

1 **Cryptic genetic variation underpins rapid adaptation to ocean acidification**

2 M. C. Bitter<sup>1\*</sup>, L. Kapsenberg<sup>2</sup>, J.-P. Gattuso<sup>3,4</sup>, and C. A. Pfister<sup>1</sup>

3 <sup>1</sup> Department of Ecology and Evolution, University of Chicago

4 <sup>2</sup> Department of Marine Biology and Oceanography, CSIC Institute of Marine Sciences,

5 Barcelona, Spain

6 <sup>3</sup> Laboratoire d'Océanographie de Villefranche, Sorbonne Université, CNRS, 181 chemin du

7 Lazaret, 06230 Villefranche-sur-mer, France

8 <sup>4</sup>Institute for Sustainable Development and International Relations, Sciences Po, 27 rue Saint

9 Guillaume, 75007 Paris, France

10 \*Corresponding author

11

12 **ORCID IDs:**

13 MCB: 0000-0001-7607-2375

14 LK: 0000-0002-7361-9061

15 CAP: 0000-003-0892-637X

16

17

18

19

20

## 21 **Abstract**

22           Global climate change has intensified the need to assess the capacity for natural  
23 populations to adapt to abrupt shifts in the environment. Reductions in seawater pH constitute a  
24 conspicuous stressor associated with increasing atmospheric carbon dioxide that is affecting  
25 ecosystems throughout the world's oceans. Here, we quantify the phenotypic and genetic  
26 modifications associated with rapid adaptation to reduced seawater pH in the marine mussel,  
27 *Mytilus galloprovincialis*. We reared a genetically diverse larval population in ambient and  
28 extreme low pH conditions (pH<sub>T</sub> 8.1 and 7.4) and tracked changes in the larval size and allele  
29 frequency distributions through settlement. Additionally, we separated larvae by size to link a  
30 fitness-related trait to its underlying genetic background in each treatment. Both phenotypic and  
31 genetic data show that *M. galloprovincialis* can evolve in response to a decrease in seawater pH.  
32 This process is polygenic and characterized by genotype-environment interactions, suggesting  
33 the role of cryptic genetic variation in adaptation to future climate change. Holistically, this work  
34 provides insight into the processes underpinning rapid evolution, and demonstrates the  
35 importance of maintaining standing variation within natural populations to bolster species'  
36 adaptive capacity as global change progresses.

37

## 38 **Introduction**

39           A fundamental focus of ecological and evolutionary biology is determining if and how  
40 natural populations can adapt to rapid changes in the environment. Recent efforts that have  
41 combined natural population censuses with genome-wide sequencing techniques have shown that  
42 phenotypic changes due to abrupt environmental shifts oftentimes occur concomitantly to  
43 signatures of selection at loci throughout the genome<sup>1-4</sup>. These studies demonstrate the

44 importance of standing genetic variation in rapid evolutionary processes<sup>5</sup>, and challenge classical  
45 population genetic theory, which assumes that most genetic variation has a small effect on fitness  
46 and that selective forces alter this variation gradually over a timescale of millennia<sup>6,7</sup>.  
47 Theoretical<sup>8,9</sup> and experimental studies<sup>10–13</sup> have further shown that rapid adaptation via standing  
48 variation is oftentimes characterized by genotype-environment interactions, in which a particular  
49 genetic background is most fit in one environment, while an alternate genetic background leads  
50 to a fitness advantage when the environment shifts<sup>14</sup>.

51         Recently, it has been suggested that genotype-environment interactions during extreme  
52 stress or exposure to novel environmental conditions are underpinned by “cryptic genetic  
53 variation”, defined as a subclass of standing genetic variation with a conditional effect, such that  
54 it becomes adaptive in a new environment<sup>14,15</sup>. The ways in which cryptic genetic variation is  
55 maintained and ultimately influences evolutionary dynamics has been explored in theory<sup>9,16</sup>. But,  
56 the relative importance of cryptic variation in nature has yet to be robustly demonstrated,  
57 particularly within the context of non-model and ecologically important species<sup>14</sup>. Even amidst  
58 its suggested ecological importance to colonization of novel habitats<sup>12</sup>, empirical validation of its  
59 presence and role is limited, and has almost exclusively focused on prokaryotic systems<sup>10,13,17</sup>  
60 and model species, such as *Drosophila* and *Arabidopsis*<sup>18,19</sup>. Still, existing empirical work has  
61 provided initial evidence that cryptic variation may allow populations to circumvent a fitness  
62 valley during an evolutionary response, thereby preventing severe population bottlenecks during  
63 rapid adaptation<sup>14</sup>. Confirming the role of cryptic genetic variation in rapid adaptation is  
64 especially relevant given the threat of global climate change, as natural populations become  
65 increasingly exposed to environmental conditions that exceed contemporary variability. The

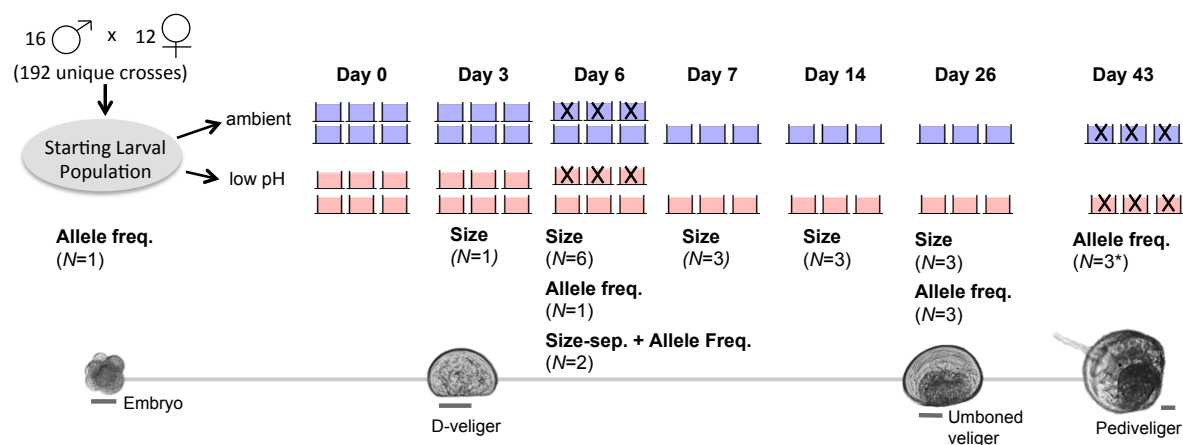
66 extent to which current levels of genetic variation will facilitate the magnitude and rate of  
67 adaptation necessary for species persistence is unclear<sup>20,21</sup>.

68 In marine systems one pertinent threat is ocean acidification, the global-scale decline in  
69 seawater pH driven by oceanic sequestration of anthropogenic carbon dioxide emissions<sup>22</sup>. The  
70 current rate of pH decline is unprecedented in the past 55 million years<sup>23</sup>, and lab-based studies  
71 have shown negative effects of expected pH conditions on a range of fitness-related traits (e.g.,  
72 growth, reproduction, and survival) across life-history stages and taxa<sup>24</sup>. Marine bivalves are one  
73 of the most vulnerable taxa to ocean acidification<sup>25,26</sup>, particularly during larval development<sup>27</sup>.  
74 The ecologically and economically valuable Mediterranean mussel, *Mytilus galloprovincialis*, is  
75 an exemplary species for studying the effects of ocean acidification on larval development. Low  
76 pH conditions reduce shell size and induce various, likely lethal, forms of abnormal larval  
77 development<sup>28,29</sup>. Sensitivity to low pH, however, can vary substantially across larvae from  
78 distinct parental crosses, suggesting that standing genetic variation could fuel an adaptive  
79 response to ocean acidification<sup>29</sup>.

80 Here, we explored the potential for, and dynamics of, rapid adaptation to ocean  
81 acidification in *M. galloprovincialis*. We tracked the phenotypic distributions and trajectories of  
82 29,400 single nucleotide polymorphisms (SNPs), from the embryo stage through larval pelagic  
83 growth and settlement in a genetically diverse larval population, within artificially imposed  
84 ambient (pH<sub>T</sub> 8.05) and extreme low pH treatment (pH<sub>T</sub> 7.4). To test for a genotype-environment  
85 interaction underpinning adaptation, we associated shell size, a trait negatively correlated with  
86 fitness-related abnormalities<sup>29</sup>, to its underlying genetic background in each pH treatment. The  
87 results presented demonstrate the capacity for natural populations to adapt to rapid  
88 environmental change, and suggest that this process will be fueled by cryptic genetic variation.

## 89 Methods Summary

90 We quantified the effects of low pH exposure on phenotypic and genetic variation  
 91 throughout development in a single population of *M. galloprovincialis* larvae (Fig. 1). Larvae  
 92 were reared in ambient and low pH and (i) shell size distributions were quantified on days 3, 6,  
 93 7, 14, and 26; (ii) SNP frequencies across the species' exome were estimated on days 6, 26, and  
 94 43; and (iii) the genetic background of shell size was determined in each treatment to assess the  
 95 presence of a genotype-environment interaction. To generate a starting larval population  
 96 representative of the standing genetic variation within a wild population of *M. galloprovincialis*,  
 97 16 males were crossed to each of 12 females, hereafter referred to as the founding individuals ( $N$   
 98 = 192 unique crosses). The resulting larval population was reared in an ambient ( $pH_T$  8.05,  $N = 6$   
 99 replicate buckets) and low pH treatment ( $pH_T$  7.4-7.5,  $N = 6$  replicate buckets). While the low  
 100 pH treatment falls outside the range of annual variability the population currently experiences  
 101 ( $pH_T \sim 7.8-8.1$ )<sup>29</sup>, and exceeds the -0.4  $pH_T$  units expected globally by 2100<sup>22</sup>, normal  
 102 development of *M. galloprovincialis* larvae can occur at this  $pH$ <sup>29</sup>. We thus expected, *a priori*,  
 103 that this value would effectively reveal the presence of cryptic variation underpinning low pH  
 104 tolerance.

