

# A cascade of destabilizations: combining *Wolbachia* and Allee effects to eradicate insect pests

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## Summary

1. The management of insect pests has long been dominated by the use of chemical insecticides, with the aim of instantaneously killing enough individuals to limit their damage. To minimize unwanted consequences, environmen-

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tally friendly approaches propose biological controls that take advantage of intrinsic demographic processes to eliminate pest populations.

2. We address the feasibility of a novel pest management strategy based on the release of insects infected with *Wolbachia*, which causes cytoplasmic incompatibilities in its host population, into a population with a pre-existing Allee effect. Successful invasion of *Wolbachia* leads to transient declines in population size, and this can theoretically trigger extinction if the population is brought below its Allee threshold.
3. We developed a stochastic population model that accounts for *Wolbachia*-induced cytoplasmic incompatibilities in addition to an Allee effect arising from mating failures at low population densities. Using our model, we identify conditions under which cytoplasmic incompatibilities and Allee effects successfully interact to drive insect pest populations toward extinction. Based on our results, we delineate control strategies based on introductions of *Wolbachia*-infected insects.
4. We extend this analysis to evaluate control strategies that implement successive introductions of two incompatible *Wolbachia* strains. Additionally, we consider methods that combine *Wolbachia* invasion with mating disruption tactics that enhance the pre-existing Allee effect.
5. We demonstrate that *Wolbachia*-induced cytoplasmic incompatibility and the Allee effect act independently from one another: the Allee effect does not modify the *Wolbachia*-invasion threshold, and cytoplasmic incompatibilities only have a marginal effect on the Allee threshold. However, the interaction

of these two processes can drive even large populations to extinction. The success of this method is amplified by the introduction of multiple *Wolbachia* cytotypes as well as the addition of mating disruption.

6. Our study provides novel and translational ideas for the use of cytoplasmic incompatibility and the Allee effect to eradicate insect pests. More generally, it points to the importance of transient dynamics, and the relevance of manipulating a cascade of destabilizations for pest management.

## Keywords

biological control; cytoplasmic incompatibility; eradication; extinction; mating disruption; transient dynamics.

## Introduction

1 Although most insect species provide crucial ecosystem services (Losey & Vaughan  
2 (2006)), a minority of taxa that we consider pests ( $\sim 1\%$ ) have an overwhelming  
3 influence on the development of population management in theory and in practice.  
4 Among the various environmentally friendly approaches that have been envisaged  
5 to control invasive species, we focus on a research avenue that proposes the ex-  
6 ploitation of Allee effects, i.e., the decrease in survival or reproduction at small  
7 population sizes and the consequent reduction in population growth (Liebhold &  
8 Bascompte (2003); Liebhold & Tobin (2008)). The central ideas surrounding these  
9 methods are twofold: management tactics could be combined in order to (1) reduce  
10 a population size down below the Allee threshold – the population size at which

11 the *per capita* growth rate decreases (a “weak” Allee effect) or becomes negative (a  
12 “strong” Allee effect) – which, in turn, increases the probability of stochastic extinc-  
13 tion, and/or (2) amplify the mechanisms underpinning a pre-existing Allee effect  
14 to increase the Allee threshold itself (Liebhold *et al.* (2016); Tobin *et al.* (2011);  
15 Suckling *et al.* (2012)). Capitalizing on Allee effects to manage undesirable species  
16 is particularly advantageous because it drives populations into extinction vortexes  
17 without needing to eliminate every last individual.

18 The idea of using Allee effects to eradicate insect pests and the subsequent de-  
19 velopment of theoretical models originate from population management of various  
20 insect species including the Oriental fruit fly, Indian meal moth, almond moths, and  
21 arguably most importantly the Gypsy moth *Lymantria dispar* (Beroza & Knipling  
22 (1972); Knipling (1970); Steiner *et al.* (1970); Sower & Whitmer (1977)). Control  
23 methods centered on the usage of biological controls as alternatives to pesticides  
24 have long been recognized as desirable (e.g. Knipling (1955); Baumhover (1955)),  
25 and they have been successfully used to control populations with pre-existing Allee  
26 effects. The Gypsy moth, for example, is an invasive forest pest in North America  
27 that triggered a major containment program to slow the spread toward the west-  
28 ern United States (Sharov *et al.* (2002a); Liebhold *et al.* (1992)). It is one of the  
29 few insect species for which both a component (mate-finding) and demographic  
30 Allee effect has been explicitly identified (Tobin *et al.* (2013, 2007); Johnson *et al.*  
31 (2006)). Mating disruption has been a major tactic used to control newly estab-  
32 lished low-density populations along the invasion front, with evidence supporting  
33 that it is more efficient as well as economically cheaper than classic treatments with  
34 the pesticide *Bacillus thuringiensis* (Sharov *et al.* (2002a,b)). This highlights the  
35 potential benefits of identifying other invasive pest species that have pre-existing

36 Allee effects and determining whether environmentally desirable forms of control  
37 may similarly be effective.

38 Several recent theoretical developments have focused on taking advantage of  
39 Allee effects to promote pest eradication (e.g. Boukal & Berec (2009); Liebhold  
40 & Bascompte (2003); Blackwood *et al.* (2012); Yamanaka & Liebhold (2009)).  
41 These models capture the underlying population dynamics of a pest and evaluate  
42 the success of population management tactics such as culling, release of sterile  
43 males, and mating disruption to determine whether these methods can create or  
44 enhance pre-existing Allee effects (Fauvergue (2013) provides a comprehensive re-  
45 view). While there is evidence that such population management strategies will  
46 be successful for populations with pre-existing Allee effects, the range of species  
47 that might benefit from these tactics may be much greater than currently known.  
48 In a meta-analysis focused on the presence of Allee effects in natural animal pop-  
49 ulations (Kramer *et al.* (2009)), terrestrial arthropods were found associated with  
50 the largest number of studies (22) and the highest proportion (77%) exhibiting an  
51 Allee effect. Mating failure at low density appeared as the most frequent mecha-  
52 nism. Additionally, Fauvergue (2013) found evidence supporting the presence of  
53 mate-finding Allee effects in 19 out of 34 published studies that investigated the  
54 interplay of population size and mating success in insects. Indirectly, the central  
55 role of Allee effects in insect population dynamics is supported by the efficiency of  
56 eradication programs based on the disruption of reproduction. Pest management  
57 based on the reduction of mating success via mass trapping, mating disruption  
58 with sex pheromones, or the release of sterile males has indeed proved successful  
59 in several instances (Knipling (1955); Baumhover (1955); Suckling *et al.* (2014,  
60 2012); Krafur (1998)).

61 In this article, we investigate *Wolbachia*-induced cytoplasmic incompatibility  
62 (CI) as a novel method for triggering reproductive failures and consequently bring-  
63 ing a pest population below its Allee threshold. *Wolbachia* are endosymbiotic bac-  
64 teria that infect at least 20% of all insect species and up to two thirds in some esti-  
65 mations (Hilgenboecker *et al.* (2008)). It has various effects on its insect hosts, the  
66 most widespread and prominent being cytoplasmic incompatibility (Stouthamer  
67 *et al.* (1999)). Under CI, matings between an infected male and a female that  
68 is either uninfected or infected with an incompatible cytotype result in offspring  
69 mortality during embryonic development. Fitness advantages of infected females  
70 as well as maternal inheritance are key features that promote invasion of *Wolbachia*  
71 into a host population: above a threshold frequency, a given *Wolbachia* strain is  
72 expected to invade until near-fixation (Barton & Turelli (2011); Hancock *et al.*  
73 (2011); Caspari & Watson (1959); Hoffmann & Turelli (1997); Turelli & Hoffmann  
74 (1991)). As a result of the associated CI and subsequent reduction in reproductive  
75 rate, *Wolbachia* invasion via the release of infected hosts is a candidate biological  
76 control agent against arthropod pests (Bourtzis (2008)).

77 In practice, there are multiple ways to implement a management strategy cen-  
78 tered on inducing CIs via introduction of *Wolbachia*. For example, similar to the  
79 use of “Sterile Insect Technique” (SIT), males bearing a *Wolbachia* strain incom-  
80 patible with that of the target population can be released in large numbers. CIs  
81 arising from the mating of females and infected males would substantially limit  
82 the total offspring in the subsequent generation, resulting in a decrease in overall  
83 population growth rate and thereby increasing the possibility of local population  
84 extinction (Laven (1967); Zabalou *et al.* (2004); Atyame *et al.* (2015)). Incom-  
85 patible males can be obtained via transfection, even between completely different

86 species of host insects (e.g. Braig *et al.* (1994)). At the population level, the  
87 underpinnings for mass-releases of incompatible males do not depart from that of  
88 SIT, for which interactions with the Allee effect have already been thoroughly an-  
89 alyzed (Boukal & Berec (2009); Yamanaka & Liebhold (2009); Fauvergue (2013);  
90 Barclay & Mackauer (1980); Barclay (1982); Berec *et al.* (2016); Lewis & Van  
91 Den Driessche (1993)).

92 An alternative management tactic using CI relies on the inoculation of a rela-  
93 tively small number of insects of both sexes with a *Wolbachia* strain incompatible  
94 with that of the target population. This method is investigated in the theoret-  
95 cal model introduced in Dobson *et al.* (2002), which combines insect population  
96 dynamics with releases of individuals infected with *Wolbachia*. During a successful  
97 invasion of *Wolbachia*, a transient reduction in the insect population size occurs.  
98 This decline results from the temporary increase in the fraction of incompatible  
99 matings, which peaks in the midst of the invasion process. Therefore, systematic  
100 introductions of different *Wolbachia* cytotypes could be applied to artificially sus-  
101 tain an unstable coexistence of multiple incompatible infections within an insect  
102 population, allowing the population size to be reduced and maintained at low levels  
103 (Dobson *et al.* (2002)).

