

# 1 **Diversity begets diversity in microbiomes**

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15

16 **Abstract**

17

18 Microbes are embedded in complex microbiomes where they engage in a wide array of  
19 inter- and intra-specific interactions<sup>1-4</sup>. However, whether these interactions are a  
20 significant driver of natural biodiversity is not well understood. Two contrasting  
21 hypotheses have been put forward to explain how species interactions could influence  
22 diversification. ‘Ecological Controls’ (EC) predicts a negative diversity-diversification  
23 relationship, where the evolution of novel types becomes constrained as available niches  
24 become filled<sup>5</sup>. In contrast, ‘Diversity Begets Diversity’ (DBD) predicts a positive  
25 relationship, with diversity promoting diversification via niche construction and other  
26 species interactions<sup>6</sup>. Using the Earth Microbiome Project, the largest standardized  
27 survey of global biodiversity to date<sup>7</sup>, we provide support for DBD as the dominant  
28 driver of microbiome diversity. Only in the most diverse microbiomes does DBD reach a  
29 plateau, consistent with increasingly saturated niche space. Genera that are strongly  
30 associated with a particular biome show a stronger DBD relationship than non-residents,  
31 consistent with prolonged evolutionary interactions driving diversification. Genera with  
32 larger genomes also experience a stronger DBD response, which could be due to a higher  
33 potential for metabolic interactions and niche construction offered by more diverse gene  
34 repertoires. Our results demonstrate that the rate at which microbiomes accumulate  
35 diversity is crucially dependent on resident diversity. This fits a scenario in which species  
36 interactions are important drivers of microbiome diversity. Further (population genomic  
37 or metagenomic) data are needed to elucidate the nature of these biotic interactions in  
38 order to more fully inform predictive models of biodiversity and ecosystem stability<sup>4,5</sup>.

39 **Main text**

40 The majority of the genetic diversity on Earth is encoded by microbes<sup>8-10</sup> and the  
41 functioning of all Earth's ecosystems is reliant on diverse microbial communities<sup>11</sup>.  
42 High-throughput 16S rRNA gene amplicon sequencing studies continue to yield  
43 unprecedented insight into the taxonomic richness of microbiomes (e.g.<sup>12,13</sup>), and abiotic  
44 drivers of community composition (e.g. pH<sup>14,15</sup>) are increasingly characterised. Although  
45 it is known that biotic (microbe-microbe) interactions can also be important in  
46 determining community composition<sup>16</sup>, comparatively little is known about how such  
47 interactions (e.g. cross-feeding<sup>1</sup> or toxin-mediated interference competition<sup>2,3</sup>) shape  
48 microbiome diversity.

49 The dearth of studies exploring how microbial interactions could influence  
50 diversification and diversity stands in marked contrast to a long research tradition on  
51 biotic controls of plant and animal diversity<sup>17,18</sup>. In an early study of 49 animal  
52 (vertebrate and invertebrate) community samples, Elton plotted the number of species  
53 versus the number of genera and observed a ~1:1 ratio in each individual sample, but a  
54 ~4:1 ratio when all samples were pooled<sup>18</sup>. He took this observation as evidence for  
55 competitive exclusion preventing related species, more likely to overlap in niche space, to  
56 co-exist. This concept, more recently referred to as niche filling or Ecological Controls  
57 (EC)<sup>5</sup> predicts speciation (or, more generally, diversification) rates to decrease with  
58 increasing standing species diversity because of diminished available niche space<sup>19</sup>. In  
59 contrast, the Diversity Begets Diversity (DBD) model predicts that when species  
60 interactions create novel niches, standing biodiversity favors further diversification<sup>6,20</sup>.  
61 For example, niche construction (i.e. the physical, chemical or biological alteration of the

62 environment) could influence the evolution of the species constructing the niche, and/or  
63 that of co-occurring species<sup>21,22</sup>.

64 Empirical evidence for the action of EC vs. DBD in natural plant and animal  
65 communities has been mixed<sup>20,23-26</sup>. Laboratory evolution experiments have sought  
66 general principles by tracking the diversification of a focal bacterial lineage in  
67 communities of varying complexity – but the results have also been varied<sup>27,28</sup>. For  
68 example, diversification of a focal *Pseudomonas* clone was favoured by increasing  
69 community diversity in the range of 0-20 species within the same genus<sup>20,29</sup> but  
70 diversification was inhibited by very diverse communities (*e.g.* hundreds or thousands of  
71 species in natural soil<sup>30</sup>). These experimental results show how interspecific competition  
72 can initially drive diversification<sup>31</sup>, and eventually inhibit diversification as niches are  
73 filled. However, these experiments were restricted to very short evolutionary time scales  
74 (*i.e.* a few dozen mutations at most) in a small number of lineages, and it is unclear if  
75 they can be generalized to natural communities evolving over longer periods, spanning  
76 multiple speciation events and large-scale genomic changes.

77 To test whether natural microbial communities conform to EC or DBD models of  
78 diversification, we used 2,000 microbiome samples from the Earth Microbiome Project  
79 (EMP), the largest available repository of biodiversity based on standardized sampling  
80 and sequencing protocols<sup>7</sup>. All samples were rarefied to 5,000 observations (counts of  
81 16S rRNA gene sequences), as diversity estimates are highly sensitive to sampling  
82 effort<sup>32</sup>. Instead of a phylogenetic approach requiring complex assumptions<sup>33,34</sup>, we use  
83 the equivalent of the Species:Genus (S:G) ratios that Elton used three quarters of a  
84 century ago<sup>18</sup> to infer bacterial diversification rates. Rather than species, we considered