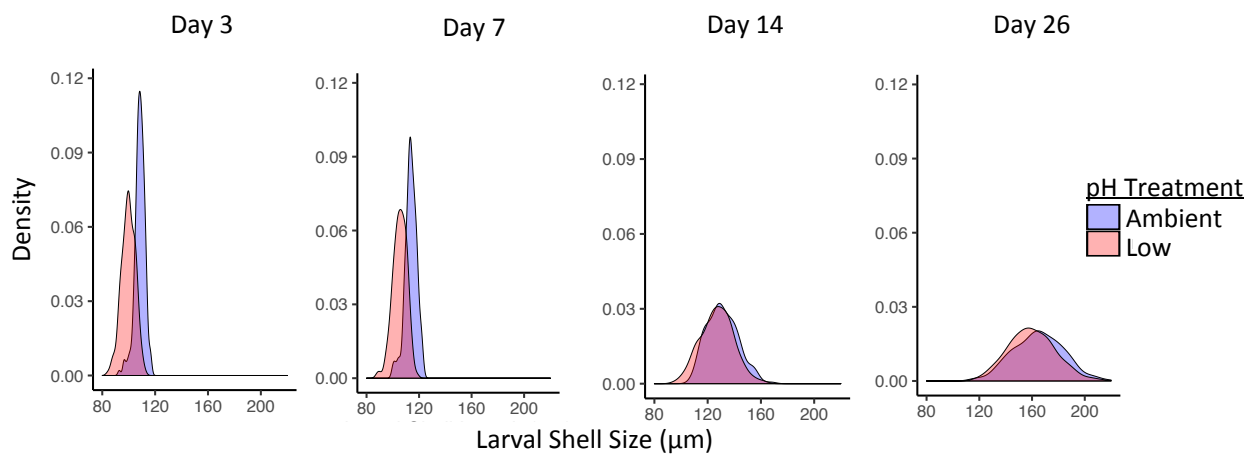


**Figure 1 | Experimental schematic depicting pictures of larvae at key developmental points, cross design, and replication and sampling strategy throughout the experiment.** Scale bar for larval pictures set at 50  $\mu$ m. Replicate buckets marked with an "X" were destructively sampled (i.e. all larvae removed/preserved) and thus absent from the experimental system on subsequent sampling days. \*Allele frequency data from two replicate buckets in the ambient treatment was generated on day 43, as the third replicate bucket to optimize protocol for sampling settled larvae

## 111 Results

### 112 *Phenotypic Trajectories*

113 As expected, shell size was significantly affected by pH treatment throughout the  
114 experiment (linear mixed effects model:  $p = 0.029$ ), and shell length of low pH larvae was 8%  
115 smaller than that of larvae reared in ambient pH on days 3 and 7. Shell length was affected by the  
116 interaction of day and treatment (linear mixed effects model:  $p < 0.001$ ), indicating treatment  
117 specific growth patterns. From days 7 to 26 the size distributions in each treatment began to  
118 converge, with larvae in low pH becoming only 2.5% smaller than those cultured in the ambient  
119 treatment by day 26 (Fig. 2).



**Figure 2 | Larval size distributions throughout the shell growing period.** Larval size was significantly affected by treatment ( $p = 0.029$ ) and the interaction of day and treatment ( $p < 0.001$ ) throughout the shell growing period.

120

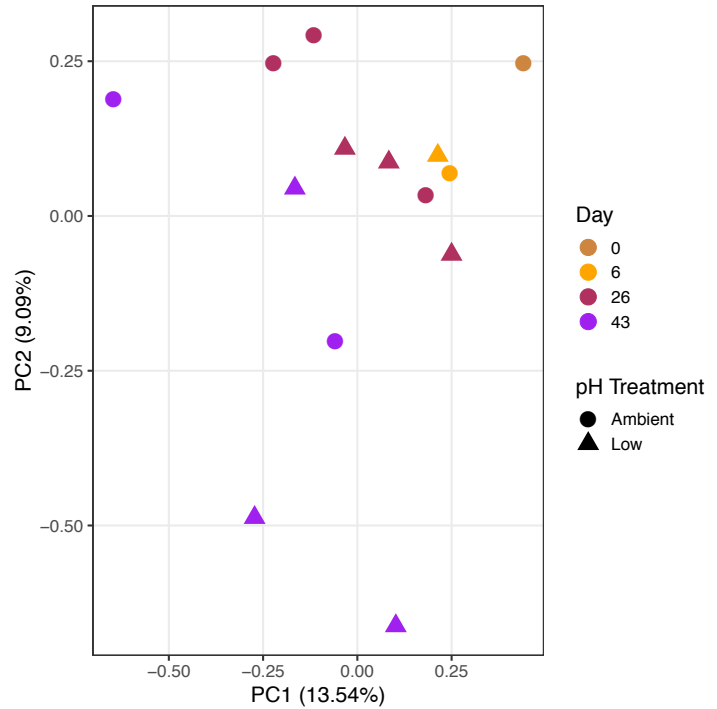
121

122

### 123 *Changes in Genetic Variation*

124 We identified 29,400 SNPs across the species exome that were present within the larval  
125 population across all sampling days and treatments. To link the observed phenotypic trends in

126 each treatment to changes in this variation,  
127 we analyzed the SNP's using principal  
128 component analysis (PCA), outlier loci  
129 identification, and a statistical metric of  
130 genomic differentiation ( $F_{ST}$ ). For both  
131 ambient and low pH treatments, all  
132 analyses indicated increasing genomic  
133 differentiation of the larval cultures away  
134 from the day 0 larval population. This  
135 trend is visually apparent in the principal  
136 component analysis (PCA), which  
137 incorporated allele frequency data from all  
138 larval samples collected during the pelagic



**Figure 3 | Principal component analysis of allele frequency data from larval samples collected throughout the course of the experiment.** Allele frequency from 29,400 SNP's were used for PCA. Sample color corresponds to day and sample shape corresponds treatment condition.

139 stage and settlement (excluding size-separated groups) (Fig. 3). At later days (e.g. days 26 and  
140 43), there is an observed increase in Euclidian distance among samples. This may be driven, at  
141 least in part, by selection-induced declines in larval population survival throughout the pelagic  
142 phase, and an associated increase in the influence of allele frequency “drift” among replicate  
143 buckets. Observations of increased larval mortality throughout the experiment (indicated via  
144 empty D-veliger shells in buckets) corroborated these trends, though we were unable to quantify  
145 larval mortality (MCB, *pers. obs.*).

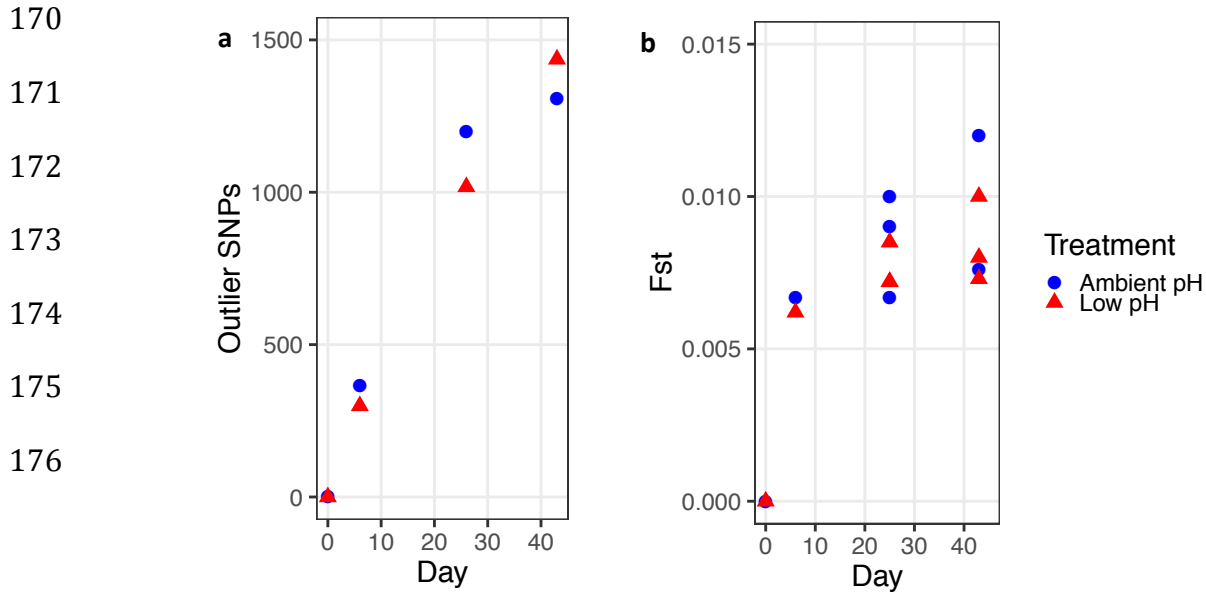
146 We identified SNPs that changed significantly in frequency (*i.e.*, outlier SNPs) between  
147 the day 0 larval population and the larvae sampled on day 6, 26, and 43 in each treatment.  
148 Outlier SNPs were identified using a rank-based approach and the observed allele frequency shift

149 probabilities generated from the Fisher's Exact and Cochran-Mantel-Haenszel tests. This  
150 analysis indicated pervasive signatures of selection in both treatments, with thousands of SNPs  
151 significantly changing in frequency throughout the course of the experiment relative to day 0  
152 (Fig 4a). As the larvae sampled on day 6 were drawn from different replicate buckets as those  
153 sampled on days 26 and 43 (Fig. 1), outlier SNPs observed on all three sampling days point to  
154 candidate loci that may be putatively under selection in each pH environment. We used these  
155 robust SNPs to generate lists of pH-specific loci or overlapping loci (genes containing outlier  
156 SNPs that were responsive in each treatment). In total, we identified 99 ambient pH-specific loci  
157 (31 annotated), 88 low pH specific-loci (29 annotated), and 63 shared loci (24 annotated) based  
158 on transcriptome provided in Moreira *et al.* (2015)<sup>30</sup> (see Supplementary File 1). Therefore, 58%  
159 of the loci exhibiting signatures of selection in the low pH treatment were unresponsive, and  
160 putatively neutral, in the ambient treatment (Fig. 6a). This finding provided an initial indication  
161 of the presence of cryptic variation in the population that may facilitate adaptation to low pH  
162 conditions.

163 Another statistical metric of genetic differentiation,  $F_{ST}$ , was used to identify changes in  
164 the magnitude of selection throughout development. We computed exome-wide (global)  
165 estimates of  $F_{ST}$  pairwise between the day 0 larval population and each available replicate bucket  
166 on all sampling days. The greatest change in  $F_{ST}$  occurred between day 0 and 6, before elevating  
167 more slowly thereafter, suggesting that the majority of selective mortality in *M. galloprovincialis*  
168 larvae occurred prior to day 6 (Fig. 4b).