104 Our goal is to determine when the latter implementation of *Wolbachia* in-  
105 troductions can drive a population to extinction in the presence of Allee effects.  
106 Specifically, we derive a theoretical model built upon Dobson *et al.*'s (2002) ap-  
107 proach of CI management that additionally accounts for Allee effects as well as  
108 environmental and demographic stochasticity. We also consider mating disruption  
109 in our model as a potential complementary tactic. We use this model to address  
110 three primary questions: (1) What is the influence of Allee effects present within

111 a host population on *Wolbachia* invasion dynamics? (2) What is the influence of  
112 cytoplasmic incompatibility on the demographic Allee effect? (3) What is the in-  
113 fluence of a combination of *Wolbachia*-induced CI, Allee effects, mating disruption,  
114 and stochasticity on the probability of host extinction?

## 115 **Methods**

### 116 **Population model**

117 In this section, we first introduce a model that considers the population dynamics  
118 in the absence of individuals infected with *Wolbachia*. Our model extends the  
119 framework introduced by Dobson *et al.* (2002) by accounting for pre-existing Allee  
120 effects, the release of pheromone sources as a method of mating disruption, as well  
121 as both demographic and environmental stochasticity.

122 Similar to Dobson *et al.* (2002), we considered populations such that the dy-  
123 namics can be modeled in discrete time with non-overlapping generations. The  
124 model explicitly tracks the total population size at each time  $t$ , given by  $N_t$ , and  
125 also tracks the distribution of infected and uninfected individuals within the pop-  
126 ulation. We note that our population model can be expressed in terms of either  
127 census size or density. However, we are considering a theoretical population not pa-  
128 rameterized to any specific species; for simplicity we hereafter refer to our model  
129 in terms of size. While the population size can take on non-integer population  
130 sizes, the stochastic model forces integer population sizes. We assume that each  
131 time step can be broken into two stages: the first (at time  $t + 0.5$ ) captures repro-  
132 duction, and the second (at time  $t + 1$ ) captures density dependent survivorship

133 of offspring to adults. The total number of offspring is given by

$$N_{t+0.5} = mN_t g_1(N_t) g_2(N_t), \quad (1)$$

134 where  $m$  is the maximum *per capita* fecundity.  $g_1(N_t)$  captures a component Allee  
135 effect that results from the failure to find mates at low densities such that

$$g_1(N_t) = \frac{N_t}{N_t + \theta}. \quad (2)$$

136 Here,  $\theta$  measures the strength of the Allee effect; a convenient interpretation of  
137 this term is that  $\theta$  represents the population size at which half of the females  
138 successfully mate. Therefore, we use the quantity  $\theta/K$  as an indicator of the  
139 intensity of the Allee effect. Further,  $g_2(N_t)$  in Eqn. 1 captures the decline in  
140 fecundity resulting from techniques to control populations via mating disruption.  
141 We assume that pheromones are maintained in the population at a fixed number  
142  $P$ ; given that there are  $F_t$  females in the population, only a fraction  $F_t/(F_t + P)$   
143 males successfully find a mate (Fauvergue (2013)). We assume a 50:50 sex ratio  
144 (i.e.  $F_t = N_t/2$ ) so that

$$g_2(N_t) = \frac{N_t}{N_t + 2P} \quad (3)$$

145 Finally, we assume that survivorship of offspring to adults is density dependent so  
146 that

$$N_{t+1} = N_{t+0.5}S_N \quad (4)$$

147 where

$$S_N = \frac{S_0}{1 + (\alpha N_{t+0.5})^\gamma} \quad (5)$$

148 where the constant  $\alpha$  is related to the carrying capacity,  $\gamma$  is related to intraspecific  
149 competition, and  $S_0$  is survivorship in the absence of intraspecific competition  
150 (Slatkin & Smith (1979)).

151 This modeling framework allows us to analytically determine the equilibrium  
152 values of carrying capacity  $K$  as well as the Allee threshold. Equilibrium values  
153 in discrete time models occur when  $N_{t+1} = N_t$ , and the equilibria  $N_-^*$  and  $N_+^*$  of  
154 our model are given by the roots of the polynomial

$$m\alpha(N^*)^2 + (1 - mS_0)N^* + \theta = 0 \quad (6)$$

155 under the simplifying assumption  $\gamma = 1$ . This expression is used to analytically  
156 determine the carrying capacity and the Allee threshold. In the Supplementary  
157 Information (S3), we show this explicitly and also perform a basic stability analysis.  
158 In all simulations, we choose our parameters to guarantee the existence of a stable  
159 equilibrium corresponding to the carrying capacity (see Supplement). Importantly,  
160 this formulation captures two general properties that are integral to insect species

161 that we are considering: there is a carrying capacity and a strong Allee effect for  
162 sufficiently high  $\theta$ . We therefore conjecture that an alternative form form density  
163 dependence that captures these properties will have qualitatively similar results.

164 Based on this deterministic framework, we build in environmental and demo-  
165 graphic stochasticity. Environmental stochasticity results from variation in the  
166 population's fecundity between years Melbourne & Hastings (2008). Therefore,  
167 we adapt the methods of Schoener *et al.* (2003) and account for environmental  
168 stochasticity by rewriting Eqn. 1 as

$$N_{t+0.5} = m_e N_t g_1(N_t) g_2(N_t), \quad (7)$$

169 where the fecundity  $m_e$  is drawn at each generation from a normal distribu-  
170 tion with mean  $m$  (that is truncated so that  $m \geq 0$ ) and a standard deviation  
171  $\sigma$ . We fix the standard deviation so that it is the square root of the mean; how-  
172 ever, a sensitivity analysis of the magnitude of the standard deviation is provided  
173 in the Supplementary Information. This analysis shows that the the extinction  
174 probability increases as the magnitude of the variability increases.

175 Demographic stochasticity, on the other hand, results from variation in fecun-  
176 dity at the individual level Melbourne & Hastings (2008). Note that the total  
177 number of individuals that successfully reproduce is given by

$$B = N_t g_1(N_t) g_2(N_t). \quad (8)$$

178 We assume that each of these individuals at a given time  $t$  reproduces with  
179 fecundity  $m_e$  (as described above), and the total number of eggs produced is a  
180 Poisson random variable Melbourne & Hastings (2008). Since the sum of inde-  
181 pendent Poisson random variables is also a Poisson random variable, the total  
182 offspring of all adults at is

$$N_{t+0.5} \sim \text{Poisson}(m_e B) \quad (9)$$

183 Finally, we include stochasticity in density dependent survivorship following  
184 Melbourne & Hastings (2008). Given that  $S_N$  (as defined in Eqn. 5) is the proba-  
185 bility that offspring survive to adults, we assume that survivorship is a binomially  
186 distributed so that

$$N_{t+1} \sim \text{Binomial}(N_{t+0.5}, S_N). \quad (10)$$

## 187 Infection dynamics

188 We consider the infection dynamics of up to two different cytotypes of *Wolbachia*  
189 (referred to as cytotypes  $X$  and  $Y$ ) and denote the number of uninfected individuals  
190 as  $W$ . Note that all variables and parameters with subscripts  $X$  (or  $Y$ ) are related  
191 to cytotype  $X$  (or  $Y$ ). This model is adapted from Dobson *et al.* (2002); therefore,  
192 we use similar notation throughout.

193 In the presence of a single cytotype of *Wolbachia*, there are only unidirec-  
194 tional cytoplasmic incompatibilities (CI); in contrast, in the presence of multiple

195 cytotypes there may be bi-directional CI. We first introduce the case of a single  
196 cytotype and then extend the model to include two cytotypes. Below we describe  
197 the mathematical formulation of the infection dynamics, and Table summarizes  
198 the processes involved.

199

### 200 *One cytotype*

201 To capture the *Wolbachia* dynamics, we first determine the proportions of  
202 infected and uninfected individuals in our population at time  $t$ . For example, if  
203 there are  $W_t$  uninfected individuals and  $X_t$  infected with cytotype  $X$  then, under  
204 the assumption that there is a 50:50 sex ratio, the fraction of females infected with  
205 cytotype  $X$  at  $t + 0.5$  is given by

$$a_t = \frac{X_t}{W_t + X_t},$$

206 Similarly, we find the fraction  $i_t$  of all males that are infected (where  $i_t = a_t$ ),  
207 the fraction  $q_t$  of all males that are uninfected, and the fraction  $c_t$  of all females  
208 that are uninfected (again note that  $q_t = c_t$ ).

209 Based on the proportions of uninfected and infected individuals in the popula-  
210 tion, we can now determine the fraction of offspring that are infected. Crosses be-  
211 tween pairs with an infected female suffer a fecundity loss due to infection ( $1 - F_X$ ),  
212 where  $F_X$  is the probability of mating success in these mating couples. Vertical  
213 transmission of *Wolbachia* occurs maternally and we assume that transmission is  
214 successful with probability  $(1 - \mu_X)$ , where  $\mu_X$  is the probability of transmission  
215 failure. In the instance of *Wolbachia*-induced CIs, crosses between infected females

216 and uninfected males in addition to crosses between infected males and infected  
217 females give rise to infected offspring. The proportion of viable offspring that are  
218 infected with cytotype  $X$  after reproduction (i.e. at time  $t + 0.5$ ) is therefore given  
219 by

$$x_{t+0.5} = a_t (1 - \mu_X) F_X (i_t + q_t) \quad (11)$$

220 where a lowercase  $x$  is used to denote proportion rather than number. Second,  
221 we can identify the proportion of viable offspring that are uninfected ( $w_{t+0.5}$ ).  
222 Uninfected individuals can arise from crosses between both uninfected females and  
223 uninfected males. Further, matings between both infected females and infected  
224 males can have viable uninfected offspring. This results from failure to vertically  
225 transmit *Wolbachia* to their offspring (i.e. with probability  $\mu_X$ ). When one type  
226 of *Wolbachia* is present within a population, then only unidirectional cytoplasmic  
227 incompatibility (CI) is possible. This type of CI occurs through matings between  
228 infected males and uninfected females. Therefore, we assume that pairings between  
229 infected males and uninfected females undergo CI and a fraction  $H_X$  survives.  
230 The proportion of viable offspring that are not infected with *Wolbachia* following  
231 reproduction is given by

$$w_{t+0.5} = (\mu_X F_X a_t + c_t) \times (i_t H_X + q_t) \quad (12)$$

232 Notice that due to cytoplasmic incompatibilities and the fecundity cost due to

233 infection with *Wolbachia*, the fraction of the total population that successfully  
234 reproduces ( $x_{t+0.5} + w_{t+0.5}$ ) is less than one. Therefore, the total number of offspring  
235 as governed by Eqn. 1 can be rewritten as

$$N_{t+0.5} = m (x_{t+0.5} + w_{t+0.5}) N_t g_1(N_t) g_2(N_t). \quad (13)$$

236 In other words, the product  $g_1(N_t)g_2(N_t)$  captures the total fraction of adults  
237 at time  $t$  who successfully find a mate, and the sum  $x_{t+0.5} + w_{t+0.5}$  is the fraction of  
238 all offspring that are viable. Finally, as described in the previous section, density  
239 dependent mortality limits the total number of adults at time  $t + 1$  (Eqn. 5).

