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# Functional redundancy in natural pico-phytoplankton communities depends on temperature and biogeography

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#### 17 Abstract:

Biodiversity affects ecosystem function, but how this relationship will pan out in a changing 18 world is still a major question in ecology. It remains especially understudied for pico-19 20 phytoplankton communities, which contribute to carbon cycles and aquatic food webs year-21 round. Observational studies show a link between phytoplankton community diversity and 22 ecosystem stability, but there is only scarce causal or empirical evidence. Here, we sampled 23 phytoplankton communities from two biogeographically distinct (but close enough to not be 24 confounded by differences in day length and precipitation) regions in the Southern Baltic 25 Sea, and carried out a series of dilution/regrowth experiments across three assay 26 temperatures. This allowed us to investigate the effects of loss of rare taxa and establish 27 causal links in natural communities between species richness and several ecologically 28 relevant traits (e.g. size, biomass production, and oxygen production), depending on sampling 29 location and assay temperature. We found that the samples' bio-geographical origin 30 determined whether and how functional redundancy changed as a function of temperature for all traits under investigation. Samples obtained from the slightly warmer and more thermally 31 32 variable regions showed overall high functional redundancy. Samples from the slightly 33 cooler, less variable, stations showed little functional redundancy, i.e. function decreased the 34 more species were lost from the community. The differences between regions were more pronounced at elevated assay temperatures. Our results imply that the importance of rare 35 36 species and the amount of species required to maintain ecosystem function even under short-37 term warming (e.g. during heat waves) may differ drastically even within geographically 38 closely related regions of the same ecosystem. 39 40 Key words: Functional redundancy, global warming, pico-phytoplankton communities, Baltic

41 Sea

#### 43 Introduction

44 When ecosystems lose species, the function and services of the ecosystem may decline. The 45 insurance hypothesis states that ecosystem stability increases with biodiversity [1]. Evidence 46 has long been emerging for this to be true across taxa and biomes, spanning heterotrophic and 47 autotrophic microbial communities to plants to metazoans, i.e. ecosystems with higher 48 biodiversity tend to be more productive on average regardless of the type of organism under 49 investigation [2-6]. In a warming world, variation in temperature may shape the biodiversity-50 ecosystem function relationship on short and long-term time scales [7,8]. A recent study has 51 experimentally examined the synergistic effects of warming and biodiversity loss on function 52 in bacterial assemblages [9], but no comparable data exist for natural phytoplankton 53 communities. As a consequence, studies on phytoplankton biodiversity and ecosystem function 54 remain largely observational [2] and usually do not consider loss of biodiversity in interaction 55 with aspects of climate change (but see [10]) or evolutionary history (but see [11] for 56 population genetics). Pico-phytoplankton contribute about 20% of phytoplankton primary 57 production [12], and unlike larger phytoplankton, they do not form blooms, but contribute to the photosynthesising foundation of aquatic ecosystems all year [13,14]. They show rapid 58 59 physiological [15,16] and evolutionary [17,18] responses to changing environments when 60 studied as single-strain cultures. There is isolated evidence that the speed and mechanism of 61 evolutionary responses in a changing environment may differ between phytoplankton strains 62 evolving in isolation and strains evolving in communities, mixed cultures, or multi-genotype 63 biofilms [19-23]. Comparable experiments have not been carried out for pico-phytoplankton 64 (but see [24]). As a consequence, we need to conduct manipulative experiments to establish a 65 better understanding of how the links between pico-phytoplankton community function and pico-phytoplankton community diversity change with temperature on short- and long-term 66 67 timescales. Since natural phytoplankton strains are notoriously difficult to grow in isolation

under laboratory conditions, assembling artificial communities from natural phytoplankton components can be an arduous task. Organisms from culture collections which have already been shown to grow well in isolation are easier to assemble into communities (e.g.[10]) but may not reflect well the complex environments that the lineages were originally sampled from. Therefore, when working with natural assemblages, serial dilution of natural samples can provide a useful tool to reduce biodiversity to such an effect that rare species are lost, and common species remain [25,26].