169





**Figure 4 | Changes in outlier SNPs and  $F_{ST}$  throughout early development.** (a) The number of outlier SNPs identified in ambient and low pH treatment throughout the experiment. The number of outliers reported was standardized by the number of replicate buckets sampled on each day in order to account for increased power associated with increased replication (b)  $F_{ST}$  between the Day 0 larval population and the larval population in each treatment through day 43.

177

178

### 179 *Size Separation*

180 The size-separation of larvae on day 6 isolated the largest 18% from the smallest 82% in

181 the ambient and the largest 21% from the smallest 79% in the low pH treatment (Fig. S1).

182 Hereafter, these groups will be referred to as the fastest and slowest growers, respectively. Shell

183 size on day 6 was significantly affected by treatment and size class (Linear Mixed Effects Model,

184  $p < 0.001$ ). PCA using allele frequency data from the day 0 starting larval population and larval

185 samples collected on day 6 revealed a strong genetic signature of size class (Fig. 5). Specifically,

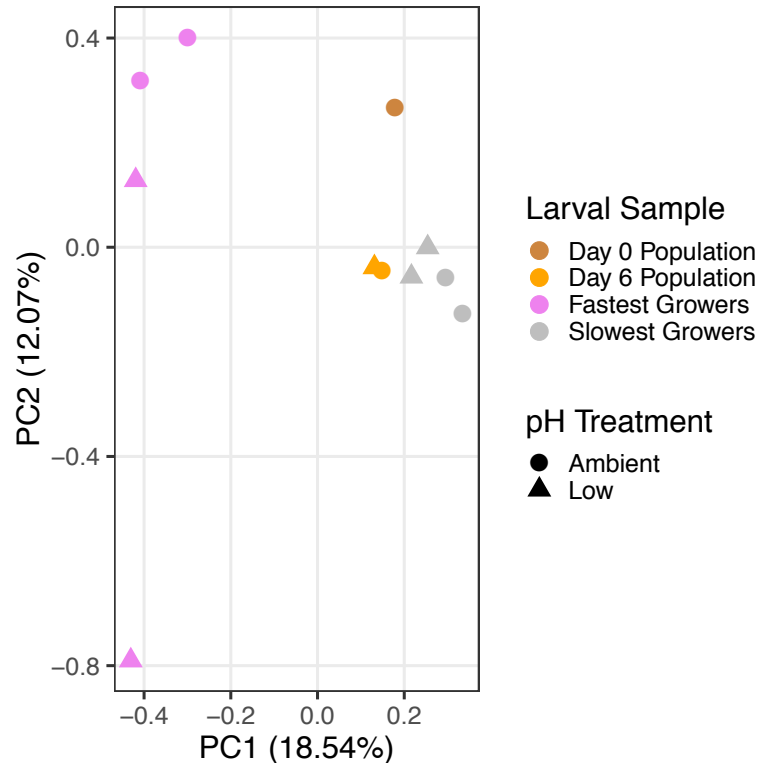
186 the fastest growers segregated along PC1 from the slowest growers in both treatments, with the

187 day 0 larval population and day 6 larval population samples (from each treatment) falling in

188 between the size-separated groups. The number of outlier SNP's differentiating the fastest and

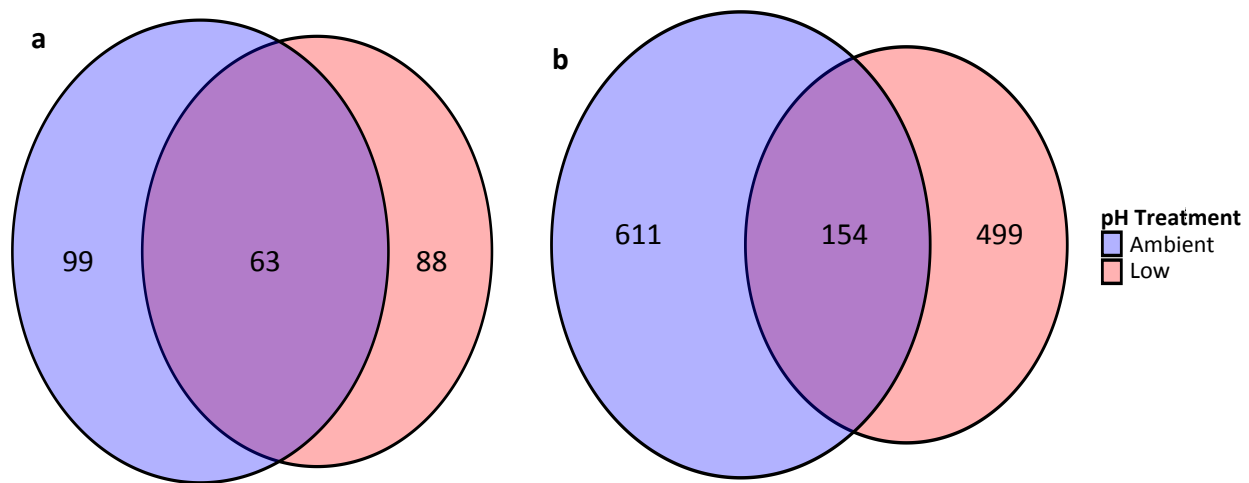
189 slowest growers in each treatment, hereafter referred to as size-selected SNPs, was comparable:

190 963 outlier SNPs identified  
191 ambient and 846 outlier SNPs  
192 identified in the low pH treatment  
193 (outliers identified using  
194 Cochran-Mantel-Haenszel test).  
195 This led to the identification of  
196 611 size-selected loci that were  
197 unique to the ambient pH  
198 treatment (225 annotated), 499  
199 size-selected loci that were  
200 unique to the low pH treatment  
201 (184 annotated), and 154 size-  
202 selected loci (51 annotated) that



**Figure 5 | Principle component analysis of SNPs examining the genomic signature of shell growth.** Allele frequency data using 29,400 SNPs identified in larval samples collected on day 6, 6, and size separated larvae from day 6 (i.e. fastest and slowest growers).

203 were shared between environments<sup>30</sup> (Supplementary File 2). Therefore, 76% of loci associated  
204 with fast shell growth in low pH were not associated with fast growth in the ambient treatment,  
205 and indicated that the fastest growers in the low pH environment come from a largely distinct  
206 genetic background (Fig 6b).  $F_{ST}$  analysis corroborated this trend, as elevated signatures of  
207 differentiation between the fastest growers in ambient and low pH, relative to differentiation  
208 between the slowest growers in ambient and low pH treatments as well as the entire day 6 larval  
209 populations in ambient and low treatments, were observed (Fig. S2). Thus, the size separation, in  
210 concert with the pH-specific signatures of selection observed throughout the entire larval period,  
211 further indicated a unique genetic background associated with fitness in the low pH treatment  
212 (Fig. 6).



**Figure 6 | Genotype-environment interactions during selection in alternate pH environments.** Venn diagrams show the extent of overlap between candidate outlier loci in ambient and low pH conditions (a) throughout the entire larval period (58% unique outlier loci in low pH treatment), and (b) between the fastest and slowest growers on day 6 (76% unique outlier loci in low pH treatment)

213

## 214 Discussion

### 215 *Concurrent shifts in larval size and genetic variation throughout development*

216 While previous work has shown strong negative effects of low pH on larval development  
217 in bivalves<sup>25,27</sup>, the results presented here suggest that standing variation within the species could  
218 facilitate rapid adaptation to ocean acidification. Observed shell length differences on days 3 and  
219 7 matched expectations for the species based on previous work (-1  $\mu\text{m}$  per 0.1 unit decrease in  
220 pH)<sup>29</sup>. However, this difference was reduced ~50% by day 14. Mechanistically, low pH

221 treatment effects on bivalve larval shell growth are driven by the limited capacity of larvae to  
222 regulate carbonate chemistry, specifically aragonite saturation state, in their calcifying space<sup>31-33</sup>.  
223 Our data show that this physiological limitation is greatest prior to day 7, after which low pH  
224 larvae were able to partially recover in size compared to larvae reared in the ambient treatment.

225 It is likely that partial recovery of shell size in low pH larvae observed by day 14 was, at  
226 least in part, driven by natural selection. We have previously shown that the smallest D-veligers  
227 in the low pH treatment display an increased prevalence of morphological abnormalities, which  
228 likely become lethal during the shell growth period<sup>29</sup>. Directional selection against this

229 phenotypic group would shift the size distribution closer to that of larvae reared in ambient  
230 conditions, as we observed. The unique genetic backgrounds of the fastest growing larvae in low  
231 pH at 6 days post-fertilization, as well as the unique outlier loci identified in the low pH  
232 environment throughout the larval period, further strengthen the notion that these phenotypic  
233 trends were rooted in changes in the larval population's underlying genetic variation.

234         Increasing genetic differentiation through time, as evidenced by PCA, outlier loci  
235 identification, and  $F_{ST}$  in both treatments further suggested the process of selection during the  
236 shell growth period. However, the observed trends in  $F_{ST}$  highlighted a developmental point of  
237 heightened selection, even before larval size distributions began to converge. Specifically, when  
238  $F_{ST}$  is scaled by duration of treatment exposure, the genomic differentiation between the day 0  
239 larval population and the larval population on day 6 was three and five times greater than that  
240 observed between the day 0 larval population and the larval population on days 26 and day 43,  
241 respectively. This suggests that a largely singular, intense selection event occurred prior to day 6  
242 and may be responsible for the majority of genetic differentiation that occurs during larval  
243 growth and settlement. We recently identified two specific early developmental processes that  
244 are sensitive to low pH conditions and occur in this timeframe<sup>29</sup>. These processes include the  
245 formation of the shell field (early trochophore stage) and the transition between growth of first  
246 and second larval shell (late trochophore stage), both of which occur within 48 hours of  
247 fertilization, resulting in a suite of size-dependent morphological abnormalities that likely  
248 become lethal during the shell growth period<sup>29</sup>. Traditionally, metamorphosis from the  
249 swimming D-veliger to the settled juvenile is regarded as the main genetic bottleneck during the  
250 development of marine bivalve larvae<sup>34</sup>. Our sampling from the embryo stage through

251 settlement, however, suggests that there is a major selection event prior to day 6 that may have  
252 an even larger effect on shaping genotypes of settled juveniles any selection thereafter.