240

241 Our parameterization of the population model is based on both the parame-  
242 terization used in Dobson *et al.* (2002) and common ranges for insect populations.  
243 For example, as noted by Dobson *et al.* (2002) the basic reproductive rate in in-  
244 sect species to which sterile insect technique has been employed ranges from 1-11  
245 (Davidson (1974)) which is consistent with our parameterization (e.g. see Fig. 2).  
246 Additionally, several parameters vary for our analysis including the strength of the  
247 Allee effect,  $\theta$ , and the initial population size. However, we note that our results  
248 are intended to assess the general qualitative behavior of *Wolbachia* introductions  
249 and consequently the actual implementation of such management tactics would  
250 require a detailed analysis and parameterization specific to the target population  
251 and cytotype.

252 *Two cytotypes*

253 In addition to releasing a single cytotype of *Wolbachia*, we consider a scenario

254 in which a second cytotpe is introduced. When two cytotpes of *Wolbachia* are  
255 present within a population, bidirectional CI occurs when a male with one cytotpe  
256 mates with a female infected with an incompatible *Wolbachia* cytotpe. Similar  
257 to the previous section, we assume that a fraction  $H_X$  (or  $H_Y$ , depending on the  
258 infection type of the male and female) of offspring survives.

259 Therefore, in the presence of two strains we rewrite Eqn. 11 as

$$x_{t+0.5} = a_t (1 - \mu_X) F_X (i_t + j_t H_Y + q_t) \quad (14)$$

260 where  $j_t$  is the fraction of males infected with cytotpe  $Y$ . Similarly, the  
261 proportion of viable offspring infected with cytotpe  $Y$  following reproduction is  
262 given by

$$y_{t+0.5} = b_t (1 - \mu_Y) F_Y (i_t H_X + j_t + q_t). \quad (15)$$

263 where  $b_t$  is the fraction of females infected with cytotpe  $Y$ . The proportion of  
264 viable uninfected offspring is now given by

$$w_{t+0.5} = (\mu_X F_X a_t + \mu_Y F_Y b_t + c_t) (i_t H_X + j_t H_Y + q_t) \quad (16)$$

265 Our parameterization for the infection dynamics is based on values that are  
266 reasonable for *Wolbachia* cytotpes (Hoffmann & Turelli (1997); Dobson *et al.*  
267 (2002); Charlat *et al.* (2005)). In the main text, we assume that fecundity loss,

268 transmission failure, and survival of CI are equal between all cytotypes. However,  
269 the Supplementary Information provides an analysis of the dynamics when the  
270 introduced cytotypes are not identical. Although our sensitivity analysis indicates  
271 that the qualitative results are robust to differences between cytotypes, practical  
272 implementation of such methods would require further analyses that are specific  
273 to cytotypes.

## 274 Results

275 In the following sections we first test our model against well-established results  
276 related to *Wolbachia* invasion as a method of model validation, establish the re-  
277 lationship between *Wolbachia* and the location of the Allee threshold, and finally  
278 evaluate the potential for the release of infected insects to control a population.

### 279 Model validation

280 We first determine whether our model captures the same features of the important  
281 earlier work (Hoffmann *et al.* (1990); Turelli & Hoffmann (1991); Hancock *et al.*  
282 (2011)). Hoffmann *et al.* (1990) derived an analytic expression for the expected  
283 equilibrium infection frequencies. After adjusting their notation to match ours and  
284 simplifying, the equilibrium infection frequency for a single cytotype of *Wolbachia*  
285  $X$  should satisfy the equation

$$(1 - H_X)(1 - \mu_X F_X)p^2 + (F_X + H_X - 2)p + 1 - F_X(1 - \mu_X) = 0.$$

286 Their work predicts that there is an unstable equilibrium, below which the  
287 invasion of the introduced cytotype is unsuccessful and above which invasion is  
288 successful. This equilibrium is therefore referred to as the *Wolbachia* invasion  
289 threshold. Initial infection frequencies above this value will increase until reaching  
290 the higher stable equilibrium that indicates a successful invasion.

291 Following Charlat *et al.* (2005), we considered invasion dynamics by estimat-  
292 ing the infection frequency at generation  $t + 1$  as a function of the frequency at  
293 generation  $t$ . Doing so allows us to create a simple graphical representation of  
294 the stable and unstable equilibria (Fig. 1A). More specifically, we compute the  
295 infection frequency between two subsequent generations with generation  $t$  on the  
296 horizontal axis and generation  $t + 1$  on the vertical axis. Equilibria occur when  
297 this curve intersects the line  $y = x$  (i.e. the infection frequency at generation  
298  $t + 1$  is the same as at generation  $t$ ). In addition to using our model to find these  
299 equilibria, we also used the analytically derived results of Hoffmann *et al.* (1990)  
300 (Fig. 1). Unless stated otherwise, the default parameter values are listed in Table  
301 .

302 As evidenced by the results shown in Fig. 1, our simulation results are consis-  
303 tent with the analytically derived equilibrium infection frequencies. This verifies  
304 that our simulations are in line with the behavior we would expect from our model  
305 and are consistent with the results observed in Charlat *et al.* (2005). However, this  
306 is not surprising given that our model makes similar assumptions on the mecha-  
307 nisms driving *Wolbachia* invasion dynamics (e.g. fecundity loss and cytoplasmic in-  
308 compatibilities). In contrast to earlier studies, our population model is dynamically  
309 different because of the inclusion of Allee effects and false pheromones. Therefore,  
310 we determined the relationship between the invasion threshold and these features

311 of the model. We found that the *Wolbachia* invasion threshold is not affected by  
312 Allee effects (Fig. 1B) nor by the application of false pheromones to the host in-  
313 sect (not shown). This is important to note because in all of our simulations and  
314 analyses, the invasion threshold does not vary as  $\theta$  and  $P$  are adjusted. Finally, we  
315 note that the invasion threshold is not affected by the parameter  $S_0$  (not shown).

## 316 **The effect of *Wolbachia* on the Allee threshold**

317 To determine the dynamical effects of the presence of *Wolbachia* infection within a  
318 population, we find the Allee threshold in insect populations both in the presence  
319 and absence of infection. In this section, we ignore stochasticity as well as the  
320 release of pheromones ( $P = 0$ ). For a given initial proportion of infected indi-  
321 viduals, we calculate the reproductive rate between two consecutive generations  
322 (i.e.  $N_{t+1}/N_t$ ) across all population sizes. The resulting reproductive rates for  
323 a population exhibiting strong Allee effects (that is, the population growth rate  
324 drops below one at low densities) are provided in Fig. 2. The equilibria for our  
325 population model occur when  $N_{t+1} = N_t$ , and there are three equilibria: the first  
326 corresponds to population extinction (stable), the second is the Allee threshold  
327 (unstable), and finally the third is the carrying capacity (stable).

328 In addition to considering the population model in the absence of *Wolbachia*-  
329 infected individuals, we calculated the reproductive rates when the initial popu-  
330 lation is comprised of 10% and 50%, respectively, infected individuals (Fig. 2).  
331 Given our parameterization, the frequency of infected individuals is chosen to lie  
332 above the invasion threshold (which is  $\sim 8.5\%$ ); therefore, this figure captures  
333 the dynamics between two consecutive generations during the replacement process

334 when the population contains the specified distribution of infected and uninfected  
335 individuals. As a consequence of cytoplasmic incompatibilities, the maximum re-  
336 productive rate decreases as the proportion of infected individuals increases and  
337 there is also an increase in the Allee threshold. Finally, there is a slight decrease  
338 in the carrying capacity that results from the fecundity loss associated with *Wol-*  
339 *bachia* infection. However, the proportion of *Wolbachia*-infected individuals has a  
340 significantly smaller effect on the location of the Allee threshold than the strength  
341 of the Allee effect itself (Fig. 2).

## 342 **Implications for population management**

### 343 *Deterministic results*

344 In this section, we characterize implications for population management through  
345 the release of *Wolbachia*-infected individuals into an insect population. As ob-  
346 served by Dobson *et al.* (2002), there should be a transient decline in the population  
347 size during the replacement of uninfected hosts by *Wolbachia*-infected individuals.  
348 Therefore, we find the magnitude of this decline in the presence of Allee effects  
349 to determine the conditions under which the replacement process brings the pop-  
350 ulation size below the Allee threshold in a deterministic setting, thereby forcing  
351 extinction. This is achieved by running our model over a range of values for the  
352 strength of the mate-finding Allee effect ( $\theta$ ) and we assume that the initial popu-  
353 lation size is at its carrying capacity (which can be found explicitly, as shown in  
354 the Supplementary Information).

355 We find the minimum population size (relative to  $K$ ) over 50 generations fol-  
356 lowing the introduction of one cytotype (Fig. 3A) and two cytotypes (Fig. 3B).