75

76 Here, we investigated the combined effects of transient warming and biogeographic history on 77 the diversity-function relationship in natural pico-phytoplankton communities. To do so, we 78 obtained pico-phytoplankton community samples during two cruises of RV ALKOR (AL505 79 and AL507) on the Southern Baltic Sea in 2018 (see Figure 1 and Table S1). Of the two basins sampled, the Kiel Bight is characterised by, on average 2 °C higher temperature than the 80 81 Bornholm Basin and unpredictable fluctuations in temperature on the timescale of days to 82 months (Santelia et al. in prep - SI document 2). In the Bornholm Basin, fluctuations in 83 temperature follow a highly predictable pattern governed by seasonality (Santelia et al. in prep 84 - SI document 2). The basins are connected through currents [27], and the regions are 85 geographically close to each other (see Figure 1) so that we can rule out confounding effects 86 introduced by e.g. differences in precipitation, day length and light intensity [28]. By assaying 87 the effect of species loss on communities across three temperatures within the range of 88 Southern Baltic Sea spring and summer temperatures [27,28], we can test the contributions of 89 long-term (comparison of basins) and short-term (assay temperatures) changes in temperature 90 on key community functions.

91

#### 93 Methods

#### 94 This is a methods summary. The full methods are available in the supporting

#### 95 information.

96 We obtained pico-phytoplankton community samples during two RV ALKOR cruises (AL505 and AL507 respectively) in 2018 (see Figure 1 and Table S1 for sampling dates and 97 98 locations) from 5m. Samples were size fractioned to obtain the pico-phytoplankton 99 community. To rule out effects of parameters other than temperature and diversity during the 100 experiment, all samples were grown in f/2 media [29] at the salinity of the sampling location. 101 Community samples grew in semi-continuous batch culture in a common garden at 18°C and 100 µmol quanta  $m^{-2} s^{-1}$  (12:12 light/dark cycle) until used for the experiment. We counted 102 103 cell numbers in all samples the flow cytometre, and the flow cytometric fingerprints also 104 allow for an estimate of phenotypic diversity or trait-level diversity [30] (based on photo-105 pigment composition and size [31, ]Figure S2 for details). Samples were then diluted in 10-106 fold dilution steps at the appropriate salinity, down to the lowest point of dilution (in theory 107 containing no more than 1 species or pico-phyoplankton per mL). Six technical replicates of each sample were left to regrow to  $10^6$  cells mL<sup>-1</sup> at the assay temperatures of 15°C, 18°C, 108 109 and 22°C. Then, we re-diluted all samples to the same density and tracked a full growth 110 curve. Cell size was obtained from the flow cytometre to calculate an estimate of pg carbon 111 per mL after [32].

112

113 Net photosynthesis rates were obtained when samples were in exponential phase, on PreSens
114 ® SDR Sensor Dish optodes at 10<sup>5</sup> cells mL<sup>-1</sup> in the measurement vials and measured oxygen
115 production for 15 minutes in the light, and respiration for 15 minutes in the dark.

117 We obtained two measures of biodiversity in our samples. One, following CTAB DNA

118 extractions [33], samples were sent for DNA- meta-barcoding at biome-id, resulting in a

119 MOTU (meta-barcoding operational taxonomic units) estimate for those samples. Two,

120 phenotypic diversity [30,31] was assessed using the parameters returned by the flow

121 cytometre.

122

#### 123 Statistical analysis

All data were analysed in the R programming environment (version 3.5.3.). To analyse the shape of the growth curves, non-linear curve fitting of a baranyi growth model [34] was carried out using the 'nlsLM' function in the R package, 'minpack.lm'(version 1.2-1). For multi-model selection, we computed small sample-size corrected AIC scores (AICc) and then compared the models by calculating delta AICc values and AICc weights using the "MuMIn" package (version 1.42-1).