253 Additional factors that may have allowed the larvae reared in led to the observed  
254 phenotypic dynamics are food-augmented acclimation and selective mortality via food  
255 competition. It has been demonstrated that increased energy availability can allow marine  
256 invertebrates to withstand pH stress<sup>35</sup> and, in the case of *Mytilus edulis*, food availability can  
257 mitigate the negative effects of ocean acidification<sup>36</sup>. This compensation, however, is unlikely in  
258 our experiment. Our algal concentrations during the period of phenotypic convergence (days 7-  
259 26) fell below optimal concentrations reported for the species<sup>37,38</sup>. The substantial delay in the  
260 duration of the pelagic phase relative to published developmental timelines for the species<sup>38,39</sup>,  
261 further demonstrates that the larvae were indeed food limited in each treatment. This food  
262 limitation may have induced intraspecific competition and facilitated the selective mortality of  
263 less fit genotypes, thereby producing the pervasive signatures of selection observed in both  
264 treatments starting on day 6. Selection via differential mortality may have been concentrated on  
265 the smallest larvae in low pH, thus driving the phenotypic convergence between treatments.  
266 Ultimately, surviving larvae in the low pH treatment were able to partially compensate for the  
267 negative effect of CO<sub>2</sub>-acidification on calcification kinetics<sup>32</sup>, though the extent to which multi-  
268 generational selection would allow the population to completely recover the offset in shell size is  
269 an important area of future research. These results bolster emerging studies that have  
270 demonstrated the importance of standing variation in shaping the pH tolerance of coastal marine  
271 species<sup>40-42</sup>.

272 *The role of cryptic genetic variation in extreme low pH adaptation*

273 We identified hundreds of loci responding to each pH treatment throughout the larval  
274 period. Notably, 58% of our candidate low pH loci were statistically unchanged in the ambient  
275 conditions. While some of this treatment disparity may be an artifact (*i.e.* false positives in the  
276 low pH or false negatives in the ambient treatment), it is unlikely that this is the case for all the  
277 unique SNPs identified. These data thus provide evidence that loci that were not critical for  
278 fitness in the ambient environment came under strong selection in the low pH treatment.

279 Associating shell growth to its underlying genetic background in each environment strengthened  
280 this conclusion. Specifically, there was an exceedingly small amount of overlap in size-selected  
281 loci between treatments (76% of size-selected loci were unique to the low pH environment) and  
282 a relative elevation of  $F_{ST}$  differentiation between the fastest growers from the low and ambient  
283 pH buckets. These patterns display a classic genotype-environment interaction in which a  
284 particular genetic background exhibits a specific trait value in one environment (e.g. accelerated  
285 shell growth in ambient pH), while an alternate genetic background leads to the same trait value  
286 when the environment shifts<sup>14</sup>. As shell growth is a direct proxy for fitness<sup>29,43</sup>, these data  
287 suggest that the most fit genotypes in ambient conditions may not be the individuals that harbor  
288 the adaptive genetic variation necessary to improve fitness in simulated ocean acidification.

289 Ultimately, our observation of genotype-environment interactions in response to a  
290 dramatic shift in seawater pH provides strong evidence that adaptation to ocean acidification  
291 may be fueled by cryptic genetic variation. The role of cryptic genetic variation during  
292 adaptation to novel and extreme environmental conditions is becoming increasingly  
293 recognized<sup>12,14,15,44,45</sup>, though its relative importance has yet to be demonstrated as clearly in  
294 nature as it has been in theory<sup>9,14,16</sup>. Our data not only suggest the role of cryptic genetic  
295 variation in rapid adaptation, but also demonstrate this phenomenon in the context of a non-

296 model species subject to global change. The economic and ecological importance of marine  
297 mussels, as well as their global exposure to declining seawater pH, highlights the need to  
298 conserve standing variation in order to allow the adaptive capacity of natural populations to play  
299 out as climate change progresses.

300 Many outlier loci in low pH were also outliers in the ambient treatment (42%). This  
301 likely represents the action of selection against recessive homozygotes within the population,  
302 termed genetic load<sup>46</sup>, and selection by the laboratory regime. The influence of genetic load has  
303 been demonstrated to induce signatures of selection in “neutral” environments in a range of  
304 highly fecund species, such as plants and marine bivalves<sup>34</sup>. This was likely amplified in the  
305 present study by our crossing scheme, which purposefully induced equal proportions of all  
306 pairwise crosses, thereby maximizing the likelihood of lethal, or less “fit”, homozygotes in the  
307 day 0 larval population. These shared signatures of selection could be further associated with  
308 selective pressures induced by the laboratory conditions, such as salinity, temperature, or the  
309 food resources, which are independent of the pH manipulation. While the laboratory conditions  
310 were designed to optimize larval growth and mirror environmental conditions at the collection  
311 site, it is unsurprising that a specific set of environmental variables (*i.e.* lab conditions) favors a  
312 subset of genetic backgrounds in a species exhibiting such high levels of heterozygosity<sup>47</sup>.

### 313 *Putative targets of selection as ocean acidification progresses*

314 The low-pH specific loci we identified (loci with outlier SNPs in every replicate bucket  
315 and across all sampling days) provide targets of natural selection as ocean acidification  
316 progresses. These included a HSPA1A gene (Swissprot ID: Q8K0U4), which encodes heat shock  
317 protein 70 (HSP70), one of a group of gene products whose expression is induced by  
318 physiological stressors and generally work to mediate/prevent protein denaturation and folding<sup>48</sup>

319 . While substantial evidence has documented the role of HSP70 in the thermal stress response  
320 across a range of taxa<sup>48</sup>, emerging transcriptomic studies have also demonstrated the protein's  
321 role in the physiological response to low pH conditions in echinoderms<sup>49</sup> and bivalves<sup>50</sup>. Another  
322 notable candidate locus is a tyrosinase-like protein (Swissprot ID: H2A0L0). Tyrosinase genes  
323 are known to influence biomineralization in marine bivalves during larval<sup>51,52</sup> and juvenile  
324 stages<sup>53</sup>, a process that is fundamentally affected by changes in seawater chemistry. While these  
325 gene expression-based studies provided initial insight into the underlying physiological  
326 responses to changes in seawater chemistry in marine bivalves, our study demonstrates the  
327 presence of underlying genetic variation within these putative loci. This provides, to our  
328 knowledge, the first documentation of standing genetic variation at functionally relevant loci  
329 within marine bivalves, and ultimately offers robust evidence for the species' capacity to adapt to  
330 extreme changes in seawater pH. We are currently investigating these candidates more intensely  
331 through a combination of comparative transcriptomics, quantitative PCR, and *in situ*  
332 hybridization in *M. galloprovincialis* larvae (Kapsenberg *et al.*, unpublished).

### 333 *Conclusions*

334       Species persistence as global climate change progresses will, in part, hinge upon their  
335 ability to evolve in response to the shifting abiotic environment<sup>21</sup>. Our data suggest that the  
336 economically and ecologically valuable marine mussel, *M. galloprovincialis*, currently has the  
337 standing variation necessary to adapt to ocean acidification, though with a potential trade-off of  
338 shell size. We have demonstrated that much of this variation may be cryptic in ambient pH  
339 conditions, yet bolsters a fitness-related trait (shell growth) when seawater pH is reduced.  
340 Ultimately, these findings support conservation efforts aimed at maintaining variation within  
341 natural populations to increase species resilience to future ocean conditions. In a broader



342 evolutionary framework, the substantial levels of genetic variation present in natural populations  
343 have puzzled evolutionary biologists for decades<sup>54</sup>. Though this study does not address the  
344 processes that maintain this variation during periods of environmental stasis, we demonstrate its  
345 utility in rapid adaptation, thereby advancing our understanding of the mechanisms by which  
346 natural populations evolve to abrupt changes in the environment.

## 347 **Methods**

### 348 *Larval cultures*

349 Mature *M. galloprovincialis* individuals were collected in September 2017 from the  
350 underside of a floating dock in Thau Lagoon (43.415°N, 3.688°E), located in Sète, France. Thau  
351 Lagoon has a mean depth of 4 m and connection to the Mediterranean Sea by three narrow  
352 channels. pH variability at the collection site during spawning season ranges from pH<sub>T</sub> 7.80 to  
353 8.10<sup>29</sup>. Mussels were transported to the Laboratoire d'Océanographie (LOV) in Villefranche-sur-  
354 Mer, France and stored in a flow-through seawater system maintained at 15.2°C until spawning  
355 was induced.

356 Within 3 weeks of the adult mussel collection, individuals were cleaned of all epibiota  
357 using a metal brush, byssal threads were cut, and mussels were warmed in seawater heated to  
358 27°C (~+12°C of holding conditions) to induce spawning. Individuals that began showing signs  
359 of spawning were immediately isolated, and allowed to spawn in discrete vessels, which were  
360 periodically rinsed to remove any potential gamete contamination. Gametes were examined for  
361 viability and stored on ice (sperm) or at 16°C (eggs). In total, gametes from 12 females and 16  
362 males were isolated to generate a genetically diverse starting larval population. To produce  
363 pairwise crosses, 150,000 eggs from each female were placed into sixteen separate vessels,  
364 corresponding to the sixteen founding males. Sperm from each male was then used to fertilize

365 the eggs in the corresponding vessel, thus eliminating the potential effects of sperm competition  
366 and ensuring that every male fertilized each female's eggs. After at least 90% of the eggs had  
367 progressed to a 4-cell stage, equal volumes from each vessel were pooled to generate the day 0  
368 larval population (~2 million individuals), from which the replicate culture buckets were seeded.  
369 100,000 individuals were added to each culture buckets ( $N = 12$ , 18 embryos  $\text{mL}^{-1}$ ). The  
370 remaining embryos were frozen in liquid nitrogen, and stored at  $-80^{\circ}\text{C}$  for DNA analysis of the  
371 day 0 larval population. Likewise, gill tissue was collected from all founding individuals and  
372 similarly stored for downstream DNA analyses. Larvae were reared at  $17.2^{\circ}\text{C}$  for 43 days.  
373 Starting on day 4, larvae were fed  $1.6 \times 10^8$  cells of *Tisochrysis lutea* daily. Beginning on day 23,  
374 to account for growth and supplement diet, larvae diet was complemented with  $0.2 \mu\text{L}$  of 1800  
375 Shellfish Diet (Reed Mariculture) (days 23-28 and day 38) and approximately  $1.6 \times 10^8$  cells of  
376 *Chaetoceros gracilis* (days 29-37 and 39-41).