85 16S rRNA gene Amplicon Sequence Variants (ASVs) as our finest taxonomic unit. We  
86 then used a range of taxonomic ratios (ASV:Genus, Genus:Family, Family:Order,  
87 Order:Class, and Class:Phylum) as proxies for diversification of a focal lineage, from  
88 shallow to deep evolutionary time, and plot these as a function of the number of non-focal  
89 lineages (Genera, Families, Orders, Classes, and Phyla, respectively) with which the focal  
90 lineage could interact. A negative relationship is consistent with the EC hypothesis,  
91 whereas a positive relationship is consistent with the DBD hypothesis (**Fig. 1**). We used  
92 generalized linear mixed models (GLMMs) to determine how the diversification of a  
93 focal lineage (*e.g.* its ASV:Genus ratio) is affected by the diversity of other lineages (*e.g.*  
94 non-focal genera) in the community. The effects of environment (as defined by the EMP  
95 Ontology ‘level 3 biomes;’ Methods) and the identity of the focal lineage were included  
96 by fitting these as random effects on the slope and intercept. We also controlled for the  
97 submitting laboratory (identified by the principal investigator) and the EMP unique  
98 sample identifier (*i.e.* if two taxa were part of the same sample). Finally, we repeated  
99 these analyses using a taxonomy-free method based on nucleotide sequence identity  
100 cutoffs (Methods).

101 The DBD model was supported across taxonomic ratios, which all had  
102 significantly positive slopes fitting the diversity-diversification relationship (**Table S1**,  
103 **Supplementary Data file 1 Section 1**), and the vast majority of slope estimates across  
104 different lineages and environments were positive (**Fig. S1**). For example, the most  
105 prevalent phylum across all samples, Proteobacteria, had significantly positive slopes  
106 when fitted with linear models in all environments, except hypersaline and non-saline  
107 sediments (**Fig. 2a**). For each taxonomic ratio, the three most prevalent taxa followed

108 positive slopes in most environments (**Fig. S2-S6**), with only a few instances of  
109 significantly negative slopes (**Fig. 2b**). The predominance of positive slopes is robust and  
110 remains after controlling for data structure and taxonomic assignment (**Fig. S7, S8**;  
111 Supplementary Text), nor are they explained by widely measured abiotic drivers (*e.g.*  
112 pH) that could simultaneously increase both diversity and diversification (**Table S2**;  
113 **Supplementary Data file 1 Section 2**; Supplementary Text). Thus, the EMP data are  
114 broadly consistent with the predictions of a DBD model.

115         The DBD hypothesis rests on the premise that species interactions drive  
116 diversification<sup>5,20</sup>. We therefore expect that lineages that are more tightly associated with  
117 a specific biome (*i.e.* long-term residents) are more likely to have had a long history of  
118 interaction with community members and thus are more likely to experience DBD than  
119 lineages that are not tightly associated with that biome (*i.e.* poorly adapted migrants or  
120 broadly adapted generalists). To test this prediction, we clustered environmental samples  
121 by their genus-level community composition using fuzzy *k*-means clustering (**Fig. 3a**),  
122 which identified three clusters: ‘animal-associated’, ‘saline’, and ‘non-saline’. The  
123 clustering included some outliers (*e.g.* plant corpus grouping with animals), but were  
124 generally intuitive and consistent with known distinctions between host-associated vs.  
125 free-living<sup>7</sup>, and saline vs. non-saline communities<sup>35</sup>. Resident genera were defined as  
126 those with a strong preference for a particular environment cluster, using indicator  
127 species analysis (permutation test,  $P < 0.05$ ; **Fig. 3a**; **Fig. S9**; **Supplementary Data file**  
128 **2**), and genera without a strong preference were considered generalists. For each  
129 environment cluster, we ran a GLMM with resident genus-level diversity (number of  
130 non-focal genera) as a predictor of diversification (ASV:Genus ratio) for residents,

131 generalists, or migrants (residents of one cluster found in a different cluster)  
132 (**Supplementary Data file 1 Section 3**). Resident diversity had no significant effect on  
133 the diversification of generalists ( $z=0.646$ ,  $P=0.518$ ;  $z=0.279$ ,  $P=0.780$ ;  $z=0.347$ ,  
134  $P=0.729$ , respectively for animal-associated, saline and non-saline clusters), but did  
135 significantly increase resident diversification ( $z=7.1$ ,  $P=1.25e-12$ ;  $z=3.316$ ,  $P=0.0009$ ;  
136  $z=7.109$ ,  $P=1.17e-12$ , respectively). Resident diversity significantly decreased migrant  
137 diversification in saline ( $z=-3.194$ ,  $P=0.0014$ ) and non-saline environment clusters ( $z=-$   
138  $2.840$ ,  $P=0.0045$ ), but had no significant effect in the animal-associated cluster ( $z=-0.566$ ,  
139  $P=0.571$ ) (**Fig. 3b**). These results suggest that diversity begets diversification among  
140 lineages sharing the same environment over a long evolutionary time period, but that this  
141 is not the case for lineages that do not consistently occur in the same microbiome and  
142 presumably interact less frequently. The diversification of migrants in a new environment  
143 might even be impeded, presumably because most niches are already occupied by  
144 residents.

145         The positive effect of diversity on diversification should eventually reach a  
146 plateau as niches, including those constructed by biotic interactions, become  
147 saturated<sup>27,30</sup>. In the animal distal gut, a relatively low-diversity biome, we observed a  
148 strong linear DBD relationship at most sequence identity ratios; in contrast, the more  
149 diverse soil biome clearly attained a plateau (**Fig. S10**). To further test the hypothesis that  
150 increasingly diverse microbiomes experience weaker DBD due to saturated niche space,  
151 we used a GLMM including the interaction between diversity and environment type as a  
152 fixed effect. We considered this model only for taxonomic ratios with evidence for  
153 significant DBD slope variation by environment (**Table S1**): Family:Order, Order:Class

154 and Class:Phylum. Consistent with our hypothesis, DBD slopes were significantly more  
155 positive in less diverse (often host-associated) biomes (**Fig. 4a, Figure S11,**  
156 **Supplementary Data file 1 Section 4**).