130

#### 131 **Results:**

132 <u>Proof of concept</u>

133 MOTU (meta-barcoding operational taxonomic units) analyses revealed that biodiversity (as 134 species richness) was slightly higher for samples from the Kiel Bight than from the 135 Bornholm Basin sampling region (see Figure 1C and Figure S1), and roughly in line with 136 previous studies on Southern Baltic Sea phytoplankton communities [35]. Further, while diversity was not reduced to a single species in the most dilute samples, MOTU data 137 confirmed that i) the most dilute samples were on average  $10^5$  times less species-rich than the 138 139 least dilute samples, ii) MOTU diversity scaled with phenotypic diversity (Figure S1) and finally, that diversity once established by dilution, did not change significantly throughout the 140 141 time of the growth curve (Figure S2)

142

143	Assay temperature and biogeography explain differences in functional redundancy
144	We consider a result to be in line with high functional redundancy, when a function - such as
145	biomass production or net photosynthesis - does not change significantly across levels of
146	species richness. This results in slopes across dilution steps (i.e. species richness) that are
147	equal or close to zero. When a function declines with species richness, we assume low
148	functional redundancy. In these cases, the slope of trait value as a function of species richness
149	is <i>positive</i> . We report the steepness of these slopes across three assay temperatures in Figure
150	2 (Table S2 for slopes, and Tables S3 to S5 for model selection and output).
151	
152	Across dilutions, temperature had a significant impact on biomass (Figure S3, likelihood ratio
153	test comparing models with and without 'temperature': $\Delta$ d.f. = 4, $\chi 2$ = 382.71, P < 0.0001,
154	Table S3). When communities lost rare species, biomass production did not change
155	significantly in the samples from the Kiel Bight (Figure 2A). Biomass production in samples
156	from the Bornholm Bight samples rapidly decreased when rare species were lost, resulting in
157	positive slopes (Figure S3, Figure 2A, likelihood ratio test comparing models with and
158	without 'region': $\Delta$ d.f. =2, $\chi$ 2 = 185.23, <i>P</i> < 0.0001, Table S3).
159	
160	Changes in biomass were driven by a change in cell number and a change in cell volume
161	across dilution steps, with a trend for smaller cells at higher species richness (Figure 2B, C,
162	Figure S4, S5, Table S2, S4). Cell volume decreased with temperature (likelihood ratio test

Figure S4, S5, Table S2, S4). Cell volume decreased with temperature (likelihood ratio test comparing models with and without 'temperature':  $\Delta d.f. = 2, \chi 2 = 382.70, P < 0.0001$ , Figure 2C, Table S4). While cells in samples from the cooler, more stable, Bornholm Basin were on average less reactive to temperature in terms of cell size, the effect of losing rare species was stronger here than in samples from the Kiel Bight and cell size rapidly increased as species 167 richness decreased (likelihood ratio test comparing models with and without 'region:  $\Delta d.f. =$ 168 2,  $\chi 2 = 223.68$ , P < 0.0001, Figure 2 C, Table S4)

169

170 Net Photosynthesis rates (NP) per cell followed a similar trend to biomass (Figure 2D, Figure 171 S6, Table S5 for model selection and output): assay temperature altered community NP 172 depending on dilution and sampling origin. Samples from the Kiel region were overall more 173 photosynthetically active than those from the Bornholm region (likelihood ratio test comparing models with and without 'region':  $\Delta d.f. = 3$ ,  $\chi 2 = 21.63$ , P < 0.0001, Figure 2D, 174 175 Figure S6, Table S5), especially at the warmer temperatures (likelihood ratio test comparing models with and without 'temperature':  $\Delta d.f. = 1$ ,  $\chi 2 = 85.47$ , P < 0.0001, Figure 2D Figure 176 177 S6, Table S5). NP in samples from the Kiel region showed no clear trend with dilution. In 178 samples from the Bornholm region, on the other hand, responses of NP to temperature were 179 overall less pronounced, but were strongly influenced by species richness, with the full 180 community samples photosynthesising nearly 1.5 times (per cell) as much as the samples 181 with the lowest species richness.