357 Here, values of zero for the minimal population size indicate that the transient  
358 reduction in population size brought the population below the Allee threshold,  
359 therefore leading to deterministic extinction. The first cytotypic is always released  
360 in the second generation, and the release of the second cytotypic is optimized so  
361 that the release occurs in the generation that causes the largest decline in popula-  
362 tion size. To implement this strategy, we assume that the longest amount of time  
363 between introductions is 25 generations. While in the main text we assume that all  
364 cytotypes have the same infection properties, this assumption is challenged in the  
365 Supplementary Information and our qualitative results are unchanged. To ensure  
366 that the introduction size is above the invasion threshold, in all simulations we  
367 assume that the introduction is large enough so that the initial infection frequency  
368 is 10%. This value lies just above the actual threshold of  $\sim 8.5\%$  resulting from  
369 our parameter values. Therefore, the number of infected individuals introduced in  
370 our simulations directly depends on the current host population size.

371 To determine the success of the releases under varying reproductive rates, we  
372 replicated the results for four different values of  $S_0$ . We observe that for all repro-  
373 ductive rates, the release of individuals infected with a second and incompatible  
374 cytotypic of *Wolbachia* leads to greater success of these methods. Importantly,  
375 we also find that the effectiveness of control via releases of infected individuals  
376 critically depends on the maximum reproductive rate of a population: for low re-  
377 productive rates, a single release is likely more effective at driving a population to  
378 extinction for smaller values of  $\theta$  relative to  $K$  (as indicated in Fig. 3).

379 In contrast, however, populations with high reproductive rates are only driven  
380 to extinction with unrealistically strong Allee effects, suggesting the need for com-  
381plementary management tactics to successfully eliminate a pest population. While

382 low reproductive rates point to greater success of *Wolbachia* introductions, how-  
383 ever, it should be noted that the maximum reproductive rates reported in the  
384 caption to Fig. 3 are in the *absence* of Allee effects. As a consequence of a low  
385 reproductive rate in the absence of Allee effects, strong enough Allee effects can  
386 bring the overall population growth rate below one. In that case, the population  
387 will be driven to extinction by Allee effects alone (this is explored further in the  
388 Supplementary Information). Because we are interested in the combined effect  
389 of Allee effects and *Wolbachia*-induced CI on population dynamics, we consider  
390 populations with relatively high reproductive rates ( $S_0 = 0.15$ ) for the remainder  
391 of the manuscript (the Supplementary Information provides a sensitivity analysis  
392 for lower reproductive rates).

393 When reproductive rates are relatively high, complimentary tactics that either  
394 decreases the population size or further increases the Allee threshold may amplify  
395 the effects of *Wolbachia* introductions. Therefore, we additionally consider the use  
396 of mating disruption through the release of sex pheromones ( $P$ ) as a supplemental  
397 management tactic (see Eqns. 3 and 1). In Fig. 3C, we fix the strength of the  
398 Allee effect relative to  $K$  so that a release of *Wolbachia*-infected individuals does  
399 not successfully bring the population below the Allee threshold. Further,  $\theta/K$  is  
400 a measure for the intensity of the Allee effect – which impacts low density popu-  
401 lations – so this value was chosen to be relatively small ( $\theta/K = 0.1$ ). Of course,  
402 this value would vary by species. We then calculate trajectories of the model to  
403 determine the minimum population size as the number of pheromone sources is  
404 varied. While the release of pheromones alone can drive the population below the  
405 Allee threshold, combining both methods is significantly more effective than either  
406 tactic alone (as shown in Fig. 3C).

407

408 *Stochastic results*

409 The analysis in the previous section was centered on the deterministic outcome  
410 of population management strategies in the presence of Allee effects. Additionally,  
411 we assumed that releases of *Wolbachia*-infected individuals as well as the imple-  
412 mentation of mating disruption occurred in populations that have reached their  
413 carrying capacity (i.e. as assumed in obtaining the results shown in Fig. 3). How-  
414 ever, the interplay of Allee effects and stochasticity can be especially important  
415 at low population sizes, when the population is at higher risk of stochastic extinc-  
416 tion. Therefore, in this section we determine the ability of *Wolbachia* and mating  
417 disruption to drive populations with variable initial population sizes to extinction  
418 in the presence of Allee effects and stochasticity.

419 To achieve this, we determine the probability of extinction based on 500 realiza-  
420 tions of the stochastic model (i.e. Eqns. 7-10) over all relevant combinations of the  
421 initial population size and strength of the Allee effect (i.e. the initial population  
422 size is at most at carrying capacity and the Allee threshold is below the carry-  
423 ing capacity). To determine the relative roles of environmental and demographic  
424 stochasticity, we simulate the model while including both types of stochasticity  
425 (bottom row of Fig. 4) as well as demographic stochasticity alone (top row of Fig.  
426 4). Further, we find the extinction probability under three scenarios: no intro-  
427 duction of *Wolbachia*-infected individuals (first column of Fig. 4), introduction of  
428 one cytotype (second column of Fig. 4), and the introduction of two incompatible  
429 cytotypes (third column of Fig. 4). As in the deterministic setting, we assume  
430 that the introduction of the first cytotype occurs at the second generation. When  
431 two cytotypes are introduced, the generation of the second release is determined

432 in the same way as it is found in the deterministic setting: the second introduc-  
433 tion is optimized for each realization so that it occurs in the generation (up to 25  
434 generations) that creates the largest transient decrease in population size resulting  
435 from the *Wolbachia* introduction. The generation of the second introduction (as it  
436 corresponds to the third column of Fig. 4C) is usually only a small number ( $\sim 2$ )  
437 of generations after the first generation. However, the number of generations be-  
438 tween releases increases as the strength of the mate-finding Allee effect decreases  
439 (see Supplementary Information). As before, each release is implemented so that  
440 the proportion of infected individuals of a given cytotype is 10% (just above the  
441 invasion threshold).

442 Similar to the deterministic results, we find that the introduction of one cyto-  
443 type of *Wolbachia* leads to a higher extinction probability of the population as the  
444 strength of the Allee effects increases (as displayed in the first and second columns  
445 of Fig. 4). Moreover, when the strength of the Allee effect is high, demographic  
446 stochasticity does not increase extinction probability significantly as compared to  
447 the deterministic setting (e.g. Fig. 3). This result follows from the relatively large  
448 initial population sizes near the Allee threshold when  $\theta$  is large. However, the  
449 presence of environmental stochasticity increases the uncertainty in whether ex-  
450 tinction will occur near the boundary between extinction and population survival  
451 (as observed in the second row of Fig. 4). As described in the previous section,  
452 the success of *Wolbachia* releases increases for lower reproductive rates. This find-  
453 ing holds in the stochastic setting (see sensitivity analysis in the Supplementary  
454 Information).

455 Interestingly, when  $\theta$  is relatively high, *Wolbachia* introductions succeed in  
456 driving population extinction independent of the initial population size. This

457 result has the important implication that the success of *Wolbachia* introductions  
458 in driving extinction do not necessarily rely on having a pest population at the  
459 initial stage, or at the front, of the invasion. Additionally, the introduction of two  
460 cytotypes is much more successful than one (Fig. 4, third panel).

461 As explored in the deterministic framework, combining *Wolbachia* introduc-  
462 tions with other methods that increase the Allee threshold (e.g. mating disruption)  
463 will likely further increase the success of the overall management strategy. This is  
464 highlighted in Fig. 4 as a result of the sensitivity of the extinction probability to  
465 the combination of *Wolbachia* introduction and mating disruption. Therefore, as  
466 in Fig. 3C, we fix  $\theta$  relative to  $K$  at a value of 0.1. With this parameterization,  
467 introductions of *Wolbachia* generally do not drive the population to extinction  
468 (with the exception of small initial population sizes). Here, we again consider the  
469 population dynamics under three different management programs: mating disrupt-  
470 tion only (first column of Fig. 5), mating disruption and the introduction of one  
471 cytotype (second column of Fig. 5), and mating disruption and the introduction  
472 of two cytotypes of *Wolbachia* (third column of Fig. 5). Additionally, we note that  
473 utilizing both mating disruption and CI is much more effective than using mating  
474 disruption alone (Fig. 5). Additionally, it is important to note that when  $\theta/K$  is  
475 fixed at 0.1, releases of infected individuals in the absence of mating disruption  
476 has little effect on the extinction probability. Therefore, these two methods can  
477 serve as complementary tactics for pest management.

## 478 Discussion

479 We investigated a population management strategy that considers *Wolbachia*-  
480 induced cytoplasmic incompatibility in the presence of Allee effects. In particular,  
481 we developed a stochastic population model, building upon the seminal approach  
482 of Dobson *et al.* (2002) and the continuously expanding body of literature inves-  
483 tigating the use of Allee effects for the eradication of pest species (Liebhold &  
484 Bascompte (2003); Tobin *et al.* (2011); Liebhold *et al.* (2016)). Our model demon-  
485 strates that the introduction of a small number of incompatible individuals into  
486 a pest population that has a strong pre-existing Allee effect can drive the pest  
487 population to extinction with no further intervention. We also demonstrate that  
488 extinction is possible for surprisingly large pest populations, and that combinations  
489 of more than one strain of *Wolbachia* and mating disruption via sex pheromones  
490 work synergistically to increase the population's extinction risk. Biological control  
491 has been studied for decades as an environmentally friendly alternative to pesti-  
492 cide use (e.g. Knippling (1955); Baumhover (1955)), and our study adds to this  
493 work by providing insight into ways that *Wolbachia* invasions can take advantage  
494 of intrinsic population processes – that is, Allee effects – to manipulate and control  
495 pest populations.