157 The Black Queen hypothesis posits that microbes embedded in complex  
158 communities can exploit the production of extracellular public goods produced by other  
159 species, resulting in selection for loss of genes encoding these goods – as long as the  
160 essential trait is not lost from the community as a whole<sup>36</sup>. Lineages that interact more  
161 frequently with other lineages through such public good exploitation would be expected  
162 to experience greater loss of function and thus greater genome reduction. These reduced  
163 genome would also be expected to experience stronger DBD, because their survival and  
164 diversification is dependent on other community members. To test this expectation, we  
165 assigned genome sizes to 576 genera for which at least one whole-genome sequence was  
166 available and added an interaction term between genome size and diversity as a fixed  
167 effect to the GLMM (Methods). Contrary to expectation, we observed a slight but  
168 significant positive effect of genome size on the slope ( $z=2.5$ ,  $P=0.01$ ; **Fig. 4b,**  
169 **Supplementary Data file 1 Section 5**). The positive relationship may even be stronger  
170 than estimated, because genus-level genome size estimates are likely quite noisy. This  
171 result supports a model in which biotic interactions (and resulting diversification) drive  
172 genome expansion (*e.g.* through the accumulation of toxin- and resistance-gene diversity  
173 during antagonistic coevolution<sup>2</sup>). Alternatively (or additionally), species with larger  
174 biosynthetic gene repertoires and greater opportunity to engage in niche construction<sup>21</sup>  
175 could be more prone to interact with other species, driving DBD.



176           Using 10 million individual marker sequences, we demonstrated a pervasive  
177 positive relationship between prokaryotic diversity and diversification, which holds  
178 across a broad range of environments and taxa. The strength of the DBD relationship  
179 dissipates with increasing microbiome diversity which might be due to niche saturation,  
180 or potentially due to the fact that highly diverse communities prevent species from  
181 reliably interacting with each other. DBD appears to be particularly strong among deeply  
182 diverged lineages (*e.g.* phyla), suggesting the importance of DBD in the ancient  
183 diversification of bacterial lineages and supporting the view that high taxonomic ranks  
184 are ecologically coherent<sup>37,38</sup>. We note that the very early stages of diversification are  
185 inaccessible at the resolution of 16S ASVs, but this could be addressed in the future using  
186 (meta-)genomic approaches. At the limited resolution of 16S sequences, we do not expect  
187 measurable diversification within an individual microbiome sample; however community  
188 diversity could still select for (as in DBD) or against (as in EC) standing diversity in a  
189 focal lineages, even if this lineage diversified before the sampled community assembled.  
190 Due to the correlational nature of our data, it is not possible to test whether the positive  
191 relationship between diversification and diversity is primarily due to the creation of novel  
192 niches via biotic interactions and niche construction<sup>22</sup>, or potentially due to increased  
193 competition leading to specialisation on underexploited resources<sup>3,29</sup>. Despite their  
194 importance in shaping microbiome diversity and community structure, abiotic factors  
195 such as pH and temperature do not appear to be driving the DB relationship; this could be  
196 further tested in studies with more extensive abiotic metadata. Regardless of the  
197 underlying mechanisms, our results demonstrate the importance of biotic interactions in  
198 shaping microbiome diversity, which has important implications for modelling and

199 predicting their function and stability<sup>4,39</sup>. The answer to the question ‘why are  
200 microbiomes so diverse?’ might in a large part be because microbiomes are so diverse<sup>25</sup>.

201

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## 209 **Author contributions**

210 Conceptualization: BJS, MV. Data curation: NM. Formal analysis: NM, MV, BJS.  
211 Funding acquisition: BJS. Investigation: NM, MV, PL, BJS. Methodology: NM, MV, PL,  
212 BJS. Resources: BJS, PL. Supervision: PL, BJS. Software: NM. Visualization: NM.  
213 Writing original draft: NM, MV, BJS. Writing - review & editing: NM, MV, PL, BJS.

214

215 **Competing interests:** none to declare.

216

217 **Data and materials availability:** All data is available from the Earth Microbiome  
218 Project (<ftp.microbio.me>), as detailed in the Methods. All computer code used for  
219 analysis are available at <https://github.com/Naima16/dbd.git>.

220

221

222 **Supplementary Materials**

223

224 Supplementary text

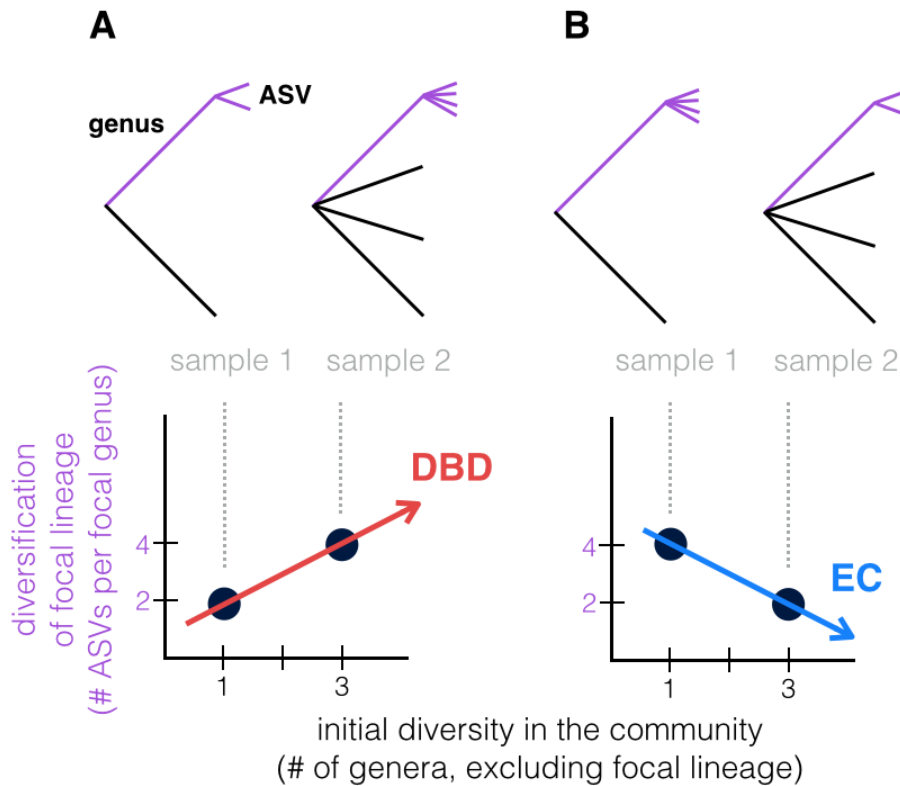
225 Methods

226 Tables S1 – S2

227 Fig S1 – S11

228 File 1. Full GLMM outputs.

229 File 2. Indicator species analysis.



230

231 **Fig. 1. Contrasting the Diversity Begets Diversity (DBD) and Ecological Controls**

232 **(EC) models of diversification.** We consider the diversification of a focal lineage as a

233 function of initial diversity present at the time of diversification.