182

#### 183 **Discussion**

Our investigations of the link between species richness and traits relevant for ecosystem 184 services showed that the steepness of the biodiversity-function slope hinged on the 185 186 communities' evolutionary histories as well as on short-term (ca 20 generations) changes in 187 temperature and fit well within the theoretical and observational general framework of biodiversity/ecosystem function studies [2-5]. The effect of temperature modulating the 188 189 strength of the diversity/function relationship was the most pronounced for samples originating 190 from a cooler, less variable region, in line with theory predicting that regions that are more 191 variable should contain a greater number of taxa with more variable tolerance thresholds [36].

193 Overall, our findings are also in accordance with the literature on experimentally assembled 194 communities in microcosms, where the researchers found that a decline in productivity was 195 more pronounced at rising temperatures in phytoplankton and bacteria ([10] and [9] 196 respectively). We make the case that this pattern of declining productivity due to species loss 197 being enhanced at higher temperatures may be conserved across environments as diverse as 198 laboratory conditions and natural community samples. This may also mean that results 199 obtained from assembling artificial communities reflect results found in natural assemblages 200 well in direction, and possibly in magnitudes of responses. We specifically find that as species 201 are lost from samples with a history of comparatively cooler average temperatures and highly 202 predictable seasonal variation, the community rapidly shows a decrease in net photosynthesis 203 rates, biomass production, and – unlike in experimental assemblages [10] - an increase in cell 204 size. Changes in the biodiversity-ecosystem function relationship were thus strongly linked to 205 size-dependent turnover in community composition, but it is unclear whether this is a cause or 206 a consequence of the other observed patterns. We can only speculate about the underlying 207 mechanisms. While it is possible that competition for resources is stronger in a mixed 208 community, and that this pressure is relieved once most species are lost from the system, our 209 samples were kept in full media, and the pattern was obvious during exponential phase already 210 (Figure S7). It is thus unlikely that competition for nutrients, light, or space, was the main 211 factor here. Neither can the pattern be explained by the larger cells' being more common in the 212 original samples (which do have slightly higher variation in cell size than very dilute samples) 213 and therefore more likely to remain present after dilution. The pattern may be driven by 214 interactions beyond competition, e.g. types of facilitation [37], or a more general trend for trait 215 variability and biodiversity being intertwined [38]. It remains to be seen to which degree the 216 patterns we observed here are also driven by interactions with their associated bacteria. The 217 latter may have different nutrient requirements that are not met in the same way by our culture 218 media, which may in turn affect how they interact with and shape the phenotypes of the 219 phytoplankton species.

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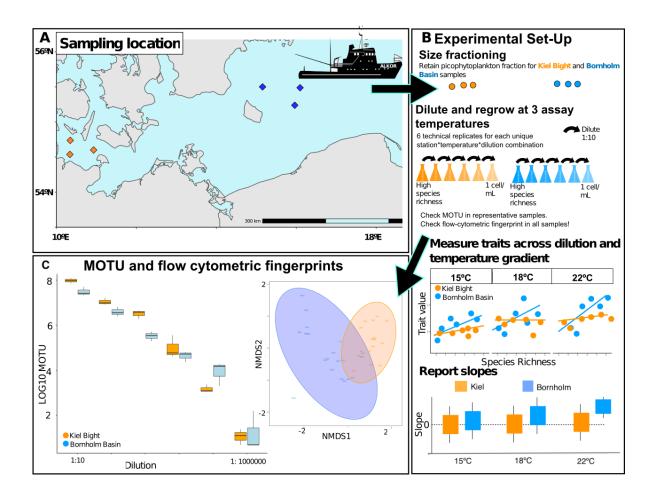
Irrespective of the reasons underlying the shift in cell size we observed here, the contribution of the very small pico-phytoplankton cells to total phytoplankton biomass in the ocean has been shown to increase with temperature [39] and changes to size distribution in scenarios where warming and loss of biodiversity interact may have unpredictable impacts on the function of phytoplankton ecosystems [40]