496 An important first step of our modeling work was to uncover the basic inter-  
497 actions between Allee effects and cytoplasmic incompatibility (CI). We show that  
498 the interactions between Allee effects and CI are weak or non-existent: the *Wol-*  
499 *bachia* invasion threshold does not depend on the strength of the Allee effect of its  
500 insect host, and the Allee threshold has a marginal decrease in the presence of CI.  
501 Therefore, invasion of a particular *Wolbachia* strain into a population only depends

502 on the critical population size (or infection frequency) above which invasion suc-  
503 ceeds in a deterministic setting (Barton & Turelli (2011)). This invasion threshold  
504 corresponds to a proportion of infected hosts above which infection spreads up  
505 to almost fixation, and is determined by parameters such as the reduction in egg  
506 hatch-rate caused by CI, the fitness costs of *Wolbachia* carriage, and the fraction  
507 of offspring that inherit the bacteria from an infected mother (Turelli (1994)).  
508 The invasion threshold found with our simulation model is consistent with that  
509 derived analytically (Turelli & Hoffmann (1991)), and unaffected by the intensity  
510 of a mate-finding Allee effect (Fig. 1B). In addition to adding validation to our  
511 model, this result holds interest because many theoretical approaches of *Wolbachia*  
512 invasion dynamics are purely genetic and consider changes in invasion frequency  
513 without considering host population dynamics. Our result follows that of Hancock  
514 *et al* (2011) in suggesting that *Wolbachia* invasion thresholds predicted analyti-  
515 cally hold for closed populations, even when, as assumed here, host reproductive  
516 rate is affected by both positive and negative density dependence.

517 In the presence of strong Allee effects a population below the Allee thresh-  
518 old will be forced to extinction in a deterministic setting, making the Allee effect  
519 a central paradigm for conservation (Deredec & Courchamp (2007); Stephens &  
520 Sutherland (1999)), invasions (Taylor & Hastings (2005)), biological control in-  
521 troduction (Fauvergue *et al.* (2007, 2012)), and as hypothesized in the present  
522 work, eradication (Tobin *et al.* (2011)). Whether an Allee effect is weak or strong  
523 (and the value of the Allee threshold in the latter case) depends on the strength  
524 of the underlying component Allee effect(s) relative to other density-dependent  
525 processes. Our simulations of various levels of cytoplasmic incompatibilities in  
526 a population with a pre-existing mate finding Allee effect suggest that the Allee

527 threshold is much less sensitive to variations in the initial frequency of *Wolbachia*-  
528 infected individuals (0-50% infected individuals) than it is to variations in the  
529 mate-finding Allee effect (Fig. 2B). *Wolbachia*-induced cytoplasmic incompatibil-  
530 ity does decrease population growth rate, as expected, but it has a minimal effect  
531 on the extinction threshold. Therefore, *Wolbachia*-induced CI may be considered  
532 a culling population management tactic where population size is temporarily de-  
533 creased as a result of cytoplasmic incompatibilities (Dobson *et al.* (2002)).

534 Despite their initial apparent independence, cytoplasmic incompatibility and  
535 the Allee effect yield interesting properties when acting in concert. Our first anal-  
536 ysis that considered the combined occurrence of Allee effects and CI in a deter-  
537 ministic context reveals that the transient decrease in population size (resulting  
538 from the invasion of an incompatible *Wolbachia* strain into an uninfected insect  
539 population) is large enough to trigger extinction when the reproductive rate of the  
540 host species is relatively low. Extinction caused by the introduction of a single  
541 *Wolbachia* cytotype in populations with higher reproductive rates is only observed  
542 for very strong Allee effects (Fig. 3A). However, the strength of the Allee effect  
543 required for extinction lowers for the introduction of an additional incompatible  
544 *Wolbachia* strain. The resulting insect extinction probability, estimated by simu-  
545 lating the model in the presence of stochasticity, confirmed the interaction between  
546 the two processes. In the absence of *Wolbachia*, we determined the population's  
547 extinction probability as it varies with its population size and the strength of the  
548 component Allee effect (Fig 4A and 4D). Introducing infected individuals results  
549 in the extinction of populations that would have persisted otherwise (i.e., a popu-  
550 lation that is above its Allee threshold can be brought to extinction). Introducing  
551 a second incompatible cytotype of *Wolbachia* increases CIs within the population

552 and, consequently, increases the extinction domain by reducing the severity of  
553 Allee effect necessary to trigger extinction (Fig 4).

554 Nonetheless, with the exception of species with low reproductive rates, our  
555 model predicts that although Allee effects and CI combine to drive populations to  
556 extinction – even in surprisingly large populations – these extinctions may occur  
557 for unrealistically severe Allee effects. For instance, after the introduction of two  
558 incompatible *Wolbachia* strains, extinction is expected when  $\theta/K \geq 0.4$ ; that is,  
559 extinction occurs if only half of all females successfully mate when the population  
560 is at 40% of the carrying capacity. Unfortunately, field estimations of mate-finding  
561 Allee effects in insects are rare, but it is probable that mating failures only occur  
562 at very low densities. For instance, in the Gypsy moth *Lymantria dispar*, mating  
563 failures occurred below a density (estimated via the rate of male captures on sex-  
564 pheromone traps) of about 4 whereas the carrying capacity was estimated around  
565 800 (Tobin *et al.* (2007, 2013)), so that estimation of  $\theta/K$  in this species could be  
566 one or two orders of magnitude lower than that yielding extinction in our model.

567 Our results indicate that cytoplasmic incompatibility management may inde-  
568 pendently be an effective strategy for populations with a mate-finding Allee ef-  
569 fect and low reproductive rates. Regardless of reproductive rate, our simulations  
570 demonstrate that eradication can be obtained via the combined introduction of  
571 *Wolbachia*-infected individuals and sex pheromone sources into an insect popu-  
572 lation subject to pre-existing Allee effect (Fig. 5). Further, eradication is not  
573 restricted to small populations, but also applies to populations that have reached  
574 carrying capacity. Our results support previous studies that highlight the poten-  
575 tial benefit of simultaneously using multiple complementary management tactics  
576 (Blackwood *et al.* (2012); Suckling *et al.* (2012)); however, the benefits and prac-

577 ticality of such methods will depend on, for example, associated economic costs  
578 and available resources.

579 It is important to note that our analysis is intended to provide insight into a  
580 broad variety of pest species; therefore, in an empirical context it is essential to  
581 perform more thorough analyses of the dynamics of *Wolbachia* invasion and mat-  
582 ing disruption using a species-specific parameterization. Additional methods for  
583 controlling a pest population should also be considered. For example, the combi-  
584 nation of *Wolbachia*-induced CI and mating disruption work synergistically, and  
585 other methods for population control such as parasitism or predation by native  
586 natural enemies may also be complementary. Additionally, while our focus was on  
587 *Wolbachia*-induced CI, there is evidence that *Wolbachia* is capable of other repro-  
588 ductive manipulations including male-killing (Dyer & Jaenike (2004); Richardson  
589 *et al.* (2016)). This suggests the existence of additional avenues for utilizing *Wol-*  
590 *bachia* in the context of pest management.

591 There is a long and prolific body of research in population dynamics that  
592 focuses on understanding the mechanisms stabilizing species near their carrying  
593 capacities (e.g. Hassell & May (1973); May & Anderson (1978); Bernstein (2000)).  
594 More recently, global climate change and the biodiversity crisis, including popu-  
595 lation declines, extinctions, or biological invasions, points towards the increasing  
596 relevance of nonequilibrium ecology (Rohde (2006)) and the biology of small popu-  
597 lations (Fauvergue *et al.* (2012)). Transient dynamics are increasingly emphasized  
598 (Hastings (2004)) and sometimes considered in the specific context of population  
599 management (Ezard *et al.* (2010); Kidd & Amarasekare (2012)). As first high-  
600 lighted by Dobson *et al.* (2002), cytotype replacement which occurs in the course  
601 a successful *Wolbachia* invasion yields a transient coexistence of incompatible in-

602 fections within a host population, and as a consequence, a transient decrease in  
603 reproductive rate. Here, the transients only last a few generations and this pertur-  
604 bation of the population's microbiome is the first step in a destabilizing cascade.  
605 We show here that the population can then be pushed toward a second step of  
606 destabilization, triggered by a mate-finding Allee effect that can be reinforced by  
607 the application of mating disruption, which potentially drives the population to  
608 extinction.

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## 613 **Authors' contributions statement**

614 JCB and XF conceived the study. JCB, XF, and RV discussed the model. JCB and  
615 RV developed, ran, and analyzed the model. JCB and XF wrote the manuscript.

## 616 **References**

617 Atyame, C.M., Cattell, J., Lebon, C., Flores, O., Dehecq, J.S., Weill, M., Gouagna,  
618 L.C. & Tortosa, P. (2015) Wolbachia-Based Population Control Strategy Target-  
619 ing *Culex quinquefasciatus* Mosquitoes Proves Efficient under Semi-Field Con-  
620 ditions. *Plos One*, **10**.

- 621 Barclay, H. & Mackauer, M. (1980) The sterile insect release method for pest  
622 control: a density-dependent model. *Environmental Entomology*, **9**, 810–817.
- 623 Barclay, H.J. (1982) The sterile release method with unequal male competitive  
624 ability. *Ecological Modelling*, **15**, 251–263.
- 625 Barton, N. & Turelli, M. (2011) Spatial waves of advance with bistable dynamics:  
626 cytoplasmic and genetic analogues of Allee effects. *The American Naturalist*,  
627 **178**, E48–E75.
- 628 Baumhover, A.H., Graham, A.J., Bitter, B.A., Hopkins, D.E., New, W.D., Dudley,  
629 F.H. & Bushland, R.C. (1955) Screw-worm control through release of sterilized  
630 flies. *Journal of Economic Entomology*, **48**, 462–466.
- 631 Berec, L., Maxin, D. & Bernhauerová, V. (2016) Male-killing bacteria as agents of  
632 insect pest control. *Journal of Applied Ecology*, **53**, 1270–1279.
- 633 Bernstein, C. (2000) Host-parasitoid models: the story of a successful failure.  
634 *Parasitoid population biology*, pp. 41–57.
- 635 Beroza, M. & Knippling, E. (1972) Gypsy moth control with the sex attractant  
636 pheromone. *Science*, **177**, 19–27.
- 637 Blackwood, J.C., Berec, L., Yamanaka, T., Epanchin-Niell, R.S., Hastings, A. &  
638 Liebhold, A.M. (2012) Bioeconomic synergy between tactics for insect eradica-  
639 tion in the presence of Allee effects. *Proceedings of the Royal Society B-Biological*  
640 *Sciences*, **279**, 2807–2815.
- 641 Boukal, D.S. & Berec, L. (2009) Modelling mate-finding Allee effects and pop-