234 **(A)** For example, sample 1 contains one non-focal genus, and two ASVs diversify within

235 the focal genus (point at  $x=1$ ,  $y=2$  in the plot). Sample 2 contains three non-focal genera,

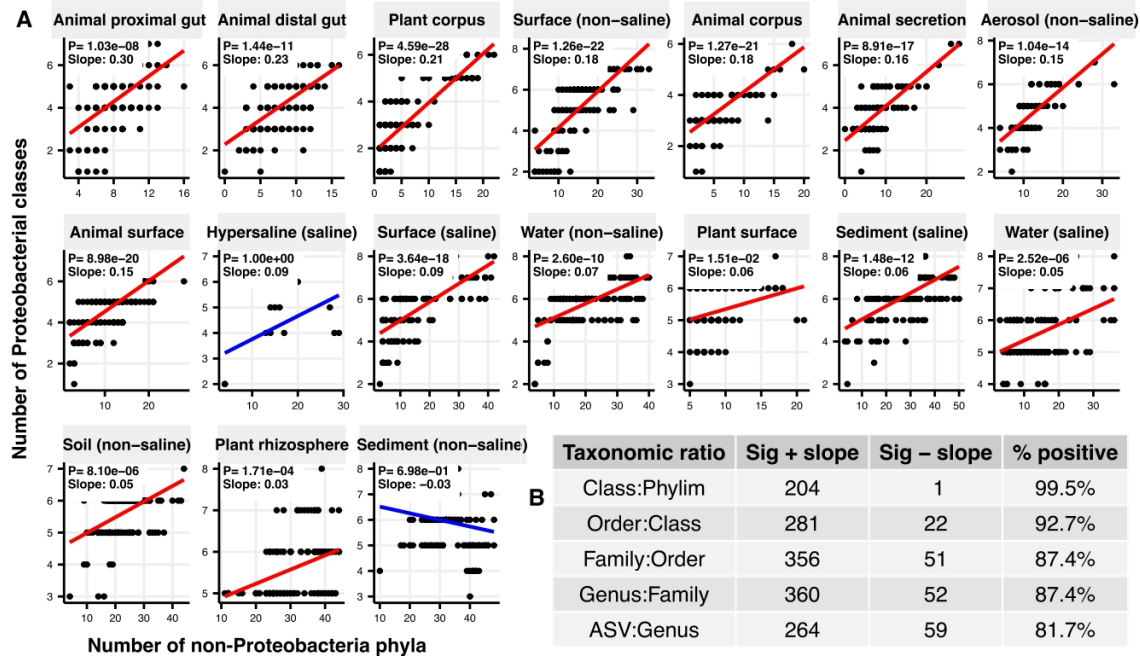
236 and four ASVs diversify within the focal genus (point at  $x=3$ ,  $y=4$ ). Tracing a line

237 through these points yields a positive slope, supporting the Diversity Begets

238 Diversification (DBD) model (red).

239 **(B)** Alternatively, a negative slope would support the Ecological Controls (EC) model

240 (blue line).



241

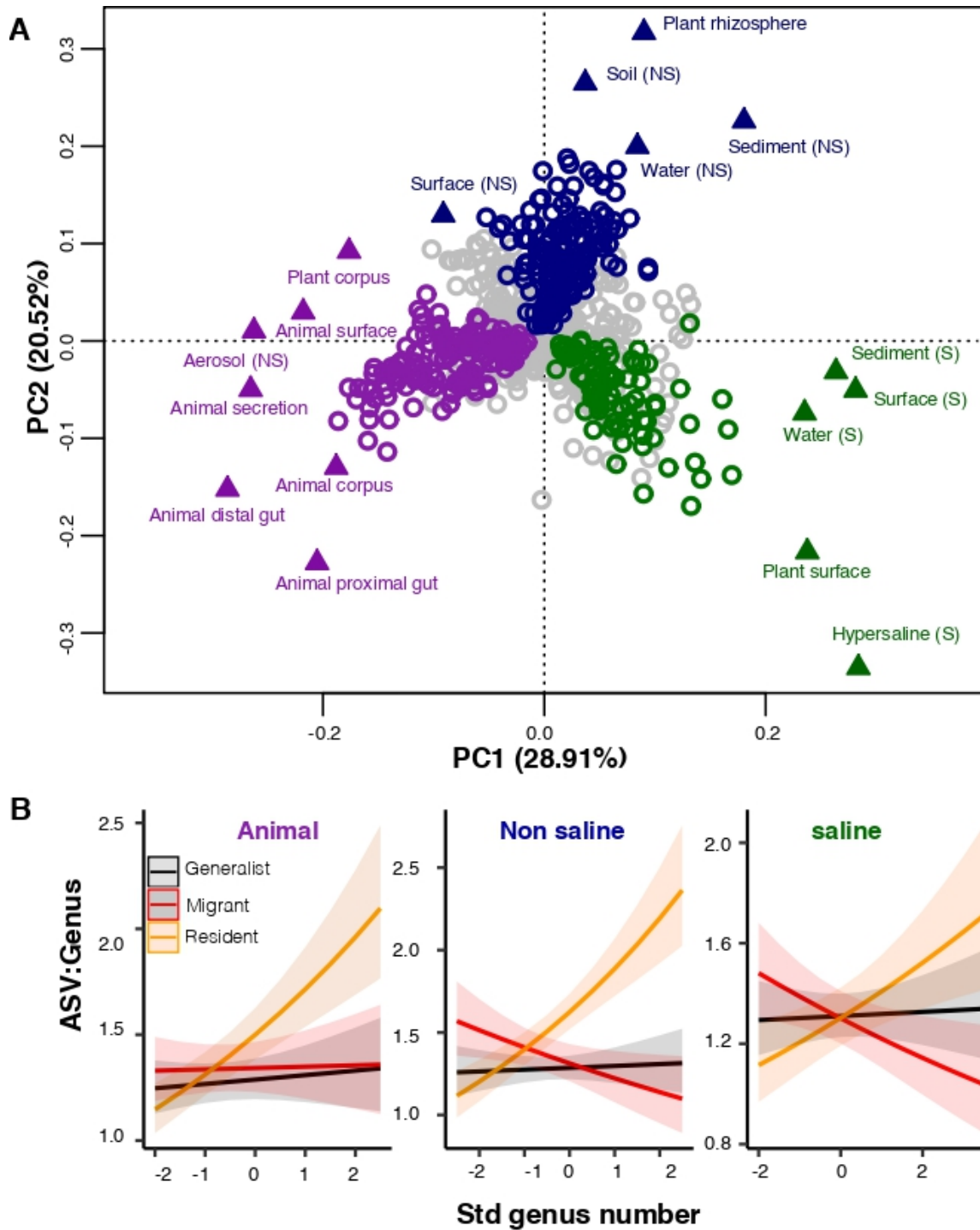
242

243 **Fig. 2. Diversification as a function of diversity across biomes in the phylum**

244 **Proteobacteria.**

245 **(A) Linear models for diversification** (the number of classes within Proteobacteria, y-  
 246 axis) as a function of diversity (the number of non-proteobacterial phyla, x-axis) in each  
 247 of the 17 environments (EMPO3 biomes). P-values are Bonferroni corrected for 17 tests.  
 248 Significant ( $P < 0.05$ ) models are shown with red trend lines; non-significant ( $P > 0.05$ )  
 249 trends are shown in blue.

250 **(B) Summary of linear model slopes across taxonomic ratios.** The number of  
 251 significant positive (+) or negative (-) slope estimates are shown for each taxonomic  
 252 ratio, summed across biomes. Significant slopes are those with  $P < 0.05$  (Bonferroni  
 253 corrected). Non-significant slope estimated are excluded.



254

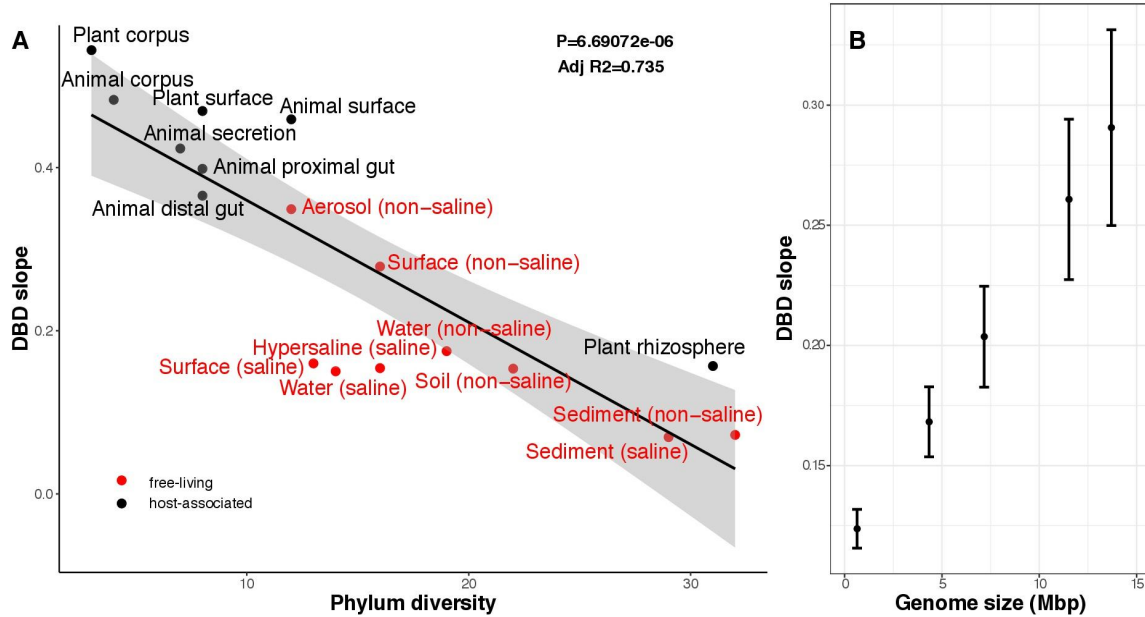
255 **Fig. 3. Diversity begets diversification in resident versus non resident genera.**

256 **(A) PCA showing genera clustering into their preferred environment clusters.**

257 Circles indicate genera and triangles indicate environments (EMPO 3 biomes). The three

258 environment clusters identified by fuzzy *k*-means clustering are: Non-saline (NS, blue),  
259 saline (S, green) and animal-associated (purple). Resident genera were identified by  
260 indicator species analysis.

261 **(B) DBD in resident versus non resident genera across environment clusters.** Results  
262 of GLMMs modeling diversification as a function of diversity in resident, migrant, or  
263 generalist groups. The x-axis shows the standardized number of non-focal resident genera  
264 (diversity); the y-axis shows the number of ASVs per focal genus (diversification).  
265 Resident focal genera are shown in orange, migrant focal genera in red, and generalist  
266 focal genera in black.



267

268

269 **Fig. 4. Ecological and evolutionary mechanisms to explain variation in the strength**  
270 **of DBD.**

271 **(A) DBD slope is higher in low-diversity (often host-associated) microbiomes.** The x-  
272 axis shows the mean number of phyla in each biome. On the y-axis, DBD slope was  
273 estimated by the GLMM predicting diversification as a function of the interaction  
274 between diversity and environment type at the Class:Phylum ratio (**Supplementary Data**  
275 **file 1 Section 4.3**). The line represents a regression line; the shaded area depicts 95%  
276 confidence limits of the fitted values.

277 **(B) Positive correlation between genome size and DBD slope.** Results are shown from  
278 a GLMM predicting diversification as a function of the interaction between diversity and  
279 genome size at the ASV:Genus ratio (**Supplementary Data file 1 Section 5**). The x-axis  
280 is genus-level genome size in Mbp (min=0.97, max=14.78); the y-axis is DBD slope (the  
281 effect of diversity on diversification). Vertical bars indicate 95% confidence limits of the  
282 fitted values.



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371 **Supplementary Materials**

372

373 **Diversity begets diversity in microbiomes**

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379

380 **Supplementary Text**

381 **Methods**

382 **Tables S1 – S2**

383 **Figures S1 – S11**

## 384 **Supplementary Text**

385           To test for any potential confounding effects of data structure or sampling bias,  
386 we sought to remove any patterns of co-occurrence between ASVs in the same sample  
387 via permutation. We took 2,000 simulated samples by selecting from the overall  
388 distribution of 155,002 unique ASVs across all samples, weighted by their abundance  
389 (total number of sequence counts). This resulted in a slightly negative diversity-  
390 diversification relationship (slope =  $-0.002$ ; Pearson correlation =  $-0.61$ ;  $P < 2.2 \times 10^{-16}$ ;  
391 **Fig. S7**), indicating that the observed positive relationships (**Table S1; Fig. 2**) are not the  
392 effect of data structure.

393           We sought to further validate the results with a taxonomy-independent approach,  
394 because not all taxonomic ranks have the same phylogenetic depth<sup>40</sup> and not all named  
395 taxa are monophyletic<sup>41</sup>. Therefore, we clustered ASVs at decreasing levels of nucleotide  
396 identity, from 100% identical ASVs down to 75% identity (roughly equivalent to phyla  
397<sup>42</sup>). We estimated diversification as the mean number of descendants per cluster (e.g.  
398 number of 100% clusters per 97% cluster) and plotted this against the total number of  
399 non focal clusters (97% identity in this example). For each of the six nucleotide  
400 divergence ratios tested, the relationship between diversity and diversification was  
401 positive (**Fig. S8**), consistent with DBD and suggesting that the taxonomic analyses were  
402 largely unbiased.

403           To exclude the possibility that our results were driven by abiotic confounders, we  
404 repeated the taxonomic analysis on a subset of 192 EMP samples for which  
405 measurements of four important abiotic drivers of diversity, temperature, pH, latitude,  
406 and elevation<sup>5,14,15,43</sup> were available. We fitted a GLMM with diversification rate as the

407 dependent variable, and with the number of non-focal lineages, the four abiotic factors  
408 and their interactions as predictors (fixed effects). As in the full dataset (**Table S1**),  
409 diversification was positively associated with diversity at all taxonomic ratios (**Table S2**).  
410 As expected, certain abiotic factors, alone or in combination with diversity, had  
411 significant effects on diversification. However, the effects of abiotic factors were always  
412 weaker than the effect of community diversity (**Table S2; Supplementary Data file 1**  
413 **Section 2**). Although only a small subset of abiotic factors was considered, this analysis  
414 suggests that the DBD trend is unlikely to be mainly driven by variation in the abiotic  
415 environment.

## 416 **Methods**

417

### 418 **16S rRNA marker data acquisition and preprocessing.**

419 16S rRNA-V4 region reads (90 bp, GreenGenes 13.8 taxonomy) along with

420 environmental data and EMPO3 designations

421 (<http://press.igsb.anl.gov/earthmicrobiome/protocols-and-standards/emp/>) were

422 downloaded from the EMP FTP server (<ftp://ftp.microbio.me>), on February 9, 2018. Sequence

423 summaries were downloaded from :

424 [ftp://ftp.microbio.me/emp/release1/otu\\_distributions/otu\\_summary.emp\\_deblur\\_90bp.sub](ftp://ftp.microbio.me/emp/release1/otu_distributions/otu_summary.emp_deblur_90bp.sub)

425 [set\\_2k.rare\\_5000.tsv](#), environmental data from :

426 [ftp://ftp.microbio.me/emp/release1/mapping\\_files/emp\\_qiime\\_mapping\\_release1.tsv](ftp://ftp.microbio.me/emp/release1/mapping_files/emp_qiime_mapping_release1.tsv), and

427 EMPO3 designations from :

428 [ftp://ftp.microbio.me/emp/release1/mapping\\_files/emp\\_qiime\\_mapping\\_subset\\_2k.tsv](ftp://ftp.microbio.me/emp/release1/mapping_files/emp_qiime_mapping_subset_2k.tsv).

429 The list of the associated 97 studies and 61 corresponding principal investigator identities

430 were downloaded from <https://www.nature.com/articles/nature24621#s1>.

431 We used the EMP ‘2000 subset’ rarefied to 5000 sequences per sample. This subset

432 contains 155 002 ASVs from 2000 samples with an even distribution across 17 natural

433 environments (EMP Ontology level 3) (Thompson et al., 2017). Based on the ASVs

434 annotations across samples, we estimated diversification for every taxonomic ratio

435 (ASV:Genus, Genus:Family, Family:Order, Order:Class and Class:Phylum), along with

436 the number of non-focal lineages (Python script, Python Version 2.7).

437

438



## 439 **Generalized Linear Mixed Models (GLMMs)**

440 All models were fitted in Rstudio (Version 1.1.442, R Version 3.5.2) using the glmer  
441 function of the lme4 package<sup>44</sup>. Data standardization (transformation to a mean of zero  
442 and a standard deviation of one) was applied to all predictors to get comparable  
443 estimates. In models with only one predictor, applying standardization resolved  
444 convergence warnings and considerably sped up the optimization. Standardization has  
445 previously been reported to improve model performance and solve convergence  
446 problems<sup>45</sup>.

447 We used likelihood-ratio tests (anova R function from stats package) as follows:  
448 1) on nested models to assess the significance of random effects (in the nested models,  
449 each effect was dropped one at a time); 2) on the full model and the null model  
450 comprising only random effects, to assess the significance of fixed effects<sup>46</sup>; 3) on the full  
451 model and the model without the interaction term, to assess the significance of  
452 interactions. All models reported here were found to be significant ( $P < 0.05$ ).

453 Diagnostic plots (plot and qqnorm R functions in base and stats packages) were  
454 checked for each model to ensure that residual homoscedasticity (homogeneity of  
455 variance) was fulfilled: no increase of the variance with fitted values and residuals were  
456 symmetrically distributed tending to cluster around the 0 of the ordinate, but with an  
457 expected pattern due to count data. Normality plots were imperfect, but they generally  
458 showed that the residuals were close to being normally distributed. The assumption of  
459 normality is often difficult to fulfill with high numbers of observations, as is the case in  
460 our models (<https://www.statisticshowto.datasciencecentral.com/shapiro-wilk-test/>), and

461 non-normality is less of concern than heteroscedastic for the validity of GLMMs  
462 ([https://bbolker.github.io/mixedmodels-misc/ecostats\\_chap.html#diagnostics](https://bbolker.github.io/mixedmodels-misc/ecostats_chap.html#diagnostics)).

463 We tested for overdispersion using the `overdisp_fun` R function available at  
464 <https://bbolker.github.io/mixedmodels-misc/glmmFAQ.html>, and found that the models  
465 were not overdispersed, but rather were underdispersed. The ratio of the sum of squared  
466 Pearson residuals to residual degrees of freedom was  $< 1$  and non-significant when tested  
467 with a chi-squared test. Given that underdispersion leads to more conservative results, we  
468 retained the GLMMs with Poisson error distribution, despite the underdispersion.  
469 (GLMM FAQ; Ben Bolker and others; 25 September 2018;  
470 <https://bbolker.github.io/mixedmodels-misc/glmmFAQ.html#underdispersion>).

471

## 472 **Taxonomy-based generalized linear mixed models**

473 The effect of diversity on diversification was tested for different environment types and  
474 lineages using generalized linear mixed models (GLMMs) fitted on the EMP dataset, for  
475 all taxonomic ratios. As the dependent variable (diversification, defined as taxonomic  
476 ratios, ASV:Genus, Genus:Family, Family:Order, Order:Class, and Class:Phylum) was a  
477 count response, we used a Poisson error distribution with a log link function. Diversity  
478 (number of non-focal lineages: non-focal Genera, Families, Orders, Classes, and Phyla),  
479 standardized to a mean of zero and a standard deviation of one, was specified as the  
480 predictor (fixed effect). We included the following random effects on the slope and  
481 intercept: lineage (Lin), environment (Env), environment nested within lineage (a lineage  
482 may be present in different environments) and lab (the principal investigator who  
483 conducted the EMP study) nested within environment (different labs sampled and

484 sequenced a given environment) (as suggested in [http://bbolker.github.io/mixedmodels-](http://bbolker.github.io/mixedmodels-misc/glmmFAQ.html)  
485 [misc/glmmFAQ.html](http://bbolker.github.io/mixedmodels-misc/glmmFAQ.html)). Defining random effects on the slope enabled us to test slope  
486 variation across groups of each categorical variable. We included the EMP unique sample  
487 ID as a random effect to control for dependencies between observations (if two taxa were  
488 part of the same sample).

489 To test for the relative effect of biotic and abiotic environmental variables on  
490 diversification across different taxonomic ratios, we used a separate GLMM, with  
491 Poisson error distribution with a log link function, for every ratio. We fitted the GLMM  
492 on a subset (~10%) of the whole dataset, 192 samples (from water: saline (19) and non-  
493 saline (44), surface: saline (42) and non-saline (19), sediment: saline (22) and non-saline  
494 (31), soil (8) and plant rhizosphere (7)), for which measurements of four key abiotic  
495 variables (temperature, pH, latitude and elevation) were available. We defined diversity  
496 and the abiotic variables as well as the interactions between diversity and every abiotic  
497 variable as predictors (fixed effects) of diversification. All predictors were standardized  
498 to a mean of zero and a standard deviation of one to obtain comparable estimates. The  
499 GLMM had the same random effects as in the previous analysis, but only on the intercept  
500 for simplicity.

501

## 502 **Nucleotide sequence identity-based analysis**

503 We defined a threshold of percent nucleotide identity between ASVs, corresponding to  
504 different taxonomic ranks (from 100% identical ASVs down to 75% identity)<sup>42</sup>. Fasta  
505 files for all samples were produced by a python script (Python Version 2.7) from the  
506 sequences summary file (otu\_summary.emp\_deblur\_90bp.subset\_2k.rare\_5000 from

507 EMP ftp server). We clustered sequences from each sample using USEARCH V9.2. We  
508 estimated diversity as the total number of clusters at a given level (*e.g.* 97% identity) and  
509 diversification as the mean number of descendent clusters (*e.g.* number of 100% clusters  
510 per 97% cluster). To describe the relationship between diversity and diversification, we  
511 tested three models: linear, quadratic and cubic (lm function in R). Model comparisons  
512 were based on the adjusted  $R^2$ .

513         We note that diversity at level  $i$  ( $d_i$ ) and diversification at level  $i+1$  ( $d_{i+1}/d_i$ ) are not  
514 independent in this analysis because  $d_{i+1}$  must be greater than or equal to  $d_i$ . To assess the  
515 effects of this non-independence on the results, we conducted permutation tests by  
516 randomizing the associations between  $d_i$  and  $d_{i+1}$ . Using 999 permutations,  $P$ -values were  
517 calculated based on how many times we observed a correlation greater than that seen in  
518 the real data (cor.test R function with kendall method). In each permutation, we  
519 recalculated the significance test (Wald  $z$ ) for the correlation in the randomized data, and  
520 then computed the  $P$ -value based on how many times we observed a  $z$  value greater than  
521 that of the original data (one tailed test because we wanted to demonstrate that the  
522 relationship was positive). At all six levels of nucleotide identity, the real data always  
523 showed a significantly stronger positive correlation when compared to permuted data ( $P$   
524 = 0.001), indicating that the DBD patterns was not an artefact of the dependence structure  
525 in the data.