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While biodiversity manipulation by dilution has limitations (e.g. diversity and species identity can never be fully disentangled, dilution introduces quasi-random differences in beta diversity), these apparent disadvantages have potential important ecological implications as they allow us to specifically investigate the importance of rare taxa, rather than random species loss.

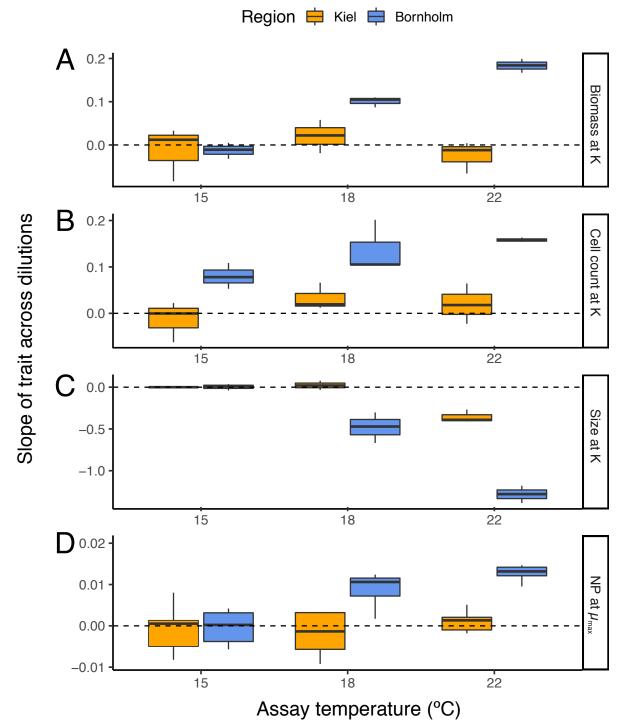
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The Baltic Sea is generally low in species due to its characteristic brackish waters. To rule out that our results are typical for regions where biodiversity (as richness) is low to begin with, it would be necessary to carry out comparable studies across systems that vary in their initial species compositions. Together, such data will further improve our understanding of the relationship between diversity and ecosystem function at the foundation of warming aquatic ecosystems.



238 239

240 Figure 1: Overview of sampling locations, experimental set-up, and assessment of species richness 241 242 via meta-barcoding and flow cytometry. A) sampling locations: We took samples in two biogeographically distinct regions of the Southern Baltic Sea: the Kiel Bight (KB, orange squares), 243 244 and the Bornholm Basin (BB, blue squares) (Table S1). B) Set up and data-processing: The picophytoplankton fraction was obtained through size fractioning on board. Upon returning to the 245 Institute for Marine Ecosystem and Fishery Science Hamburg (IMF), we established six technical 246 replicates per community for each dilution across the three assay temperatures (15°C, 18°C, and 247 22°C). We report the slopes of three traits across the dilution steps for each region\*assay 248 temperature combination. C) Proof-of-concept via meta-barcoding and flow cytometry: To test our 249 experimental design, we obtained MOTUs (Meta-barcoding Operational Taxonomic Units) and flow 250 cytometric fingerprints (based on size and photopigment composition). While there were some 251 regional differences in initial MOTU composition, those were overall not significant. (Figures S1 and 252 S2).