- 642      ulations dynamics, with applications in pest control. *Population Ecology*, **51**,  
643      445–458.
- 644      Bourtzis, K. (2008) *Wolbachia-Based technologies for insect pest population control*,  
645      vol. 627 of *Advances in Experimental Medicine and Biology*, pp. 104–113.
- 646      Braig, H.R., Guzman, H., Tesh, R.B. & Oneill, S.L. (1994) Replacement of the  
647      natural wolbachia symbiont of *Drosophila simulans* with a mosquito counterpart.  
648      *Nature*, **367**, 453–455.
- 649      Caspari, E. & Watson, G. (1959) On the evolutionary importance of cytoplasmic  
650      sterility in mosquitoes. *Evolution*, **13**, 568–570.
- 651      Charlat, S., Calmet, C., Andrieu, O. & Mercot, H. (2005) Exploring the Evolution  
652      of Wolbachia Compatibility Types A Simulation Approach. *Genetics*, **170**, 495–  
653      507.
- 654      Davidson, G. (1974) *Genetic Control of Insect Pests*. Academic Press Inc.
- 655      Deredec, A. & Courchamp, F. (2007) Importance of the Allee effect for reintro-  
656      ductions. *Ecoscience*, **14**, 440–451.
- 657      Dobson, S.L., Fox, C.W. & Jiggins, F.M. (2002) The effect of Wolbachia-induced  
658      cytoplasmic incompatibility on host population size in natural and manipulated  
659      systems. *Proceedings of the Royal Society B-Biological Sciences*, **269**, 437–445.
- 660      Dyer, K.A. & Jaenike, J. (2004) Evolutionary stable infection by a male-killing  
661      endosymbiont in *Drosophila innubila*. *Genetics*, **168**, 1443–1455.
- 662      Ezard, T.H.G., Bullock, J.M., Dalglish, H.J., Millon, A., Pelletier, F., Ozgul, A. &  
663      and Koons, D.N. (2010) Matrix models for a changeable world: the importance

664 of transient dynamics in population management. *Journal of Applied Ecology*,  
665 **47**, 515–523.

666 Fauvergue, X. (2013) A review of mate-finding Allee effects in insects: from in-  
667 dividual behavior to population management. *Entomologia Experimentalis et*  
668 *Applicata*, **146**, 79–92.

669 Fauvergue, X., Malausa, J.C., Giuge, L. & Courchamp, F. (2007) Invading par-  
670 asitoids suffer no Allee effect: a manipulative field experiment. *Ecology*, **88**,  
671 2392–2403.

672 Fauvergue, X., Vercken, E., Malausa, T. & Hufbauer, R.A. (2012) The biology  
673 of small, introduced populations, with special reference to biological control.  
674 *Evolutionary applications*, **5**, 424–443.

675 Hancock, P.A., Sinkins, S.P. & Godfray, H.C.J. (2011) Population dynamic models  
676 of the spread of Wolbachia. *The American Naturalist*, **177**, 323–333.

677 Hassell, M.P. & May, R.M. (1973) Stability in insect host-parasite models. *The*  
678 *Journal of Animal Ecology*, pp. 693–726.

679 Hastings, A. (2004) Transients: the key to long-term ecological understanding?  
680 *Trends in Ecology & Evolution*, **19** pp. 39–45.

681 Hilgenboecker, K., Hammerstein, P., Schlattmann, P., Telschow, A. & Werren, J.H.  
682 (2008) How many species are infected with Wolbachia?—a statistical analysis of  
683 current data. *FEMS microbiology letters*, **281**, 215–220.

684 Hoffmann, A.A. & Turelli, M. (1997) *Cytoplasmic incompatibility in insects*. Influ-

685 ential passengers: inherited microorganisms and arthropod reproduction. Ox-  
686 ford University Press, Oxford, UK.

687 Hoffmann, A.A., Turelli, M. & Harshman, L.G. (1990) Factors affecting the dis-  
688 tribution of cytoplasmic incompatibility in *Drosophila simulans*. *Genetics*, **126**,  
689 933–948.

690 Johnson, D.M., Liebhold, A.M., Tobin, P.C. & Bjornstad, O.N. (2006) Allee effects  
691 and pulsed invasion by the gypsy moth. *Nature*, **444**, 361–363.

692 Kidd, D. & Amarasekare, P. (2012) The role of transient dynamics in biological  
693 pest control: insights from a host–parasitoid community. *Journal of Animal*  
694 *Ecology*, **81**, 47–57.

695 Knippling, E.F. (1955) Possibilities of insect control or eradication through the use  
696 of sexually sterile males. *Journal of Economic Entomology*, **48**, 459–462.

697 Knippling, E.F. (1970) Suppression of pest Lepidoptera by releasing partially sterile  
698 males a theoretical appraisal. *Bioscience*, **20**, 465–470.

699 Krawfsur, E.S. (1998) Sterile insect technique for suppressing and eradicating insect  
700 populations: 55 years and counting. *Journal of Agricultural Entomology*, **15**,  
701 303–317.

702 Kramer, A.M., Dennis, B., Liebhold, A.M. & Drake, J.M. (2009) The evidence for  
703 Allee effects. *Population Ecology*, **51**, 341–354.

704 Laven, H. (1967) Eradication of *Culex pipiens fatigans* through Cytoplasmic In-  
705 compatibility. *Nature*, **216**, 383–384.

- 706 Lewis, M. & Van Den Driessche, P. (1993) Waves of extinction from sterile insect  
707 release. *Mathematical Biosciences*, **116**, 221–247.
- 708 Liebhold, A. & Bascompte, J. (2003) The Allee effect, stochastic dynamics and  
709 the eradication of alien species. *Ecology Letters*, **6**, 133–140.
- 710 Liebhold, A.M., Halverson, J.A. & Elmes, G.A. (1992) Gypsy moth invasion in  
711 North America - a quantitative analysis. *Journal of Biogeography*, **19**, 513–520.
- 712 Liebhold, A.M., Berc, L., Brockerhoff, E.G., Epanchin-Niell, R.S., Hastings, A.,  
713 Herms, D.A., Kean, J.M., McCullough, D.G., Suckling, D.M., Tobin, P.C. & Ya-  
714 manaka, T. (2016) Eradication of Invading Insect Populations: From Concepts  
715 to Applications. *Annual Review of Entomology*, **61**, 335–352.
- 716 Liebhold, A.M. & Tobin, P.C. (2008) Population ecology of insect invasions and  
717 their management. *Annual Review of Entomology*, **53**, 387–408.
- 718 Losey, J.E. & Vaughan, M. (2006) The economic value of ecological services pro-  
719 vided by insects. *Bioscience*, **56**, 311–323.
- 720 May, R.M. & Anderson, R.M. (1978) Regulation and Stability of Host-Parasite  
721 Population Interactions: II. Destabilizing Processes. *Journal of Animal Ecology*,  
722 **47**, 249–267.
- 723 Melbourne, B.A. & Hastings, A. (2008) Extinction risk depends strongly on factors  
724 contributing to stochasticity. *Nature*, **454**, 100–103.
- 725 Richardson, K.M., Schiffer, M., Griffin, P.C., Lee, S.F. & Hoffman, A.A. (2016)  
726 Tropical *Drosophila pandora* carry *Wolbachia* infections causing cytoplasmic in-  
727 compatibility or male killing. *Evolution*, **70**, 1791–1802.

- 728 Klaus, K. Nonequilibrium Ecology (2006) Nonequilibrium Ecology. Cambridge  
729 University Press, Cambridge, UK.
- 730 Schoener, T.W., Clobert, J., Legendre, S. & Spiller, D.A. (2003) Life-history mod-  
731 els of extinction: a test with island spiders. *The American Naturalist*, **162**,  
732 558–573.
- 733 Sharov, A.A., Leonard, D., Liebhold, A.M., Roberts, E.A. & Dickerson, W. (2002a)  
734 “Slow The Spread”: A National Program to Contain the Gypsy Moth. *Journal*  
735 *of Forestry*, **100**, 30–35.
- 736 Sharov, A.A., Leonard, D., Liebhold, A.M. & Clemens, N.S. (2002b) Evalua-  
737 tion of Preventive Treatments in Low-Density Gypsy Moth Populations Using  
738 Pheromone Traps. *Journal of Economic Entomology*, **95**, 1205–1215.
- 739 Slatkin, M. & Smith, J.M. (1979) Models of coevolution. *Quarterly Review of*  
740 *Biology*, pp. 233–263.
- 741 Sower, L.L. & Whitmer, G.P. (1977) Population growth and mating success of In-  
742 dian meal moths and almond moths in the presence of synthetic sex pheromone.  
743 *Environmental Entomology*, **6**, 17–20.
- 744 Steiner, L., Hart, W., Harris, E., Cunningham, R., Ohinata, K. & Kamakahi,  
745 D. (1970) Eradication of the oriental fruit fly from the Mariana Islands by the  
746 methods of male annihilation and sterile insect release. *Journal of Economic*  
747 *Entomology*, **63**, 131–135.
- 748 Stephens, P.A. & Sutherland, W.J. (1999) Consequences of the Allee effect for

749 behaviour, ecology and conservation. *Trends in ecology & evolution*, **14**, 401–  
750 405.

751 Stouthamer, R., Breeuwer, J.A.J. & Hurst, G.D.D. (1999) *Wolbachia pipientis*:  
752 Microbial manipulator of arthropod reproduction. *Annual Review of Microbiol-*  
753 *ogy*, **53**, 71–102.

754 Suckling, D.M., Stringer, L.D., Stephens, A.E.A., Woods, B., Williams, D.G.,  
755 Baker, G. & El-Sayed, A.M. (2014) From integrated pest management to in-  
756 tegrated pest eradication: technologies and future needs. *Pest Management*  
757 *Science*, **70**, 179–189.