### 377 *Larval sampling*

378 We strategically sampled larvae throughout the experiment to observe phenotypic and  
379 genetic dynamics across key developmental events, including the trochophore to D-veliger  
380 transition (day 6), the shell growth period (day 26), and the metamorphosis from D-veligers to  
381 settlement (day 43). On day 6 of the experiment, larvae were sampled from three of the six  
382 replicate buckets per treatment. A subset of larvae ( $N = 91-172$ ) from each bucket was isolated to  
383 obtain shell length distributions of larvae reared in the two treatments. The remaining larvae  
384 (~25,000) were separated by shell size using a series of six Nitex mesh filters ( $70 \mu\text{m}$ ,  $65 \mu\text{m}$ ,  $60$   
385  $\mu\text{m}$ ,  $55 \mu\text{m}$ ,  $50 \mu\text{m}$ , and  $20 \mu\text{m}$ ; Figure S1) and frozen at  $-80^{\circ}\text{C}$ . The smallest size group  
386 contained larvae arrested at the trochophore stage, and therefore unlikely to survive. The  
387 remaining five size classes isolated D-veligers from the smallest to the largest size. The shell

388 length distribution of the larvae was used to inform, *a posteriori*, which combination of size  
389 classes would produce groups of the top 20% and bottom 80% of shell sizes from each treatment.  
390 The relevant size groups from two replicates per treatment were then pooled for downstream  
391 DNA analysis of each phenotypic group. For the third replicate, *a posteriori*, all size groups were  
392 pooled in order to compute the allele frequency distribution from the entire larval population in  
393 each treatment on day 6. This sample was incorporated into analyses of remaining replicate  
394 buckets, which were specifically used to track shifts in phenotypic and genetic dynamics  
395 throughout the remainder of the larval period in each treatment.

396       Following size separation on day 6, the remaining replicate buckets (N = 3 per treatment)  
397 were utilized to track changing phenotypic and allele frequency distributions in the larval  
398 population through settlement. Larvae were sampled for size measurements on day 3 (n = 30-36),  
399 day 7 (N = 38-71), day 14 (N = 37-104), and day 26 (N = 49-112). Also on day 26, an additional  
400 ~1,000 larvae per replicate were frozen and stored at -80°C pending DNA analysis. Finally, on  
401 day 43, settled individuals were sampled from each bucket (settlement was first observed on day  
402 40 in all buckets). Treatment water was removed, and culture buckets were washed three times  
403 with FSW to remove unsettled larvae. Individuals that remained attached to the walls of the  
404 bucket were frozen and stored at -80°C for DNA analysis.

#### 405 *Culture system and seawater chemistry*

406       Larvae were reared in a temperature-controlled sea table (17.2°C) and 0.35 µm filtered  
407 and UV-sterilized seawater (FSW), pumped from 5 m depth in the bay of Villefranche. Two  
408 culture systems were used consecutively to rear the larvae, both of which utilized the additions of  
409 pure CO<sub>2</sub> gas for acidification of FSW. First, from days 0-26 the larvae were kept in a flow-  
410 through seawater pH-manipulation system described in Kapsenberg *et al.* (2017)<sup>55</sup>. Briefly,

411 seawater pH ( $\text{pH}_T$  8.05 and  $\text{pH}_T$  7.4) was controlled in four header tanks using a glass pH  
412 electrode feedback system (IKS aquastar) and pure  $\text{CO}_2$  gas addition and constant  $\text{CO}_2$ -free air  
413 aeration. Two header tanks were used per treatment to account for potential header tank effects.  
414 Each header tank supplied water to three replicate culture buckets (drip rate of  $2 \text{ L h}^{-1}$ ), fitted  
415 with a motorized paddle and Honeywell Durafet pH sensors for treatment monitoring (see  
416 Kapsenberg et al 2017 for calibration methods).

417 On day 27 of the experiment, the flow-through system was stopped due to logistical  
418 constraints and treatment conditions were maintained, in the same culture buckets, using water  
419 changes every other day. For water changes 5 L of treatment seawater (70% of total volume) was  
420 replaced in each culture using FSW pre-adjusted to the desired pH treatment. Seawater pH in  
421 each culture bucket was measured daily, and before and after each water change.

422 All pH measurements (calibration of Durafets used from day 0-26 and monitoring of  
423 static cultures from day 27-43) were conducted using the spectrophotometric method and  
424 purified *m*-cresol dye and reported on the total scale ( $\text{pH}_T$ )<sup>56</sup>. Samples for total alkalinity ( $A_T$ )  
425 and salinity were taken from the header tanks every 2-3 days from days 0-26 and daily during the  
426 remainder of the experiment.  $A_T$  was measured using an open cell titration on Metrohm  
427 Titrando 888<sup>56</sup>. Accuracy of  $A_T$  measurements was determined using comparison to a certified  
428 reference material (Batch #151, A. Dickson, Scripps Institution of Oceanography) and ranged  
429 between  $-0.87$  and  $5.3 \mu\text{mol kg}^{-1}$ , while precision was  $1.23 \mu\text{mol kg}^{-1}$  (based on replicated  
430 samples,  $n = 21$ ). Aragonite saturation and  $p\text{CO}_2$  were calculated using pH and  $A_T$  measurements  
431 and the *seacarb* package in R<sup>57</sup> with dissociation constants  $K_1$  and  $K_2$ <sup>58</sup>,  $K_f$ <sup>59</sup> and  $K_s$ <sup>60</sup>. Seawater  
432 chemistry results are presented in the electronic supplementary, Tables S1-S2.

433 *Shell Size Analysis*

434 Shell size was determined as the maximum shell length parallel to the hinge using  
435 brightfield microscopy and image analysis in ImageJ software. All statistical analyses were  
436 conducted in R (v. 3.5.3)<sup>61</sup>. As larval shell length data did not pass normality tests (Shapiro-Wilk  
437 test), shell-size was log-transformed to allow parametric statistical analysis. We tested the effect  
438 of day, treatment, and the interaction of the two using linear mixed effects models, with day and  
439 treatment as fixed effects and replicate bucket as a random effect (*lmer*). Effects of treatment and  
440 size class on log-transformed shell length from size-separated larvae were also analyzed using a  
441 linear-mixed effect model. Size class, treatment, and their interaction were fixed effects, while  
442 larval bucket was a random effect.

#### 443 *DNA extraction and exome sequencing*

444 We implemented exome capture, a reduced-representation sequencing approach, to  
445 identify SNPs and their frequency dynamics throughout the course of the experiment. Exome  
446 capture targets the protein-coding region of the genome, and thus increases the likelihood that  
447 identified polymorphisms have functional consequences<sup>62</sup>. Genomic DNA from each founding  
448 individual and larval sample was extracted using the EZNA Mollusc Extraction Kit, according to  
449 manufacturer's protocol. DNA was quantified with a Qubit, and quality was determined using  
450 agarose gel, Nanodrop (260/280), and TapeStation analysis.

451 Genomic DNA was hybridized to a customized exome capture array designed and  
452 manufactured by Arbor Biosciences (Ann Arbor, Michigan) and using the species transcriptome  
453 provided in Moreira et al. (2015). Specifically, in order to design a bait set appropriate for  
454 capture of genomic DNA fragments, 90-nucleotide probe candidates were tiled every 20  
455 nucleotides across the target transcriptome contigs. These densely-tiled candidates were  
456 MEGABLASTed to the *Mytilus galloprovincialis* draft genome contigs available at NCBI

457 (GCAA\_001676915.1\_ASM167691v1\_genomic.fna), which winnowed the candidate list to only  
458 baits with detected hits of 80 nucleotides or longer. After predicting the hybrid melting  
459 temperatures for each near-full-length hit, baits were further winnowed to those with at most two  
460 hybrids of 60°C or greater estimated melting temperature in the *M. galloprovincialis* genome.  
461 This collection of highly specific baits with near-full-length hits to the draft genome were then  
462 down-sampled to a density of roughly one bait per 1.9 kbp of the final potential target space, in  
463 order to broadly sample the target while still fitting within our desired number of myBaits kit  
464 oligo limit. The final bait set comprises 100,087 oligo sequences, targeting 94,668 of the original  
465 121,572 transcript contigs.

466       Genomic DNA from each sample was subject to standard mass estimation quality  
467 control, followed by sonication using a QSonica QR800 instrument and SPRI-based dual size-  
468 selection to a target modal fragment length of 350 nucleotides. Following quantification, 300 ng  
469 total genomic DNA was taken to library preparation using standard Illumina Truseq-style end  
470 repair and adapter ligation chemistry, followed by six cycles of indexing amplification using  
471 unique eight nucleotide dual index primer pairs. For target enrichment with the custom myBaits  
472 kit, 100 ng of each founder-derived library were combined into two pools of 14 libraries each,  
473 whereas 450 ng of each embryonic and larval-pool derived library were used in individual  
474 reactions. After drying the pools or individual samples using vacuum centrifugation to 7 µL  
475 each, Arbor followed the myBaits procedure (v. 4) using the default conditions and overnight  
476 incubation to enrich the libraries using the custom probe set. After reaction cleanup, half (15 µL)  
477 of each bead-bound enriched library was taken to standard library amplification for 10 cycles  
478 using Kapa HiFi polymerase. Following reaction cleanup with SPRI, each enriched library or  
479 library pool was quantified using qPCR, indicating yields between 30 and 254 ng each.

480           The captured libraries were sequenced at the University of Chicago Genomics Core  
481 Facility on three lanes of Illumina HiSeq 4000 using 150-bp, paired-end reads. The captured  
482 adult libraries were sequenced on an individual lane, while the twenty-two, pooled larval  
483 samples were split randomly between the remaining two lanes. Average coverage for founding  
484 individuals was 40x, while average coverage in pooled samples was 100x.

#### 485 *Read Trimming and Variant Calling*

486           Raw DNA reads were filtered and trimmed using Trimmomatic<sup>63</sup> and aligned to the  
487 species reference transcriptome provided in Moreira *et al.* (2015) using bowtie2<sup>64</sup>. Variants in  
488 the founding individuals were identified using the Genome Analysis Toolkit's<sup>65</sup> Unified  
489 Genotyper. These variants were filtered using VCFTOOLS<sup>66</sup> with the following specifications:  
490 Minor Allele Frequency of 0.05, Minimum Depth of 10x, and a Maximum Variant Missing of  
491 0.75. The resulting .vcf files provided a list of candidate bi-allelic polymorphisms to track at  
492 each time point, treatment, and phenotypic group in the larval samples. Accordingly, GATK's  
493 Haplotype Caller was used to identify these candidate polymorphisms within each larval  
494 alignment file, and the resulting .vcf was filtered using VCFTOOLS and the following  
495 specifications: Minor Allele Frequency of 0.01, Minimum Depth of 50x, and Maximum Depth of  
496 450x. Only variants that passed quality filtering and were identified in all larval samples (i.e.  
497 each day, treatment, and phenotypic group) were retained for downstream analyses. This process  
498 resulted in a candidate SNP list of 29,400 variants. Allele frequencies for each variant were  
499 computed as the alternate allelic depth divided by total coverage at the locus.

#### 500 *Allele Frequency Analysis*

501           To explore how the allele frequency of the 29,400 SNPs changed in each environment  
502 throughout the course of the experiment, we used a combination of PCA, outlier loci

503 identification tests, and a statistical test of genomic differentiation ( $F_{ST}$ ). We visualized patterns  
504 of genetic variation throughout the experiment with PCA (*prcomp* function in R). Prior to PCA,  
505 the allele frequency matrix (the generation of which is provided in the previous section) was  
506 centered and scaled using the *scale* function in R. Only larval samples that encompassed the full  
507 phenotypic distribution within a particular bucket were included in this analysis. In other words,  
508 the rows of the allele frequency matrix corresponding to larval samples that were selectively  
509 segregated based on shell size were removed, and PCA was run using the day 0 larval population  
510 and larval samples collected from each treatment on days 6, 26, and 43. A separate PCA was  
511 then implemented using data from the day 0 larval population and all day 6 larval samples,  
512 which included discrete size groups from each treatment. This analysis thus explicitly examined  
513 a genomic signature of the individuals that were phenotypically distinct.

514 We next sought to identify the presence, number, and treatment-level overlap of genetic  
515 variants that significantly changed in frequency between larval samples. Specifically, Fisher's  
516 Exact test (FET) and the Cochran-Mantel-Haenszel (CMH) test were used to generate  
517 probabilities of observed allele frequency changes, using the package *Popoolation*<sup>67</sup> in R. P-  
518 values for each SNP were converted to q-values in the R package *qvalue*<sup>68</sup>, and outliers were  
519 identified as those SNPs with a q-value <0.01. The FET was used to identify outliers between the  
520 day 0 larval population and the day 6 larval populations in each treatment (no treatment  
521 replicates were available for this comparison). For all remaining allele frequency comparisons  
522 (in which replicate bucket information was available), the CMH test was used to identify  
523 consistent allele frequency shifts among replicate buckets. We used this test to identify  
524 significant allele frequency changes between the day 0 larval population and the day 26  
525 treatment replicates (N = 3), the day 0 larval population and the settled individuals treatment



526 replicates (ambient pH treatment: N = 2, low pH treatment: N = 3), and between the top 20% and  
527 bottom 80% of growers in each treatment (N = 2).

528 To provide a third, independent metric of genomic change in the larval population  
529 throughout the experiment, we computed the  $F_{ST}$  statistic for a series of comparisons.  
530 Specifically, we implemented a methods-of-moments estimator of  $F_{ST}$  from Pool-seq data in an  
531 analysis of variance framework, as described in Hivert et al. (2018)<sup>69</sup> (*poolFstat* package). A  
532 global (exome-wide)  $F_{ST}$  statistic was computed pairwise between the day 0 larval population  
533 and the day 6 ambient and low pH larvae replicate buckets, day 26 ambient and low pH larvae  
534 replicate buckets, and settled individuals from all replicate buckets in ambient and low pH.  $F_{ST}$   
535 was also computed to compare differentiation between phenotypic groups (top 20% and bottom  
536 80% of growers) on day 6. Estimates of pool sizes were based on the known day 0 larval  
537 population size, the known volume of larvae sampled, and the observed declines in larval culture  
538 density. Input pool sizes were as follows: 100,000 for day 0 larval population; 25,000 for  
539 ambient and low pH larval samples on day 6; 5,250 for top 20% of growers in ambient  
540 conditions on day 6; 4,250 for top 20% of growers in low pH conditions on day 6; 19,750 for  
541 bottom 80% of growers in ambient conditions 20,750 for bottom 80% of growers in low pH  
542 conditions on day 6; 2,500 larval samples on day 26; 250 for settled individuals collected on day  
543 43. Equal densities within each treatment were assumed as no discernable difference in mortality  
544 between treatments was observed (MCB, *pers. obs.*).

#### 545 *Gene Identification/Ontologies*

546 We next sought to explore the biological pathways that were associated with survivorship  
547 in each pH treatment and/or size group during the experiment. To accomplish this, we indexed  
548 the outlier loci that contained variants with significant frequency changes using the annotated

549 transcriptome provided in Moreira *et al.* (2015). Their annotation utilized NCBI's nucleotide and  
550 non-redundant, Swissprot, KEGG, and COG databases, thus providing a thorough survey of  
551 potential genes and pathways associated with our candidate SNPs. We generated gene lists for  
552 pH-specific loci, which were identified as loci that showed signatures of selection on all  
553 sampling days and were unique to each environment. We also generated a candidate gene list for  
554 loci that exhibited shared signatures of selection in each treatment. These lists thus only contain  
555 robust candidate loci (loci identified as outliers in multiple independent replicates), with  
556 potentially strong effect sizes (loci identified as outliers at multiple developmental stages).  
557 Lastly, we used the Moreira *et al.* (2015) annotation to explore the genes that exhibited  
558 signatures of selection for shell growth in ambient and low pH conditions, as well as shared  
559 signatures of selection for shell size in each treatment.

560

561 **Acknowledgements:** We thank Samir Alliouane for extensive technical assistance during the  
562 completion of the experiment. We would also like to thank Angelica Miglioli for experimental  
563 assistance, Régis Lasbleiz for microalgal supply for larval feeding, Jacob Enk at Arbor  
564 Biosciences for guidance during exome capture, and the University of Chicago Genomics Core  
565 Facility for sequencing assistance. Lastly, we thank D. Rice and T. Price for insightful comments  
566 on the manuscript.

567 **Funding:** This research was supported by the National Science Foundation Graduate Research  
568 Fellowship Program under Grant No. 1746045 to MCB and NSF OCE-1521597 to LK. MCB  
569 was supported by Department of Education Grant No. 200A150101. LK was also supported by  
570 the European Commission Horizon 2020 Marie Skłodowska-Curie Action (No. 747637).

571 Research funding was provided by the France and University of Chicago Center FAACTs award  
572 to CAP and MCB.

573 **Author Contributions:** MCB conceived and designed the experiment with inputs from LK,  
574 JPG, and CAP. MCB and LK performed the experiment. MCB completed molecular lab work,  
575 with exome capture and sequencing assistance from Arbor Biosciences and the University of  
576 Chicago Genomics Core. MCB completed all bioinformatics, statistical, and computational  
577 analyses. MCB wrote the manuscript with inputs from LK, JPG, and CAP.

578  
579 **References**

- 580 1. Umina, P. A., Weeks, A. R., Kearney, M. R., McKechnie, S. W. & Hoffmann, A. A. A Rapid  
581 Shift in a Classic Clinal Pattern in *Drosophila* Reflecting Climate Change. *Science* **308**,  
582 691–693 (2005).
- 583 2. Rodríguez-Trelles, F., Tarrío, R. & Santos, M. Genome-wide evolutionary response to a  
584 heat wave in *Drosophila*. *Biol. Lett.* **9**, 20130228 (2013).
- 585 3. Bergland, A. O., Behrman, E. L., O'Brien, K. R., Schmidt, P. S. & Petrov, D. A. Genomic  
586 Evidence of Rapid and Stable Adaptive Oscillations over Seasonal Time Scales in  
587 *Drosophila*. *PLOS Genet.* **10**, e1004775 (2014).
- 588 4. Campbell-Staton, S. C. *et al.* Winter storms drive rapid phenotypic, regulatory, and  
589 genomic shifts in the green anole lizard. *Science* **357**, 495–498 (2017).
- 590 5. Barrett, R. D. H. & Schluter, D. Adaptation from standing genetic variation. *Trends Ecol.*  
591 *Evol.* **23**, 38–44 (2008).
- 592 6. Orr, H. A. The genetic theory of adaptation: a brief history. *Nat. Rev. Genet.* **6**, 119–127  
593 (2005).

- 594 7. Messer, P. W., Ellner, S. P. & Hairston, N. G. Can Population Genetics Adapt to Rapid  
595 Evolution? *Trends Genet.* **32**, 408–418 (2016).
- 596 8. Via, S. & Lande, R. Genotype-Environment Interaction and the Evolution of Phenotypic  
597 Plasticity. *Evolution* **39**, 505–522 (1985).
- 598 9. Hermisson, J. & Wagner, G. P. The Population Genetic Theory of Hidden Variation and  
599 Genetic Robustness. *Genetics* **168**, 2271–2284 (2004).
- 600 10. Hartl, D. L. & Dykhuizen, D. E. Potential for selection among nearly neutral allozymes  
601 of 6-phosphogluconate dehydrogenase in *Escherichia coli*. *Proc. Natl. Acad. Sci.* **78**,  
602 6344–6348 (1981).
- 603 11. Vieira, C. *et al.* Genotype-Environment Interaction for Quantitative Trait Loci  
604 Affecting Life Span in *Drosophila melanogaster*. *Genetics* **154**, 213–227 (2000).
- 605 12. McGuigan, K., Nishimura, N., Currey, M., Hurwit, D. & Cresko, W. A. Cryptic Genetic  
606 Variation and Body Size Evolution in Threespine Stickleback. *Evolution* **65**, 1203–1211  
607 (2011).
- 608 13. Hayden, E. J., Ferrada, E. & Wagner, A. Cryptic genetic variation promotes rapid  
609 evolutionary adaptation in an RNA enzyme. *Nature* **474**, 92–95 (2011).
- 610 14. Paaby, A. B. & Rockman, M. V. Cryptic genetic variation: evolution’s hidden  
611 substrate. *Nat. Rev. Genet.* **15**, 247–258 (2014).
- 612 15. Gibson, G. & Dworkin, I. Uncovering cryptic genetic variation. *Nat. Rev. Genet.* **5**,  
613 681–690 (2004).
- 614 16. Masel, J. Cryptic Genetic Variation Is Enriched for Potential Adaptations. *Genetics*  
615 **172**, 1985–1991 (2006).

