

1 **Title: Competition in depleting resource environments shapes the thermal response of**  
2 **mosquito population fitness**

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33 **ABSTRACT**

- 34 1. A population's maximal growth rate ( $r_m$ ) depends on the survivorship, development,  
35 and reproduction of its individuals. In ectotherms, these (functional) traits respond  
36 predictably to temperature, which provides a basis for predicting how climatic  
37 warming could affect natural populations, including disease vectors and the diseases  
38 they transmit.
- 39 2. Such predictions generally arise from mathematical models that incorporate the  
40 temperature-dependence of traits (thermal performance curves) measured under  
41 laboratory conditions. Therefore, the accuracy of these predictions depends on the  
42 relevance of lab-measured trait thermal performance curves to natural conditions.  
43 However, the joint effect of temperature and resource availability—another key  
44 limiting environmental factor in nature—on traits is largely unknown.
- 45 3. We investigated how larval competition for ecologically-realistic depleting resources  
46 affects the thermal performance of  $r_m$  and its underlying life history traits in the  
47 disease vector in *Aedes aegypti*. We show that competition at food concentrations  
48 below a certain threshold drastically depresses  $r_m$  across the entire temperature range,  
49 causes it to peak at a lower temperature, and narrows the breadth of temperatures over  
50 which  $r_m$  is positive (the thermal niche breath).
- 51 4. This resource-dependence of the thermal performance curve of  $r_m$  is driven primarily  
52 by the fact that competition delays development and increases juvenile mortality. This  
53 is compounded by reduced size at maturity, which in turn decreases adult lifespan and  
54 fecundity.
- 55 5. These results show that intensified larval competition in depleting resource  
56 environments can significantly affect the temperature-dependence of  $r_m$  by  
57 modulating the thermal responses of underlying traits in a predictable way. This has  
58 important implications for forecasting the effects of climate change on population  
59 dynamics in the field of not just disease vectors, but holometabolous insects in  
60 general.

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62 Key words: ectotherm, climatic warming,  $r_m$ , population-level fitness, thermal performance  
63 curve, resource depletion, competition, vector-borne disease.

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## 67 INTRODUCTION

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69 Environmental change is predicted to affect the spatiotemporal distributions of ectotherms  
70 (Deutsch et al. 2008), including disease vectors (Mordecai et al. 2020, WHO 2020). For  
71 example, recent studies suggest that climatic warming may increase the thermal suitability for  
72 Zika virus transmission, leading to 1.3 billion more people being at risk of exposure by 2050  
73 (Ryan et al. 2021). Such predictions often arise from disease transmission models that  
74 incorporate thermal performance curves (TPCs) for vector life history traits, such as juvenile  
75 development and mortality, which together define the TPC of maximal population growth  
76 rate ( $r_m$ , a measure of population fitness; Savage et al. 2004).

77

78 Typically, trait-level TPCs are obtained from data that come from vector populations reared  
79 under optimal food conditions in the laboratory. Yet, in nature insect populations experience  
80 spatiotemporal variation in resource availability. Indeed, very few empirical studies have  
81 examined how the  $r_m$  TPC may be shaped by intraspecific competition (but see Johnson et al.  
82 2016, Mallard et al. 2020). In particular, the effect of competition for depleting resources on  
83 the  $r_m$  TPC through its effects on underlying traits has been rarely studied, although resource  
84 depletion in adult and larval habitats is arguably a common scenario in nature (Ostfeld and  
85 Keesing 2000, Wright 1983, Yee and Juliano 2012). For example, resource depletion should  
86 occur at a faster rate with rising temperatures because per capita energy requirements for  
87 somatic maintenance and growth increase with temperature (Kooijman 2000, West et al.  
88 2001).

89

90 As competition intensifies with increasing temperature and resource depletion rates, the  
91 growth, development, and survival of increasing numbers of individuals are bound to be  
92 compromised. These trait-level effects are then expected to propagate through the stage-  
93 structured population dynamics to change the shape of the  $r_m$  TPC (Amarasekare and Savage  
94 2012, Huey and Kingsolver 2019). This is because  $r_m$  is essentially proportional to the  
95 difference between biomass gained through consumption and that lost to respiration and  
96 mortality (Savage et al. 2004). Resource limitation and competition would be expected to  
97 decrease  $r_m$  across temperatures as they would both undermine biomass intake and elevate  
98 biomass loss. Furthermore, if the rate of biomass loss increases faster than any increase in  
99 biomass gain with temperature, the thermal optimum of  $r_m$  ( $T_{opt}$ ) may also shift downwards  
100 (García-Carreras et al. 2018). For the same reason, the range of temperatures over which  $r_m$  is

101 positive (the thermal niche width) may become narrower. As a result, species range  
102 boundaries could contract due to the combined effects of climatic warming and intensified  
103 competition. Conversely, concurrent increases in temperature and resource availability with  
104 climatic warming would optimise  $r_m$ , promoting the invasion of tropical taxa into temperate  
105 habitats (Amarasekare and Simon 2020).

106

107 Studies on how intraspecific larval competition affects the  $r_m$  TPC are rare, but recent  
108 empirical work provides some insights into how competition may affect fitness traits. For  
109 example, a study of the bordered plant bug suggests that intraspecific competition between  
110 adults for resources is strongest at temperatures that are optimal for reproduction (Johnson et  
111 al. 2016). Other studies suggest that the effects of competition on individual growth rates,  
112 fecundity and body size may change as the thermal environment and population dynamics  
113 change (Laws and Belovsky 2010, Mallard et al. 2020). However, despite these advances, the  
114 question of whether and how larval competition in depleting resource environments can alter  
115 the shape of the  $r_m$  TPC remains open.

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117 Addressing this question is particularly important for semiaquatic insects, such as  
118 mosquitoes. This is because such species are expected to be mainly regulated by larval  
119 competition for limited resources (Dye 1984), and have juvenile stages that are restricted to  
120 small, ephemeral aquatic habitats that are susceptible to infrequent resource inputs  
121 (Arrivillaga and Barrera 2004, Barrera et al. 2006, Yee and Juliano 2012). Here, we  
122 investigate the effects of intraspecific competition on the  $r_m$  TPC by exposing *Ae. aegypti*  
123 larvae to a realistic range of temperatures and resource concentration levels. We show that, in  
124 this species, intensified larval competition at low resource concentration can significantly  
125 change the shape of the  $r_m$  TPC through its effects on underlying traits in a predictable and  
126 general way.

127

## 128 **METHODS**

129

130 To investigate the effects of temperature and larval competition for food on mosquito life  
131 history, we employed a  $5 \times 4$  factorial design comprised of five temperatures (22, 26, 32, 34,  
132 and 36°C) and resource concentration levels (0.183, 0.367, 0.550 and 0.733 mg ml<sup>-1</sup>). These  
133 resource levels simulate a situation where juvenile habitats do not receive any additional  
134 resource during the timescale relevant to our single generation study. Ensuring constant

135 resource concentration was important to isolate the effects of competition in depleting  
136 resource environments. The experiment was carried out in two consecutive blocks.  
137 Temperatures were randomly selected to determine the order in which they were to be tested.  
138 The first block comprised of 22, 34 and 36°C, the second comprised of 26 and 32°C. We  
139 selected our temperature range to span the minimum and maximum temperatures  
140 (approximately 22 to 34°C) that this strain of *Ae. aegypti* is likely to experience in the wild  
141 between May (the onset of mosquito season) and November (Fort Meyer, FL; Arguez et al.  
142 2012). We extended our range to 36°C to determine the upper critical thermal limit for this  
143 strain.

144

145 Batches of approximately 800 *Ae. aegypti* (F16-19) eggs were randomly assigned to one of  
146 the five experimental temperatures and immersed in plastic tubs containing 300 ml of tap  
147 water. Each tub was provided with a pinch of powdered fish food (Cichlid Gold®, Hikari,  
148 Kyrin Food Industries Ltd., Japan) to stimulate overnight hatching. The tubs were then  
149 submerged in water baths (Grant Instruments: JAB Academy) set at either 22, 26, 32, 34, or  
150 36°C. Water baths were situated in a 20°C climate-controlled insectary with a 12L:12D  
151 photoperiod and 30 minutes of gradual transition of light levels to simulate sunrise and  
152 sunset.

153

154 On the following day, first instar larvae were separated into cohorts of 50 and held in tubs  
155 containing 300 ml of water. We created three replicate tubs per treatment (150 individuals  
156 treatment<sup>-1</sup>). Resource concentration levels were attained by adding 55, 110 and 220 mg to  
157 the tubs, respectively. Prior testing of these levels resulted in high levels of juvenile mortality  
158 due to fouling at the higher temperature × resource levels. Therefore, for all treatments, half  
159 of the assigned food quantity was provided on Day 1. The remaining half was provided on  
160 Day 4 after any remaining food from Day 1 had been sieved and rearing waters had been  
161 replenished with fresh tap water. After Day 4, resource levels were not adjusted but water  
162 volumes were adjusted, as required.

163

#### 164 **Fitness calculation**

165

166 To calculate  $r_m$ , we used our life history trait data to parameterise stage-structured matrix  
167 projection models (eqn. 1; Caswell 1989), which describe change in a population over time:

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$$\mathbf{N}_{t+1} = \mathbf{M}\mathbf{N}_t, \quad (1)$$

169  
170 where  $\mathbf{N}_t$  is a vector of abundances in the stage classes at time  $t$  and  $\mathbf{M}$  is the population  
171 projection matrix. The first row of  $\mathbf{M}$  is populated by daily fecundity (the number of female  
172 offspring produced per female at age  $i$ ). The sub-diagonal of  $\mathbf{M}$  is populated with the  
173 probabilities of survival from age  $i$  to age  $i+1$ . Multiplying the transition matrix ( $\mathbf{M}$ ; eqn. 1)  
174 and stage-structured population size vector ( $\mathbf{N}_t$ ; eqn. 1) sequentially across time intervals  
175 yields the stage-structured population dynamics. Once the stable stage distribution of the  
176 abundance vector is reached, the dominant eigenvalue of the system is the finite population  
177 rate of increase ( $\lambda$ ) (Caswell 1989). Then, the intrinsic rate of population growth is

$$r_m = \log(\lambda).$$

179  
180 This is a population's inherent capacity to reproduce, and therefore a measure of population-  
181 level fitness (Birch 1948, Cole 1954, Savage et al. 2004). Negative  $r_m$  values indicate decline  
182 and positive ones, growth. The projection matrices were built and analysed using the  
183 `popbio` R package (R Core Team 2018, Stubben and Milligan 2007).

184

## 185 **Model parameterisation**

186

### 187 *Immature development time and immature and adult survival proportions*

188 Matrix survival elements (the sub-diagonal of the matrix  $\mathbf{M}$ ; eqn. 1) were populated with  
189 continuous survival proportions estimated using the Kaplan-Meier survival function in the  
190 `survival` R package (R Core Team 2018, Therneau 2020). We assumed life stage duration  
191 (i.e. larva-to-pupa-to-adult) was the mean duration of transitioning into and out of that stage,  
192 and a fixed age of adult emergence at the mean age of emergence. Adult survival elements  
193 were populated with the Kaplan-Meier proportions. Hatching-to-adult development times  
194 were calculated by recording the day and time that egg eclosion, pupation and adult  
195 emergence occurred for each individual. Upon pupation, mosquitoes were held in individual  
196 falcon tubes containing 5 ml of tap water. This enabled pupa-to-adult development durations  
197 and the lifespans of individual starved adults to be recorded. Adult lifespan was recorded in  
198 the absence of food, which forces adults to metabolise the nutritional reserves accumulated  
199 during larval development. Therefore, starved adult lifespan should increase with size at  
200 emergence, so it is a useful indicator of the carry over effects of temperature and competition  
201 in the larval habitat (Agnew et al. 2002, Briegel 1990).

## 202 *Daily fecundity rate*

203 The use of scaling relationships between fecundity and size is common in predictions of  
204 population growth in *Aedes* mosquitoes (Focks et al. 1993, Juliano 1998). A detailed  
205 description of our method for estimating fecundity is provided in the supplementary material  
206 (Fig. S1). Briefly, we measured individual dry mass, and estimated lifetime fecundity using  
207 previously published datasets on the temperature-dependent scaling between mass and wing  
208 length (van den Heuvel 1963), and wing length and fecundity (Briegel 1990, Farjana and  
209 Tuno 2012). Daily fecundity rate is required for the matrix projection models (eqn. 1), so we  
210 divided lifetime fecundity by lifespan and multiplied by 0.5 (assuming a 1:1 male-to-female  
211 offspring ratio) to give temperature-specific individual daily fecundity. Later, we show that  
212 this much variation in the scaling of fecundity does not qualitatively change our results.

213

## 214 **Parameter sensitivity**

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216 We used the delta method to approximate 95% confidence intervals (CIs) to account for how  
217 uncertainty in survival and fecundity estimates is propagated through to the  $r_m$  estimate  
218 (Caswell 1989, Skalski et al. 2007). This method requires the standard errors of the survival  
219 and fecundity element estimates. For survival, we used the standard errors estimated by the  
220 Kaplan-Meier survival function in the `survival` R package. For fecundity, we calculated  
221 the standard errors of the mean daily fecundity rates (Table S2) for each treatment using the  
222 `Rmisc` R package (Hope 2013). As an additional sensitivity analysis, we recalculated fitness  
223 using the upper and lower 95% CIs of the exponents for the scaling of size and lifetime  
224 fecundity (Fig. S3).

225

## 226 **Elasticity analysis**

227

228 We used elasticities to quantify the proportional contributions of individual life history traits  
229 to  $r_m$ . Elasticity,  $e_{ij}$ , measures the proportional effect on  $\lambda$  of an infinitesimal change in an  
230 element of  $\mathbf{M}$  (eqn. 1) with all other elements held constant (the partial derivative) (Caswell  
231 et al. 1984, de Kroon et al. 1986). This partial derivative of  $\lambda$ , with respect to each element of  
232  $\mathbf{M}$ , is  $s_{ij} = \partial\lambda/\partial a_{ij} = v_i w_j$  with the dot product  $\langle \mathbf{w}, \mathbf{v} \rangle = 1$ . Here,  $\mathbf{w}$  is the dominant right  
233 eigenvector (the stage distribution vector of  $\mathbf{M}$ ),  $\mathbf{v}$  is the dominant left eigenvector (the  
234 reproductive value vector of  $\mathbf{M}$ ), and  $a_{ij}$  is the  $i \times j^{\text{th}}$  element of  $\mathbf{M}$ . Elasticities can then be

235 calculated using the relationship:  $e_{ij} = a_{ij}/\lambda \times s_{ij}$ . Multiplying an elasticity by  $\lambda$  gives the  
236 absolute contribution of its corresponding  $a_{ij}$  to  $\lambda$  (Caswell et al. 1984, de Kroon et al. 1986).  
237 Absolute contributions for juvenile and adult elements were summed and changed  
238 proportionally to quantify the sensitivity of  $r_m$  to these traits.

239

## 240 **Statistical analyses**

241

242 All statistical analyses were conducted using R (R Core Team 2018). The trait data (adult  
243 lifespan and body size) were normally distributed, so we used full factorial linear models  
244 (LM) to determine the significance of each predictor on the thermal response of each of these  
245 traits. The development time data were nonlinear, positive, and right skewed, so we used a  
246 generalized linear model (GLM) with gamma distribution and identity link functions  
247 (predictor effects were considered additive). Replicate was included in all regression models  
248 as a fixed effect. We investigated the effect of larval competition on the temperature-  
249 dependency of daily per capita juvenile mortality rate by fitting an exponential function to the  
250 survival data with R package `flexsurv` (Jackson 2016). The final mortality model was  
251 obtained by dropping terms from the full model (consisting of temperature  $\times$  resource level +  
252 replicate as fixed effect predictors). If removing a term did not improve model fit ( $\Delta AIC >$   
253  $-2$ ), it was removed from the full model (Table S3).

254

## 255 **Quantifying $r_m$ 's thermal performance curve**

256

257 To determine how competition affected the shape of the  $r_m$  TPC, we fitted several  
258 mathematical models that allow for negative values at both cold and hot extremes, including  
259 polynomial models using linear regression, as well as non-linear models with non-linear least  
260 squares (NLLS) using the `rTPC` R package (Padfield et al. 2020). Overall, the Lactin2 (Lactin  
261 et al. 1995) and Kamykowski (Kamykowski and McCollum 1986) models were equally best-  
262 fitting according to the Akaike Information Criterion (AIC) (Table S4). From these, we  
263 picked the Kamykowski model because it was better at describing the estimated  $r_m$  at our  
264 lowest resource level. This model is defined as

265

$$r_m(T) = a(1 - e^{-b(T - T_{\min})})(1 - e^{-c(T_{\max} - T)}),$$

266



267 where  $T$  ( $^{\circ}\text{C}$ ), and  $T_{\max}$  and  $T_{\min}$  are the high and low temperatures at which  $r_m$  becomes  
268 negative, respectively, and  $a$ ,  $b$ , and  $c$ , are shape parameters without any biological meaning.  
269 Bootstrapping was used to calculate 95% prediction bounds for each  $r_m$  TPC (Padfield et al.  
270 2020) and confidence intervals (CIs) around its  $T_{\text{opt}}$ , as well as the thermal niche width ( $T_{\max}$   
271  $- T_{\min}$ ).

272

## 273 RESULTS

274

275 All trait responses varied significantly with temperature and resource level, with a significant  
276 interaction between the two environmental variables (Fig. 1, Tables S1, S2). Therefore, the  
277 realised effect of temperature on trait responses was consistently and significantly mediated  
278 by the how the depletion of resources at low resource availability increased the negative  
279 effects of intraspecific competition.

280

281 Competition at low resource concentration ( $0.183 \text{ mg ml}^{-1}$ ) increased the negative effect of  
282 increased temperature on juvenile mortality rate (Fig. 1a, Table S2). As temperatures  
283 increased from  $22$  to  $34^{\circ}\text{C}$ , non-overlapping 95% credible intervals indicate that juvenile  
284 mortality rate was significantly higher at low resource concentration than at the intermediate  
285 resource concentration ( $0.367 \text{ mg ml}^{-1}$ ). At  $0.183 \text{ mg ml}^{-1}$ , it increased from  $0.05$  at  $22^{\circ}\text{C}$  to  
286  $0.14 \text{ individual}^{-1} \text{ day}^{-1}$  at  $34^{\circ}\text{C}$ , whereas, at  $0.367 \text{ mg ml}^{-1}$ , it increased from  $0.04$  to  $0.05$   
287  $\text{individual}^{-1} \text{ day}^{-1}$  across this temperature range.

288

289 Development time varied significantly with the interaction between temperature and resource  
290 concentration (ANOVA;  $F_{9, 2.24} = 13.44$ ,  $p < 0.001$ , Table S1). Development time decreased  
291 with temperature at all resource levels, but the decrease with temperature was greater at the  
292 low resource level than at higher resource levels due to intensified competition. At  $0.183 \text{ mg}$   
293  $\text{ml}^{-1}$ , development time decreased from  $18.30$  days at  $22^{\circ}\text{C}$  to  $8.26$  days at  $34^{\circ}\text{C}$ .  
294 Development time at the higher resource levels decreased from approximately  $13.50$  days at  
295  $22^{\circ}\text{C}$  to  $\sim 7.50$  days at  $34^{\circ}\text{C}$  (Table S2).

296

297 Intensified competition at low resource concentration resulted in significant variation in size  
298 at maturity (mass,  $\text{mg}$ ) between resource levels (ANOVA;  $F_{9, 0.92} = 24.26$ ,  $p < 0.001$ , Table  
299 S1). Adult size decreased both at warmer temperatures and at lower food concentration,  
300 though the decrease with temperature was greater at higher resource levels than at the low

301 resource level. At low resource concentration, size decreased by 0.13 mg as temperatures  
302 increased from 22 to 34°C, while at the highest resource concentration (0.733 mg ml<sup>-1</sup>), size  
303 decreased by 0.26 mg (Fig 1c, Table S2).

304

305 Adult lifespan varied significantly with the interaction between temperature and resource  
306 concentration (ANOVA;  $F_{9, 699.60} = 7.96$ ,  $p < 0.001$ , Table S1). Lifespan was greatest at 0.733  
307 mg ml<sup>-1</sup>, where it was 11.24 days at 22°C. It increased to 11.65 days at 26°C, and then  
308 decreased to 4.68 days at 34°C. In contrast, at low resource concentration, lifespan decreased  
309 from 8.00 days at 22°C to 3.76 days at 34°C mg (Fig 1d, Table S2).

310

311 At all resource levels, predicted daily fecundity rate increased with temperature (Table S2),  
312 though the increase was greater at low resource concentration than at higher resource levels.

313 At low resource concentration, fecundity increased with temperature from 1.77 eggs  
314 individual<sup>-1</sup> day<sup>-1</sup> at 22°C to 6.87 eggs individual<sup>-1</sup> day<sup>-1</sup> at 34°C. At the higher resource  
315 levels, fecundity increased from ~2 eggs at 22°C to ~5 eggs individual<sup>-1</sup> day<sup>-1</sup> at 34°C.

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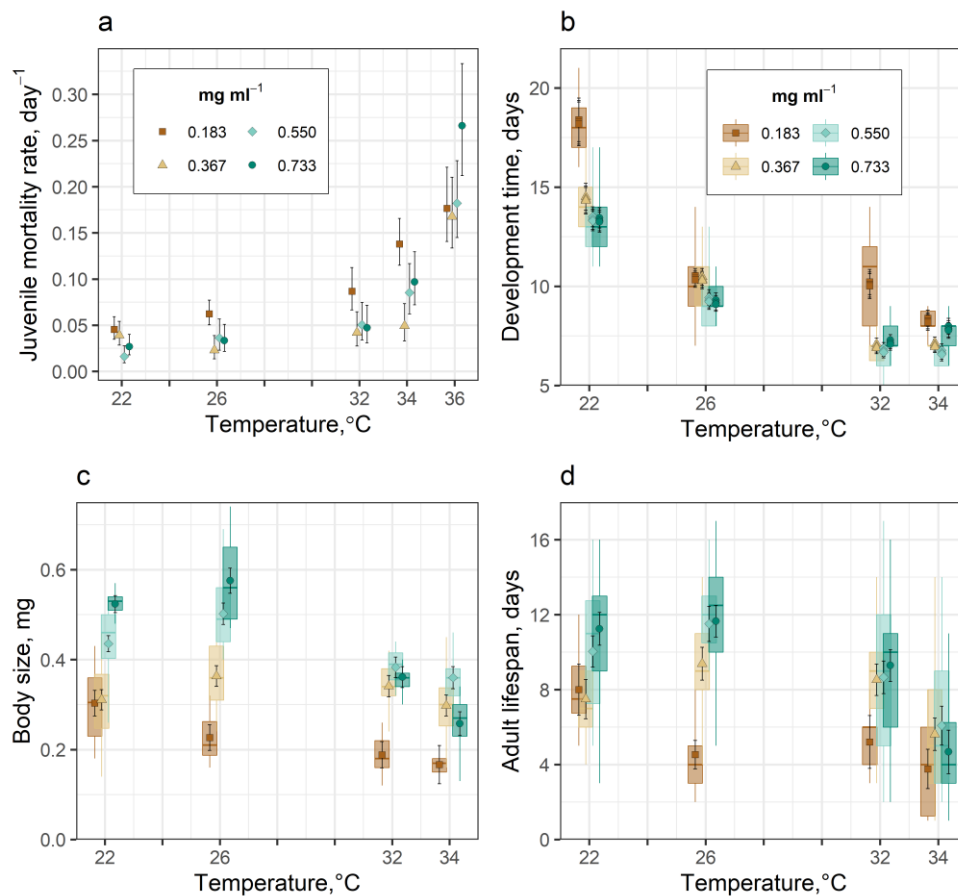
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338 **Fig. 1 | The effect of intraspecific larval competition on fitness traits in *Ae. aegypti* with 95%**

339 **credible intervals (CIs). a,** Intensified competition at low resource concentration (0.183 mg ml<sup>-1</sup>)

340 increased the negative effect of increased temperature on juvenile mortality. The CIs for 0.183 mg ml<sup>-1</sup>

341 and 0.367 mg ml<sup>-1</sup> do not overlap, indicating a significant effect of competition on this trait's

342 thermal response between these treatments, as temperatures increased from 22 to 34°C. **b,**

343 Development time decreased with temperature at all resource levels, but, at most temperatures, it was

344 significantly extended by intensified competition at low resource concentration. **c,** As temperatures

345 increased from 22°C, intensified competition at 0.183 mg ml<sup>-1</sup> significantly reduced size at

346 emergence. **d,** As temperatures increased from 22 to 32°C, intensified competition at 0.183 mg ml<sup>-1</sup>

347 significantly reduced adult lifespan. Symbols denote the regression estimated means with 95% CIs

348 calculated from the standard errors (Table S2) for the resource levels at each temperature. The

349 resulting ANOVAs of the regressions for each trait are presented in Table S1. Boxplot horizontal lines

350 represent medians. Lower and upper hinges are the 25th and 75th percentiles. Upper whiskers extend

351 from the hinge to the largest value no further than 1.5 × inter-quartile range (IQR) from the hinge. The

352 lower whisker extends from the hinge to the smallest value at most 1.5 × IQR of the hinge.

353

354 **Population fitness**

355 At all resource levels,  $r_m$  responded unimodally to temperature. However, intensified  
356 competition at low resource concentration (0.183 mg ml<sup>-1</sup>) significantly depressed  $r_m$  across  
357 the entire temperature range (Fig. 2a) and caused it to peak at a significantly lower  
358 temperature than at our intermediate resource concentration (0.367 mg ml<sup>-1</sup>; Fig. 2b, Table 1).  
359 Intensified competition at low resource concentration also significantly narrowed the thermal  
360 niche width for  $r_m$  compared to when competition was relatively relaxed due to higher  
361 resource levels (Fig. 2a, Table 1).

362

363 At low resource concentration,  $r_m$  was negative until temperatures increased to 23.3°C (95%  
364 credible interval [CI]: 22.4–24.9°C, Fig. 2, Table 1). At this resource level,  $r_m$  reached a peak  
365 of 0.05 at its  $T_{opt}$  (26.6°C); it then declined to negative growth at 30.1°C ([CI]: 28.6–31.2°C).  
366 The breadth of  $r_m$ 's thermal niche width at low resource concentration was 6.8°C ([CI]:  
367 3.7–8.8°C).

368

369 In contrast, at the intermediate food level (0.367 mg ml<sup>-1</sup>),  $r_m$  became positive as temperatures  
370 increased to 18.8°C ([CI]: 17.1–20.2°C), and it was maximal at 33.0°C (0.24, Fig. 2, Table  
371 1). At 0.367 mg ml<sup>-1</sup>,  $r_m$  declined to negative growth at 35.4°C ([CI]: 35.4–35.5°C). The  
372 thermal niche width for  $r_m$  at this resource level was 16.6°C ([CI]: 15.3–18.4°C).  
373 Overlapping CIs indicate that the predicted differences between the intermediate resource  
374 level and the higher resource levels (0.550 and 0.733 mg ml<sup>-1</sup>) in  $r_m$  at  $T_{opt}$ ,  $T_{opt}$ , and the  
375 thermal niche width were non-significant (Fig 2, Table 1).

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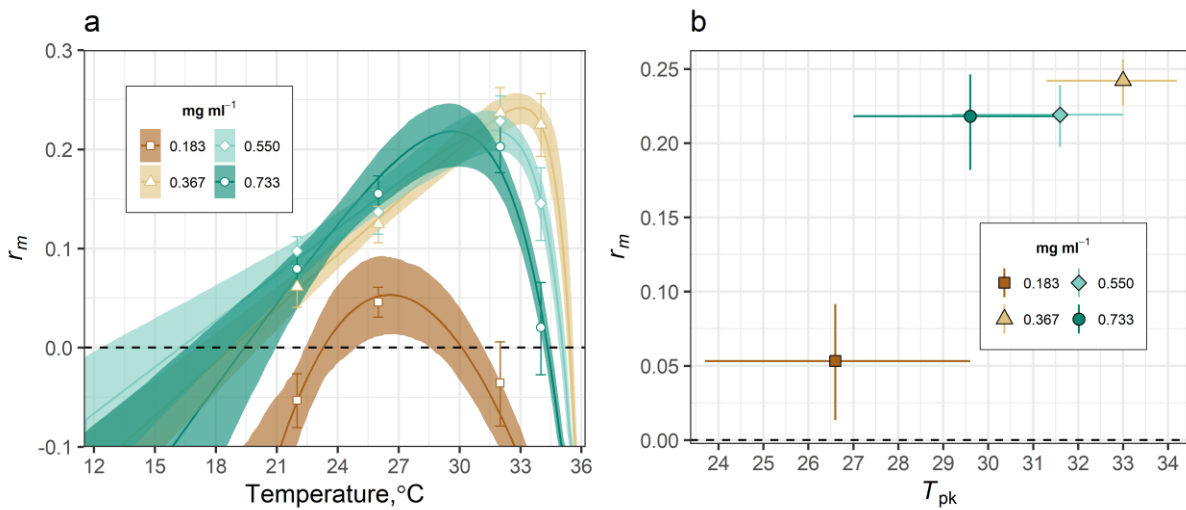
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390 **Fig 2 | The effect of larval competition on the thermal response of population-level *Ae. aegypti***

391 **fitness ( $r_m$ ) with bootstrapped 95% prediction bounds. a,** Intensified competition at low resource

392 concentration ( $0.183 \text{ mg ml}^{-1}$ ) significantly depressed  $r_m$  across the entire temperature range and

393 narrowed its thermal niche width compared to the higher resource levels (non-overlapping 95%

394 prediction bounds, Table 1). Symbols denote matrix projection estimates with 95% confidence

395 intervals. **b,** Competition at  $0.183 \text{ mg ml}^{-1}$  significantly (non-overlapping 95% confidence intervals)

396 lowered maximal  $r_m$  and caused it to peak at a significantly lower temperature than at the intermediate

397 resource level ( $0.367 \text{ mg ml}^{-1}$ ). Predicted  $r_m T_{opt}$  at  $0.183 \text{ mg ml}^{-1}$  indicates that competition could

398 decrease  $r_m T_{opt}$  by  $6.4^\circ\text{C}$ , when compared to the intermediate resource level ( $0.367 \text{ mg ml}^{-1}$ , Table 1).

399

Resource concentration ( $\text{mg ml}^{-1}$ )	$r_m$ at $T_{opt}$ ( $\pm 95\%$ CI)	$T_{opt}$ ( $^\circ\text{C}$ ) (95% CI)	$T_{min}$ ( $^\circ\text{C}$ ) (95% CI)	$T_{max}$ ( $^\circ\text{C}$ ) (95% CI)	Thermal niche width ( $^\circ\text{C}$ ) (95% CI)
0.183	$0.05 \pm 0.04$	26.6 (23.7 - 29.6)	23.3 (22.4 - 24.9)	30.1 (28.6 - 31.2)	6.8 (3.7 - 8.8)
0.367	$0.24 \pm 0.02$	33.0 (31.3 - 34.2)	18.8 (17.1 - 20.2)	35.4 (35.4 - 35.7)	16.6 (15.2 - 18.6)
0.550	$0.22 \pm 0.02$	31.6 (29.2 - 33)	16.2 (12.4 - 18.6)	35.1 (35.0 - 35.3)	18.8 (16.4 - 22.9)
0.733	$0.22 \pm 0.03$	29.6 (27.0 - 31.5)	19.4 (16.6 - 21.0)	34.3 (34.2 - 34.6)	14.9 (13.2 - 18)

400

401 **Table 1 | Parameter estimates of the Thermal Performance Curves of  $r_m$  by resource**

402 **concentration.** Non-overlapping 95% Confidence Intervals (CIs) indicate that intensified competition

403 at low resource concentration ( $0.183 \text{ mg ml}^{-1}$ ) significantly depressed maximal growth ( $r_m$  at  $T_{opt}$ )

404 compared to the higher resource levels. Intensified competition at  $0.183 \text{ mg ml}^{-1}$  caused a significant

405 decrease in  $r_m T_{opt}$  compared to  $r_m T_{opt}$  at  $0.367 \text{ mg ml}^{-1}$ . Competition at  $0.183 \text{ mg ml}^{-1}$  also caused a

406 significantly narrower thermal niche width compared to the higher resource levels.

407 **Elasticity analysis**

408

409 Juvenile survival was the most important contributor to  $r_m$  (Fig. S4). For example, at low  
410 resource concentration at 26°C, a 0.5 proportional increase in juvenile survival would  
411 increase rate of increase from 0.046 to 0.063 (Fig. S4d). In contrast, for the same treatment, a  
412 proportional increase of the same magnitude for adult survival would increase  $r_m$  from 0.046  
413 to 0.050 (Fig. S4e), and fecundity would increase  $r_m$  from 0.046 to 0.048 (Fig. S4f). This  
414 underlines how the temperature-dependence of  $r_m$  derives mainly from how intensified  
415 competition at low resource concentration impacts juvenile mortality and development,  
416 which determine the number of reproducing individuals and the timing of reproduction,  
417 respectively. Fecundity and adult survival, on the other hand, have relatively negligible  
418 effects on  $r_m$ , which suggests that the carry over effect of reduced size at maturity on  $r_m$  is  
419 relatively weak.

420

421 **DISCUSSION**

422

423 Using a trait-based approach, laboratory experiments and mathematical modelling, we have  
424 shown that intraspecific larval competition in depleting resource environments can  
425 significantly change the shape of the  $r_m$  thermal performance curve. Differences between the  
426 thermal response of underlying  $r_m$  traits at low resource concentration versus the higher  
427 resource levels resulted in a marked divergence of the temperature-dependence of  $r_m$  between  
428 the four resource levels (Fig. 2).

429

430 Intensified competition at low resource concentration (0.183 mg ml<sup>-1</sup>) significantly depressed  
431  $r_m$  across the entire temperature range, caused a significant decrease in  $r_m T_{opt}$  compared to  
432 the intermediate food level (0.367 mg ml<sup>-1</sup>), and significantly narrowed the thermal niche  
433 width compared to the higher resource levels (Fig. 2, Table 1). This indicates that population  
434 fitness becomes increasingly, and non-linearly constrained by larval competition as  
435 temperatures and resource depletion rates increase. This suggests that intraspecific  
436 competition is strongest at temperatures that are optimal for population growth.

437

438 The elasticity analysis shows that the primary mechanism underlying the divergent  
439 temperature-dependence of  $r_m$  across resource levels is decreased juvenile survival at low  
440 resource concentration (Fig. S2). This effect derived from how intensified larval competition

441 at low resource concentration increased juvenile mortality rate and development time, which  
442 decreased population-level reproductive output and delayed the onset of reproduction.

443

444 Fecundity and adult lifespan had comparatively negligible effects on  $r_m$ , which suggests that  
445 the carry over effect of reduced size at maturity on  $r_m$  is relatively weak. For example, at low  
446 resource concentration, body size, adult lifespan and lifetime fecundity were greater at 22°C  
447 than at 26°C, yet fitness at 26°C was predicted to be over 200% higher (Figs. 1, S2a and b,  
448 S3a and b, Table S2). This is because the increase in temperature from 22 to 26°C decreased  
449 development time at 26°C (Fig. 1a, b). This increased juvenile survival at 26°C, leading to  
450 increased fitness as greater numbers of individuals could contribute to population growth  
451 through reproductive output. This is consistent with general studies of ectotherm fitness  
452 (Amarasekare and Coutinho 2013, Huey and Berrigan 2001), including mosquitoes (Huxley  
453 et al. 2020). This key finding implies that predictions about the effect of climatic warming on  
454 ectotherms, including disease vectors and transmission (which are generally from populations  
455 under high- or optima resource supply) likely underestimate the effect of temperature on  
456 development time and juvenile survival, and overestimate effects of temperature on lifespan  
457 and fecundity.

458

459 The trait-level responses of our higher resource concentration treatments correspond with  
460 studies that have synthesised laboratory-derived trait responses to temperature to estimate  
461 vector fitness and disease transmission. In these studies, the juvenile development rate of  
462 most mosquito vectors is expected to increase from  $\sim 0.07 \text{ day}^{-1}$  at 22°C to  $\sim 0.14 \text{ day}^{-1}$  at  
463 32°C (Mordecai et al. 2019). In the present study, development rate (1/development time;  
464 Fig. 1b, Table S2) increased by a similar margin. For example, development rate at 0.550 mg  
465  $\text{ml}^{-1}$  increased from  $0.07 \text{ day}^{-1}$  at 22°C to  $0.15 \text{ day}^{-1}$  at 32°C. In contrast, at low resource  
466 concentration, we found juvenile development rate increased from  $0.05 \text{ day}^{-1}$  at 22°C to  $0.12$   
467  $\text{day}^{-1}$  at 32°C (Fig. 1b, Table S2). Although these differences in juvenile development rate  
468 may appear small, we show that they can have dramatic effects on the temperature-  
469 dependency of  $r_m$  when combined with the negative impact of intensified competition on  
470 juvenile survival (Fig. 1a, Table S2).

471

472 Juvenile mortality rate increased significantly with temperature and it was consistently higher  
473 at low resource concentration (Fig. 1a) than at higher resource levels. This is probably  
474 because somatic maintenance costs increase with metabolic rate (Kooijman 2000), which

475 cannot be met below a threshold resource supply level. Intensified competition at low  
476 resource concentration is likely to have prevented some individuals from meeting this  
477 increased energy demand, which explains why juvenile mortality rates were highest at 32 and  
478 34°C at low-resource concentration (except at 36°C where no individuals survived at all)  
479 where the energy supply-demand deficit was expected to be the largest.

480

481 Because we show that the  $r_m$  TPC significantly changes shape in response to larval  
482 competition,  $r_m$  TPCs that are measured without competition effects (e.g. Deutsch et al. 2008,  
483 Vasseur et al. 2014) maybe less useful for predicting the consequences of future climatic  
484 warming on holometabolous insects than currently thought. More specifically, the profound  
485 effect of intensified larval competition at low resource concentration on the temperature-  
486 dependence of  $r_m$  is likely to substantially alter predictions about the temperature-dependence  
487 of vector-borne disease transmission. This underlines the importance of considering the  
488 effects of larval competition in depleting resource environments, in particular, when  
489 predicting the responses of ectothermic disease vectors to climatic warming.

490

491 In nature, populations likely experience intermittent resource depletion (Ostfeld and Keesing  
492 2000, Wright 1983, Yee and Juliano 2012), but very few studies have examined how  
493 intraspecific competition in such environments can affect the  $r_m$  TPC. For example, in  
494 mosquitoes, field and laboratory studies have examined how resource availability and larval  
495 density interact to affect fitness (Gilpin and McClelland 1979, Gimnig et al. 2002, Heath et  
496 al. 2020, Walsh et al. 2013), while others have investigated the effect of resource availability  
497 and larval density separately (Romeo Aznar et al. 2018, Shapiro et al. 2016). However, none  
498 of these studies have considered environmental temperature. On the other hand, studies that  
499 have considered temperature have not considered resource depletion (Huxley et al. 2020), or  
500 how the effects of temperature and resource availability together affect fitness through traits  
501 (Couret et al. 2014, Padmanabha et al. 2012). As such, our study addresses an important gap  
502 by showing that the interactive effects of temperature and resource depletion should also be  
503 considered in studies of ectotherm population dynamics.

504

505 Since resource depletion can mediate the regulatory effects of larval competition on the  
506 temperature-dependence of  $r_m$ , it is also important to determine the temperature-dependence  
507 of resource availability itself (Huey and Kingsolver 2019). For example, recent studies show  
508 that mosquitoes can be reared exclusively on cultures of *Asaia* bacteria (Chouaia et al. 2012,



509 Souza et al. 2019), that microbiota at mosquito breeding sites is spatially heterogenous (Hery  
510 et al. 2021), and that larval exposure to microbial variability can affect adult mosquito life  
511 history traits (Dickson et al. 2017). Microbial growth rates also increase with temperature to  
512 some optimum, so the concentration of food in the environment may also increase (Craine et  
513 al. 2010, Cross et al. 2015, Smith et al. 2019).

514

515 If resource availability increases with climatic warming, its regulatory effect on population  
516 growth could be relaxed through increased juvenile development and adult recruitment rates.  
517 Indeed, increased food availability with warming could contribute to the expansion of disease  
518 vectors and other invasive insect species into regions that were previously prohibitive by  
519 broadening  $r_m$ 's thermal niche width (Amarasekare and Simon 2020). On the other hand,  
520 evidence from our high resource concentration treatments (e.g. a lower  $T_{opt}$  at 0.733 than at  
521  $0.367 \text{ mg ml}^{-1}$ ) may suggest that warming could have a negative impact on population growth  
522 by causing resources to be overabundant, which could lead to eutrophication and hypoxia in  
523 aquatic environments (Liikanen et al. 2002).

524

525 Alternatively, if climate change reduces the quantity of food available to ectotherms, their  
526 spatiotemporal ranges could contract (Huey and Kingsolver 2019, Lister and Garcia 2018).  
527 This is because, as we have shown here, intensified competition at low resource  
528 concentration prevents  $r_m$  from being positive at lower temperatures, decreases its optimal  
529 temperature, and forces it to become negative at lower temperatures. In this way, the effects  
530 of rising temperatures on insect populations could be especially pernicious, if food  
531 availability is simultaneously reduced (Huey and Kingsolver 2019).

532

533 We did not measure the effect of temperature and resource concentration on fecundity  
534 directly, but used the size-scaling of this trait to estimate this effect. This is because most of  
535 the effect of intensified larval competition at low resource concentration, is expected to affect  
536 adult mosquitoes indirectly by reducing size at emergence and lifespan (Briegel 1990,  
537 Steinwascher 1982). Despite these assumptions, we show that substantial under- or  
538 overestimation of fecundity by our size-scaling predictions and the use of starved adult  
539 lifespans, would not affect our main conclusions. This is because predicted fitness was  
540 relatively insensitive to these traits (Figs. S2, S3).

541

542 Although the increased negative carry over effects of temperature and competition on adult  
543 traits may have had a relatively weak impact on fitness compared to juvenile traits,  
544 temperature  $\times$  resource concentration interactions may have important effects on other  
545 components of vector-borne disease transmission (Parham et al. 2015). For example, smaller  
546 individuals may bite more frequently, which can increase transmission probability (Scott et  
547 al. 2000). Also, larval nutrition (Shapiro et al. 2016) and temperature (Shapiro et al. 2017)  
548 can independently influence within-vector parasite development, but future studies could  
549 consider how the combined effects of temperature and resource availability affect this, and  
550 other important transmission traits.

551

552 Studies on the effects of climatic change on resource availability have recently shown how  
553 warming could lead to the collapse of consumer populations due to increased phenological  
554 asynchrony with resource populations (Simmonds et al. 2020). However, while climatic  
555 warming is identified as the most important threat to insect populations (Halsch et al. 2021),  
556 very few field studies have considered how it may indirectly affect food availability.  
557 Therefore, our study underlines the importance developing accurate methods to measure how  
558 effective temperature-dependent fitness in the field is affected by the temperature-dependence  
559 of resource availability. Our study also underlines the importance of the effects of larval  
560 competition on the temperature-dependence of population-level fitness of an important  
561 disease vector. In doing so, our findings suggest that current projections of how climatic  
562 warming affects vector-borne disease transmission may prove inaccurate because they  
563 generally fail to consider how resource availability can mediate the effects of larval  
564 competition on the temperature-dependence of population fitness, and therefore ultimately,  
565 the effects of climatic variation on abundance dynamics in the field.

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