758 Suckling, D.M., Tobin, P.C., McCullough, D.G. & Herms, D.A. (2012) Combining  
759 Tactics to Exploit Allee Effects for Eradication of Alien Insect Populations.  
760 *Journal of Economic Entomology*, **105**, 1–13.

761 Taylor, C.M. & Hastings, A. (2005) Allee effects in biological invasions. *Ecology*  
762 *Letters*, **8**, 895–908.

763 Tobin, P.C., Onufrieva, K.S. & Thorpe, K.W. (2013) The relationship between  
764 male moth density and female mating success in invading populations of *Ly-*  
765 *mantria dispar*. *Entomologia Experimentalis Et Applicata*, **146**, 103–111.

766 Tobin, P.C., Berec, L. & Liebhold, A.M. (2011) Exploiting Allee effects for man-  
767 aging biological invasions. *Ecology Letters*, **14**, 615–624.

768 Tobin, P.C., Whitmire, S.L., Johnson, D.M., Bjornstad, O.N. & Liebhold, A.M.  
769 (2007) Invasion speed is affected by geographical variation in the strength of  
770 Allee effects. *Ecology Letters*, **10**, 36–43.

- 771 Turelli, M. (1994) Evolution of incompatibility-inducing microbes and their hosts.  
772 *Evolution*, pp. 1500–1513.
- 773 Turelli, M. & Hoffmann, A.A. (1991) Rapid spread of an inherited incompatibility  
774 factor in California *Drosophila*. *Nature*, **353**, 440–442.
- 775 Yamanaka, T. & Liebhold, A.M. (2009) Spatially implicit approaches to under-  
776 stand the manipulation of mating success for insect invasion management. *Pop-*  
777 *ulation Ecology*, **51**, 427–444.
- 778 Zabalou, S., Riegler, M., Theodorakopoulou, M., Stauffer, C., Savakis, C. &  
779 Bourtzis, K. (2004) Wolbachia-induced cytoplasmic incompatibility as a means  
780 for insect pest population control. *Proceedings of the National Academy of Sci-*  
781 *ences of the United States of America*, **101**, 15042–15045.

782 **Tables**

Parameter	Description	Value	Source
$N_0$	initial population size	varies	
$m, m_e$	<i>per capita</i> fecundity	25	
$\sigma^*$	standard deviation of fecundity	5	
$S_0^*$	survivorship in absence of competition	varies	
$\alpha$	related to carrying capacity	0.00002	[1]
$\gamma$	related to intraspecific competition	1	[1]
$\mu_X^*, \mu_Y$	maternal transmission failure	0.03	[1]
$F_X^*, F_Y$	relative fecundity of infected individuals	0.95	[1]
$H_X, H_Y$	proportion of offspring surviving CI	0.05	[2]
$\theta$	strength of Allee effect	varies	
$P$	number of pheromone sources	varies	

Table 1: List of model parameters. Unless otherwise stated, all figures use these parameter values. Parameter values with a “\*” have associated sensitivity analyses (as discussed in the main text) in the Supplementary Information. [1] refers to the reference Dobson *et al.* 2002, and [2] refers to Charlat *et al.* (2005).

$\begin{array}{c} \text{♀} \\ \text{♀} \end{array} \backslash \begin{array}{c} \text{♂} \\ \text{♂} \end{array}$	$W$	$X$	$Y$
$W$	$W: c_t q_t$ $X: 0$ $Y: 0$	$W: H_X c_t i_t$ $X: 0$ $Y: 0$	$W: H_Y c_t j_t$ $X: 0$ $Y: 0$
$X$	$W: \mu_X F_X a_t q_t$ $X: (1 - \mu_X) F_X a_t q_t$ $Y: 0$	$W: \mu_X F_X H_X a_t i_t$ $X: (1 - \mu_X) F_X a_t i_t$ $Y: 0$	$W: \mu_X F_X H_Y a_t j_t$ $X: (1 - \mu_X) F_X H_Y a_t j_t$ $Y: 0$
$Y$	$W: \mu_Y F_Y b_t q_t$ $X: 0$ $Y: (1 - \mu_Y) F_Y b_t q_t$	$W: \mu_Y F_Y H_X b_t i_t$ $X: 0$ $Y: (1 - \mu_Y) F_Y H_X b_t i_t$	$W: \mu_Y F_Y H_Y b_t j_t$ $X: 0$ $Y: (1 - \mu_Y) F_Y b_t j_t$

Table 2: Summary of *Wolbachia* transmission and its effects on reproduction in its host population. The first column states the maternal *Wolbachia* cytotype and the first row states the paternal *Wolbachia* cytotype. Each box in the table corresponding to a particular pairing between a female and male provides that proportion of the offspring from that pair that are uninfected ( $W$ ), infected with cytotype  $X$ , and infected with cytotype  $Y$ .

## 783 Figure captions

784 **Figure 1.** (A) Verification that our model accurately predicts the invasion thresh-  
785 old as analytically determined in Hoffman *et al.* 1990. Here, we ignore Allee effects,  
786 stochasticity, and assume  $P = 0$ . The thick black curve is the frequency of infected  
787 individuals at time  $t + 1$  given the frequency at  $t$ ; equilibria occur when this curve  
788 and the gray line (which corresponds to the case that the frequency at generations  
789  $t$  and  $t + 1$  are equal) intersect. Note that when the black curve lies above the gray  
790 line, then the infection frequency is increasing; similarly, the infection frequency is  
791 decreasing when the black curve falls below the gray line. The dashed lines indi-  
792 cate the analytically predicted equilibrium. The smaller intersection is an unstable  
793 equilibrium that defines the invasion threshold: i.e. individuals introduced at a  
794 frequency higher than this threshold will successfully invade the population and  
795 approach the higher stable equilibrium. (B) Demonstration that the Allee effect  
796 does not change the invasion threshold (dashed line is our model prediction, solid  
797 line is the analytically predicted non-zero, stable equilibrium). In this figure, we  
798 use more extreme values for parameters related to CI to more clearly demonstrate  
799 the location of the invasion threshold (specifically,  $\mu_X = 0.2$ ,  $H_X = 0.1$ ).

800

801 **Figure 2.** Reproductive rate as a function of population size when  $\theta = 100$  (gray)  
802 and  $\theta = 1500$  (red). Values above one correspond to population growth, and val-  
803 ues below one correspond to decline. The populations corresponding to the solid  
804 lines have no *Wolbachia*-infected individuals, populations with dashed lines have  
805 10% of the population infected, and dash-dotted lines have 50% of the population  
806 infected at generation  $t$ .

807

808 **Figure 3.** Deterministic results. (A) single introduction; (B) two introductions.  
809 Plot displays the minimum population size relative to  $K$  over 50 generations as-  
810 suming that  $N_0 = K$ . The solid line has  $S_0 = 0.08$  (maximum reproductive  
811 rate of 2 in the absence of AE, as in Dobson *et al.* (2002)), the dashed line has  
812  $S_0 = 0.15$  (maximum reproductive rate of 3.75 in the absence of AE), dash-dotted  
813 line has  $S_0 = 0.2$  (maximum reproductive rate of 5 in the absence of AE), and  
814 dotted line has  $S_0 = 0.25$  (maximum reproductive rate of 6.25 in absence of AE).  
815 (C) Results when  $\theta$  relative to  $K$  is fixed 0.1 (as displayed in A and B) with  
816  $S_0 = 0.15$ , and instead varies the number of pheromone sources  $P$  relative to  $K$ .  
817 The black line corresponds to the case with no *Wolbachia*-infected individuals,  
818 dashed corresponds to a single release, and dash-dotted corresponds to a release of  
819 two cytotypes. In all plots, each release is created so the initial infection frequency  
820 of that cytotype is 10%. The first release is at generation one, and the second is  
821 at generation six.

822

823 **Figure 4.** Using  $S_0 = 0.15$ , the colors of each plot represent the extinction prob-  
824 ability for a given parameter combination based on 500 realizations of the model.  
825 In each plot, the initial population size and the strength of the Allee effect  $\theta$  rela-  
826 tive to  $K$  are varied. We note that the carrying capacity of the population in the  
827 absence of Allee effects is 5500 with these parameters; therefore, the introduction  
828 sizes – which adjust the population size so that there is a 10% infection frequency  
829 – do not exceed  $\sim 612$  insects. Top row: demographic stochasticity only. Bottom  
830 row: both demographic and environmental stochasticity. First column: no intro-  
831 duction. Second column: introduction such that infection frequency is at 10%.

832 Third column: two subsequent introductions, both of which ensure the infection  
833 frequency is 10% for each cytotype at time of introduction (see Supplementary In-  
834 formation for generation of second introduction). Gray line is the Allee threshold  
835 (i.e. initial populations below the gray line go to extinction in the deterministic  
836 model).

837

838 **Figure 5.** Using  $S_0 = 0.15$  and fixing  $\theta/K = 0.1$ , the colors of each plot represent  
839 the extinction probability for a given parameter combination based on 500 real-  
840 izations of the model. In each plot, the initial population size and the number of  
841 pheromone sources ( $P$ ) relative to  $K$  are varied. Top row: demographic stochas-  
842 ticity only. Bottom row: both demographic and environmental stochasticity. First  
843 column: no introduction. Second column: introduction such that infection fre-  
844 quency is at 10%. Third column: two subsequent introductions.