- 616 17. Dykhuizen, D. & Hartl, D. L. Selective neutrality of 6PGD allozymes in *E. coli* and the  
617 effects of genetic background. *Genetics* **96**, 801–817 (1980).
- 618 18. Sangster, T. A. *et al.* HSP90-buffered genetic variation is common in *Arabidopsis*  
619 *thaliana*. *Proc. Natl. Acad. Sci.* **105**, 2969–2974 (2008).
- 620 19. Queitsch, C., Sangster, T. A. & Lindquist, S. Hsp90 as a capacitor of phenotypic  
621 variation. *Nature* **417**, 618–624 (2002).
- 622 20. Grant, P. R. *et al.* Evolution caused by extreme events. *Phil Trans R Soc B* **372**,  
623 20160146 (2017).
- 624 21. Hoffmann, A. A. & Sgrò, C. M. Climate change and evolutionary adaptation. *Nature*  
625 **470**, 479–485 (2011).
- 626 22. Hoegh-Guldberg, O. *et al.* In: *Climate Change 2014: Impacts, Adaptation, and*  
627 *Vulnerability. Part B: Regional Aspects. Contribution of Working Group II to the Fifth*  
628 *Assessment Report of the Intergovernmental Panel on Climate Change.* (Cambridge  
629 University Press, Cambridge, United Kingdom and New York, NY, USA).
- 630 23. Hönisch, B. *et al.* The Geological Record of Ocean Acidification. *Science* **335**, 1058–  
631 1063 (2012).
- 632 24. Kroeker, K. J. *et al.* Impacts of ocean acidification on marine organisms: quantifying  
633 sensitivities and interaction with warming. *Glob. Change Biol.* **19**, 1884–1896 (2013).
- 634 25. Gazeau, F. *et al.* Impacts of ocean acidification on marine shelled molluscs. *Mar. Biol.*  
635 **160**, 2207–2245 (2013).
- 636 26. Thomsen, J., Haynert, K., Wegner, K. M. & Melzner, F. Impact of seawater carbonate  
637 chemistry on the calcification of marine bivalves. *Biogeosciences* **12**, 4209–4220 (2015).

- 638 27. Kurihara, H. Effects of CO<sub>2</sub>-driven ocean acidification on the early developmental  
639 stages of invertebrates. *Mar. Ecol. Prog. Ser.* **373**, 275–284 (2008).
- 640 28. Ventura, A., Schulz, S. & Dupont, S. Maintained larval growth in mussel larvae  
641 exposed to acidified under-saturated seawater. *Sci. Rep.* **6**, 23728 (2016).
- 642 29. Kapsenberg L. *et al.* Ocean pH fluctuations affect mussel larvae at key developmental  
643 transitions. *Proc. R. Soc. B Biol. Sci.* **285**, 20182381 (2018).
- 644 30. Moreira, R. *et al.* RNA-Seq in *Mytilus galloprovincialis*: comparative transcriptomics  
645 and expression profiles among different tissues. *BMC Genomics* **16**, 728 (2015).
- 646 31. Waldbusser, G. G. *et al.* A developmental and energetic basis linking larval oyster  
647 shell formation to acidification sensitivity. *Geophys. Res. Lett.* **40**, 2171–2176 (2013).
- 648 32. Waldbusser, G. G. *et al.* Saturation-state sensitivity of marine bivalve larvae to ocean  
649 acidification. *Nat. Clim. Change* **5**, 273–280 (2015).
- 650 33. Melzner, F., Thomsen, J., Ramesh, K., Hu, M. Y. & Bleich, M. Mussel larvae modify  
651 calcifying fluid carbonate chemistry to promote calcification. *Nat. Commun.* **8**, 1709  
652 (2017).
- 653 34. Launey, S. & Hedgecock, D. High Genetic Load in the Pacific Oyster *Crassostrea gigas*.  
654 *Genetics* **159**, 255–265 (2001).
- 655 35. Pansch, C., Schaub, I., Havenhand, J. & Wahl, M. Habitat traits and food availability  
656 determine the response of marine invertebrates to ocean acidification. *Glob. Change Biol.*  
657 **20**, 765–777 (2014).
- 658 36. Thomsen, J., Casties, I., Pansch, C., Körtzinger, A. & Melzner, F. Food availability  
659 outweighs ocean acidification effects in juvenile *Mytilus edulis*: laboratory and field  
660 experiments. *Glob. Change Biol.* **19**, 1017–1027 (2013).

- 661 37. Phillips, N. E. Effects of Nutrition-Mediated Larval Condition on Juvenile  
662 Performance in a Marine Mussel. *Ecology* **83**, 2562–2574 (2002).
- 663 38. Pettersen, A. K., Turchini, G. M., Jahangard, S., Ingram, B. A. & Sherman, C. D. H.  
664 Effects of different dietary microalgae on survival, growth, settlement and fatty acid  
665 composition of blue mussel (*Mytilus galloprovincialis*) larvae. *Aquaculture* **309**, 115–124  
666 (2010).
- 667 39. Carl, C., Poole, A. J., Williams, M. R. & Nys, R. de. Where to Settle—Settlement  
668 Preferences of *Mytilus galloprovincialis* and Choice of Habitat at a Micro Spatial Scale.  
669 *PLOS ONE* **7**, e52358 (2012).
- 670 40. Kelly, M. W., Padilla-Gamiño, J. L. & Hofmann, G. E. Natural variation and the capacity  
671 to adapt to ocean acidification in the keystone sea urchin *Strongylocentrotus*  
672 *purpuratus*. *Glob. Change Biol.* **19**, 2536–2546 (2013).
- 673 41. Pespeni, M. H. *et al.* Evolutionary change during experimental ocean acidification.  
674 *Proc. Natl. Acad. Sci.* **110**, 6937–6942 (2013).
- 675 42. Thomsen, J. *et al.* Naturally acidified habitat selects for ocean acidification-tolerant  
676 mussels. *Sci. Adv.* **3**, e1602411 (2017).
- 677 43. Allen, J. D. Size-Specific Predation on Marine Invertebrate Larvae. *Biol. Bull.* **214**, 42–  
678 49 (2008).
- 679 44. Le Rouzic, A. & Carlborg, Ö. Evolutionary potential of hidden genetic variation.  
680 *Trends Ecol. Evol.* **23**, 33–37 (2008).
- 681 45. McGuigan, K. & Sgrò, C. M. Evolutionary consequences of cryptic genetic variation.  
682 *Trends Ecol. Evol.* **24**, 305–311 (2009).

- 683 46. Turner, J. R. G. & Williamson, M. H. Population Size, Natural Selection and the  
684 Genetic Load. *Nature* **218**, 700 (1968).
- 685 47. Romiguier, J. *et al.* Comparative population genomics in animals uncovers the  
686 determinants of genetic diversity. *Nature* **515**, 261–263 (2014).
- 687 48. Feder, M. E. & Hofmann, G. E. HEAT-SHOCK PROTEINS, MOLECULAR CHAPERONES,  
688 AND THE STRESS RESPONSE: Evolutionary and Ecological Physiology. *Annu. Rev. Physiol.*  
689 **61**, 243–282 (1999).
- 690 49. Moya, A. *et al.* Rapid acclimation of juvenile corals to CO<sub>2</sub>-mediated acidification by  
691 upregulation of heat shock protein and Bcl-2 genes. *Mol. Ecol.* **24**, 438–452 (2015).
- 692 50. Cummings, V. *et al.* Ocean Acidification at High Latitudes: Potential Effects on  
693 Functioning of the Antarctic Bivalve *Laternula elliptica*. *PLOS ONE* **6**, e16069 (2011).
- 694 51. Huan, P., Liu, G., Wang, H. & Liu, B. Identification of a tyrosinase gene potentially  
695 involved in early larval shell biogenesis of the Pacific oyster *Crassostrea gigas*. *Dev. Genes*  
696 *Evol.* **223**, 389–394 (2013).
- 697 52. Yang, B. *et al.* Functional analysis of a tyrosinase gene involved in early larval shell  
698 biogenesis in *Crassostrea angulata* and its response to ocean acidification. *Comp.*  
699 *Biochem. Physiol. B Biochem. Mol. Biol.* **206**, 8–15 (2017).
- 700 53. Hüning, A. K. *et al.* Impacts of seawater acidification on mantle gene expression  
701 patterns of the Baltic Sea blue mussel: implications for shell formation and energy  
702 metabolism. *Mar. Biol.* **160**, 1845–1861 (2012).
- 703 54. Lewontin, R. C. & Hubby, J. L. A Molecular Approach to the Study of Genic  
704 Heterozygosity in Natural Populations. II. Amount of Variation and Degree of