254 255 Figure 2: Steepness of slopes for traits across dilution gradients (see Figures S3 to S6 and 256 Table S2 for full slopes). The dashed line indicates a slope of 0, indicating high functional 257 redundancy, i.e. the trait does not change significantly as rare species are lost. A positive slope indicates low functional redundancy, i.e. function decreases as species are lost. A 258 259 negative slope also indicates that function changes with species richness, and specifically 260 shows that a trait value decreases as species richness increases. A: Biomass produced at carrying capacity is stable across dilution steps and temperatures for samples from the Kiel 261 Bight (high functional redundancy), however, there is low functional redundancy in samples 262 263 from the Bornholm Basin, especially under warming. **B:** Cell count (mL<sup>-1</sup>) at carrying capacity. Cell count followed the same pattern as overall biomass. C: Cell size at carrying 264

265 capacity. Average cell diameters were independent from dilution at 15°C for samples from 266 either region, but cells became smaller faster with temperature in the full communities as compared to low species richness samples. This effect was more pronounced for samples 267 268 from the Bornholm Basin. D: Net Photosynthesis (NP) across dilution steps during 269 exponential growth showed similar patterns as the biomass response. Orange for Kiel Bight 270 samples, blue for Bornholm Basin samples. The boxplots are displayed as is standard, with 271 the girdle band indicating the median, and the whiskers extending to the 25th and 75th 272 percentile. n=6 for each unique combination. This plot only displays significant parameters, 273 i.e. samples from different seasons have been pooled. 274 275 276 **Conflict of interest:** The authors declare no conflict of interest. 277 278 Author contributions: DZ carried out the experiments. LL and MS retrieved, prepared, and 279 maintained the phytoplankton cultures. ES conceived the experiment, supervised laboratory 280 work and handled data analysis. All authors contributed equally to writing the manuscript. 281 282 Acknowledgements: We would like to thank Margarethe Nowicki, Richard Klinger, Jens-283 Peter Hermann, and the captain and crew of RV ALKOR for support at sea (cruises AL505 284 and AL507), and Stefanie Schnell for technical assistance in the laboratory at IMF Hamburg. 285 Samuel Barton, Moritz Aehle, and Paula Franze assisted with the metabolism measurements. 286 287 References 288 289 Yachi, S. & Loreau, M. 1999 Biodiversity and ecosystem productivity in a 1. 290 fluctuating environment: the insurance hypothesis. Proceedings of the National 291 Academy of Sciences 96, 1463–1468. (doi:10.1073/pnas.96.4.1463) 292 2. Ptacnik, R., Solimini, A. G., Andersen, T., Tamminen, T., Brettum, P., Lepistö, 293 L., Willén, E. & Rekolainen, S. 2008 Diversity predicts stability and resource 294 use efficiency in natural phytoplankton communities. Proc. Natl. Acad. Sci. U.S.A. 105, 5134–5138. (doi:10.1073/pnas.0708328105) 295 García-Palacios, P., Gross, N., Gaitán, J. & Maestre, F. T. 2018 Climate 296 3. 297 mediates the biodiversity-ecosystem stability relationship globally. Proc. Natl. 298 Acad. Sci. U.S.A. 115, 8400–8405. (doi:10.1073/pnas.1800425115) Tilman, D., Reich, P. B. & Knops, J. M. H. 2006 Biodiversity and ecosystem 299 4. 300 stability in a decade-long grassland experiment. *Nature Publishing Group* 441, 629-632. 301 302 5. Vallina, S. M., Cermeno, P., Dutkiewicz, S., Ioreau, M. & Montoya, J. M. 2017 303 Phytoplankton functional diversity increases ecosystem productivity and 304 stability. *Ecological Modelling* **361**, 184–196. (doi:10.1016/j.ecolmodel.2017.06.020) 305 306 6. Awasthi, A., Singh, M., Soni, S. K., Singh, R. & Kalra, A. 2014 Biodiversity 307 acts as insurance of productivity of bacterial communities under abiotic perturbations. The ISME Journal 8, 2445-2452. 308 309 Haaland, T. R. & Botero, C. A. 2019 Alternative responses to rare selection 7.

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