3176–3192.
- HAWKINS, B. A., R. FIELD, H. V. CORNELL, D. J. CURRIE, J. F. GUEGAN, D. M. KAUFMAN, J. T. KERR, G. G. MITTELBACH, T. OBERDORFF, O. ', E. M. BRIEN, E. E. PORTER, and J. R. G. TURNER. 2003. Energy, water, and broad-scale geographic patterns of species richness. *Ecology* 84: 3105–3117.
- HENDERSON, A. 2002. *Evolution and Ecology of Palms*. New York Botanical Garden Pr Dept, Bronx, NY.
- HENDERSON, A., G. GALEANO, and R. BERNAL. 2019. *Field Guide to the Palms of the Americas*. Princeton University Press.
- JOHN, R., J. W. DALLING, K. E. HARMS, J. B. YAVITT, R. F. STALLARD, M. MIRABELLO, S. P. HUBBELL, R. VALENCIA, H. NAVARRETE, M. VALLEJO, and R. B. FOSTER. 2007. Soil nutrients influence spatial distributions of tropical tree species. *PNAS* 104: 864–869.
- JONES, M. M., S. FERRIER, R. CONDIT, G. MANION, S. AGUILAR, and R. PÉREZ. 2013. Strong congruence in tree and fern community turnover in response to soils and climate in central Panama. *Journal of Ecology* 101: 506–516.
- JONES, M. M., B. SZYSKA, and M. KESSLER. 2011. Microhabitat partitioning promotes plant diversity in a tropical montane forest RID A-3605-2009. *Glob. Ecol. Biogeogr.* 20: 558–569.
- JONES, M. M., H. TUOMISTO, D. BORCARD, P. LEGENDRE, D. B. CLARK, and P. C. OLIVAS. 2008. Explaining variation in tropical plant community composition: influence of environmental and spatial data quality. *Oecologia* 155: 593–604.
- KRAFT, N. J. B., R. VALENCIA, and D. D. ACKERLY. 2008. Functional Traits and Niche-Based Tree Community Assembly in an Amazonian Forest. *Science* 322: 580–582.
- KRISTIANSEN, T., J.-C. SVENNING, W. EISERHARDT, D. PEDERSEN, H. BRIX, S. KRISTIANSEN, M. KNADEL, C. GRÁNDEZ, and H. BALSLEV. 2012. Environment versus dispersal in the assembly of western Amazonian palm communities. *Journal of Biogeography* 39: 1318–1332.
- LEFCHECK, J. S. 2016. piecewiseSEM: Piecewise structural equation modelling in r for ecology, evolution, and systematics. *Methods in Ecology and Evolution* 7: 573–579.

- LEGENDRE, P., and E. D. GALLAGHER. 2001. Ecologically meaningful transformations for ordination of species data. *Oecologia* 129: 271–280.
- LLOYD, J., and J. A. TAYLOR. 1994. On the Temperature Dependence of Soil Respiration. *Functional Ecology* 8: 315–323.
- LOMOLINO, M. V. 2008. Elevation gradients of species-density: historical and prospective views. *Global Ecology and Biogeography* 10: 3–13.
- MCCAIN, C. M., and J.-A. GRYTNES. 2010. Elevational Gradients in Species Richness. *In* eLS. American Cancer Society. Available at: <https://onlinelibrary.wiley.com/doi/abs/10.1002/9780470015902.a0022548> [Accessed November 5, 2020].
- MONTOYA, F., and M. MORAES R. 2014. Palmeras utilizadas por los indígenas Yuracaré del Territorio Indígena Parque Nacional Isiboro-Sécure (Cochabamba, Bolivia). *Revue d'ethnoécologie* 5.
- MORAES, M., J. SARMIENTO, and E. OVIEDO. 1995. Richness and uses in a diverse palm site in Bolivia. *Biodiversity & Conservation* 4: 719–727.
- MORAES R., M. 2015. Actualización de la lista de especies nativas de Arecaceae para Bolivia. *Revista de la Sociedad Boliviana de Botánica* 8: 17–26.
- MUÑOZ, G., K. TRØJELSGAARD, and W. D. KISSLING. 2019. A synthesis of animal-mediated seed dispersal of palms reveals distinct biogeographical differences in species interactions. *Journal of Biogeography* 46: 466–484.
- MYERS, J. A., J. M. CHASE, I. JIMÉNEZ, P. M. JØRGENSEN, A. ARAUJO-MURAKAMI, N. PANIAGUA-ZAMBRANA, and R. SEIDEL. 2013. Beta-diversity in temperate and tropical forests reflects dissimilar mechanisms of community assembly. *Ecology Letters* 16: 151–157.
- NETTESHEIM, F. C., M. L. GARBIN, P. H. M. RAJÃO, D. S. D. ARAUJO, and C. E. V. GRELLE. 2018. Environment is more relevant than spatial structure as a driver of regional variation in tropical tree community richness and composition. *Plant Ecology & Diversity* 11: 27–40.
- OKSANEN, J., F. G. BLANCHET, M. FRIENDLY, R. KINDT, P. LEGENDRE, D. MCGLINN, P. R. MINCHIN, R. B. O'HARA, G. L. SIMPSON, and P. SOLYMOS. 2016. *vegan: Community Ecology Package*. R package version 2.4-3. Vienna: R Foundation for Statistical Computing.[Google Scholar].
- PANIAGUA-ZAMBRANA, N., R. HURTADO, and M. M. R. Palmeras y usos: Especies de Bolivia y la región Available at: [https://www.academia.edu/43812207/Palmeras\\_y\\_usos\\_Especies\\_de\\_Bolivia\\_y\\_la\\_regi%C3%B3n](https://www.academia.edu/43812207/Palmeras_y_usos_Especies_de_Bolivia_y_la_regi%C3%B3n) [Accessed November 5, 2020].

- PHILLIPS, O. L., P. N. VARGAS, A. L. MONTEAGUDO, A. P. CRUZ, M. E. C. ZANS, W. G. SANCHEZ, M. YLI-HALLA, and S. ROSE. 2003. Habitat association among Amazonian tree species: a landscape-scale approach. *Journal of Ecology* 91: 757–775.
- PINHEIRO, J., D. BATES, S. DEBROY, D. SARKAR, S. HEISTERKAMP, and B. VAN WILLIGEN. 2020. nlme: Linear and Nonlinear Mixed Effects Models Available at: <https://CRAN.R-project.org/package=nlme> [Accessed November 5, 2020].
- PINTAUD, J.-C., G. GALEANO, H. BALSLEV, R. BERNAL, F. BORCHSENIUS, E. FERREIRA, J.-J. DE GRANVILLE, K. MEJÍA, B. MILLÁN, and M. MORAES. 2008. Las palmeras de América del Sur: diversidad, distribución e historia evolutiva. *Revista peruana de biología* 15: 7–30.
- POULSEN, A. D., H. TUOMISTO, and H. BALSLEV. 2006. Edaphic and floristic variation within a 1-ha plot of lowland Amazonian rain forest. *Biotropica* 38: 468–478.
- PRADA, C. M., A. MORRIS, K. M. ANDERSEN, B. L. TURNER, P. CABALLERO, and J. W. DALLING. 2017. Soils and rainfall drive landscape-scale changes in the diversity and functional composition of tree communities in premontane tropical forest. *Journal of Vegetation Science* 28: 859–870.
- RAHBEK, C. 1995. The elevation gradient of species richness: a uniform pattern? *Ecography* 18: 200–205.
- SCHLINDWEIN, G., A. TONIETTO, A. D. ABICHEQUER, A. C. DE AZAMBUJA, B. B. LISBOA, L. K. VARGAS, G. SCHLINDWEIN, A. TONIETTO, A. D. ABICHEQUER, A. C. DE AZAMBUJA, B. B. LISBOA, and L. K. VARGAS. 2017. Pindo Palm fruit yield and its relationship with edaphic factors in natural populations in Rio Grande do Sul. *Ciência Rural* 47. Available at: [http://www.scielo.br/scielo.php?script=sci\\_abstract&pid=S0103-84782017000200404&lng=en&nrm=iso&tlng=en](http://www.scielo.br/scielo.php?script=sci_abstract&pid=S0103-84782017000200404&lng=en&nrm=iso&tlng=en) [Accessed November 5, 2020].
- SESNIE, S. E., B. FINEGAN, P. E. GESSLER, and Z. RAMOS. 2009. Landscape-Scale Environmental and Floristic Variation in Costa Rican Old-Growth Rain Forest Remnants. *Biotropica* 41: 16–26.
- SVENNING, J.-C. 2001. Environmental heterogeneity, recruitment limitation and the mesoscale distribution of palms in a tropical montane rain forest (Maquipucuna, Ecuador). *Journal of Tropical Ecology* 97–113.
- TELLO, J. S., M. J. MACIA, G. ARELLANO, L. CAYOLA, and A. F. FUENTES. 2018. Madidi Project Full Dataset (Version 2.1). Available at: <https://zenodo.org/record/4280178#.X7bIqFBMFPY> [Accessed November 19, 2020].
- TELLO, J. S., J. A. MYERS, M. J. MACÍA, A. F. FUENTES, L. CAYOLA, G. ARELLANO, M. I. LOZA, V. TORREZ, M. CORNEJO, T. B. MIRANDA, and P. M. JØRGENSEN. 2015. Elevational Gradients in  $\beta$ -Diversity Reflect Variation in the Strength of Local Community Assembly Mechanisms across Spatial Scales. *PLoS ONE* 10: e0121458.
- TELLO, J. S., and R. D. STEVENS. 2010. Multiple environmental determinants of regional species richness and effects of geographic range size. *Ecography* 33: 796–808.

- TUOMISTO, H., A. D. POULSEN, K. RUOKOLAINEN, R. C. MORAN, C. QUINTANA, J. CELI, and G. CANAS. 2003. Linking floristic patterns with soil heterogeneity and satellite imagery in Ecuadorian Amazonia. *Ecological Applications* 13: 352–371.
- TUOMISTO, H., K. RUOKOLAINEN, and M. YLI-HALLA. 2003. Dispersal, Environment, and Floristic Variation of Western Amazonian Forests. *Science* 299: 241–244.
- TURNER, B. L., T. BRENES-ARGUEDAS, and R. CONDIT. 2018. Pervasive phosphorus limitation of tree species but not communities in tropical forests. *Nature* 555: 367–370.
- VENABLES, W. N., and B. D. RIPLEY. 2002. *Modern Applied Statistics with S* 4th ed. Springer-Verlag, New York Available at: <https://www.springer.com/gp/book/9780387954578> [Accessed November 5, 2020].
- ZONA, S., and A. HENDERSON. 1989. A review of animal-mediated seed dispersal of palms. *Selbyana* 6–21.