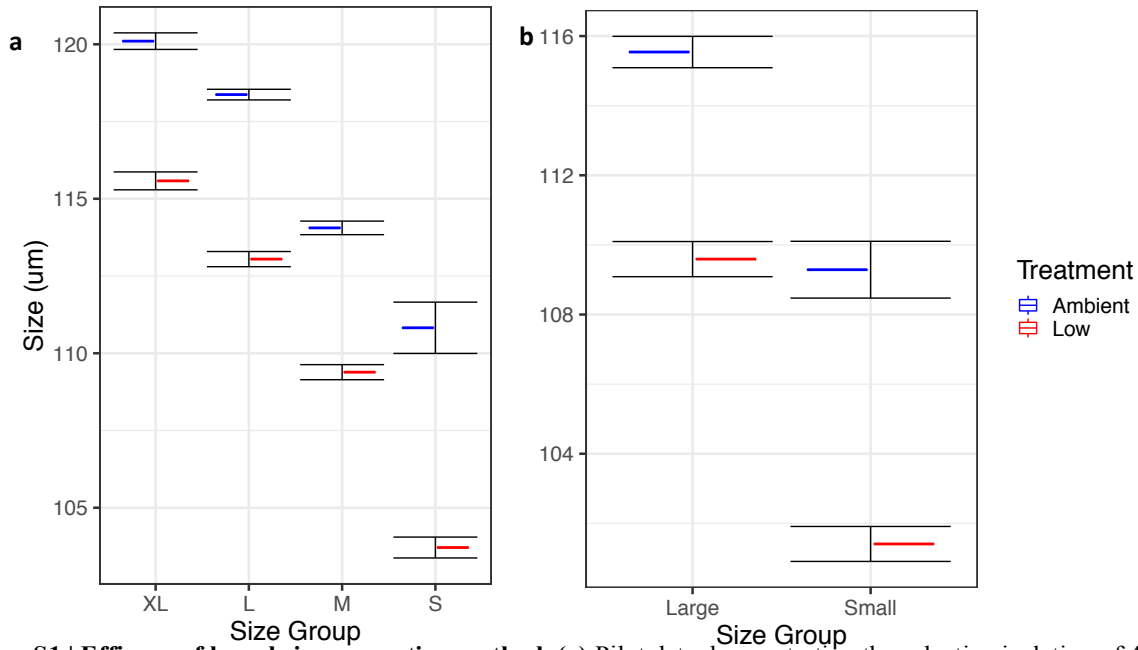


- 705 Heterozygosity in Natural Populations of DROSOPHILA PSEUDOOBSCURA. *Genetics* **54**,  
706 595–609 (1966).
- 707 55. Kapsenberg, L. *et al.* Advancing Ocean Acidification Biology Using Durafet® pH  
708 Electrodes. *Front. Mar. Sci.* **4**, (2017).
- 709 56. Dickson, A. G., Sabine, C. L. & Christian, J. R. *Guide to Best Practices for Ocean CO2*  
710 *Measurements*. (PICES Special Publication, 2007).
- 711 57. Gattuso, J.-P., Epitalon, J.-M. & Lavigne, H. seacarb: Seawater Carbonate Chemistry. R  
712 package version 3.2.2.
- 713 58. Lueker, T. J., Dickson, A. G. & Keeling, C. D. Ocean pCO<sub>2</sub> calculated from dissolved  
714 inorganic carbon, alkalinity, and equations for K<sub>1</sub> and K<sub>2</sub>: validation based on laboratory  
715 measurements of CO<sub>2</sub> in gas and seawater at equilibrium. *Mar. Chem.* **70**, 105–119  
716 (2000).
- 717 59. Perez, F. F. & Fraga, F. The pH measurements in seawater on the NBS scale. *Mar.*  
718 *Chem.* **21**, 315–327 (1987).
- 719 60. Dickson, A. G. Standard potential of the reaction: AgCl(s) + 12H<sub>2</sub>(g) = Ag(s) +  
720 HCl(aq), and the standard acidity constant of the ion HSO<sub>4</sub><sup>-</sup> in synthetic sea water  
721 from 273.15 to 318.15 K. *J. Chem. Thermodyn.* **22**, 113–127 (1990).
- 722 61. R Core Team. R: a language and environment for statistical computing.
- 723 62. Cosart, T. *et al.* Exome-wide DNA capture and next generation sequencing in  
724 domestic and wild species. *BMC Genomics* **12**, 347 (2011).
- 725 63. Bolger, A. M., Lohse, M. & Usadel, B. Trimmomatic: a flexible trimmer for Illumina  
726 sequence data. *Bioinforma. Oxf. Engl.* **30**, 2114–2120 (2014).

- 727 64. Langmead, B. & Salzberg, S. L. Fast gapped-read alignment with Bowtie 2. *Nat.*  
728 *Methods* **9**, 357–359 (2012).
- 729 65. Van der Auwera, G. A. *et al.* From FastQ data to high confidence variant calls: the  
730 Genome Analysis Toolkit best practices pipeline. *Curr. Protoc. Bioinforma.* **43**, 11.10.1-33  
731 (2013).
- 732 66. Danecek, P. *et al.* The variant call format and VCFtools. *Bioinformatics* **27**, 2156–  
733 2158 (2011).
- 734 67. Kofler, R. *et al.* PoPoolation: A Toolbox for Population Genetic Analysis of Next  
735 Generation Sequencing Data from Pooled Individuals. *PLOS ONE* **6**, e15925 (2011).
- 736 68. Storey, J. D., Bass, A. J., Dabney, A. & Robinson, D. qvalue: Q-value estimation for false  
737 discovery rate control. (2019).
- 738 69. Hivert, V., Leblois, R., Petit, E. J., Gautier, M. & Vitalis, R. Measuring Genetic  
739 Differentiation from Pool-seq Data. *Genetics* **210**, 315–330 (2018).
- 740  
741  
742  
743  
744  
745  
746  
747  
748  
749  
750  
751  
752  
753  
754  
755  
756  
757

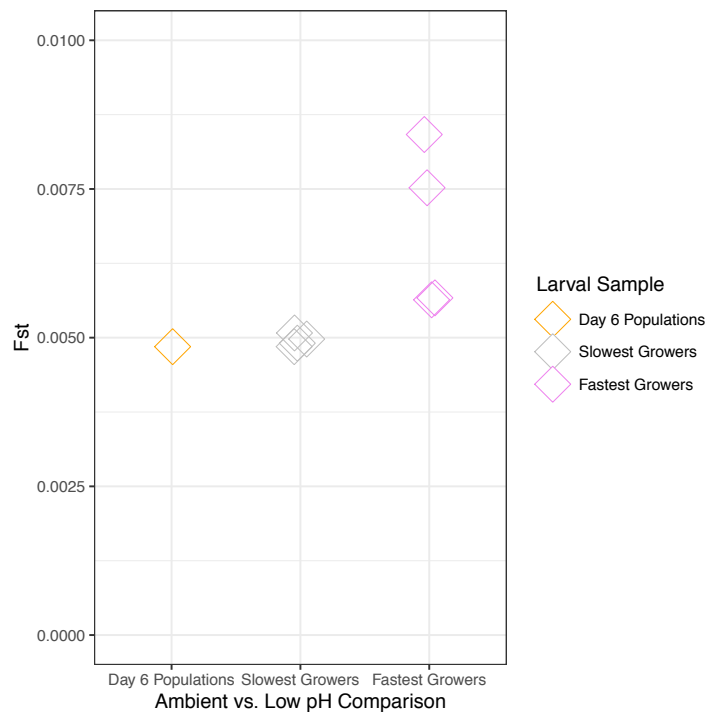
758  
759

### Supplemental Information



**Figure S1 | Efficacy of larval size separation method. (a)** Pilot data demonstrating the selective isolation of 4 consecutively smaller groups of larvae (mean shell length +/- SE) in ambient and low pH cultured larvae ( $N = 53-324$  per size group; XL = extra large; L = large; M = medium; S = small). **(b)** Shell length of larvae (mean +/- SE) in largest (top 20 %) and smallest (bottom 80 %) individuals in ambient and low pH conditions from present study ( $N = 62 - 177$  per size group).

760  
761  
762  
763  
764  
765  
766  
767  
768  
769  
770  
771  
772  
773  
774  
775  
776  
777  
778



**Figure S2 | Genomic differentiation of fastest and slowest growers in ambient and low pH.** Exome wide  $F_{ST}$  computed pairwise between ambient and low pH replicate buckets for the entire larval population and the size selected larvae isolated on day 6 (Day 6 Populations:  $N=1$ , Slowest and Fastest Growers comparisons:  $N = 4$  pairwise comparisons).

779  
780  
781

782 **Table S1.** Carbonate chemistry for flow-through experimental system used on days 0-26, during which each of four  
 783 header tanks (two per treatment) each distributed pH adjusted seawater to three replicate buckets. Low/Amb\_1/2  
 784 correspond to treatment replicates drawing water from separate header tanks, thus one replicate bucket per header  
 785 tank is represented in the table. Time-series pH and temperature data were generated using in autonomous sensor  
 786 were generated in each representative replicate bucket, and average values (+/- SD) are presented. Aragonite  
 787 saturation ( $\Omega_a$ ) and  $p\text{CO}_2$  were computed using average pH and AT for each representative replicate. Alkalinity (AT)  
 788 and salinity samples were generated from discrete samples taken from each of the four header tanks every other day.  
 789

pH treatment	pH <sub>T</sub>	$\Omega_a$	$p\text{CO}_2$ ( $\mu\text{atm}$ )	AT ( $\mu\text{mol/kg}$ )	T ( $^{\circ}\text{C}$ )	SAL (ppt)
Low_1	7.43 (+/- 0.03)	0.85	2127	2565 (n = 11)	17.3 (+/-0.1)	37.6 (+/-0.1) (n = 11)
Low_2	7.43 (+/- 0.03)	0.85	2130	2569 (n = 11)	17.2 (+/-0.1)	37.6 (+/-0.1) (n = 11)
Amb_1	8.01 (+/- 0.01)	3.32	378	2565 (n = 11)	17.2 (+/-0.1)	37.6 (+/-0.1) (n = 11)
Amb_2	8.01 (+/- 0.01)	3.31	378	2561 (n = 11)	17.2 (+/-0.1)	37.6 (+/-0.1) (n = 11)

790  
 791  
 792

793 **Table S2.** Carbonate chemistry generated from static cultures used to rear larvae from days 27-43. Low/Ambient\_#  
 794 correspond to each remaining replicate bucket during this portion of the experiment. Discrete measurements of pH,  
 795 total alkalinity, temperature, and salinity were taken daily. Aragonite saturation ( $\Omega_a$ ) and  $p\text{CO}_2$  were computed using  
 796 average pH and AT for each replicate.

pH Treatment	pH <sub>T</sub>	$\Omega_a$	$p\text{CO}_2$ ( $\mu\text{atm}$ )	AT ( $\mu\text{mol/kg}$ )	T ( $^{\circ}\text{C}$ )	SAL (ppt)
Low_1	7.53 (+/-0.10)	1.1	1675	2576 (+/-13)	17.3(+/-0.2)	37.4(+/-0.2)
Low_2	7.53 (+/-0.09)	1.0	1716	2576 (+/-13)	17.3(+/-0.2)	37.4(+/-0.2)
Low_3	7.54 (+/-0.10)	1.1	1633	2576 (+/-13)	17.3(+/-0.2)	37.4(+/-0.2)
Ambient_1	8.08 (+/-0.10)	3.2	403	2579 (+/-13)	17.3(+/-0.2)	37.4(+/-0.2)
Ambient_2	8.01 (+/-0.07)	2.8	488	2579 (+/-13)	17.3(+/-0.2)	37.4(+/-0.2)
Ambient_3	8.00 (+/-0.08)	2.8	501	2579 (+/-13)	17.3(+/-0.2)	37.4(+/-0.2)

797  
 798  
 799  
 800  
 801  
 802  
 803  
 804  
 805  
 806