845

Figure 1:

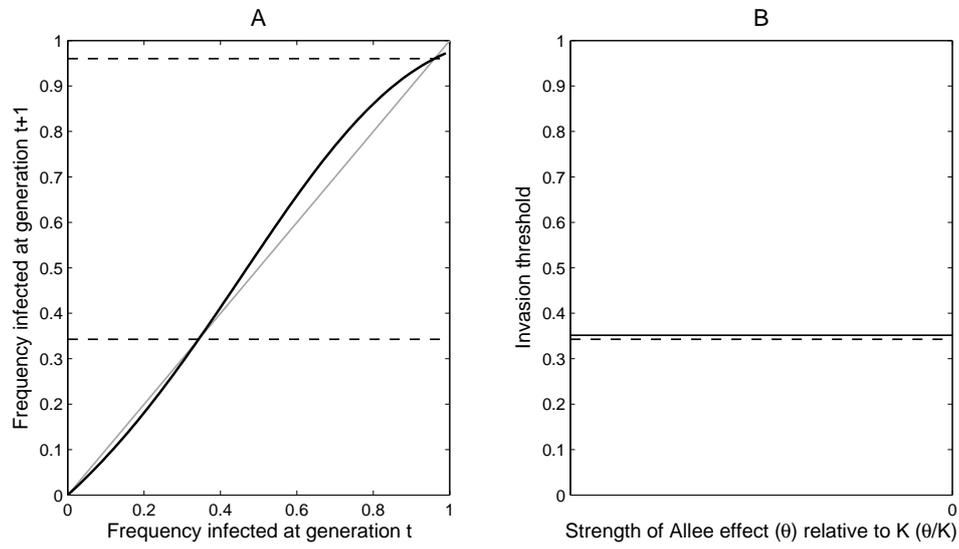


Figure 2:

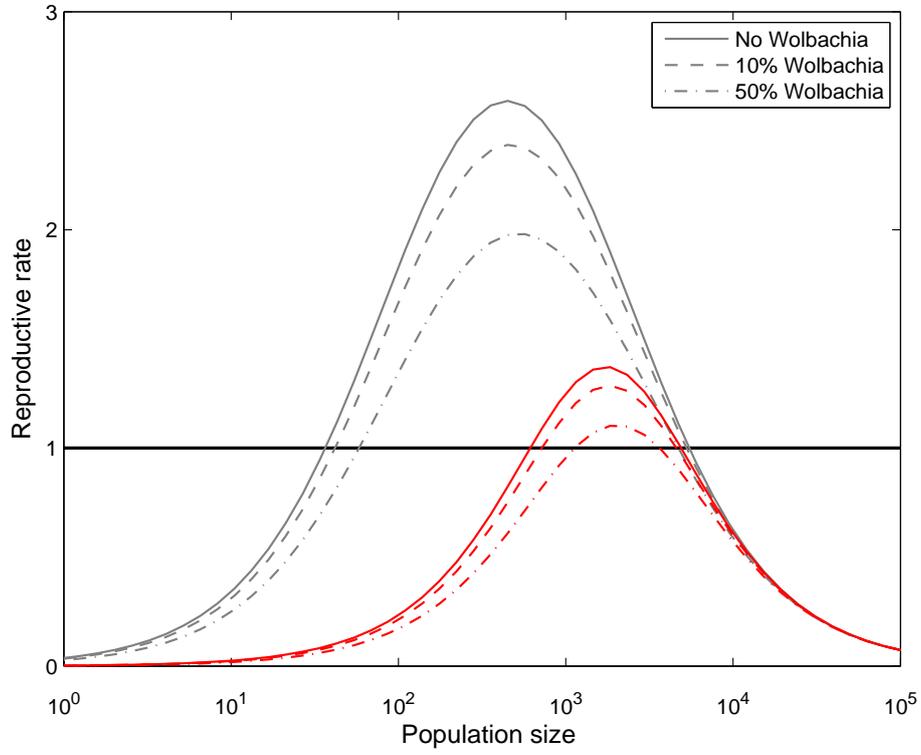


Figure 3:

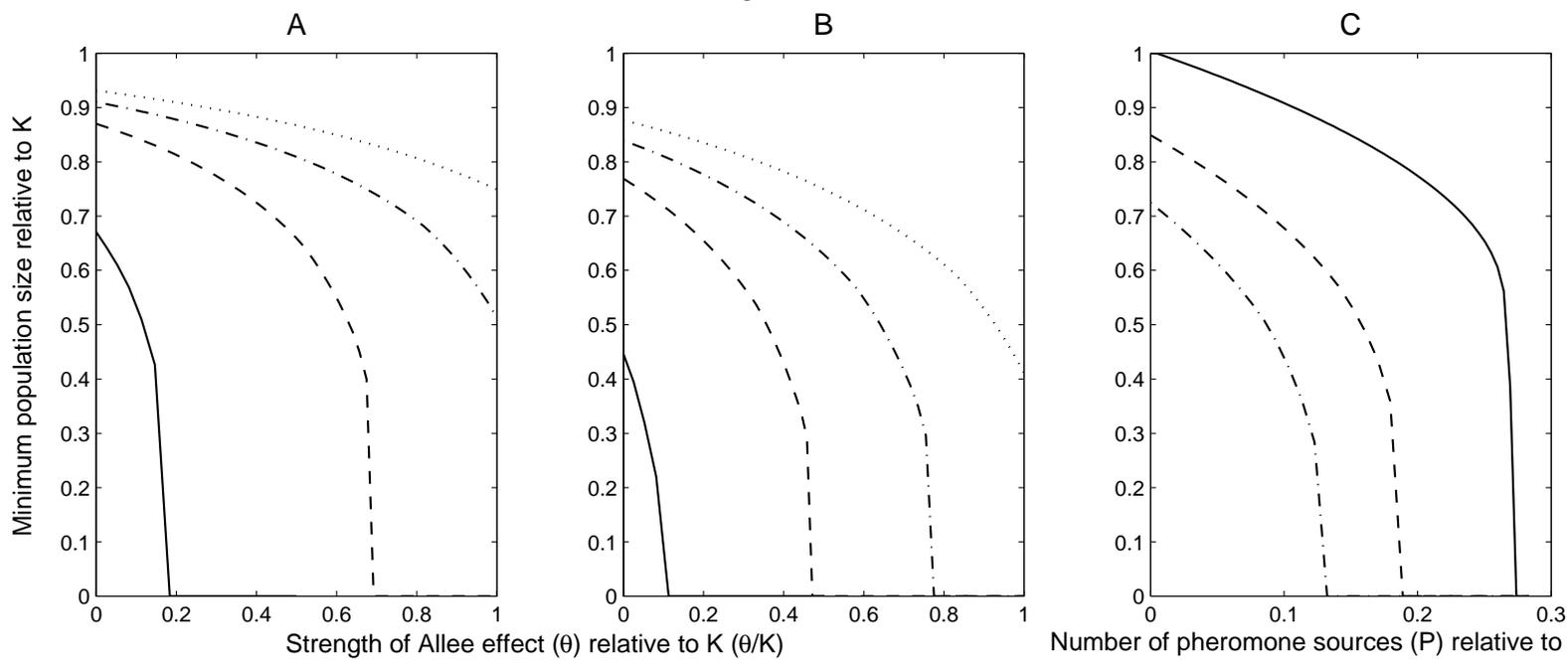


Figure 4:

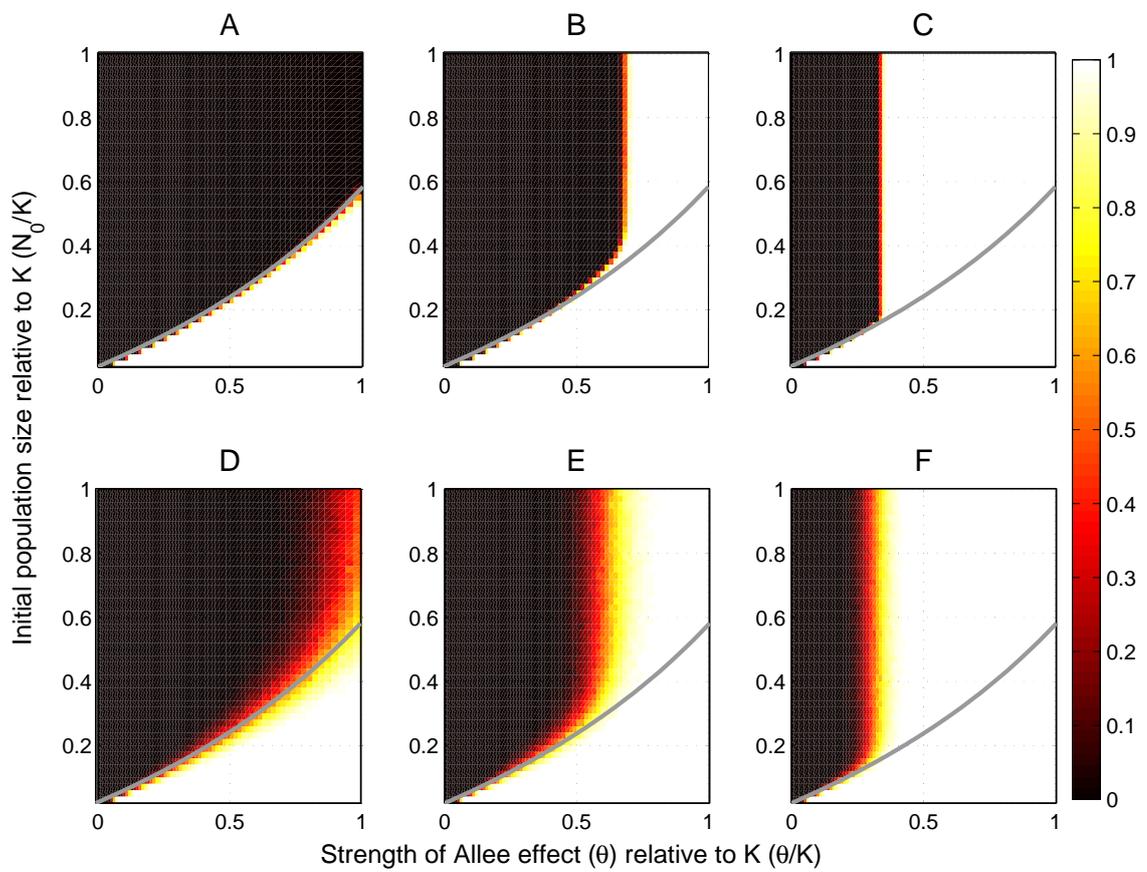


Figure 5:

