

# 1           **Cooperation and spatial self-organization determine ecosystem** 2                           **function for polysaccharide-degrading bacteria**

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10

## 11   **Abstract**

12    The recycling of particulate organic matter (POM) by microbes is a key part of the global carbon  
13    cycle, one which is mediated by the extracellular hydrolysis of polysaccharides and the production  
14    of public goods that can trigger social behaviors in bacteria. Despite the potential importance of  
15    these microbial interactions, their role in regulating of ecosystem function remains unclear. In this  
16    study, we developed a computational and experimental model system to address this challenge  
17    and studied how POM depolymerization rate and its uptake efficiency –two main ecosystem  
18    function parameters– depended on social interactions and spatial self-organization on particle  
19    surfaces. We found an emergent trade-off between rate and efficiency resulting from the  
20    competition between oligosaccharide diffusion and cellular uptake, with low rate and high  
21    efficiency being achieved through cell-to-cell cooperation between degraders. Bacteria  
22    cooperated by aggregating in cell-clusters of ~10-20µm, where cells were able to share public  
23    goods. This phenomenon, which was independent of any explicit group-level regulation, led to  
24    the emergence of critical cell concentrations below which degradation did not occur, despite all  
25    resources being available in excess. By contrast, when particles were labile and turnover rates  
26    were high, aggregation promoted competition and decreased the efficiency of carbon utilization.

27 Our study shows how social interactions and cell aggregation determine the rate and efficiency  
28 of particulate carbon turnover in environmentally relevant scenarios.

29

30 **Keywords:** marine microbes, cooperation, particulate organic matter, public goods, spatial  
31 organization

32

### 33 **Significance Statement**

34 Microorganisms can cooperate by secreting public goods that benefit local neighbors, however,  
35 the impact of cooperation on ecosystem functions remains poorly constrained. We here pair  
36 computation and experiment to show that bacterial cooperation mediates the degradation of  
37 polysaccharide particles recalcitrant to hydrolysis in aquatic environments. On particle surfaces,  
38 cooperation emerges through the self-organization of cells into ~10-20 $\mu$ m clusters that promote  
39 cooperative uptake of hydrolysis products. The transition between cooperation and competition  
40 in aggregates is mitigated by individual cell behaviors such as motility and chemotaxis, that  
41 promote reorganization on the particle surface. When cooperation is required, the degradation of  
42 recalcitrant biopolymers can only take place when degraders exceed a critical cell concentration,  
43 underscoring the importance of microbial interactions for ecosystem function.

44

### 45 **Introduction**

46 The microbial breakdown of complex polysaccharides is a key ecosystem process that enables  
47 the recycling of carbon from plant and animal detritus into global biogeochemical cycles and is a  
48 relevant process in all heterotrophic microbial ecosystems, from animal guts (1–3) to soils (4, 5)  
49 and oceans (6–8). A key feature of these polysaccharides is their insoluble nature: a large fraction  
50 is found in particles at the scale of 100  $\mu$ m, that require both surface colonization and extracellular  
51 hydrolysis to be degraded (9, 10). On particle surfaces cells can attach and grow in close  
52 proximity, increasing the opportunity for microbial interactions to impact the ecosystem process.

53 One particularly relevant type of interaction in this context is cell-cell cooperation mediated by the  
54 sharing of public goods such extracellular enzymes and hydrolysis products (11). However, the  
55 extent to which these interactions take place and impact bacterial growth in the environment  
56 remains unclear. Previous work on cooperative interactions has largely focused on the opportunity  
57 that public goods open for exploitative populations to invade (so-called cheaters), and less on the  
58 environmental and physiological conditions that enable cooperation to take place or the potential  
59 role of cooperative behavior on ecosystem processes. From the perspective of ecosystem  
60 modeling, efforts to incorporate the role of microbes in organic matter degradation equate  
61 microbial activity with enzymatic activity without considering the role of population-level  
62 phenomena such as cooperation. In this paper we reveal how microbial social interactions can  
63 impact relevant ecosystem parameters.

64 The extent to which social interactions mediated by public goods play a relevant role in  
65 ecosystem function is highly dependent on how public goods diffuse (12, 13). In a three-  
66 dimensional aqueous environment like the ocean, if cells are too far apart only a minuscule  
67 fraction of the public goods are recovered by neighbors, while the rest is lost to the environment.  
68 In contrast, if cells are sufficiently proximal to each other and the resource is limiting, growth  
69 kinetics can be cooperative, meaning that the per-capita growth rate is positively dependent on  
70 the density of degrader cells (14). This logic suggests that cooperation should be accompanied  
71 by the emergence of spatial patterns, such as cell patches. If the cooperative effects in these  
72 patches are strong, critical population density thresholds might emerge below which degradation  
73 cannot support population growth (14, 15). Less recognized is the contribution of individual cell  
74 behaviors, such as surface attachment, chemotaxis, and biofilm formation, on the ability of cells  
75 to find those critical densities by aggregating into cell patches. Therefore, in order to begin to  
76 understand the role of social interactions in natural systems, we need to take into account the  
77 physical constraints of the micro-environment and how populations interact with these constraints  
78 through their behavior.

79 To quantify the impact of bacterial social interactions and spatial behavior on ecosystem  
80 function, we focus on two main parameters: the speed at which polymers are hydrolyzed and  
81 converted to soluble oligosaccharides, that is, the turnover rate (16, 17), and the POM uptake  
82 efficiency, which is the fraction of the dissolved oligosaccharide that can be taken up by cells and  
83 converted into biomass. To study the role of social interactions and spatial behavior on ecosystem  
84 function, we developed a computational and experimental model of the colonization of insoluble  
85 particulate polysaccharides by marine heterotrophic bacteria. The individual-based model (18,  
86 19) simulates the functional traits of individual cells: chemotactic movement, particle attachment  
87 and detachment, the secretion of enzymes, oligosaccharide uptake and growth. The experimental  
88 system validates computational predictions in a chitin-degrading bacterial strain isolated from the  
89 coastal ocean, and clarifies the role of physiological parameters on social interactions (17). We  
90 leveraged the computational model to study the relationship between degradation rate and POM  
91 uptake efficiency and how emergent bacterial behaviors influence their ability to degrade  
92 recalcitrant particles, and we tested some of our predictions using our experimental model of  
93 chitin colonization. Our work demonstrates that cell-cell cooperation is critical for the degradation  
94 of complex biomaterials, implying that the degradation of recalcitrant polysaccharides can be  
95 bacteria-limited. Moreover, cell-density thresholds that determine the onset of cooperative growth  
96 depend strongly on individual cell behavior, in particular those behaviors that regulate the  
97 residence time of bacteria on particles.

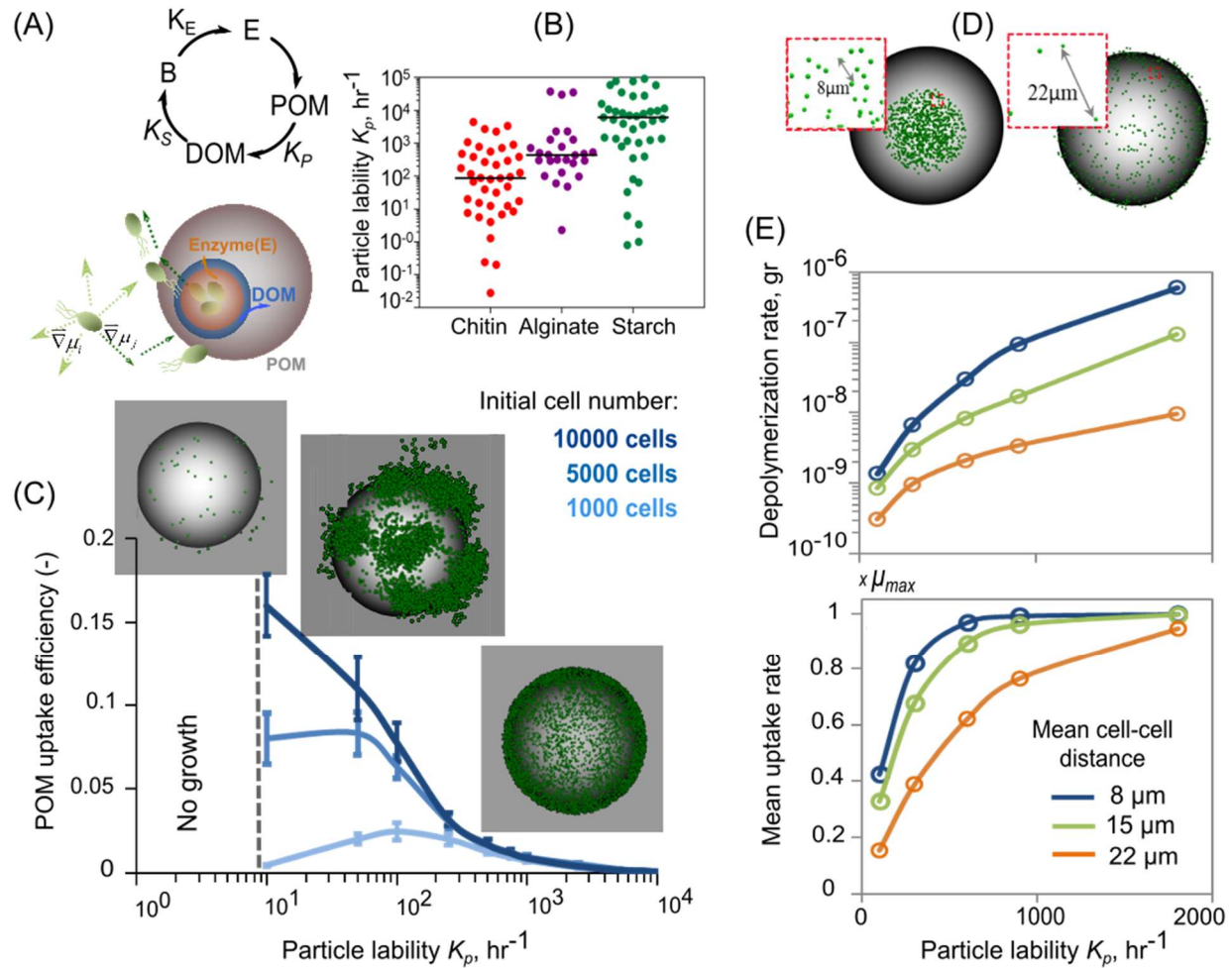
98

## 99 **Results**

100 We modeled the dynamics of cell colonization, enzyme secretion and growth (Figure 1A) using  
101 an individual based model to describe cells, coupled to a reaction-diffusion framework to describe  
102 enzymes and oligosaccharides. In the model, bacterial cells that attached to the surface of a  
103 polysaccharide particle broadcast enzymes that reacted with the surface of the particle, releasing  
104 oligosaccharides to which non-attached cells could chemotax. Cellular uptake of oligosaccharides

105 followed Monod kinetics (20) and cells were allowed to divide after a certain quota of  
106 oligosaccharide is consumed (19) (see Methods and Supplementary Information for a detailed  
107 description and Table S1 for the parameters). This individual-based approach allowed us to  
108 modulate traits such as chemotaxis or particle-attachment rate, and measure their impact on the  
109 carbon uptake rate on a cell by cell basis.

110 A crucial parameter of our model was the “particle lability”,  $K_p$ , which defined how many  
111 grams of oligosaccharide were released per gram of enzyme acting on the polysaccharide surface  
112 per unit of time.  $K_p$  was a compound parameter that resulted from the product of the catalytic  
113 activity of the enzyme,  $k_{cat}$  and the recalcitrance of the substrate. This parameter played a central  
114 role because it determined the maximum degradation rate and controlled the nutrient supply rate  
115 of bacteria. A survey of hydrolysis rate values reported in the literature revealed that the particle  
116 lability,  $K_p$  can exhibit significant variation across natural environments and microbial enzymes.  
117  $K_p$  varied by more than 6 orders of magnitude within glycosyl hydrolase families- a trend that held  
118 true among different substrate types such as chitin, alginate and starch (Figure 1B). This led us  
119 to ask how variation in particle lability,  $K_p$  affected population growth dynamics and the rate –  
120 efficiency relation of POM depolymerization.



121

122 **Figure 1. POM uptake efficiency is regulated by an emergent rate-yield trade-off.** A)  
 123 Conceptual representation of enzyme secretion by bacterial cells, breakdown of polysaccharide  
 124 substrates to oligosaccharides by enzymes and uptake of oligosaccharides by bacterial cells.  
 125 Schematic representation of trait based modeling of microbial dispersal and colonization on the  
 126 particle is shown (below). B) The distribution range of particle lability ( $K_p$ ) from natural polymeric  
 127 carbohydrates (Chitin, Alginate, Starch). The data are for various bacterial species with their  
 128 corresponding abiotic conditions (species name, substrates and environmental conditions are  
 129 represented in supplementary Table S2) (data are from Brenda database: [www.brenda-enzymes.org](http://www.brenda-enzymes.org)).  
 130 The solid line indicates the mean value of the particle lability. C) POM uptake  
 131 efficiency as a function of particle lability and initial population size. Dashed line indicates the no  
 132 growth zone. Microbial population assembly on the particle for three levels of particle lability  
 133 (recalcitrant I:  $K_p:1hr^{-1}$ , semi-labile II:  $K_p:100hr^{-1}$  and labile III:  $K_p:1000hr^{-1}$ ) are shown. Green dots  
 134 show individual cells on the particle. Simulations are performed for a range of initial cell densities  
 135 and 1% detachment is allowed. Half saturation,  $K_s$  is assumed 0.1mg/L. D) The cell spatial  
 136 distribution on particles for scenarios with 8 and 22 $\mu m$  mean cell distances are shown. E)  
 137 Depolymerization and mean uptake rates for a range of mean cell-cell distance are represented  
 138 as a function of particle lability. The simulations are initialized with placing individual cells  
 139 uniformly on the particle to meet their corresponding mean cell-cell distance for neighboring cells.  
 140 No detachment is allowed in simulations. Initial cell number was 1000 cells.  
 141

142 Our results revealed that there is an emergent negative relationship between the rates of  
143 depolymerization and growth, and the POM uptake efficiency of the particle-associated bacterial  
144 population (Figure 1C) (21, 22). This emergent rate – efficiency trade-off was a consequence of  
145 the diffusion of oligosaccharide in a three dimensional environment where soluble products that  
146 were not taken up by cells in the vicinity of the particle were lost. At high values of  $K_p$ ,  
147 oligosaccharides are produced in excess of  $K_s$ , the half-saturation constant of the Monod growth  
148 function, and therefore cells approached their maximum growth rate. However, high  $K_p$  also led  
149 to ~99% loss of oligosaccharide to diffusion (~1% recovery), which reduced the theoretical  
150 biomass yield of the population and the POM uptake efficiency. For comparison, if the system  
151 was closed, as in a laboratory reactor, POM uptake efficiency could theoretically reach 100%  
152 because dissolved oligosaccharides would accumulate (Figure S1). However, natural  
153 environments are rarely, if ever, closed and diffusive losses are likely to limit POM uptake  
154 efficiency in nature, given low particle densities, (Table S3). Moreover, adding the movement of  
155 fluids around particles by convective flow to the model (23–25), further increased the loss rate of  
156 oligosaccharide to the bulk environment, and reduced POM uptake efficiency from 10% to 2%  
157 (Figure S2-S5). Although the exact value of POM uptake efficiency could depend on substrate  
158 affinity ( $1/K_s$ ) and on cell numbers (the more cells that can capture oligosaccharides the higher  
159 uptake efficiency), the trade-off between rate and efficiency held for different physiological  
160 parameters (Figure S6). Taken together, these results suggest that in natural environments most  
161 public goods are lost and that the competition between diffusion and uptake should lead to a  
162 tradeoff between POM uptake efficiency and its turnover rate. Quantifying the uptake efficiency  
163 for natural marine particles revealed that the maximum efficiency barely exceeded 7% at the  
164 optimum particle lability ( $K_p \sim 100\text{hr}^{-1}$ ) for the highest of particle-associated cell density observed  
165 ( $\sim 2.32 \times 10^7$ ) (Table S3).

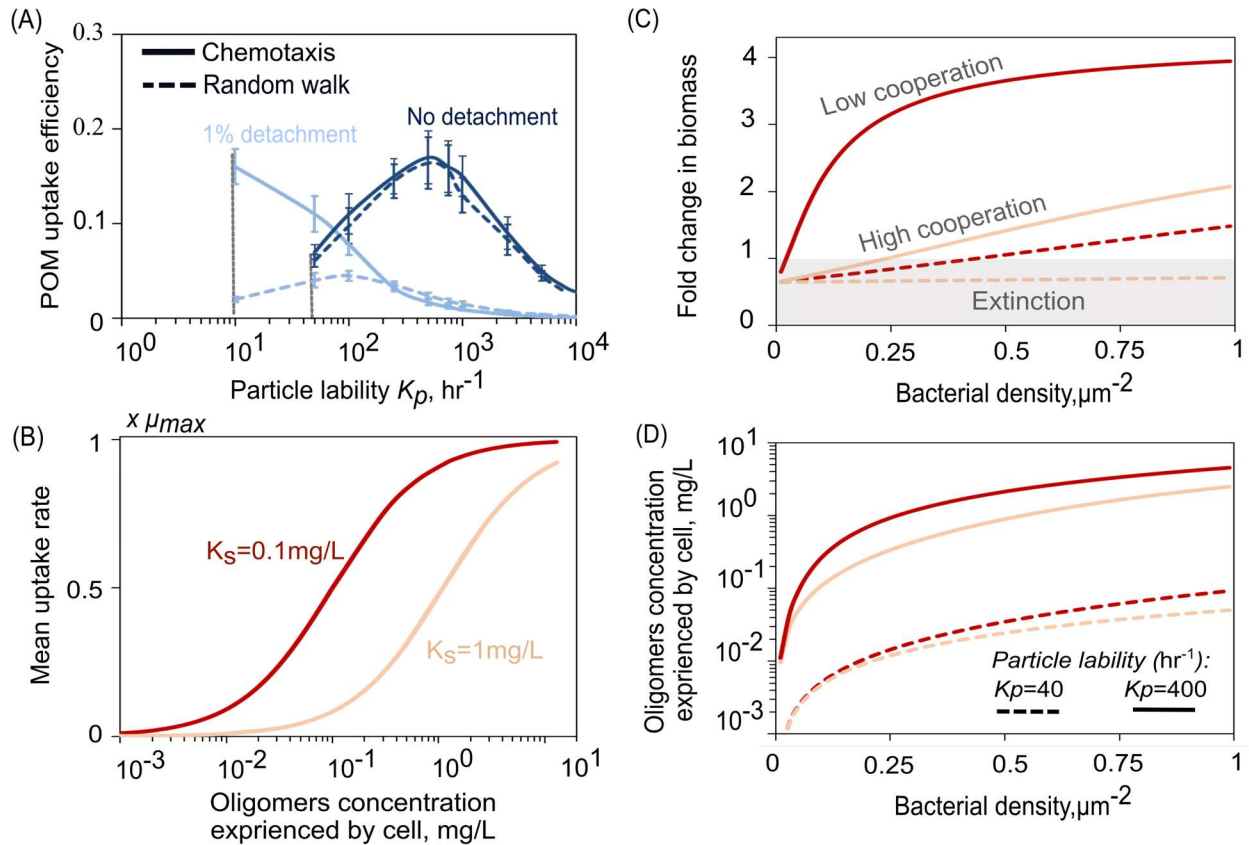
166 Surprisingly, we found that the high POM uptake efficiency observed at low  $K_p$  (recalcitrant  
167 particles/low enzymatic activity per cell) was mediated by the aggregation of cells into micro-scale

168 patches on the particle surface, a phenomenon that was not hardcoded in the model but emerged  
169 from the interplay between diffusion, cell behavior and growth (Figure 1C and Figure S7). Within  
170 these patches, cells grew cooperatively by sharing oligosaccharides that would otherwise be lost  
171 to diffusion, which increased the per capita growth rate and POM uptake efficiency up to a density  
172 of 0.3 cells/ $\mu\text{m}^2$  (Figure S8). To characterize the spatial density dependence, we performed  
173 simulations to quantify particle depolymerization and mean growth rates as a function of the inter-  
174 cell distance (Figure 1D). Our analysis showed that dense spacing (a nearest neighbor distance  
175 of 8  $\mu\text{m}$ ) promoted cooperation by sharing of oligosaccharides, but only when particles were  
176 recalcitrant and the oligosaccharide production rate was slow ( $K_p < 100 \text{hr}^{-1}$ ) (Figure 1E). More  
177 precisely, when the amount of oligosaccharide available to cells fell near  $K_s$ , the half-saturation of  
178 the Monod growth curve, an increase in the local concentration of oligosaccharide due to cell-cell  
179 aggregation increased the per capita growth rate. In contrast, at high  $K_p$  ( $\sim 2000 \text{hr}^{-1}$ ),  
180 oligosaccharides quickly accumulated and the uptake rate was decoupled from the spatial  
181 organization of the cells on the particle, since there were enough resources for cells to grow at  
182 their maximal rate ( $[C] \gg K_s$ ) (Figure 1E). Under these conditions, there is no benefit to  
183 aggregation and even cells spaced 22  $\mu\text{m}$  apart reached their maximum oligosaccharide uptake  
184 rate (Figure 1E).

185 In our model, cell detachment and reattachment from the particle surface was a critical  
186 behavior that enabled the formation of patches and the degradation of recalcitrant particles. On  
187 recalcitrant particles ( $K_p = 10-100 \text{hr}^{-1}$ ) 1% detachment significantly increased the particle  
188 degradation rate and its uptake efficiency (Figure 2A), and also increased the mean carbon  
189 uptake rate by a factor of 5 (Figure S9A), compared to a non-detaching population. This allowed  
190 populations to survive on recalcitrant particles that might otherwise not sustain growth and drive  
191 the population to extinction (Figure S9B). Without chemotaxis, random motility alone still allowed  
192 detaching populations to grow on more recalcitrant particles than non-detaching populations, but  
193 at  $\sim 1/6$  the POM uptake efficiency (Figure 2A) and  $\sim 1/10$  the rate of biomass accumulation (Figure



194 S9B). This was due to the fact that with chemotactic motility most cells had access to the same  
195 of hydrolysis products emanating from cell patches, with the distribution of carbon uptake rates  
196 for individual cells displaying a tight peak near the maximum uptake rate ( $\mu \sim 0.8\mu_{\max}$ ) (Figure  
197 S9C). Our model thus suggests that detachment and chemotaxis enhance POM uptake efficiency  
198 under nutrient-limited conditions ( $[C] \sim K_s$ ) by enabling the formation of patches where cells  
199 cooperate by sharing public goods. This also implies that the onset of cooperation is dependent  
200 on the individual strain physiology. Organisms that had either a high affinity for oligosaccharides  
201 (low  $K_s$ ) or a high hydrolytic activity saturated their growth at low oligosaccharide concentrations  
202 (Figure 2B), circumventing the need to cooperate Figure (2C-D). In contrast, organisms with a  
203 low uptake affinity for the public good, or organisms with a low per-cell rate of hydrolysis, such as  
204 those that tether enzymes to their membrane, had a higher need to cooperate with other cells  
205 (Figure 2D). Therefore, although there was a general trend to increase cooperation as particles  
206 became harder to degrade (Figure 2D), traits such as motility, surface detachment rates,  
207 substrate affinity or enzyme localization determined determine the exact onset of cooperation for  
208 each population.  
209

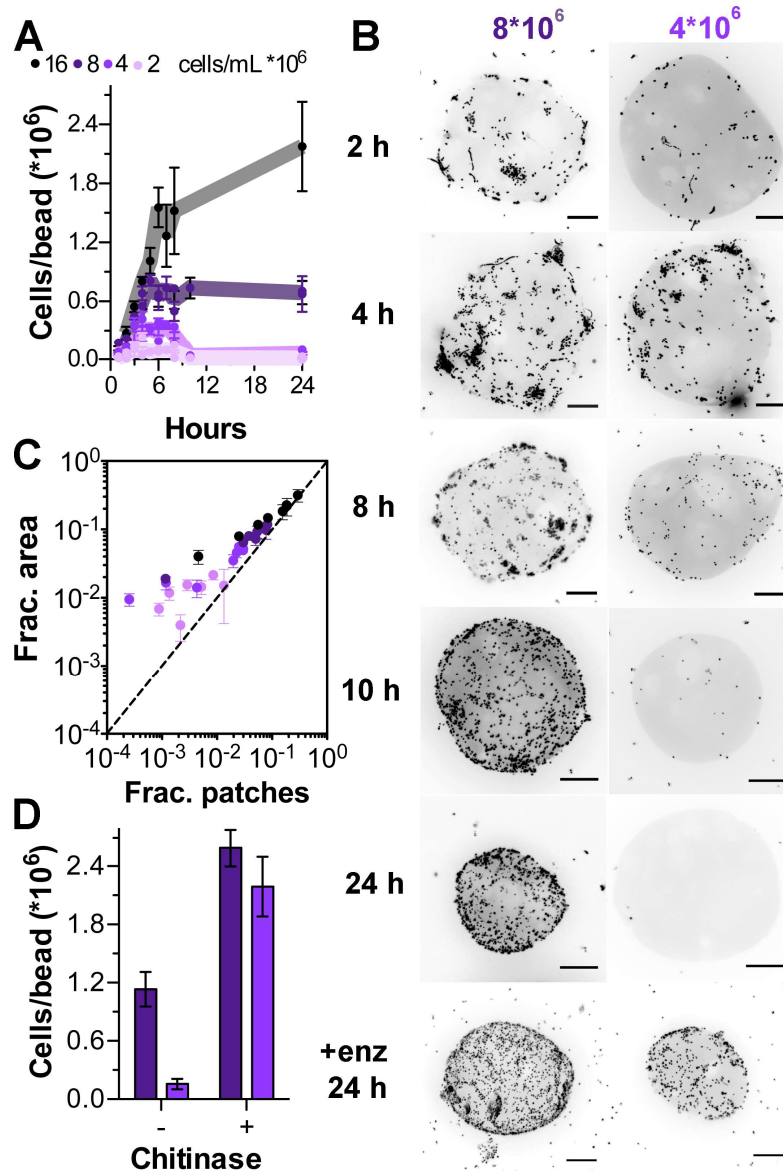


210  
 211 **Figure 2. Bacterial cooperation enhances carbon uptake rate.** A) Effects of individual cells  
 212 attachment/detachment frequency to/from particles on POM uptake efficiency. Solid line shows  
 213 the simulation results with chemotactic behavior compared to simulations with only random walk  
 214 shown with dashed line. The computational results are shown for the initial cell number of  
 215  $10^7$  cells/ml and the particle size was set to  $200 \mu\text{m}$ . The results are shown for simulations after  
 216 10 hours. No bacterial growth is expected for  $K_p$  values below the grey dashed line. B) Monod  
 217 growth kinetics represents the mean cellular uptake rate of oligosaccharides as a function of  
 218 oligosaccharide concentration. C) The fold of change in biomass as a function of initial bacterial  
 219 density is shown for simulations of bacterial colonization on a single particle after 20 hours. D) The  
 220 effects of bacterial cell density on produced oligosaccharide concentration, experienced by  
 221 individual cells. The simulations are performed for a particle with constant radius of  $200 \mu\text{m}$  and  
 222 initial cell density of  $0.3 \text{ cell}/\mu\text{m}^2$ . Labile ( $K_p = 400 \text{ hr}^{-1}$ ) vs. recalcitrant ( $K_p = 40 \text{ hr}^{-1}$ ) particles are  
 223 considered with two relatively low ( $K_S = 1 \text{ mg/L}$ ) and high ( $K_S = 0.1 \text{ mg/L}$ ) affinity to substrates uptake  
 224 by bacterial cells.  
 225

226 To experimentally validate our prediction that cell-cell cooperation drives the degradation  
 227 of hard-to-degrade polysaccharides, we turned our attention to *Psychromonas* sp., *psych6C06*, a  
 228 marine isolate that had previously been enriched from coastal seawater on model chitin  
 229 particles(10). The strain readily degrades chitin hydrogel in  $\sim 30$  hours (17) and encodes at least  
 230 eight predicted chitinases, or glycosyl hydrolase family 18 and 19 homologs, but no other families

231 of glycosyl hydrolases, leading us to conclude that the strain is representative of a chitin specialist.  
232 We reasoned that if cooperative growth kinetics played a role in this system, we would observe a  
233 strong dependency between the initial number of cells that can colonize the particle and the  
234 growth of the population. In particular, we would expect a critical cell density below which the  
235 population is unable to form the patches required to degrade particles, revealing that the  
236 degradation process is bacteria-limited.

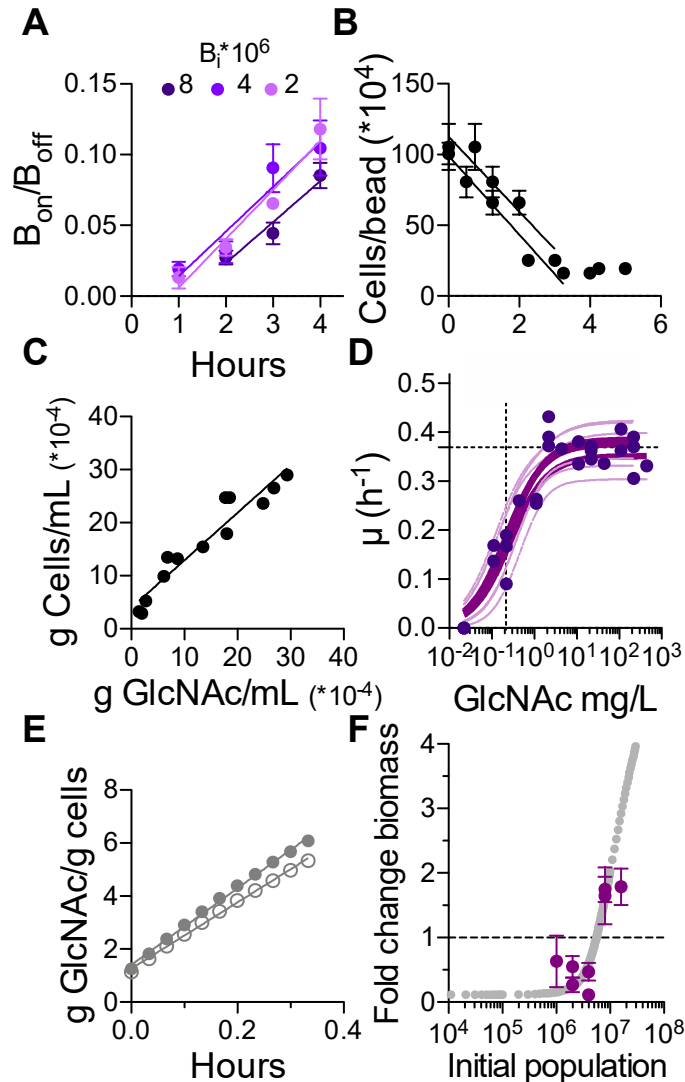
237 In agreement with this prediction, *psych6C06* displayed a strong density dependence  
238 when growing on hydrogel chitin beads, in the form of a critical cell density below which  
239 degradation never occurred (Figure 3A). Interestingly, we observed that colonization involved the  
240 formation of cell patches, in agreement with the model results (Figure 3B). At concentrations just  
241 below the threshold critical cell density, we saw that cells that initially colonized the particle were  
242 not able to persist. Populations that persisted did so by forming cell patches (Figure 3C). We  
243 artificially increased  $K_p$  by adding exogenous chitinase to supply 776  $\mu\text{g/h}$  GlcNAc. Consistent  
244 with individual-based model results (Figure 2D), the addition of the exogenous enzyme activity  
245 lowered the cell density-dependent threshold for colonization of the chitin hydrogel beads (Figure  
246 3D). In addition, the broadcast chitinases led to a more uniform distribution of *psych6C06* cells on  
247 the chitin hydrogel bead, a state that was morphologically distinct from the patchy colonization  
248 achieved at 24 h *psych6C06* without enzyme (Figure 3B, Figure S10C) and similar to the  
249 simulation results at high  $K_p$  (Figure 2D).



250  
 251 **Figure 3. Chitinase limitation drives psych6C06 to form growth-promoting surface-**  
 252 **associated clusters.** A) An Allee effect emerges for populations of psych6C06 over 24 h. Data  
 253 points are combined from three experimental replicates. Error bars are SEM from at least 6  
 254 individual measurements of colonization density on chitin beads. Lines represent the mean  
 255 trajectory for each cell density. B) Representative images showing colonization density of initial  
 256 populations of cells below ( $4 \times 10^6$  cells/mL), and above ( $8 \times 10^6$  cells/mL) the colonization threshold  
 257 at timepoints during colonization. Scale bars are 20  $\mu$ m. C) The fraction of cell area (Frac. area)  
 258 on chitin hydrogel beads that exists in patches >2000 cells (Frac. patches). Dashed line indicates  
 259 the limit where all cell area exists in patches ( $x=y$ ). Raw data analyzed are the same as 3A and  
 260 3B. D) The addition of exogenous chitinase enables smaller populations of cells to colonize chitin  
 261 hydrogel particles. Dark purple bars, initial density  $8 \times 10^6$  cells/mL; light purple bars, initial density  
 262  $4 \times 10^6$  cells/mL. Colonization density was assessed after 24 h. Bars are SEM from at least 5  
 263 measurements. The amount of chitinase added to the medium supplies 776  $\mu$ g/h GlcNAc.  
 264

265 To obtain a more mechanistic understanding of the critical threshold phenomenon we took  
266 a bottom-up approach by predicting the threshold cell density based on measurements of the  
267 relevant physiological and behavioral parameters of *psych6C06*. We calculated particle  
268 attachment and detachment rates by quantifying cell density on particles (Figure 4A-B). Our  
269 measurements revealed rapid attachment and detachment rates (attachment rate  $0.03 \text{ h}^{-1}$ ,  
270 detachment rate  $0.26 \text{ h}^{-1}$ ) that were proportional to the density of cells off and on the particle,  
271 respectively suggesting that the population undergoes frequent rearrangement on the surface of  
272 the particle. These observations echo the rearrangement observed in the individual based  
273 simulations (Figure 2A). We measured  $K_p$  for *psych6C06*, using the fluorescent substrate 4-  
274 Methylumbelliferyl (MUF)-*N*-acetyl- $\beta$ -D-glucosaminide, which detects the release of GlcNAc from  
275 chitin. We noted that very little chitinase activity was detected in the culture supernatant of  
276 *psych6C06*, but robust activity was associated with the cells themselves (Figure 4C, Figure S10A)  
277 indicating that enzymes were membrane bound. Using MUF-conjugated substrates with different  
278 cleavage specificities, we determined that most of *psych6C06* chitinase activity was derived from  
279 exochitinase, which releases GlcNAc as a product (Figure S10B). Thus, we assessed the GlcNAc  
280 to biomass conversion factor (the biomass yield) of this strain by direct measurement of sugar  
281 consumption and cell density in exponentially growing cultures (Figure 4D). We measured of the  
282 growth rate of *psych6C06* on a range of GlcNAc concentrations, and used these substrate-limited  
283 growth measurements to derive  $\mu_{max}$  and  $K_s$  from a fit of the Monod growth equation(20) (Figure  
284 4E). We used the diffusive loss of oligosaccharides predicted by our 3D simulations to estimate  
285 GlcNAc loss, and assumed that all cells on the surface experienced the same concentration (no  
286 local gradients). Using these measurements, we parameterized a simple version of the individual  
287 based model written to describe population-level growth dynamics on a surface where cells can  
288 attach, detach, and grow as a function of the hydrolyzed product concentration (see Methods for  
289 a full description of the analytical bottom-up model). Using this simplified model, with no free  
290 parameters, we studied how the initial cell density determines the population colonization rate

291 and the growth of bacteria on the particle surface. We found a remarkable quantitative agreement  
 292 between model and experiments, with critical thresholds predicted between initial densities of  
 293  $5 \cdot 10^6$  and  $10^7$  cells/mL (Figure 4F). Our analysis thus shows that individual cell level physiology  
 294 and behavior together with diffusion regulate the onset of social degradation of POM.  
 295



296  
 297 **Figure 4. Physiological traits predict an Allee effect for chitin-degrading strain psych6C06**  
 298 **in a diffusive environment.** A) Rates of attachment to chitin hydrogel beads, measured for three  
 299 different initial cell densities per mL ( $B_i$ ). Error is SEM from a minimum of 10 measurements and  
 300 lines are fits of the data to a linear regression, where the slope ( $B_{on}/B_{off}$ ) gives the rate of cell  
 301 accumulation on the particle (for  $B_i=8 \cdot 10^6$ ,  $B_{on}/B_{off} = 2.9 \cdot 10^{-2} \pm 6.9 \cdot 10^{-3}$ ;  $B_i=4 \cdot 10^6$ , for  $B_{on}/B_{off}$   
 302  $= 3.2 \cdot 10^{-2} \pm 6.6 \cdot 10^{-3}$ ;  $B_i=2 \cdot 10^6$ , for  $B_{on}/B_{off} = 3.4 \cdot 10^{-2} \pm 5.0 \cdot 10^{-3}$ ,  $\pm$  standard error). B) Rates

303 of detachment from chitin hydrogel beads, measured for two replicates 24 h after colonization by  
304 an initial population of  $8 \times 10^6$  cells/mL. Error is SEM from a minimum of 11 measurements and  
305 lines are fits of the data to a linear regression. Data from the first 3.5 h of the experiment were fit,  
306 and the slope of each line gives the rate at which cells leave the particle. Two independent  
307 replicates are shown (slope replicate 1=  $-2.8 \times 10^5 \pm 3.4 \times 10^4$  cells/bead/hour; slope replicate 2= -  
308  $2.7 \times 10^5 \pm 5.1 \times 10^4$  cells/bead/hour,  $\pm$  standard error). C) Yield of psych6C06 cells grown on  
309 GlcNAc as the sole carbon source. Data points represent individual measurements. Line is a  
310 linear regression of the data, where the slope ( $0.89 \pm 0.08$  g cells / g GlcNAc) gives the fractional  
311 yield. D) Dependency of psych6C06 growth rate on substrate availability. Three biological  
312 replicates are shown. Measurements of growth rate at nine different GlcNAc concentrations were  
313 fit to the Monod growth model  $\mu = \mu_{max} * S / (K_s + S)$ , where  $\mu$  is the observed exponential growth  
314 rate,  $S$  is the concentration of GlcNAc, to find the maximum growth rate ( $\mu_{max}$ ) and half-saturation  
315 constant ( $K_s$ ). Bold lines indicates fit for each biological replicate, dashed lines indicate 95%  
316 confidence intervals. For replicates 1-3 respectively,  $\mu_{max} = 0.35 \pm 0.02$ ,  $0.39 \pm 0.02$ ,  $0.37 \pm 0.02$   
317 and  $K_s = 1.2 \pm 0.5$ ,  $1.1 \pm 0.3$ ,  $1.0 \pm 0.32$ . E) The rate of chitin hydrolysis to GlcNAc by cell-  
318 associated psych6C06 chitinase ( $K_p$ ) was measured using fluorescent substrates for  
319 exponentially growing cells at two different densities (OD 0.1 open circles and OD 0.4 filled  
320 circles). Linear regression was fit to each replicate  $K_p$  (OD 0.1)=  $12.5 \pm 0.2$  g GlcNAc/g cells/h,  
321  $K_p$  (OD 0.4)=  $14.5 \pm 0.3$  g GlcNAc/g cells/h. The average of the two replicates was used as the  
322 estimate for psych6C06  $K_p$ . F) Prediction of psych6C06 biomass yield after 24 h of growth on  
323 chitin hydrogel beads with different initial population densities (cells/mL). The analytical model is  
324 based on the parameters measured in 3A-E (grey points). Dashed horizontal line indicates no  
325 biomass change. Purple data points indicate the experimentally measured change in biomass  
326 observed for the indicated initial population densities. Error bars represent SD for measurements  
327 from at least 3 replicates.  
328

## 329 Discussion

330 Despite the key regulatory role of microbes in carbon cycling (17, 26, 27), linking the micro-scale  
331 physiology and behavior of bacteria to carbon flux models remains elusive. Here, we show that in  
332 conditions where diffusion limits oligosaccharide accumulation, like in the ocean, the breakdown  
333 of particulate polysaccharides is subject to population density-dependent effects. These effects  
334 are driven by three key physiological parameters: the affinity of cells for hydrolyzed  
335 oligosaccharide, the rate of polysaccharide hydrolysis, and the amount of exchange on/off the  
336 particle surface that define a tradeoff between the rate of polysaccharide degradation and  
337 biomass yield. These features contrast with a common assumption of carbon flux models: that  
338 model heterotrophic cells consume nutrients at rates comparable to the consumption of simple  
339 dissolved substrates by laboratory-adapted model organisms in a well-mixed systems (27–29). In  
340 particular, our computational and experimental results highlight that the frequency at which cells

341 exchange on and off the polysaccharide surface has a large impact on cell carbon uptake. This  
342 effect is emerged not only by setting the threshold population density that is achieved by the initial  
343 population without growth, but by reconfiguring the arrangement of surface-associated cells in  
344 ways that maximize oligosaccharide uptake and growth. In addition, we show that the emergence  
345 of microbial aggregates on recalcitrant particles increases the chance of survival for populations  
346 of bacteria by enhancing the local dissolved carbon production rate and cell uptake rates. The  
347 benefit of aggregate formation is dependent on size: structures that are too large or dense to  
348 support the maximum uptake rate of individual cells promote competition rather than cooperation  
349 (30). In natural ecosystems, aggregate formation and dispersal is the rule rather than an exception  
350 (31–34) and cells are likely to carry adaptations to enhance aggregation/dispersal on particle  
351 surfaces, for instance by regulating chemotactic movement or the expression of biofilm  
352 components such as adhesins and matrix proteins (35, 36). Thus, micro-scale interactions could  
353 significantly affect the rates of POM turnover in the environment, underscoring the need to  
354 incorporate these interactions into models of carbon cycling.

355 Our computational and experimental system also highlights the importance of social  
356 collective behavior and spatial self-organization on community fitness and survival. Previous  
357 studies have shown that secretion of public goods (enzyme) favors the formation of patchy  
358 microbial aggregates by enhancing cooperative behavior (11, 14, 15, 37). Notably, many bacteria  
359 actively regulate enzyme secretion and other group behaviors at the level of transcription (38–  
360 40), and cell-density-dependent transcription factors such as quorum signal receptors are capable  
361 of sensing both changes to the environment and to cell density (41). While our simulations reveal  
362 that physiology alone is sufficient to explain patch formation, further studies are required to  
363 evaluate the contribution of such regulation on the group behaviors that may facilitate patch  
364 formation and dispersal by *psych6C06*. Our results also suggest that the benefit of clustering is  
365 not universal and is instead dependent on physiology: the POM uptake efficiency and growth rate  
366 of cells with high affinity for substrates suffers in the context of an aggregate, while aggregation



367 is optimal for strains with lower substrate affinity and lower rates of polysaccharide hydrolysis  
368 (11). This observation highlights the fact that in systems with potential for spatial organization –  
369 that is, most systems outside the lab, the balance between cooperation and competition can be  
370 delicate and modulated by the intersection of physical processes with microbial physiology.

371

## 372 **Methods**

### 373 **Individual based model of individual cell behavior and physiology**

374 The mathematical model represents metabolism, surface interaction and flagellar motility of  
375 individual cells in 3D space in the presence of chemical gradients. We introduce an individual-  
376 based model (42, 43) to quantify single-cell interactions with organic particles by abstracting the  
377 structural heterogeneities of natural POM into a mathematically simpler spherical shape, while  
378 preserving some key physical and chemical processes associated with POM degradation. A  
379 spherical organic particle of 200  $\mu\text{m}$  radius is simulated such that it remains static in the middle  
380 of an aqueous volume ( $\sim 1\text{mm}^3$ ). While natural organic matter aggregates may show various  
381 shapes and chemical compositions, we modeled particles as perfect spheres made of a single  
382 type of insoluble linear polysaccharides such as chitin, alginate, or cellulose. This computational  
383 model is inspired by experimental model systems used to study community assembly on marine  
384 POM (10, 17). The particle's size and its surface chemistry are assumed to be unchanged during  
385 particle degradation: only the particle density changes over time to satisfy mass conservation.  
386 This assumption is consistent with experimental observations that have shown no significant  
387 change in organic particle size during microbial degradation until the final stages of collapse(17).  
388 We simulated a scenario where an isogenic population of cells is allowed to colonize and degrade  
389 a particle with a defined volume. The simulations were started with zero oligosaccharides and the  
390 particle was considered to be the sole carbon source.

391 To take into account the fact that cells might regulate their enzymatic activity, the model  
392 limits enzyme secretion to two scenarios: when cells adhere to the particle surface or when the

393 rate of oligosaccharide supply exceeds the maintenance threshold. Importantly, our simulations  
394 ensure mass conservation between total carbon uptake, growth and loss of oligosaccharides.  
395 Individual cells are initialized as a uniform random distribution in the aqueous volume, and are  
396 allowed to disperse following gradients of chemo-attractant (in this case, oligosaccharide). The  
397 cells can consume the oligosaccharide, grow, and divide to new daughter cells and experience a  
398 range of local conditions. A full derivation of the mathematical expressions and steps used for  
399 modeling of microbial growth, dispersal and enzyme secretion can be found in the Supplemental  
400 Information.

401 **Experimental methods:** Strain *psych6C06* was previously isolated from an enrichment of  
402 nearshore coastal seawater (Nahant, MA, USA) for surface-associated chitin degrading microbial  
403 communities (10, 17). The strain was maintained as colonies on Marine Broth 2216 (Difco  
404 279110) with 1.5% agar (BD 214010). To establish exponential growth without hysteresis, we  
405 modified a culturing protocol previously developed for *Escherichia coli* K12(44), and grew cells on  
406 a defined seawater medium with the *N*-acetyl-D-glucosamine (GlcNAc) at concentrations  
407 indicated. Chitin hydrogel beads (NEB) were washed and diluted to 200-250 particles per mL with  
408 size range from 40 to 100  $\mu\text{m}$  in diameter. The beads were rotated end over end at 21-25  $^{\circ}\text{C}$ . The  
409 density of inoculated cells was set to be at an  $A_{600}$  of 0.01, diluted from 20 mM GlcNAc minimal  
410 medium cultures prepared as described above. To visualize particles and their surface-associated  
411 bacteria, 200  $\mu\text{l}$  subsamples were stained with the DNA-intercalating dye SYTO9 (Thermo Fisher,  
412 S34854) at a 1:285 dilution of the stock in 96-well plates with optically clear plastic bottoms (VWR  
413 10062-900).

414 Cell density measurements (Absorbance at 600 nm,  $A_{600}$ ) of exponentially-growing cells  
415 were used to measure the maximum cellular growth rate, and plating was used to measure growth  
416 under GlcNAc limitation, from which we derived the half-saturation constant. GlcNAc depletion  
417 was measured during growth using the dinitrosalicylic acid reagent method(45), and the depletion  
418 rate was used to calculate the biomass yield (see Supplemental Information)

419 Chitinase activity was quantified using Methylumbelliferyl(MUF)-conjugated substrates N,N'-  
420 diacetyl- $\beta$ -D-chitobioside, N-acetyl- $\beta$ -D-glucosaminide, and  $\beta$ -D-N,N',N''-triacetylchitotriose  
421 (Sigma CS1030). Microscopy was performed on micro-confocal high-content imaging system  
422 (ImageXpress Micro Confocal, Molecular Devices), using the 60  $\mu$ m pinhole spinning disk mode.  
423 Fluorescent signal was visualized with a LED light cube (Lumencore Spectra X light engine), and  
424 bandpass filters (ex 482/35 nm em 538/40 nm dichroic 506 nm), with a 40x objective (Nikon Ph  
425 2 S Plan Fluor ELWD ADM 0.60 NA cc 0-2 mm, correction collar set to 1.1), and a sCMOS  
426 detector (Andor Zyla). Image analysis was performed in MATLAB (release 2018a). Briefly, image  
427 stacks were split in half and a maximum intensity projection was obtained for each half. The low  
428 level of fluorescent signal associated with free dye in the hydrogel particles was used to define  
429 an intensity threshold suitable to create a binary mask for the particle projections. A mask of the  
430 cells within the beads was then defined using their brighter fluorescence intensity. We used this  
431 segmentation to quantify the total surface area occupied by the cells on the bead, and to quantify  
432 the total surface area occupied by patches (areas where cells contact other cells  $>10 \mu\text{m}^2$ ).

433

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439

#### 440 **References:**

- 441 1. Flint HJ, Scott KP, Duncan SH, Louis P, Forano E (2012) Microbial degradation of complex  
442 carbohydrates in the gut. *Gut Microbes* 3(4). doi:10.4161/gmic.19897.
- 443 2. Martens EC, et al. (2011) Recognition and degradation of plant cell wall polysaccharides  
444 by two human gut symbionts. *PLoS Biol* 9(12). doi:10.1371/journal.pbio.1001221.

- 445 3. Cuskin F, et al. (2015) Human gut Bacteroidetes can utilize yeast mannan through a selfish  
446 mechanism. *Nature* 517(7533):165–169.
- 447 4. Paul EA (2016) The nature and dynamics of soil organic matter: Plant inputs, microbial  
448 transformations, and organic matter stabilization. *Soil Biol Biochem* 98:109–126.
- 449 5. Liang C, Schimel JP, Jastrow JD (2017) The importance of anabolism in microbial control  
450 over soil carbon storage. *Nat Microbiol* 2(8). doi:10.1038/nmicrobiol.2017.105.
- 451 6. Dang H, Lovell CR (2016) Microbial surface colonization and biofilm development in marine  
452 environments. *Microbiol Mol Biol Rev* 80(1):91–138.
- 453 7. Turner JT (2015) Zooplankton fecal pellets, marine snow, phytodetritus and the ocean's  
454 biological pump. *Prog Oceanogr* 130:205–248.
- 455 8. Grossart HP, Ploug H (2001) Microbial degradation of organic carbon and nitrogen on  
456 diatom aggregates. *Limnol Oceanogr* 46(2):267–277.
- 457 9. D'Ambrosio L, Ziervogel K, MacGregor B, Teske A, Arnosti C (2014) Composition and  
458 enzymatic function of particle-associated and free-living bacteria: a coastal/offshore  
459 comparison. *Isme J* 8:2167.
- 460 10. Datta MS, Sliwerska E, Gore J, Polz M, Cordero OX (2016) Microbial interactions lead to  
461 rapid micro-scale successions on model marine particles. *Nat Commun* 7(May):11965.
- 462 11. Drescher K, Nadell CD, Stone HA, Wingreen NS, Bassler BL (2014) Solutions to the public  
463 goods dilemma in bacterial biofilms. *Curr Biol* 24(1):50–55.
- 464 12. Azam F, Malfatti F (2007) Microbial structuring of marine ecosystems. *Nat Rev Microbiol*  
465 5(10):782–791.
- 466 13. Gore J, Youk H, van Oudenaarden A (2009) Snowdrift game dynamics and facultative  
467 cheating in yeast. *Nature* 459(7244):253–256.
- 468 14. Ratzke C, Gore J (2016) Self-organized patchiness facilitates survival in a cooperatively  
469 growing *Bacillus subtilis* population. *Nat Microbiol* 1(5):16022.
- 470 15. H. Koschwanez J, R. Foster K, W. Murray A (2011) Sucrose utilization in budding yeast as

- 471 a model for the origin of undifferentiated multicellularity. *PLoS Biol* 9(8):e1001122.
- 472 16. Newell R, Lucas M, Linley E (2007) Rate of degradation and efficiency of conversion of  
473 phytoplankton debris by marine micro-organisms. *Mar Ecol Prog Ser*.  
474 doi:10.3354/meps006123.
- 475 17. Enke TN, Leventhal GE, Metzger M, Saavedra JT, Cordero OX (2018) Microscale ecology  
476 regulates particulate organic matter turnover in model marine microbial communities. *Nat*  
477 *Commun* 9(1):2743.
- 478 18. Kreft JU, Booth G, Wimpenny JWT (1998) BacSim, a simulator for individual-based  
479 modelling of bacterial colony growth. *Microbiology* 144(12):3275–3287.
- 480 19. Ebrahimi A, Or D (2015) Hydration and diffusion processes shape microbial community  
481 organization and function in model soil aggregates. *Water Resour Res* 51:9804–9827.
- 482 20. Monod J (1949) The growth of bacterial cultures. *Annu Rev Microbiol* 3(1):371–394.
- 483 21. Rieck A, Herlemann DPR, Jürgens K, Grossart HP (2015) Particle-associated differ from  
484 free-living bacteria in surface waters of the baltic sea. *Front Microbiol* 6(DEC).  
485 doi:10.3389/fmicb.2015.01297.
- 486 22. Whitman WB, Coleman DC, Wiebe WJ (1998) Prokaryotes: The unseen majority. *Proc Natl*  
487 *Acad Sci* 95(12):6578–6583.
- 488 23. Battin TJ, et al. (2008) Biophysical controls on organic carbon fluxes in fluvial networks.  
489 *Nat Geosci* 1(2):95–100.
- 490 24. Stocker R, Seymour JR, Samadani A, Hunt DE, Polz MF (2008) Rapid chemotactic  
491 response enables marine bacteria to exploit ephemeral microscale nutrient patches. *Proc*  
492 *Natl Acad Sci*. doi:10.1073/pnas.0709765105.
- 493 25. Taylor JR, Stocker R (2012) Trade-offs of chemotactic foraging in turbulent water. *Science*  
494 (80-) 338(6107):675–679.
- 495 26. Kallenbach CM, Frey SD, Grandy AS (2016) Direct evidence for microbial-derived soil  
496 organic matter formation and its ecophysiological controls. *Nat Commun* 7.

- 497 doi:10.1038/ncomms13630.
- 498 27. Georgiou K, Abramoff RZ, Harte J, Riley WJ, Torn MS (2017) Microbial community-level  
499 regulation explains soil carbon responses to long-term litter manipulations. *Nat Commun*  
500 8(1). doi:10.1038/s41467-017-01116-z.
- 501 28. Manzoni S, Schaeffer SM, Katul G, Porporato A, Schimel JP (2014) A theoretical analysis  
502 of microbial eco-physiological and diffusion limitations to carbon cycling in drying soils. *Soil*  
503 *Biol Biochem* 73:69–83.
- 504 29. Schimel JP, Weintraub MN (2003) The implications of exoenzyme activity on microbial  
505 carbon and nitrogen limitation in soil: A theoretical model. *Soil Biol Biochem* 35(4):549–  
506 563.
- 507 30. Ebrahimi A, Schwartzman J, Cordero OX (2019) Multicellular behavior enables cooperation  
508 in microbial cell aggregates. *bioRxiv*:626481.
- 509 31. Massalha H, Korenblum E, Malitsky S, Shapiro OH, Aharoni A (2017) Live imaging of root–  
510 bacteria interactions in a microfluidics setup. *Proc Natl Acad Sci* 114(17):4549 LP – 4554.
- 511 32. Bar-Zeev E, Berman-Frank I, Girshevitz O, Berman T (2012) Revised paradigm of aquatic  
512 biofilm formation facilitated by microgel transparent exopolymer particles. *Proc Natl Acad*  
513 *Sci* 109(23):9119 LP – 9124.
- 514 33. Vidal A, et al. (2018) Linking 3D soil structure and plant-microbe-soil carbon transfer in the  
515 rhizosphere. *Front Environ Sci* 6:9.
- 516 34. Boeddinghaus RS, Nunan N, Berner D, Marhan S, Kandeler E (2015) Do general spatial  
517 relationships for microbial biomass and soil enzyme activities exist in temperate grassland  
518 soils? *Soil Biol Biochem* 88:430–440.
- 519 35. Ackermann M (2015) A functional perspective on phenotypic heterogeneity in  
520 microorganisms. *Nat Rev Microbiol*. doi:10.1038/nrmicro3491.
- 521 36. Niederdorfer R, Besemer K, Battin TJ, Peter H (2017) Ecological strategies and metabolic  
522 trade-offs of complex environmental biofilms. *npj Biofilms Microbiomes* 3(1):21.

- 523 37. Nadell CD, Drescher K, Foster KR (2016) Spatial structure, cooperation and competition  
524 in biofilms. *Nat Rev Microbiol* 14(9):589–600.
- 525 38. Goo E, et al. (2012) Bacterial quorum sensing, cooperativity, and anticipation of stationary-  
526 phase stress. *Proc Natl Acad Sci* 109(48):19775–19780.
- 527 39. Jemielita M, Wingreen NS, Bassler BL (2018) Quorum sensing controls *Vibrio cholerae*  
528 multicellular aggregate formation. *Elife* 7. doi:10.7554/eLife.42057.
- 529 40. DeAngelis KM, Lindow SE, Firestone MK (2008) Bacterial quorum sensing and nitrogen  
530 cycling in rhizosphere soil. *FEMS Microbiol Ecol* 66(2):197–207.
- 531 41. Cornforth DM, et al. (2014) Combinatorial quorum sensing allows bacteria to resolve their  
532 social and physical environment. *Proc Natl Acad Sci* 111(11):4280–4284.
- 533 42. Kreft JU, Picioreanu C, Wimpenny JWT, Van Loosdrecht MCM (2001) Individual-based  
534 modelling of biofilms. *Microbiology* 147(11):2897–2912.
- 535 43. Ebrahimi AN, Or D (2014) Microbial dispersal in unsaturated porous media: Characteristics  
536 of motile bacterial cell motions in unsaturated angular pore networks. *Water Resour Res*  
537 50(9):7406–7429.
- 538 44. Scott M, Gunderson CW, Mateescu EM, Zhang Z, Hwa T (2010) Interdependence of cell  
539 growth and gene expression: Origins and consequences. *Science (80- )* 330(6007):1099–  
540 1102.
- 541 45. Miller GL (1959) Use of dinitrosalicylic acid reagent for determination of reducing sugar.  
542 *Anal Chem* 31(3):426–428.

543

## Supplementary Information for

### **Cooperation and spatial self-organization determine ecosystem function for polysaccharide-degrading bacteria**

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#### **This PDF file includes:**

Supplementary methods

Figs. S1 to S10

Tables S1 to S3

References for SI reference citations



## Supplementary Methods

### Individual based modeling procedure

#### - Growth and division

The individual-based model assumes that an individual cell doubles to two identical cells after a certain amount of carbon is taken up (1, 2). In this model, a Monod-kinetic parameterization(3) gives the carbon uptake kinetics of individual cells  $v_i^s$  and their biomass accumulation:

$$v_i^s = V_{\max} B_i \frac{[C]}{[C] + K_s} \quad (1)$$

where  $B_i$  is the cell dry mass of the individual cell,  $i$ , and the maximum rate of substrate uptake is:

$V_{\max} = \mu_{\max} / Y_{\max}$  (maximum specific growth rate / growth yield).  $K_s$  is the half-saturation constant for dissolved oligosaccharide. We assume that dissolved oligosaccharides  $[C]$  are the primary limiting substrate for growth, and that all other nutrients (e.g., sources of phosphate and nitrogen) available at non-rate limiting levels.

For cell  $i$ , the actual biomass accumulation ( $\frac{dB_i}{dt} = v_i^s Y_{\max}$ ) and resources consumed ( $\frac{dB_i}{dt} = B_i m_i Y_{\max}$ ) for maintenance,  $m_i$  are both assumed to be proportional to the cell's dry mass,

$B_i$ , therefore, the net growth in the cell biomass ( $\mu_{net}$ ) can be described as follows:

$$\mu_{net} = \left( \frac{v_i^s}{B_i} - m_i \right) Y_{\max} \quad (2)$$

The minimum volume of the individual cell at division ( $V_{d,min}$ ) is estimated from the descriptive Donachie model (1, 2)

$$V_{d,min} = 2\bar{V}_u / 1.433 \quad (3)$$

where  $\bar{V}_u$  is median volume of the individual cell. If the individual cell volume becomes larger than  $V_{d,min}$ , the parent cell divides to form two new daughter cells. The individual cells are assumed to

be cylindrical. It should be noted that the actual growth kinetics of an individual cell could be limited due to substrate availability within the corresponding mathematical mesh grid ( $M_s$ ) at each time step ( $\Delta t$ ), therefore if the uptake driven by Eq. 1 is higher than the available substrate in a given

mesh grid ( $M_s < \sum_n \Delta t \cdot v_i^s$ ) then the available substrate is equally shared among the number of individuals inhabiting the same grid ( $v_s^{tot} = M_s / n$ ,  $n$ : number of individuals within the same mesh grid). The biological parameters of the kinetic microbial model are summarized in Table S2.

### **- Microbial active movement**

Microbial movement is modeled as flagellar motility: a self-propulsive force guided by chemotaxis that is driven by local substrate gradients (4, 5). The model describes the chemotactic movement of an individual bacterium as a biased-random walk where the flagella propel a straight “run” by rotational movement of the motor and then reverse the rotation to switch direction with a probability to change direction (“tumble”) set by surrounding gradients of chemo-attractants. To apply chemotactic movement to a single cells at each time step, the sensitivity of microbial chemoreceptors towards higher substrate concentrations are described using a receptor model that uses the specific growth rate as the chemotactic potential (6, 7). Thus in the biased random walk of an individual bacterium, the probability of transition  $P_{move,i}$  in a tumbling event into a new (more favorable) direction from the current direction (previous “run”) is expressed quantitatively as(1, 6):

$$p_i(t) = \frac{\exp(\alpha_i \nabla \mu_j(t))}{\sum_{j=1}^m \exp(\alpha_i \mu_j(t))}, \quad (4)$$

$$\alpha = \frac{X_0}{2\mu_{max} v_{1D}} \quad (5)$$

Where  $\alpha$  is the factor for the chemotactic motion derived from chemotactic sensitivity coefficient of individual cells to chemo-attractant(8),  $X_0$ , bacterial maximum swimming velocity,  $v_{1D}$  and the growth rate in response to local gradients,  $\mu_{max}$ .  $m$  is the number of directions that an individual cell could sense the local gradients in 3D surrounding environment. To insure homogenous directional choices, the summation is evaluated with a relatively high number of possible directions ( $m=20$ ) for each individual cells, but still minimize the computational burden when scaling up to systems of thousands of individual cells. The distance in which bacterial cells sense the local chemotactic gradient is set equal to the cell length (1  $\mu\text{m}$ ). Periodic boundary conditions are assumed for bacterial cells that pass outside of the external boundaries of the system.

#### - **Bacteria-particle interactions**

Bacterial cells are allowed to attach to the particle surface when they stochastically encounter a particle. The model assumes that bacterial cells attach to the particle upon surface encounter, but that a set detachment probability for each cell allows them to detach from the particle. This probability is invariant over the span of time that cells are surface-associated, which contrasts with many characterized mechanisms by which bacteria form irreversible contact with surfaces(9). To determine the effect of this simplifying assumption on the behavior of the model, we simulated a wide range of detachment probabilities, ranging from no detachment to a relatively high detachment rate (1% detachment probability per second).

#### - **Enzymatic activity**

Cells in the model can broadcast extracellular enzymes. These enzymes are produced at a constant rate that is proportional to the biomass of the cell, and the enzymes can diffuse into the bulk environment. Enzymes that come into contact with the particle subsequently hydrolyze polysaccharide and release diffusible oligosaccharide at the particle surface. The production of enzyme is assumed to be activated when a bacterial cell adheres to the particle, or when cells take up enough oligosaccharide to support cell maintenance. That is, the process of enzyme

production preserves mass conservation. We modeled the rate that an individual cell  $i$  broadcasts enzymes ( $S_{E,i}$ ) as a conditional linear function to its biomass,  $B_i$ :

$$\begin{cases} S_{E,i} = K_{enz} \cdot B_i & (v_i^{act} - B_i m_i) > 0 \\ S_{E,i} = 0 & (v_i^{act} - B_i m_i) \leq 0 \end{cases} \quad (6)$$

where  $v_i^{act}$  is the actual uptake rate by the individual cell and  $m$  is the maintenance rate. The diffusion of the enzyme is then solved by considering a source term, equivalent to the total production rate of enzymes from a corresponding cubic mesh grid ( $S_{E,P} = \sum_{i=1:n} S_{E,i}$ ) and a sink term on the particle source that gives the rate at which enzymes adhere ( $S_{A,P}$ ).

#### - Oligosaccharide diffusion-reaction in physical domain

The model explicitly simulates the loss of oligosaccharides to the bulk environment due to the presence of diffusion or flow. This is implemented by absorbing conditions at the boundaries of the simulated domain in which oligosaccharides that arrive at a boundary are lost to the bulk environment with no accumulation. This is a relevant assumption for many aquatic and terrestrial ecosystems, in contrast to soil ecosystems under dry conditions that may impose restrictions on substrate diffusivity or to batch culture where substrates accumulate in culture vessels.

The transport and uptake of depolymerized oligosaccharides are modelled based on Fick's law of diffusion and mass conservation. We modeled a 3D cubic volume around a single particle and assumed that diffusion of oligosaccharides is the main mode of mass transport. The reaction-diffusion equation is then numerically solved by the finite-difference method, assuming a regular cubic mesh discretization(10):

$$\frac{\partial C_s}{\partial t} = D_s \nabla^2 C_s - S_s + S_{dp} \quad (7)$$

where  $C_s$  is the concentration of oligosaccharide (the product of enzymatic hydrolysis),  $D_s$  is the diffusion coefficient of the oligosaccharide and  $S_s$  is the oligosaccharide consumption rate (the

sink term) due to microbial uptake.  $S_s$  is a summation of the individual uptake rates ( $v^s$ ), of all cells within a cubic mesh grid ( $S_s = \sum_{i=1:n} v_i^s$ ,  $n$  is the number of individuals in the mesh grid) (10).  $S_{dp}$  is the production rate of oligosaccharide given from Eq. 11. The mesh grid size is chosen to be approximately 10  $\mu\text{m}$  and the time step used for simulating oligosaccharide diffusion is the same as that used for chemotactic movement for computational simplicity. For each time step, the Dirichlet boundary conditions are applied for the particle surface and the external boundaries of the cubic volume around the particle. The concentration of oligosaccharide at the external boundaries is set to zero to create an absorbing boundary condition (eliminate the accumulation of oligosaccharide). This external boundary condition is similar to that used for enzyme diffusion (below). A convective term is added in the case where oligosaccharide transport processes are modeled around sinking particles.

- **Enzyme diffusion and decay:**

The diffusion of the enzyme is solved by considering a source and sink term, similar to Eq. 7:

$$\frac{\partial C_E}{\partial t} = D_E \nabla^2 C_E + S_{E,P} + S_{A,P} \quad (8)$$

here  $C_E$  is the concentration of broadcast enzyme and  $D_E$  is the enzyme diffusion coefficient. Both values can be measured experimentally. In this study, we used an empirical model that decomposes the diffusion coefficient into components describing the viscosity  $\eta$  and temperature  $T$  of the medium, and molecular weight of the molecule,  $M_w$   $D_E = 1.7 \times 10^{-7} T / M_w^{0.41} \eta$  ( $\text{cm}^2/\text{s}$ ) (11).

In Eq. 7, the adhesion of enzyme to the particle acts as a sink term at the particle surface boundary that is equivalent to the total amount of enzyme that arrives at the particle surface (

$S_{A,P} = \int_r^{r+\Delta L} C_E dA$ ,  $\Delta L$ ). The total enzyme accumulation on the particle ( $S_{E,surf}$ ) is then the integral

of the accumulation rate  $S_{A,P}$  minus the decay rate,  $S_{E,D}$ :

$$S_{E,surf.} = \int_0^t S_{A,P} dt - S_{E,D} \quad (9)$$

Enzymes are assumed to decay with first order kinetics, so  $S_{E,D}$  is a function of the amount of enzyme adsorbed to the particle surface (12–14):

$$S_{E,D} = K_l S_{E,surf.} \quad (10)$$

where  $K_l$  is defined as the enzyme decay coefficient. The depolymerization rate of polysaccharide ( $S_{dp}$ ) to oligosaccharide is therefore a function of the particle-adsorbed enzyme ( $S_{E,surf.}$ ) with a linear empirical relationship:

$$S_{dp} = K_p S_{E,surf.} \quad (11)$$

where  $K_p$  is the particle lability, defined as a lumped parameter that resembles biopolymer and enzyme biochemistries. That is,  $K_p$  is a combined term to express difficulty of the particle to degrade relative to the activity of the enzyme produced by the cells.

#### - **Oligosaccharide transport in presence of particle sinking**

Where we model sinking in our simulations, we assume a constant 1D speed rate along the water column,  $u$  (see Figure S2). For simplicity, we assume that particle sinking creates a laminar flow around the particle with the same speed as the particle sinking rate. To address the effect of flow on dissolved carbon and enzyme transport, we added advection term to diffusion model for dissolved carbon and enzyme transports (Eq. 7 and 8) and expressed as:

For dissolved carbon:

$$\frac{\partial C_s}{\partial t} + u \cdot \nabla C_s = D_s \nabla^2 C_s - S_s + S_{dp} \quad (12)$$

For enzyme:

$$\frac{\partial C_E}{\partial t} + u \cdot \nabla C_E = D_E \nabla^2 C_E + S_{E,P} + S_{A,P} \quad (13)$$

where sink and source terms for dissolved carbon ( $S_s, S_{dp}$ ) and enzyme ( $S_s, S_{dp}$ ) are all expressed similar to equations 7 and 8, respectively. Similar to the stationary particle simulations, the mesh grid size was chosen to be approximately 10  $\mu\text{m}$  and the time step of calculations was assumed to be similar to the time interval between chemotactic tumbles (5 seconds). At each time step, Dirichlet boundary conditions are applied to the particle surface to the external boundaries of the volume around the particle. At the particle surface a constant flux of oligosaccharide production is considered as given by Eq. 11, as a function of the particle-adsorbed enzyme concentration. As above, the external boundary concentration of oligosaccharide is set to zero. A finite difference method was applied to solve Eq. 12 and 13 simultaneously at each time step within the individual-based model.

To model longer time scales of particle sinking in the water column, we implemented a spatial algorithm to only model the effective zone around the particle in each time scale, instead of modeling the whole water column for each time step. Based on analyzing the concentration gradient of dissolved carbon around the particle, we chose the effective zone as the region with concentration gradient above a threshold value (5% of maximum chemical gradient).

### **‘Population-level’ analytical bottom-up model**

We developed a simple quantitative model to predict the fold change in biomass based on the measurable physiological and behavioral features of marine bacterial isolates. The model is parametrized based on the rate of enzyme production and activity (degradation rate) to predict the total of oligosaccharide production,  $M$  over time, given by a first order kinetics:

$$\frac{\partial M}{\partial t} = K_{p,m} B_p + V_{\max} B_p \frac{[C]}{[C] + K_s} - \lambda M \quad (14)$$

where  $K_{p,m}$  is the rate of oligosaccharide production per biomass per hour that is derived from enzymes bound to the cell membrane.  $B_p$  is the particle-associated fraction of the total biomass.

$V_{\max} = \frac{\mu_{\max}}{Y}$  is the maximum uptake rate, defined as the ratio of maximum growth rate of bacteria,  $\mu_{\max}$  to the yield of substrate conversion to biomass,  $Y$  (experimentally measured).  $\lambda$  is the fraction of monomers that are lost to the bulk environment and the value is assumed from what we obtained from simulations individual based model, and from previous reports in the literature about the inefficiency of oligosaccharide recovery from hydrolysis(15). Note that as the enzyme is tethered to the cells, no diffusion for the enzyme is assumed.

The particle-associated biomass production rate is represented based on the combination of Monod-type growth kinetics and attachment-detachment frequencies:

$$\frac{\partial B_p}{\partial t} = \mu_{\max} B_p \frac{[C]}{[C] + K_s} + \alpha_a B_F - \alpha_d B_p \quad (15)$$

where  $B_F$  is the fraction of free living biomass in the system,  $[C]$  is the oligosaccharide concentration, calculated from the ratio of oligosaccharide mass,  $M$  to the volume.  $K_s$  is the half-saturation constant, experimentally measured in Figure 4.  $\alpha_a$  and  $\alpha_d$  are attachment and detachment rates, respectively, measured in Figure 4.

The change in free-living biomass,  $B_F$  is derived from the frequency of attachment and detachment, assuming that free-living cells do not themselves grow:

$$\frac{\partial B_F}{\partial t} = \alpha_d B_p - \alpha_a B_F \quad (16)$$

## Experimental methods:

**Culture conditions:** Strain *psych6C06* was previously isolated from an enrichment of nearshore coastal seawater (Nahant, MA, USA) for surface-associated chitin degrading microbial communities (16, 17). The strain was maintained as colonies on Marine Broth 2216 (Difco 279110)



with 1.5% agar (BD 214010). To establish exponential growth without hysteresis, we modified a culturing protocol previously developed for *Escherichia coli* K12(18). Briefly, single colonies were picked and transferred to 3 mL liquid Marine Broth 2216 and incubated at 25 °C, shaking at 150 rpm on a VWR DS-500 orbital shaker to establish seed cultures. Seed cultures were harvested after ~5 hours by centrifugation for 1 min at 5000 rcf (Eppendorf 5415D, Rotor F45-24-11). The supernatant was discarded and serial dilution of the cells were used to establish pre-cultures in pH 8.2 minimal media supplemented with 20 mM *N*-acetylglucosamine (GlcNAc). The core minimal medium contained the major ions present in seawater, plus vitamins and trace minerals per L: 20 g NaCl, 3 g MgCl<sub>2</sub>·6H<sub>2</sub>O, 0.15 g CaCl<sub>2</sub>·2H<sub>2</sub>O, 0.05 g KCl, 2.1 mg FeSO<sub>4</sub>·7H<sub>2</sub>O, 30 µg H<sub>3</sub>BO<sub>3</sub>, 100 µg MnCl<sub>2</sub>·4H<sub>2</sub>O, 190 µg CoCl<sub>2</sub>·6H<sub>2</sub>O, 2.2 µg NiSO<sub>4</sub>·6H<sub>2</sub>O, 2.7 µg CuSO<sub>4</sub>, 144 µg ZnSO<sub>4</sub>·7H<sub>2</sub>O, 36 µg Na<sub>2</sub>MoO<sub>4</sub>·2H<sub>2</sub>O, 25 µg NaVO<sub>3</sub>, 25 µg NaWO<sub>4</sub>·2H<sub>2</sub>O, 2.5 µg SeO<sub>2</sub>, 100 µg riboflavin, 30 µg D-biotin, 100 µg thiamine-HCl, 100 µg L-ascorbic acid, 100 µg Ca-D-pantothenate, 100 µg folate, 100 µg nicotinate, 100 µg 4-aminobenzoic acid, 100 µg pyridoxine HCl, 100 µg lipoic acid, 100 µg nicotinamide adenine dinucleotide, 100 µg thiamine pyrophosphate, and 10 µg cyanocobalamin. In addition to the carbonate buffer present in the core minimal medium (reflecting natural seawater buffering capacity), we added 50 mM HEPES buffer pH 8.2 to control for the effects of heterotrophic metabolism on pH. We also supplemented the core medium with 10 mM NH<sub>4</sub>Cl, 1 mM Na<sub>2</sub>PO<sub>4</sub>, and 1 mM Na<sub>2</sub>SO<sub>4</sub> to create a carbon-limited minimal medium. Where appropriate, we supplemented with the *N*-acetyl-D-glucosamine (GlcNAc, concentrations indicated). Following overnight growth at 25 °C, the cell density was measured in 1 cm cuvettes by absorbance measurements at 600 nm ( $A_{600}$ ) using a Genesys 20 spectrophotometer. Under these conditions,  $A_{600} 1.0 = 8 \times 10^8$  cells/mL, measured by serial dilution and plating.

### **Measurement of Monod growth parameters**

To measure the maximum growth rate on *N*-acetylglucosamine, cells were prepared as described above, then diluted to a starting OD of 0.01 in 4 mL of fresh 20 mM GlcNAc minimal medium in a 10 mL vented polystyrene tube. OD measurements were taken over time, and the linear

regression was fit to the plot of  $\ln(\text{OD})$  vs. time. The slope of this regression is equivalent to the growth rate. To measure the half-saturation constant, cells were diluted from a OD 1.0 by a factor of  $10^5$  into carbon-free minimal medium, and 100  $\mu\text{l}$  of this dilution was used to inoculate medium containing a GlcNAc at different concentrations (221, 110, 22, 11, 2.2, 1.1, 0.22, 0.11, and 0.02 mg/L). At 2 h time intervals, 100  $\mu\text{l}$  of this culture was plated onto MB 1.5% agar plates. The resultant colonies were counted, and the change in colony numbers over time was used to derive the growth rate. The maximum growth rate measured by this method was the same as that measured using the OD-based approach. Growth rate was plotted against carbon concentration (S), and the experimentally measured  $\mu$  and S were fit to the Monod growth equation ( $\mu = \mu_{\text{max}}(S/(S+K_{\text{sat}}))$ ) to derive parameters  $\mu_{\text{max}}$  and  $K_{\text{sat}}$  using a least squares fit with a maximum of 1000 iterations. The biomass yield during growth on GlcNAc was derived from direct measurement of cell density ( $A_{600}$ ) and GlcNAc depletion using the dinitrosalicylic acid reagent method to colorimetrically quantify reducing sugars in cell-free media(19). The grams of GlcNAc/mL depleted from the media was plotted against the grams of cells produced, at timepoints covering three population doublings ( $A_{600}$  0.1-0.8). The mass of an individual cell was assumed to be 19 fg: a value which we derived from scaling the measured mass of individual *E. coli* with 60 min doubling time (220 fg)(20), scaled by the growth rate of *psych6C06* (0.35), and also by a factor of 4 to reflect a linear increase in biomass per cell with osmolarity between M9 and seawater (21), divided by the ratio of *E. coli* to *psych6C06* volume (16:1).

### **Measurement of protein production and enzymatic activity**

To collect secreted protein and cell-associated protein, cells were grown in large batches. Cells were prepared for inoculation as described above, and inoculated at an initial density of OD 0.01 into 250 mL Erlenmeyer flasks containing 150 mL of 20 mM GlcNAc minimal medium. The flasks were grown with 150 rpm shaking at 25 °C. Periodically, 25 mL of culture was removed from the flasks, and centrifuged at 5000 rcf for 10 min at 4 °C. Sampling was stopped when the culture flask volume reached 75 mL, past which point culture growth rate was affected by volume. The

supernatant was collected and filtered through 0.2  $\mu\text{m}$  Sterivex filters (EMD Millipore), at which point protease inhibitor (Roche cOMplete) was added. The pellet was immediately frozen at  $-20$  until quantification. Supernatant was concentrated  $\sim 50$  x in Amicon Ultra 3 kDa centricon tubes (EMD Millipore), and rinsed twice with 12 mL of carbon-free minimal media to remove small molecules and other potentially inhibitory compounds. The protein abundance was quantified by measuring absorbance at 280 nm, with a 340 nm pathlength correction on a Nanodrop spectrophotometer (Thermo Scientific). The quantification was calibrated using standards of from proteins of known concentration (BSA, chitinase). Chitinase activity was quantified using Methylumbelliferyl(MUF)-conjugated substrates N,N'-diacetyl- $\beta$ -D-chitobioside, N-acetyl- $\beta$ -D-glucosaminide, and  $\beta$ -D-N,N',N''-triacetylchitotriose (Sigma CS1030). Briefly, 2.5  $\mu\text{L}$  of a 20  $\mu\text{g}/\text{mL}$  stock of each substrate in DMSO was added to 197.25  $\mu\text{L}$  of concentrated protein or crude cell lysate in chitinase assay buffer (carbon-free minimal medium, pH 8.2 with no vitamins, ammonium or phosphate). The amount of cell lysate was normalized prior to assay. The accumulation of fluorescence (ex 360-20/em 450-20) was monitored on a Tecan Spark at 25  $^{\circ}\text{C}$  with a cooling module, by measuring fluorescence signal accumulation at 2-minute intervals with continuous shaking at 54 rpm with 6 mm amplitude. Serial dilutions of an unconjugated 4-Methylumbelliferone standard were used to establish the linear range of the instrument and to convert fluorescence intensity into mg/mL of released oligosaccharide.

### **Bacterial colonization on particles**

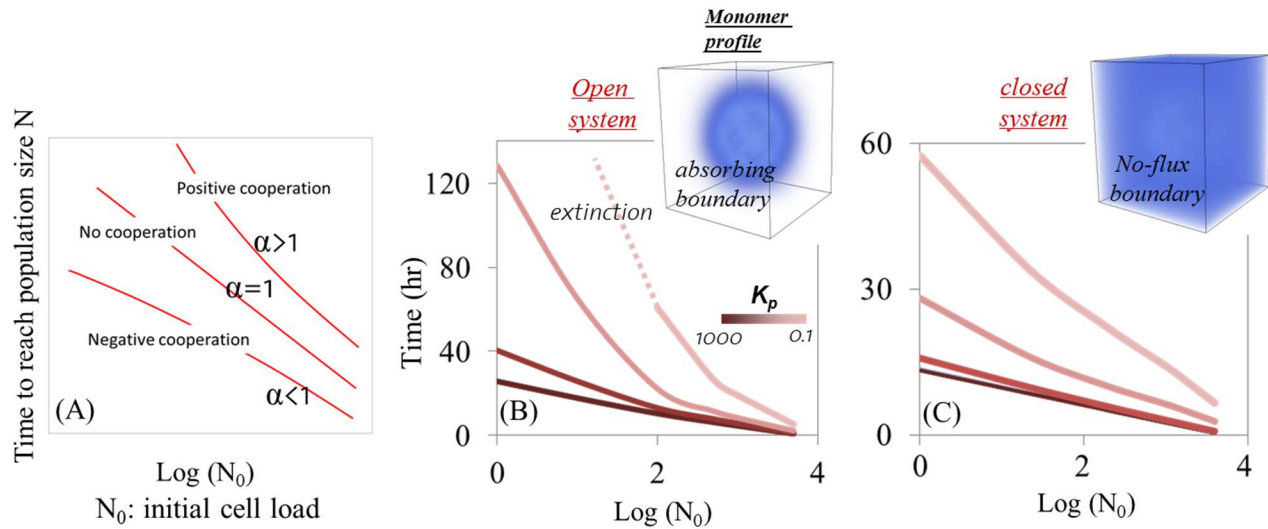
Precultures with absorbance between 0.1-0.3 were prepared as described above and used to colonize magnetic chitin hydrogel beads. To prepare the beads, 500  $\mu\text{L}$  of bead slurry was washed 3 times with carbon-free minimal media using magnetic pulldown. The washed beads were further diluted 1:3 and used to fill 15 mL conical tubes. The particle density in the tubes was counted, and all replicates contained between 200-250 particles per mL with size range from 40 to 100  $\mu\text{m}$  in diameter. This density of beads is consistent with previous studies of community assembly(17),

and provides the equivalent of about 100  $\mu\text{M}$  of the monomeric unit of chitin, *N*-acetylglucosamine (GlcNAc) to the system. The beads were rotated end over end, so that they fell through the medium due to gravity and remained constantly suspended. Because of the rotation, we were unable to continuously observe the beads and instead sub-sampled the population and made individual measurements of multiple beads at each sampled timepoint. The density of inoculated cells was set to be at an  $A_{600}$  of 0.01, diluted from 20 mM GlcNAc minimal medium cultures prepared as described above. A vertical wheel (Stuart S3B, 10" diameter wheel) was used to rotate the 15 mL tubes at 5 rpm at room temperature (21-25 °C) with overhead rotation. To visualize particles and their surface-associated bacteria, 200  $\mu\text{l}$  subsamples were stained with the DNA-intercalating dye SYTO9 (Thermo Fisher, S34854) at a 1:285 dilution of the stock in 96-well plates with optically clear plastic bottoms (VWR 10062-900). To avoid evaporation from the wells, sterile self-adhesive sealing films were used to seal the 96-well plates.

### **Confocal microscopy and image processing:**

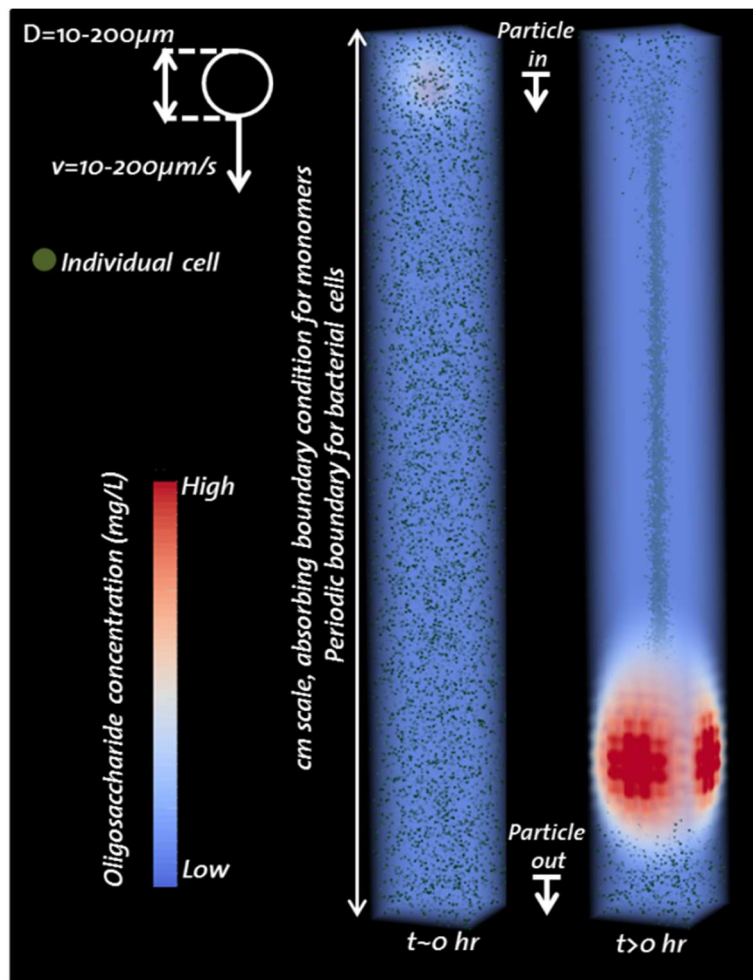
Microscopy was performed on micro-confocal high-content imaging system (ImageXpress Micro Confocal, Molecular Devices), using the 60  $\mu\text{m}$  pinhole spinning disk mode. Fluorescent signal was visualized with a LED light cube (Lumencore Spectra X light engine), and bandpass filters (ex 482/35 nm em 538/40 nm dichroic 506 nm), with a 40x objective (Nikon Ph 2 S Plan Fluor ELWD ADM 0.60 NA cc 0-2 mm, correction collar set to 1.1), and a sCMOS detector (Andor Zyla). To visualize individual particles, particles were manually centered in the field of view and then 100  $\mu\text{m}$  image stacks sampled at Nyquist were acquired in the Z plane using MetaXpress software (version revision 31201). Image analysis was performed in MATLAB (release 2018a). Briefly, image stacks were split in half and a maximum intensity projection was obtained for each half. The low level of fluorescent signal associated with free dye in the hydrogel particles was used to define an intensity threshold suitable to create a binary mask for the particle projections. A mask of the cells within the beads was then defined using their brighter fluorescence intensity. We used

this segmentation to quantify the total surface area occupied by the cells on the bead, and to quantify the total surface area occupied by patches (areas where cells contact other cells  $>10 \mu\text{m}^2$ ).

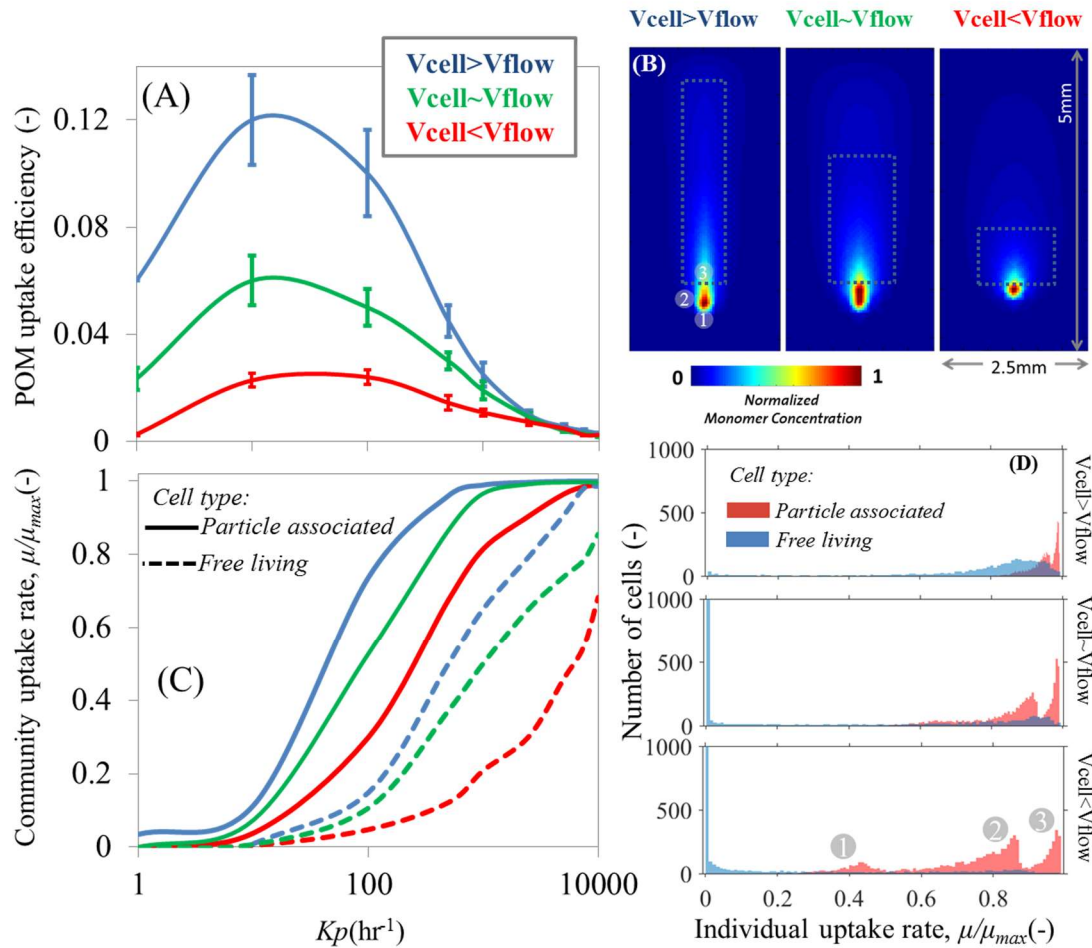


**Figure S1.** Microbial cooperative behavior affected by batch (closed system with no flux boundary conditions for bacterial cells, dissolved carbon and enzymes) vs. open (absorbing boundary conditions for dissolved carbon and enzyme and periodic boundary for bacterial cells) systems. (A) Cooperative behavior is described as dependency of bacterial growth rate to cell initial load ( $N_0$ ). It is defined based on divergence from exponential growth behavior ( $\text{Log}(\frac{N}{N_0}) = \beta t^\alpha$ ) in which

$\alpha$  is 1 for exponential growth and  $\alpha > 1$  shows positive cell dependency of growth rate (Cooperative growth). (B) and (C) growth rate dependency to initial cell load for open and closed systems, respectively. Dashed line indicates the extinction of bacterial community for cell load lower than a threshold value. Inset images indicate the dissolved carbon concentration profiles for their corresponding scenarios. Simulations are done for a particle size of  $200\mu\text{m}$  and results are plotted for the time that allows reaching to population size of  $10^5$  cells. Our simulations reveal that diffusion in an open system may lead to the emergence of spatial self-organization and cooperative growth among individual cells (defined as positive dependency of growth rate to cell number). However, in a closed environment, like a batch culture environment (often the case for experimental designs) this need not be the case. With limited diffusion, substrates produced from particulate biopolymers could eventually accumulate and be fully consumed (assuming that no other factors limit growth). Such closed conditions lead to a stepwise relationship between POM uptake efficiency and particle lability where above a threshold lability ( $K_p$ ), POM uptake efficiency reaches to 100% and below that no degradation/uptake is expected.

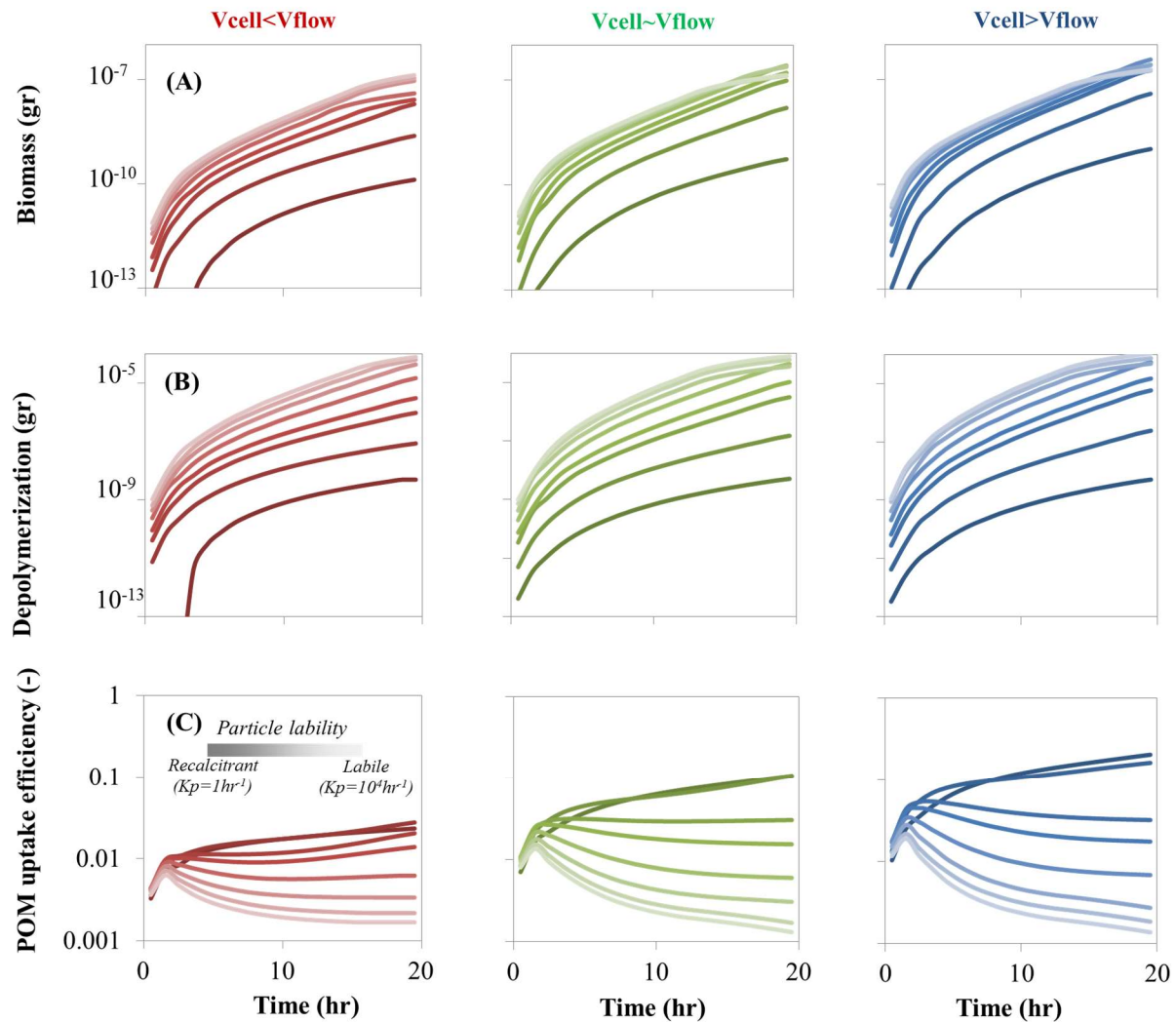


**Figure S2. Model description for particle sinking, oligosaccharide profile and individual cell distributions.** Individual cells are uniformly distributed over the water column. Each individual is assigned dispersal and enzymatic functions.

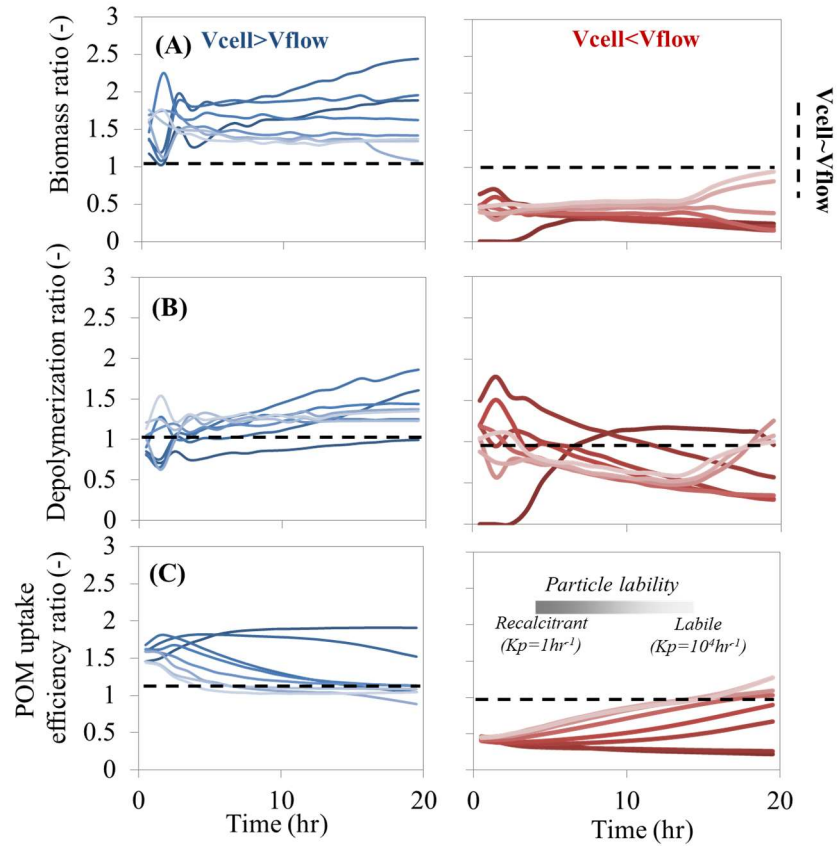


**Figure S3. Particle sinking affects local gradients of dissolved organic matter and POM uptake efficiency of particle associated and free-living bacterial cells.** (A) POM uptake efficiency for a wide range of particle lability affected by the ratio of cell velocity ( $V_{cell}$ ) to particle sinking speed ( $V_{flow}$ ). (B) Spatial patterns of dissolved carbon concentration around the particle. Concentrations are normalized by the maximum local concentrations for each scenario. (C) Mean uptake rate at community level presented for particle associated and free-living bacterial cells. (D) Individual cell uptake rate frequencies. The cell velocity is kept constant at  $10\mu\text{m/s}$  and particle sinking speed is changed from 5 (top) to 10 (middle) and 20 (bottom)  $\mu\text{m/s}$  to evaluate the effects of various ratios. The numbers in grey circle at the peaks of histogram for scenario  $V_{cell} < V_{flow}$  corresponds to the individual cells spatial location around the particle in Figure S8B. The results are shown for simulations after 10 hours. Initial cell number is about  $10^5$  cells per ml uniformly distributed along the depth of the water column.

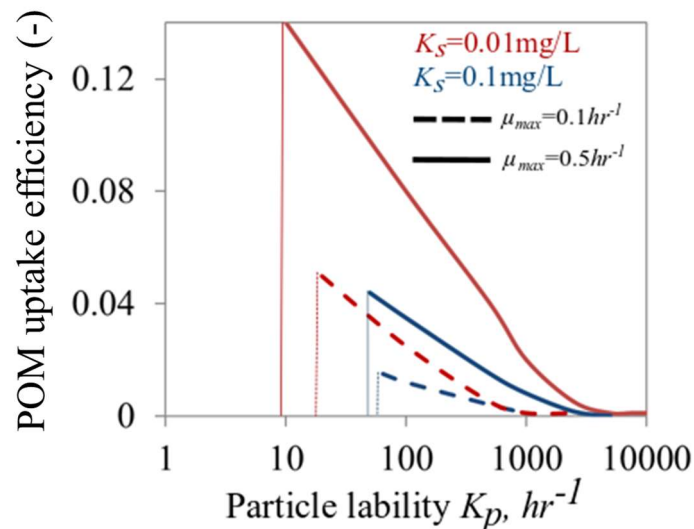




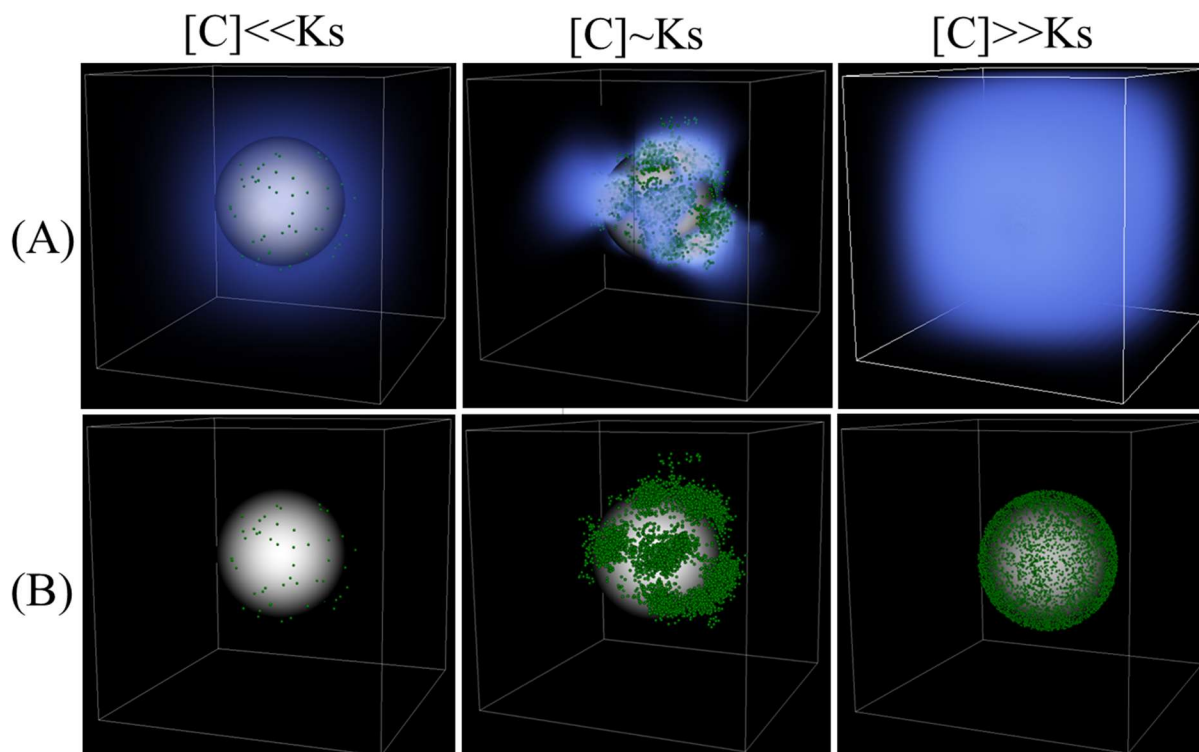
**Figure S4.** Particle sinking affects biomass accumulation (A) and depolymerization (B) rates and uptake efficiency (C). The simulations are performed for 3 ratios of cell to particle sinking speeds. The results are represented for various polymer labilities from labile ( $K_p = 10000 \text{ hr}^{-1}$ ) to recalcitrant ( $K_p = 1 \text{ hr}^{-1}$ ). The cell velocity is kept constant at  $10 \mu\text{m/s}$  and particle sinking speed is changed from 5 to 10 and  $20 \mu\text{m/s}$  to evaluate the effects of various ratios. The results are shown for simulations after 10 hours from initializing the simulations. Initial cell number is  $\sim 10^5$  cells per milliliter volume.



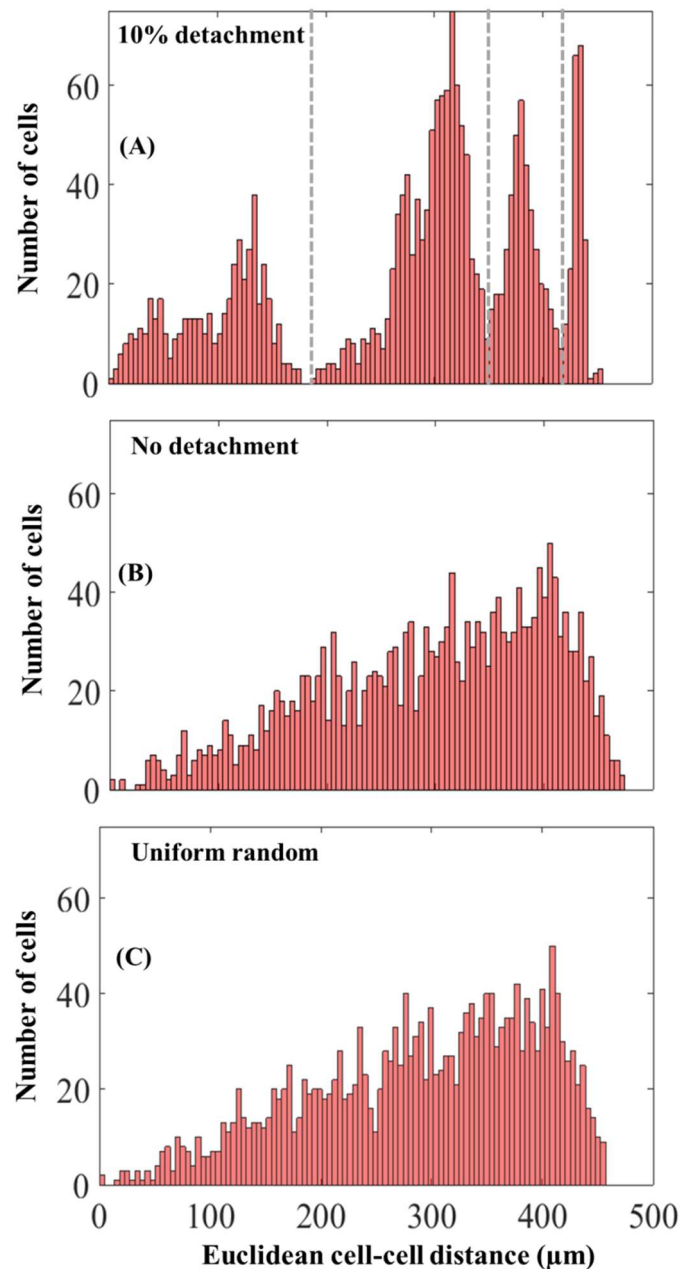
**Figure S5.** Comparing the effects of particle sinking on biomass accumulation (A) and depolymerization (B) rates and uptake efficiency (C) against reference sinking speed. The reference scenario is assumed to be when cell velocity and particle sinking speed are equal ( $V_{cell} \sim V_{flow}$ ). Ratios are calculated by dividing the biomass accumulation, depolymerization and POM uptake efficiency from Figure S9 for two scenarios of particle sinking ( $V_{cell} > V_{flow}$  &  $V_{cell} < V_{flow}$ ) with reference scenario.



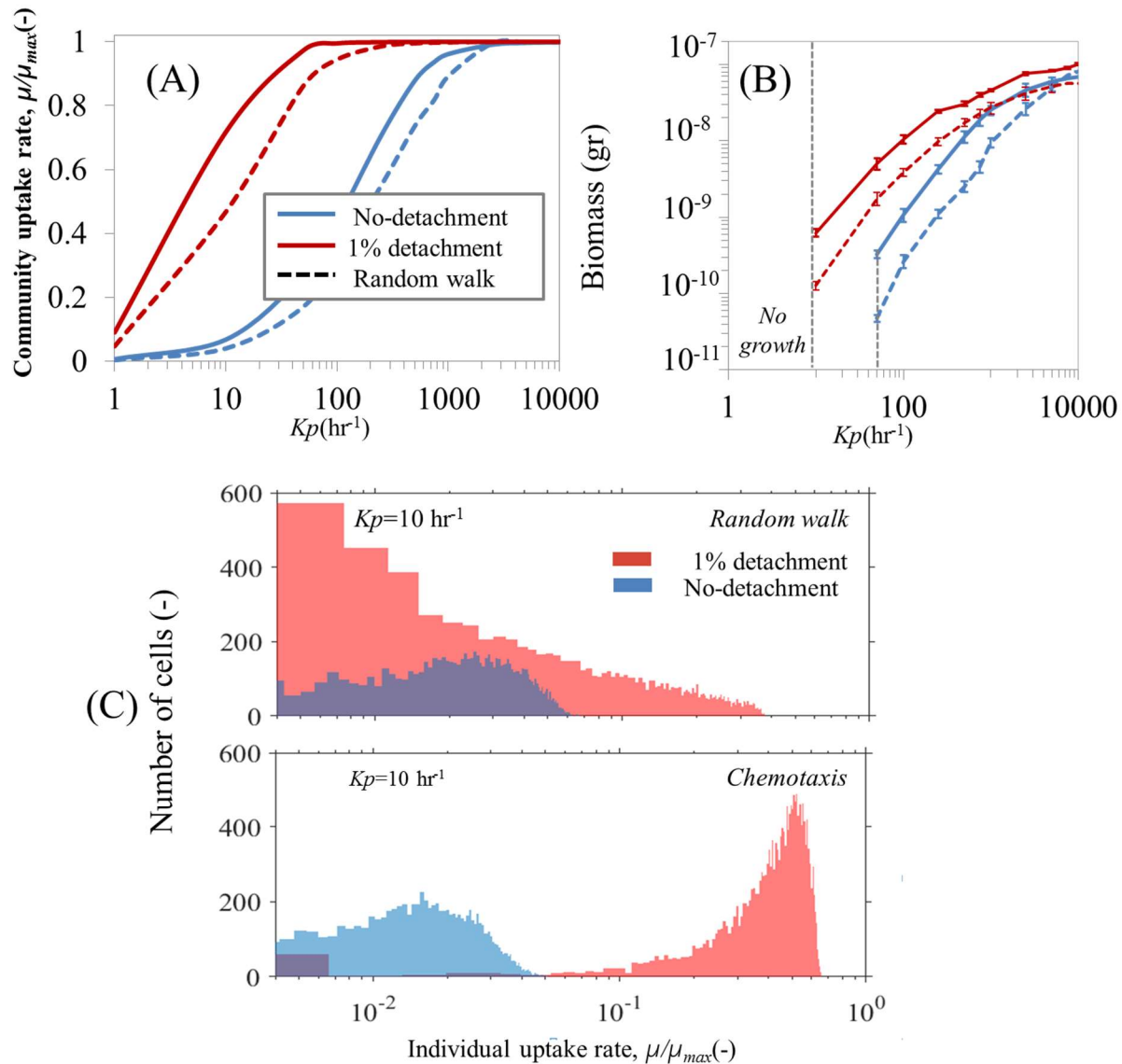
**Figure S6. Physiological properties of bacterial cells (maximum growth rate  $\mu$ , affinity to substrate  $K_s$ ) affect POM uptake efficiency.** POM uptake efficiency as a function of particle lability is shown for high ( $\mu=0.5hr^{-1}$ ) and low ( $\mu=0.1hr^{-1}$ ) maximum growth rates. The results are shown for two levels of affinity to substrate (0.1 and 0.01 mg/L). Vertical lines show the threshold particle lability below which no particle degradation and growth/substrate uptake is observed.



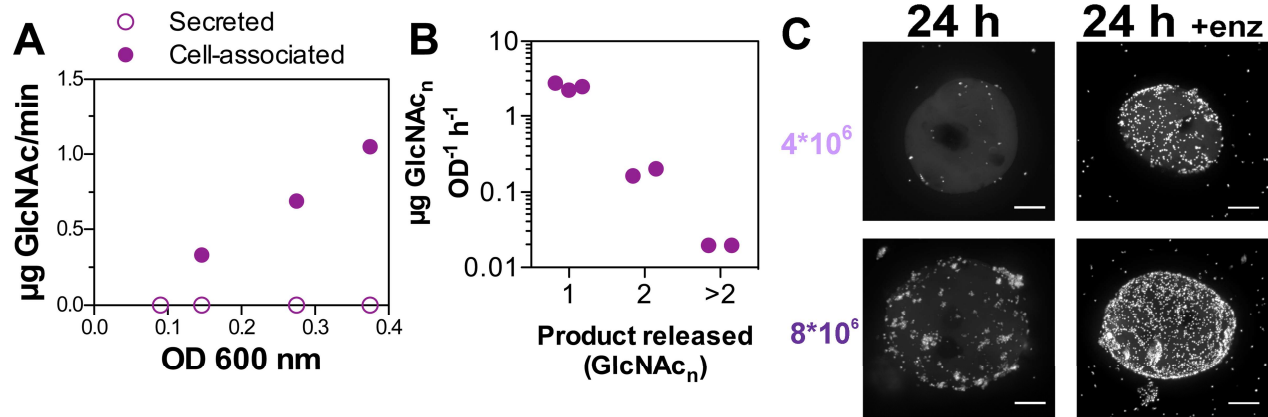
**Figure S7. Spatial self-organization and cooperative growth allow degradation of recalcitrant organic particles.** (A) Dissolved carbon concentration profile (red high and blue low concentrations) projected on microbial population assembly on the particle for three levels of particle lability (recalcitrant I:  $K_p: 1 \text{ hr}^{-1}$ , semi-labile II:  $K_p: 100 \text{ hr}^{-1}$  and labile III:  $K_p: 1000 \text{ hr}^{-1}$ ). Green dots show individual cells and (B) only shows the bacterial cell colonization on the particle. For visualization purposes, carbon concentrations below a certain threshold are not shown (below 1% of maximum concentration). A particle size of  $200 \mu\text{m}$  is considered and initial cell density was assumed to be 1000 cells.



**Figure S8. Detachment changes in the distribution of cell-cell distances within particle associated populations.** Distribution shown for simulations with 10% detachment (A) or without detachment (B). The simulations were run for semi-recalcitrant particles ( $K_p=100 \text{ hr}^{-1}$ ), corresponding to the simulations in Figure 2B with an initial cell density of 1000 and after 10 hours. The dashed grey lines separate single patches. (C) A hypothetical scenario with a uniform random distribution of bacterial cells on the particle surface is shown for comparison. In this comparison, the cell-cell distance was calculated between a reference single cell and randomly selected 2000 cells on the particle.



**Figure S9. Microbial dispersal strategies regulate POM uptake efficiency from organic particle.** (A) Effects of individual cells chemotactic behavior vs. random walk and attachment/detachment frequency to/from particles on Mean carbon uptake rates, represented as a function of particle lability ( $K_p$ ). The mean uptake rate is normalized by the maximum uptake rate of individual cells imposed by their physiological properties. (B) Biomass accumulation after 10 hours. Grey dashed line indicates extinction zone of bacterial population (no biomass accumulation). (C) Histogram of individual cell uptake rates as represented by the number of cells colonizing the particle. The results are shown only for  $K_p=10 hr^{-1}$ . The initial cell number was 10000 cells and the particle size was set to 200  $\mu m$ . The results are shown for simulations after 10 hours.



**Figure S10. Chitinase production by psych6C06 is cell-associated, and exochitinase accounts for most of the activity.** A) Activity of psych6C06 chitinases in culture supernatant, and cell-associated fractions. B) Activity of cell-associated chitinase for three different chitinase substrates (1, exochitinase-specific; 2, chitobiosidase-specific; 3, endochitinase specific). C) Representative images showing the effect of enzyme addition (Figure 3D) on the colonization pattern and density of psych6C06. Images were acquired 24 h after inoculation of chitin beads with  $4 \times 10^6$  cells/mL (light purple, below colonization threshold), or  $8 \times 10^6$  cells/mL (dark purple, above colonization threshold). Scale bars are 20  $\mu\text{m}$ .

**Table S1.** Physiological parameters for microbial growth, metabolism and nutrient concentrations in the individual-based model.

<i>Parameters</i>	<i>Values (Units)</i>
$\mu_{\max}$ : maximum growth rate ( $hr^{-1}$ )	0.2-0.5*
$K_s$ : half saturation (mg/L)	0.01-0.1*
$Y_{\max}$ : growth yield (gr dry mass/gr substrate)	0.5*
cell maintenance	$0.1 \mu_{\max}^{\Gamma}$
cell size ( $\mu m$ )	$1^{\Gamma}$
$\rho$ : cell density ( $mg L^{-1}$ )	$2.9 \times 10^5^{\Gamma}$
$V_{ID}$ : cell velocity at bulk solution ( $\mu m/s$ )	$10^*$
$V_u$ : median cell volume (fl)	$0.4^{\Gamma}$
$K_{enz.}$ : enzyme production (gr enzyme/ gr biomass $hr^{-1}$ )	$0.05^*$
Tumbling frequency ( $s^{-1}$ )	$0.1^{\dagger}$
$X_0$ : chemotactic sensitivity( $mm^2/hr$ )	$446^T$

$^{\Gamma}$  (2)

$^{\dagger}$  (22)

$^T$  (8)

\*Model assumption



**Table S2.** The distribution range of the catalytic activity of various enzymes  $k_{cat}$ . as a measure for particle lability from natural polymeric carbohydrates (Chitin, Alginate, Starch). The data are for various bacterial species with their corresponding abiotic conditions (species name, substrates and environmental conditions)

**Starch ( $\alpha$ -amylase):**

<b>Turnover Number [1/S] , <math>k_{cat}</math>.</b>	<b>Substrate</b>	<b>Species</b>	<b>Commentary</b>	<b>Literature</b>
140	Pullulan	<i>Anabaena sp.</i>	wild-type enzyme	(23)
268	maltotriose	<i>Anabaena sp.</i>	pH 4.5, 90°C, recombinant enzyme	(24)
283.3	beta-cyclodextrin	<i>Anabaena sp.</i>	-	(25)
8.73	4,6-ethylidene-[G7]-p-nitrophenyl-[G1]- $\alpha$ -D-maltoheptaoside	<i>Bacillus licheniformis</i>	wild-type, $k_{cat}/K_m$ : 770000 1/s/M, 25°C, pH 6.0	(26)
34	4,6-ethylidene-[G7]-p-nitrophenyl-[G1]- $\alpha$ -D-maltoheptaoside	<i>Bacillus licheniformis</i>	wild-type, $k_{cat}/K_m$ : 140000 1/s/M, 37°C, pH 4.5	(26)
205	4,6-ethylidene-[G7]-p-nitrophenyl-[G1]- $\alpha$ -D-maltoheptaoside	<i>Bacillus licheniformis</i>	wild-type, $k_{cat}/K_m$ : 1420000 1/s/M, 37°C, pH 6.5	(26)
0.156	starch	<i>Bacillus megaterium</i>	15°C, pH 8.3	(27)
9.8	starch	<i>Bacillus megaterium</i>	mutant enzyme A53S, in the presence of 5 mM $Ca^{2+}$ , in 20 mM Tris-HCl buffer (pH 7.2), at 30°C	(28)
19.9	starch	<i>Bacillus megaterium</i>	wild type enzyme, in the presence of 5 mM $Ca^{2+}$ , in 20 mM Tris-HCl buffer (pH 7.2), at 30°C	(28)
0.025	soluble starch	<i>Bacillus sp.</i>	-	(25)
91.4	starch	<i>Bacillus sp.</i>	pH 5.6, 40°C	(29)
98.3	soluble starch	<i>Bacillus sp.</i>	-	(30)
31.1	beta-cyclodextrin	<i>Fusicoccum sp.</i>	pH 5.6, 40°C	(29)
196	beta-cyclodextrin	<i>Fusicoccum sp.</i>	pH 4.5, 90°C, recombinant enzyme	(24)
3420	soluble potato starch	<i>Fusicoccum sp.</i>	at 30°C in 10 mM MOPS buffer (pH 7.0) with 5 mM $CaCl_2$	(31)
0.02	soluble starch	<i>Geobacillus thermoleovorans</i>	in 50 mM sodium phosphate buffer (pH 6.5) at 75°C	(32)
120	starch	<i>Pseudoalteromonas haloplanktis</i>	wild-type enzyme	(23)

392	4,6-ethylidene-[G7]-p-nitrophenyl-[G1]-alpha-D-maltoheptaoside	<i>Pseudoalteromonas haloplanktis</i>	15°C, pH 7.2	(33)
675	4,6-ethylidene-[G7]-p-nitrophenyl-[G1]-alpha-D-maltoheptaoside	<i>Pseudoalteromonas haloplanktis</i>	pH not specified in the publication, 25°C	(34)
697	4,6-ethylidene-[G7]-p-nitrophenyl-[G1]-alpha-D-maltoheptaoside	<i>Pseudoalteromonas haloplanktis</i>	25°C, pH 7.2	(33)
25.7	starch	<i>Pyrococcus furiosus</i>	wild type enzyme, in the absence of Ca <sup>2+</sup> , in 20 mM Tris-HCl buffer (pH 7.2), at 30°C	(28)
29.4	beta-cyclodextrin	<i>Pyrococcus furiosus</i>	calculated as low molecular weight products	(35)
228	acarbose	<i>Pyrococcus furiosus</i>	pH 4.5, 90°C, recombinant enzyme	(24)
241	acarbose	<i>Pyrococcus furiosus</i>	pH 4.5, 90°C, recombinant enzyme	(24)
1940	Dextrin	<i>Pyrococcus furiosus</i>	at 30°C in 10 mM MOPS buffer (pH 7.0) with 5 mM CaCl <sub>2</sub>	(31)
130	glycogen	<i>Pyrococcus sp.</i>	from rat liver, calculated as low molecular weight products	(35)
230	acarbose	<i>Pyrococcus sp.</i>	short chain amylose, calculated as low molecular weight products	(35)
2280	acarbose	<i>Pyrococcus sp.</i>	pH 6.5, 90°C	(36)
3000	starch	<i>Pyrococcus sp.</i>	calculated as low molecular weight products	(37)
4180	acarbose	<i>Pyrococcus sp.</i>	pH 6.5, 90°C	(36)
4680	starch	<i>Pyrococcus sp.</i>	pH 6.5, 90°C	(36)
1457	starch	<i>Streptomyces megasporus</i>	pH not specified in the publication, 55°C	(34)
0.085	Pullulan	<i>Thermoactinomyces vulgaris</i>	pH 5.6, 40°C	(29)
67	starch	<i>Thermoactinomyces vulgaris</i>	pH 4.5, 90°C, recombinant enzyme	(24)
153	4,6-ethylidene-[G7]-p-nitrophenyl-[G1]-alpha-D-maltoheptaoside	<i>Thermobifida fusca</i>	pH not specified in the publication, 25°C	(34)
1.6	maltotriose	<i>Thermotoga neapolitana</i>	-	(35)
34	4-nitrophenyl alpha-D-maltohexaoside	<i>Thermotoga neapolitana</i>	in 50 mM sodium phosphate buffer (pH 6.5) at 75°C	(32)
37.2	maltopentaose	<i>Thermotoga neapolitana</i>	pH 5.6, 40°C	(29)

1920	rabbit glycogen	<i>Thermotoga neapolitana</i>	at 30°C in 10 mM MOPS buffer (pH 7.0) with 5 mM CaCl <sub>2</sub>	(31)
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## Alginate lyase:

### guluronate-specific alginate lyase

<b>Turnover Number</b> <b>[1/s], <math>k_{cat}</math>.</b>	<b>Substrate</b>	<b>Species</b>	<b>Commentary</b>	<b>Literature</b>
936	alginate	<i>Flavobacterium sp.</i>	pH 8.5, 30°C	(38)
872.8	sodium alginate	<i>Flavobacterium sp.</i>	recombinant His-tagged enzyme, pH 8.5, 45°C	(38)
32	alginate	<i>Pseudomonas aeruginosa</i>	pH 7.1, 25°C	(39)
32	poly-(beta-(1->4)-D-mannuronan)	<i>Pseudomonas aeruginosa</i>	pH 7.1, 25°C	(39)
1.5	acetylated poly-(beta-(1->4)-D-mannuronan)	<i>Pseudomonas aeruginosa</i>	pH 7.1, 25°C	(39)
1.2	acetylated alginate	<i>Pseudomonas aeruginosa</i>	pH 7.1, 25°C	(39)
56.9	sodium alginate	<i>Saccharophagus degradans</i>	pH 7.5, 30°C, wild-type enzyme	(40)
19.51	sodium alginate	<i>Zobellia galactanivorans</i>	with L-guluronate content of 33.3%, pH 7.5, 30°C, recombinant enzyme	(41)
17.89	sodium alginate	<i>Zobellia galactanivorans</i>	with L-guluronate content of 52.6%, pH 7.5, 30°C, recombinant enzyme	(41)
12.66	sodium alginate	<i>Zobellia galactanivorans</i>	with L-guluronate content of 66.7%, pH 7.5, 30°C, recombinant enzyme	(41)

### mannuronate-specific alginate lyase

<b>Turnover Number</b> <b>[1/S], <math>k_{cat}</math>.</b>	<b>Substrate</b>	<b>Species</b>	<b>Commentary</b>	<b>Literature</b>
2.42	alginate	<i>Pseudomonas aeruginosa</i>	protein PA1167, pH 7.5, 37°C	(42)
748	alginate	<i>Pseudomonas aeruginosa</i>	AlgL	(42)
7.6	hepta-alpha1,4-L-guluronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
13.2	hepta-beta1,4-D-mannuronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
7.9	hexa-alpha1,4-L-guluronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
11.9	hexa-beta1,4-D-mannuronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)

7.8	nona-alpha1,4-L-guluronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
10.8	nona-beta1,4-D-mannuronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
7.3	octa-alpha1,4-L-guluronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
11.1	octa-beta1,4-D-mannuronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
6.2	penta-alpha1,4-L-guluronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
7.4	penta-beta1,4-D-mannuronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(42)
0.057	sodium alginate	<i>Saccharophagus degradans</i>	pH 7.5, 30°C, H415A	(40)
56.9	sodium alginate	<i>Saccharophagus degradans</i>	pH 7.5, 30°C, wild-type enzyme	(40)
2.5	tetra-alpha1,4-L-guluronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(43)
3.2	tetra-beta1,4-D-mannuronate	<i>Alteromonas sp.</i>	pH 8.5, 30°C	(43)

### Chitin (Chitinase)

Turnover Number [1/S], $k_{cat}$ .	Substrate	Species	Commentary	Literature
110.6	4-methylumbelliferyl-N,N',N''-triacetylchitotriose	<i>Vibrio parahaemolyticus</i>	pH 7, 37°C, VpChiA	(44)
13.94	4-methylumbelliferyl-N,N'-diacetylchitobiose	<i>Vibrio parahaemolyticus</i>	pH 7, 37°C, VpChiA	(44)
0.383	p-nitrophenol-beta-D-N,N'-diacetylchitobioside	<i>Vibrio harveyi</i>	70°C, pH 7	(45)
2.9	chitohexaose	<i>Vibrio harveyi</i>	pH 5.5, 37°C, wild-type enzyme	(46)
2.37	colloidal chitin	<i>Vibrio harveyi</i>	pH 7.0, 40°C	(47)
1.2	chitin	<i>Vibrio harveyi</i>	pH 5.5, 37°C, colloidal chitin, wild-type enzyme	(46)
0.21	GlcNAc $\beta$ (1-4)GlcNAc $\beta$ (1-4)GlcNAc $\beta$ (1-4)GlcNAc $\beta$ (1-4)GlcNAc	<i>Vibrio harveyi</i>	wild-type, 37°C	(48)
0.19	GlcNAc $\beta$ (1-4)GlcNAc $\beta$ (1-4)GlcNAc $\beta$ (1-4)GlcNAc $\beta$ (1-4)GlcNAc	<i>Vibrio harveyi</i>	wild-type, 37°C	(48)

	4)GlcNAc $\beta$ (1-4)GlcNAc			
0.032	p-nitrophenol- $\beta$ -D-mannopyranoside	<i>Vibrio harveyi</i>	-	(45)
0.005	chitin	<i>Thermococcus chitonophagus</i>	-	(45)
67.4	4-methylumbelliferyl- $\beta$ -D-N,N',N''-triacetylchitotrioside	<i>Serratia marcescens</i>	pH 6.5, isozyme ChiA	(49)
56.8	4-methylumbelliferyl- $\beta$ -D-N,N',N''-triacetylchitotrioside	<i>Serratia marcescens</i>	pH 6.5, isozyme ChiB	(49)
2	4-methylumbelliferyl- $\beta$ -D-N,N',N''-triacetylchitotrioside	<i>Serratia marcescens</i>	pH 6.5, isozyme ChiC	(49)
2.2	$\beta$ -chitin	<i>Serratia marcescens</i>	pH 6.1, 37°C	(50)
1.7	$\beta$ -chitin	<i>Serratia marcescens</i>	pH 8.0, 37°C; pH 8.0, 37°C	(50)
6.7	4-nitrophenyl chitobioside	<i>Pyrococcus furiosus</i>	pH 4.8, 50°C, wild-type enzyme	(51)
0.1	colloidal chitin	<i>Penicillium ochrochloron</i>	wild-type, 37°C	(48)
9.6	4-nitrophenyl N,N'-diacetyl- $\beta$ -D-chitobioside	<i>Paenibacillus thermoaerophilus</i>	pH 7.0, 60°C	(52)
0.99	4-nitrophenyl N,N'-diacetylchitobiose	<i>Paenibacillus sp.</i>	recombinant enzyme, pH 5.5, 37°C	(53)
83.46	chitin	<i>Paenibacillus sp.</i>	pH 5.0, 50°C	(54)
19.54	4-nitrophenyl- $\beta$ -1,4-N,N'-diacetyl-chitobiose	<i>Moritella marina</i>	pH 5.0, 28°C, purified recombinant enzyme	(55)
11.88	4-nitrophenyl- $\beta$ -1,4-N,N'-diacetyl-chitobiose	<i>Moritella marina</i>	pH 5.0, 20°C, purified recombinant enzyme	(55)
7.33	4-nitrophenyl- $\beta$ -1,4-N,N'-diacetyl-chitobiose	<i>Moritella marina</i>	pH 5.0, 15°C, purified recombinant enzyme	(55)
5.46	4-nitrophenyl- $\beta$ -1,4-N,N'-diacetyl-chitobiose	<i>Moritella marina</i>	pH 5.0, 10°C, purified recombinant enzyme	(55)
4.33	4-nitrophenyl- $\beta$ -1,4-N,N'-diacetyl-chitobiose	<i>Moritella marina</i>	pH 5.0, 5°C, purified recombinant enzyme	(55)
0.79	4-nitrophenyl- $\beta$ -1,4-N,N'-diacetyl-chitobiose	<i>Moritella marina</i>	pH 5.0, 0°C, purified recombinant enzyme	(55)
0.5	chitohexaose	<i>Coccidioides immitis</i>	30°C	(56)
18	4-methylumbelliferyl-N,N'-diacetylchitobiose	<i>Chromobacterium sp.</i>	pH 7.0, 37°C, wild-type enzyme	(57)

0.43	4-nitrophenyl-chitobiose	<i>Bacillus licheniformis</i>	wild-type, strain DSM13, pH 6.0, 37°C	(58)
0.31	4-nitrophenyl-chitobiose	<i>Bacillus licheniformis</i>	pH 6.0, 37°C	(59)
0.31	4-nitrophenyl-chitobiose	<i>Bacillus licheniformis</i>	wild-type, strain DSM8785, pH 6.0, 37°C	(58)
0.17	4-nitrophenyl-chitobiose	<i>Bacillus licheniformis</i>	wild-type, strain DSM13, pH 3.0, 37°C	(58)
0.14	4-nitrophenyl-chitobiose	<i>Bacillus licheniformis</i>	wild-type, strain DSM8785, pH 3.0, 37°C	(58)
22.4	4-methylumbelliferyl-tri-N-acetylchitotriose	<i>Bacillus circulans</i>	37°C, pH 6, wild-type	(60)
26.9	reduced chitopentaose	<i>Bacillus circulans</i>	37°C, pH 6, wild-type	(60)
2.15	p-nitrophenyl-N,N'-diacetylchitobiose	<i>Bacillus circulans</i>	pH 5.0, 30°C, wild-type enzyme	(61)
9.55	carboxymethyl chitin	<i>Bacillus circulans</i>	37°C, pH 6, wild-type	(60)
0.0007	4-methylumbelliferyl GlcNAc $\beta$ (1-4)GlcNAc $\beta$ (1-4)GlcNAc	<i>Aeromonas caviae</i>	pH 6.0, 50°C, recombinant His-tagged wild-type enzyme	(62)
0.006	4-methylumbelliferyl-GlcNAc $\beta$ (1->4)GlcNAc	<i>Aeromonas caviae</i>	pH 6.0, 50°C, recombinant His-tagged wild-type enzyme	(62)

**Table S3.** Particle associated cell density at different water solutions and seasons in the Baltic Sea. The data are extracted from Rieck et al., 2015 (69). The data on particle size range and number of particles per one litter of solution are used from field studies (70) to estimate the number of cells per single particles across observed size ranges. The simulations are performed for three initial cell densities for approximately the mean particle lability observed for natural biopolymers ( $K_p \sim 100\text{hr}^{-1}$ , Figure 1B) and POM uptake efficiency is estimated after 10hours. The simulation data are used to predict uptake efficiency from natural marine snow.

	<u>Summer</u>			<u>Winter/fall</u>			<u>Simulations (this study)</u>		
	Marine	Mesohaline	Oligohaline	Marine	Mesohaline	Oligohaline			
<b>Particle-associated cell density [cells L<sup>-1</sup>](69)</b>	0.26 ( $\pm 0.06$ ) $\times 10^8$	0.57 ( $\pm 0.13$ ) $\times 10^8$	0.76 ( $\pm 0.21$ ) $\times 10^8$	0.12 ( $\pm 0.069$ ) $\times 10^8$	0.13 ( $\pm 0.08$ ) $\times 10^8$	0.74 ( $\pm 0.03$ ) $\times 10^8$	1000*	5000*	10000*
<b>Particle size range (mm)(69)</b>	0.2-7(mean 0.5)	0.2-7(mean 0.5)	0.2-7(mean 0.5)	0.2-7(mean 0.5)	0.2-7(mean 0.5)	0.2-7(mean 0.5)	0.2	0.2	0.2
<b>Number of particles per L(70)</b>	$\leq 28$ (mean 6.4)	$\leq 28$ (mean 6.4)	$\leq 28$ (mean 6.4)	$\leq 28$ (mean 6.4)	$\leq 28$ (mean 6.4)	$\leq 28$ (mean 6.4)	----	----	----
<b>Total cell fraction per particle</b>	0.2mm:1.8%; 7mm:98.2%	0.2mm:1.8%; 7mm:98.2%	0.2mm:1.8%; 7mm:98.2%	0.2mm:1.8%; 7mm:98.2%	0.2mm:1.8%; 7mm:98.2%	0.2mm:1.8%; 7mm:98.2%	----	----	----
<b>Cells per particle</b>									
<b>0.2 mm</b>	7.80E+04	1.71E+05	2.28E+05	3.60E+04	3.90E+04	2.32E+05	8.83E+03 <sup>‡</sup>	2.10E+05 <sup>‡</sup>	6.32E+05 <sup>‡</sup>
<b>7 mm</b>	8.51E+07	1.87E+08	2.49E+08	3.93E+07	4.26E+07	2.42E+08	----	----	----
<b>Estimated POM Uptake efficiency</b>									
<b>0.2 mm</b>	0.041	0.057	0.061	0.03	0.032	0.07	0.025	0.064	0.08

<sup>‡</sup> cell numbers after 10hours

\*initial cell number for simulation

## References for Supplementary Information

1. Ebrahimi AN, Or D (2014) Microbial dispersal in unsaturated porous media: Characteristics of motile bacterial cell motions in unsaturated angular pore networks. *Water Resour Res* 50(9):7406–7429.
2. Kreft JU, Booth G, Wimpenny JWT (1998) BacSim, a simulator for individual-based modelling of bacterial colony growth. *Microbiology* 144(12):3275–3287.
3. Monod J (1949) The Growth of Bacterial Cultures. *Annu Rev Microbiol* 3(1):371–394.
4. Mitchell JG, Kogure K (2006) Bacterial motility: links to the environment and a driving force for microbial physics. *FEMS Microbiol Ecol* 55(1):3–16.
5. Son K, Brumley DR, Stocker R (2015) Live from under the lens: Exploring microbial motility with dynamic imaging and microfluidics. *Nat Rev Microbiol* 13(12):761–775.
6. Kim M, Or D (2016) Individual-based model of microbial life on hydrated rough soil surfaces. *PLoS One* 11(1):e0147394.
7. Hillen T, Painter KJ (2009) A user's guide to PDE models for chemotaxis. *J Math Biol* 58(1–2):183–217.
8. Ahmed T, Stocker R (2008) Experimental verification of the behavioral foundation of bacterial transport parameters using microfluidics. *Biophys J* 95(9):4481–4493.
9. Berne C, Ducret A, Hardy GG, Brun Y V. (2015) Adhesins involved in attachment to abiotic surfaces by Gram-negative bacteria. *Microbiol Spectr*. doi:10.1128/microbiolspec.mb-0018-2015.
10. Ebrahimi A, Or D (2018) On upscaling of soil microbial processes and biogeochemical fluxes from aggregates to landscapes. *J Geophys Res Biogeosciences*. doi:10.1029/2017JG004347.
11. Grunwald P (1989) Determination of effective diffusion coefficients - an important parameters for the efficiency of immobilized biocatalysts. *Biochem Educ*. doi:10.1016/0307-4412(89)90018-6.
12. Schimel JP, Weintraub MN (2003) The implications of exoenzyme activity on microbial carbon and nitrogen limitation in soil: A theoretical model. *Soil Biol Biochem* 35(4):549–563.
13. Drake JE, et al. (2013) Stoichiometry constrains microbial response to root exudation- insights from a model and a field experiment in a temperate forest. *Biogeosciences* 10(2):821–838.
14. Ebrahimi A, Or D (2017) Mechanistic modeling of microbial interactions at pore to profile scale resolve methane emission dynamics from permafrost soil. *J Geophys Res*



- Biogeosciences* 122(5):1216–1238.
15. Gore J, Youk H, van Oudenaarden A (2009) Snowdrift game dynamics and facultative cheating in yeast. *Nature* 459(7244):253–256.
  16. Enke TN, Leventhal GE, Metzger M, Saavedra JT, Cordero OX (2018) Microscale ecology regulates particulate organic matter turnover in model marine microbial communities. *Nat Commun* 9(1):2743.
  17. Datta MS, Sliwerska E, Gore J, Polz M, Cordero OX (2016) Microbial interactions lead to rapid micro-scale successions on model marine particles. *Nat Commun* 7(May):11965.
  18. Scott M, Gunderson CW, Mateescu EM, Zhang Z, Hwa T (2010) Interdependence of cell growth and gene expression: Origins and consequences. *Science (80- )* 330(6007):1099–1102.
  19. Miller GL (1959) Use of dinitrosalicylic acid reagent for determination of reducing sugar. *Anal Chem* 31(3):426–428.
  20. Dennis PP, Bremer H (2008) Modulation of chemical composition and other parameters of the cell at different exponential growth rates. *EcoSal Plus* 3(1). doi:10.1128/ecosal.5.2.3.
  21. Baldwin WW, Myer R, Kung T, Anderson E, Koch AL (1995) Growth and Buoyant Density of *Escherichia coli* at Very Low Osmolarities. *J Bacteriol* 177(1), 235-237.
  22. Kiørboe T, Grossart HP, Ploug H, Tang K (2002) Mechanisms and rates of bacterial colonization of sinking aggregates. *Appl Environ Microbiol* 68(8):3996–4006.
  23. Mizuno M, et al. (2004) The crystal structure of *Thermoactinomyces vulgaris* R-47  $\alpha$ -amylase II (TVA II) complexed with transglycosylated product. *Eur J Biochem.* doi:10.1111/j.1432-1033.2004.04183.x.
  24. Yang SJ, et al. (2004) Enzymatic analysis of an amylolytic enzyme from the hyperthermophilic archaeon *Pyrococcus furiosus* reveals its novel catalytic properties as both an  $\alpha$ -amylase and a cyclodextrin-hydrolyzing enzyme. *Appl Environ Microbiol.* doi:10.1128/AEM.70.10.5988-5995.2004.
  25. Champreda V, Kanokratana P, Sriprang R, Tanapongpipat S, Eurwilaichitr L (2007) Purification, biochemical characterization, and gene cloning of a new extracellular thermotolerant and glucose tolerant maltooligosaccharide-forming  $\alpha$ -amylase from an endophytic ascomycete *Fusicoccum* sp. BCC4124 . *Biosci Biotechnol Biochem.* doi:10.1271/bbb.70198.
  26. Liu YH, Lu FP, Li Y, Wang JL, Gao C (2008) Acid stabilization of *Bacillus licheniformis* alpha amylase through introduction of mutations. *Appl Microbiol Biotechnol.* doi:10.1007/s00253-008-1580-5.
  27. Vester JK, Glaring MA, Stougaard P (2015) An exceptionally cold-adapted alpha-amylase

- from a metagenomic library of a cold and alkaline environment. *Appl Microbiol Biotechnol*. doi:10.1007/s00253-014-5931-0.
28. Ghollasi M, Khajeh K, Naderi-Manesh H, Ghasemi A (2010) Engineering of a bacillus  $\alpha$ -amylase with improved thermostability and calcium independency. *Appl Biochem Biotechnol*. doi:10.1007/s12010-009-8879-2.
  29. Kato S, et al. (2007) Molecular cloning and characterization of an  $\alpha$ -amylase from *Pichia burtonii* 15-1. *Biosci Biotechnol Biochem*. doi:10.1271/bbb.70407.
  30. Uma Maheswar Rao JL, Satyanarayana T (2007) Purification and characterization of a hyperthermostable and high maltogenic  $\alpha$ -amylase of an extreme thermophile *Geobacillus thermoleovorans*. *Appl Biochem Biotechnol*. doi:10.1007/s12010-007-0017-4.
  31. Reyes-Sosa FM, Molina-Heredia FP, De La Rosa MA (2010) A novel  $\alpha$ -amylase from the cyanobacterium *Nostoc sp.* PCC 7119. *Appl Microbiol Biotechnol*. doi:10.1007/s00253-009-2191-5.
  32. Park KM, et al. (2010) Characterization of an exo-acting intracellular  $\alpha$ -amylase from the hyperthermophilic bacterium *Thermotoga neapolitana*. *Appl Microbiol Biotechnol*. doi:10.1007/s00253-009-2284-1.
  33. D'Amico S, Gerday C, Feller G (2003) Temperature adaptation of proteins: Engineering mesophilic-like activity and stability in a cold-adapted  $\alpha$ -amylase. *J Mol Biol*. doi:10.1016/j.jmb.2003.07.014.
  34. Cipolla A, Delbrassine F, Da Lage JL, Feller G (2012) Temperature adaptations in psychrophilic, mesophilic and thermophilic chloride-dependent alpha-amylases. *Biochimie*. doi:10.1016/j.biochi.2012.05.013.
  35. De Mot R, Verachtert H (1987) Purification and characterization of extracellular  $\alpha$ -amylase and glucoamylase from the yeast *Candida antarctica* CBS 6678. *Eur J Biochem*. doi:10.1111/j.1432-1033.1987.tb11175.x.
  36. Tachibana Y, Fujiwara S, Takagi M, Imanaka T (1997) Cloning and expression of the 4- $\alpha$ -glucanotransferase gene from the hyperthermophilic archaeon *Pyrococcus sp.* KOD1, and characterization of the enzyme. *J Ferment Bioeng*. doi:10.1016/S0922-338X(97)81134-8.
  37. Dey S, Agarwal SO (1999) Characterization of a thermostable  $\alpha$ -amylase from a thermophilic *Streptomyces megasporus* strain SD12. *Indian J Biochem Biophys*.
  38. Huang L, et al. (2013) Characterization of a new alginate lyase from newly isolated *Flavobacterium sp.* S20. *J Ind Microbiol Biotechnol*. doi:10.1007/s10295-012-1210-1.
  39. Farrell EK, Tipton PA (2012) Functional characterization of AlgL, an alginate lyase from *Pseudomonas aeruginosa*. *Biochemistry*. doi:10.1021/bi301425r.
  40. Park D, Jagtap S, Nair SK (2014) Structure of a PL17 family alginate lyase demonstrates

- functional similarities among exotype depolymerases. *J Biol Chem*. doi:10.1074/jbc.M113.531111.
41. Thomas F, et al. (2013) Comparative characterization of two marine alginate lyases from *Zobellia galactanivorans* reveals distinct modes of action and exquisite adaptation to their natural substrate. *J Biol Chem*. doi:10.1074/jbc.M113.467217.
  42. Yamasaki M, et al. (2004) Structure and function of a hypothetical *Pseudomonas aeruginosa* protein PA1167 classified into family PL-7: A novel alginate lyase with a  $\beta$ -sandwich fold. *J Biol Chem*. doi:10.1074/jbc.M402466200.
  43. Iwamoto Y, et al. (2001) Purification and characterization of bifunctional alginate lyase from *Alteromonas sp.* Strain No. 272 and its action on saturated oligomeric substrates. *Biosci Biotechnol Biochem* 65(1):133–142.
  44. Chuang HH, Lin FP (2007) New role of C-terminal 30 amino acids on the insoluble chitin hydrolysis in actively engineered chitinase from *Vibrio parahaemolyticus*. *Appl Microbiol Biotechnol* 76(1):123–133.
  45. Andronopoulou E, Vorgias CE (2003) Purification and characterization of a new hyperthermostable, allosamidin-insensitive and denaturation-resistant chitinase from the hyperthermophilic archaeon *Thermococcus chitonophagus*. *Extremophiles*. doi:10.1007/s00792-002-0294-3.
  46. Pantoom S, Songsiririthigul C, Suginta W (2008) The effects of the surface-exposed residues on the binding and hydrolytic activities of *Vibrio carchariae* chitinase A. *BMC Biochem*. doi:10.1186/1471-2091-9-2.
  47. Patil NS, Waghmare SR, Jadhav JP (2013) Purification and characterization of an extracellular antifungal chitinase from *Penicillium ochrochloron* MTCC 517 and its application in protoplast formation. *Process Biochem*. doi:10.1016/j.procbio.2012.11.017.
  48. Suginta W, Pantoom S, Prinz H (2009) Substrate binding modes and anomer selectivity of chitinase A from *Vibrio harveyi*. *J Chem Biol*. doi:10.1007/s12154-009-0021-y.
  49. Synstad B, et al. (2008) Expression and characterization of endochitinase C from *Serratia marcescens* BJL200 and its purification by a one-step general chitinase purification method. *Biosci Biotechnol Biochem*. doi:10.1271/bbb.70594.
  50. Hamre AG, Eide KB, Wold HH, Sørli M (2015) Activation of enzymatic chitin degradation by a lytic polysaccharide monooxygenase. *Carbohydr Res*. doi:10.1016/j.carres.2015.02.010.
  51. Tsuji H, et al. (2010) Kinetic and crystallographic analyses of the catalytic domain of chitinase from *Pyrococcus furiosus*- the role of conserved residues in the active site. *FEBS J*. doi:10.1111/j.1742-4658.2010.07685.x.

52. Ueda J, Kurosawa N (2015) Characterization of an extracellular thermophilic chitinase from *Paenibacillus thermoaerophilus* strain TC22-2b isolated from compost. *World J Microbiol Biotechnol.* doi:10.1007/s11274-014-1754-5.
53. Itoh T, et al. (2014) Overexpression, purification, and characterization of *Paenibacillus* cell surface-expressed chitinase ChiW with two catalytic domains. *Biosci Biotechnol Biochem.* doi:10.1080/09168451.2014.891935.
54. Singh AK, Chhatpar HS (2011) Purification and characterization of chitinase from *Paenibacillus* sp. D1. *Appl Biochem Biotechnol.* doi:10.1007/s12010-010-9116-8.
55. Stefanidi E, Vorgias CE (2008) Molecular analysis of the gene encoding a new chitinase from the marine psychrophilic bacterium *Moritella marina* and biochemical characterization of the recombinant enzyme. *Extremophiles.* doi:10.1007/s00792-008-0155-9.
56. Fukamizo T, Sasaki C, Schelp E, Bortone K, Robertus JD (2001) Kinetic properties of chitinase-1 from the fungal pathogen *Coccidioides immitis*. *Biochemistry.* doi:10.1021/bi001537s.
57. Park SK, Kim CW, Kim H, Jung JS, Harman GE (2007) Cloning and high-level production of a chitinase from *Chromobacterium* sp. and the role of conserved or nonconserved residues on its catalytic activity. *Appl Microbiol Biotechnol.* doi:10.1007/s00253-006-0614-0.
58. Songsiriritthigul C, Pesatcha P, Eijsink VGH, Yamabhai M (2009) Directed evolution of a *Bacillus* chitinase. *Biotechnol J.* doi:10.1002/biot.200800258.
59. Songsiriritthigul C, Lapboonrueng S, Pechsrichuang P, Pesatcha P, Yamabhai M (2010) Expression and characterization of *Bacillus licheniformis* chitinase (ChiA), suitable for bioconversion of chitin waste. *Bioresour Technol.* doi:10.1016/j.biortech.2010.01.036.
60. Watanabe T, et al. (2003) Aromatic residues within the substrate-binding cleft of *Bacillus circulans* chitinase A1 are essential for hydrolysis of crystalline chitin. *Biochem J.* doi:10.1042/bj20030419.
61. Suginta W, Songsiriritthigul C, Kobdaj A, Opassiri R, Svasti J (2007) Mutations of Trp275 and Trp397 altered the binding selectivity of *Vibrio carchariae* chitinase A. *Biochim Biophys Acta - Gen Subj.* doi:10.1016/j.bbagen.2007.03.012.
62. Lin FP, et al. (2009) Effects of C-terminal amino acids truncation on enzyme properties of *Aeromonas caviae* D1 chitinase. *Arch Microbiol.* doi:10.1007/s00203-008-0451-x.
63. Rieck A, Herlemann DPR, Jürgens K, Grossart HP (2015) Particle-associated differ from free-living bacteria in surface waters of the baltic sea. *Front Microbiol* 6(DEC). doi:10.3389/fmicb.2015.01297.
64. Möller KO, et al. (2012) Marine snow, zooplankton and thin layers: Indications of a trophic

link from small-scale sampling with the Video Plankton Recorder. *Mar Ecol Prog Ser* 468:57–69.