526         The effect of diversity on diversification was also tested across different  
527 environments analysed separately. We modelled this relationship with linear, quadratic  
528 and cubic fits, and compared those models based on the adjusted  $R^2$ .  
529

530 **DBD among residents of the same environment**

531 We clustered the environmental samples based on their genus-level community  
532 composition using fuzzy *k*-means clustering. Fuzzy clustering is a version of non-  
533 hierarchical clustering, where each cluster is a fuzzy set of all biomes and greater  
534 membership values indicates higher confidence in the allocation pattern to the cluster.  
535 The clustering (cmeans function, package e1071 in R) was done on the ‘hellinger’  
536 transformed data (decostand function, package vegan in R). To identify resident genera to  
537 each cluster, we used indicator species analysis<sup>47</sup> as implemented in the indval function  
538 (labdsv R package). Indicators are genera found mostly in a certain environment group  
539 and present in the majority of environments of that group. The indicator value (indval  
540 index) of a genus is (maximum=1) if the genus is observed in only one environmental  
541 cluster and in all samples belonging to that cluster. We defined residents as genera with  
542 indval indices between 0.4 and 0.9, with permutation test  $P < 0.05$ . Genera not been  
543 associated with any cluster were considered generalists. We used principal component  
544 analysis (PCA) to visualize clustering and indicator genera (rda function, vegan R  
545 package). We then ran a separate GLMM for each environmental cluster, with resident  
546 genus-level diversity (number of non-focal genera) as a predictor of diversification  
547 (ASV:Genus ratio) for resident, migrant (residents of one cluster found in a different  
548 cluster) and generalist genera. The fixed effect was specified as the interaction between  
549 diversity and a factor defining the genus-cluster association (with three levels: resident,  
550 migrant and generalist). Random effects on intercept and slope were kept as in the  
551 previous GLMMs.  
552

553 **DBD variation across biomes**

554 We tested the variation of DBD slope across different environments by defining  
555 environment (EMPO 3 biome type) as fixed effect. We fitted a GLMM with the  
556 interaction between diversity and environment type as a predictor of diversification. The  
557 main effects of diversity and environment individually were not included for model  
558 simplicity and we sought to look at the effect of the interaction alone  
559 (diversity\*environment). All other random effects on intercept and slope were kept as in  
560 the previous GLMMs. DBD variation across environments was tested for Family:Order,  
561 Order:Class and Class:Phylum taxonomic ratios, as DBD slope variation by environment  
562 was statistically significant (likelihood-ratio test) for these ratios (**Table S1**).

563

564 **Genome size analysis**

565 We chose a subset of genera represented by one or more sequenced genomes in the NCBI  
566 microbial genomes database  
567 (<https://www.ncbi.nlm.nih.gov/genome/browse#!/prokaryotes/>). For these genera, a  
568 representative genome size was assigned by selecting the genome with the lowest number  
569 of scaffolds (if no closed genomes were available). If multiple genomes were available,  
570 sequenced to the same level of completion, the largest genome size was used. We fitted a  
571 GLMM on the subset of data with known genome size (576 genera) with the interaction  
572 between diversity and genome size as a predictor of diversification (ASV:Genus). All the  
573 other random effects on intercept and slope were kept as in the previous GLMMs.

574

575

576 **Code availability**

577 All computer code used for analysis are archived on the github repository

578 <https://github.com/Naima16/dbd.git>.

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600



601 **Supplementary Tables**

602

603 **Table S1. Diversity has a positive effect on diversification across taxonomic ratios.**

604 The GLMMs showed statistically significant positive effect of diversity on

605 diversification. Each row reports the effect of diversity on diversification, as well as its

606 standard deviation, Wald z-statistic for its effect size and the corresponding *P*-value (left

607 section), or standard deviation on the slope for the significant random effects (right

608 section). SE=standard error, Env=environment type, Lin=lineage type, Lab=Principal

609 Investigator ID, Sample=EMP Sample ID. Interactions are denoted as ‘\*’. n.s.=not

610 significant (likelihood-ratio test).

	Slope (fixed effects)				Standard deviation on the slope (random effects)				
	Diversity	SE	z	<i>P</i>	Env	Lin	Lin*Env	Env*Lab	Sample
<b>ASV: Genus</b>	0.091	0.016	5.792	6.95e-09	n.s.	0.074	0.142	0.114	0.067
<b>Genus: Family</b>	0.047	0.008	5.911	3.41e-09	n.s.	0.071	0.07	0.039	n.s.
<b>Family: Order</b>	0.119	0.017	7.001	2.54e-12	0.023	0.094	0.092	0.106	n.s.
<b>Order: Class</b>	0.109	0.020	5.447	5.13e-08	0.05	0.141	0.078	0.051	n.s.
<b>Class: Phylum</b>	0.272	0.043	6.341	2.29e-10	0.119	0.174	0.119	0.114	n.s.

611

612

613 **Table S2. Diversity has a stronger effect than abiotic factors on diversification.**

614 Results are shown from GLMMs with diversity, four abiotic factors (temperature,  
 615 elevation, pH, and latitude), and their interactions with diversity, as predictors of  
 616 diversification. Random effects on the intercept included environment, lineage, lab ID  
 617 and sample ID. Results are summarized as the coefficient (slope)±standard error (for  
 618 fixed effects). Temp=temperature, Lat=latitude, Elev=elevation. Interactions denoted as  
 619 ‘\*’. Significant terms (Wald test) are shown in bold: \*\*\* $P < 2.2e-16$ ; \*\* $P < 0.01$ , \* $P < 0.05$ .  
 620 Random effects are not shown.

621  
 622

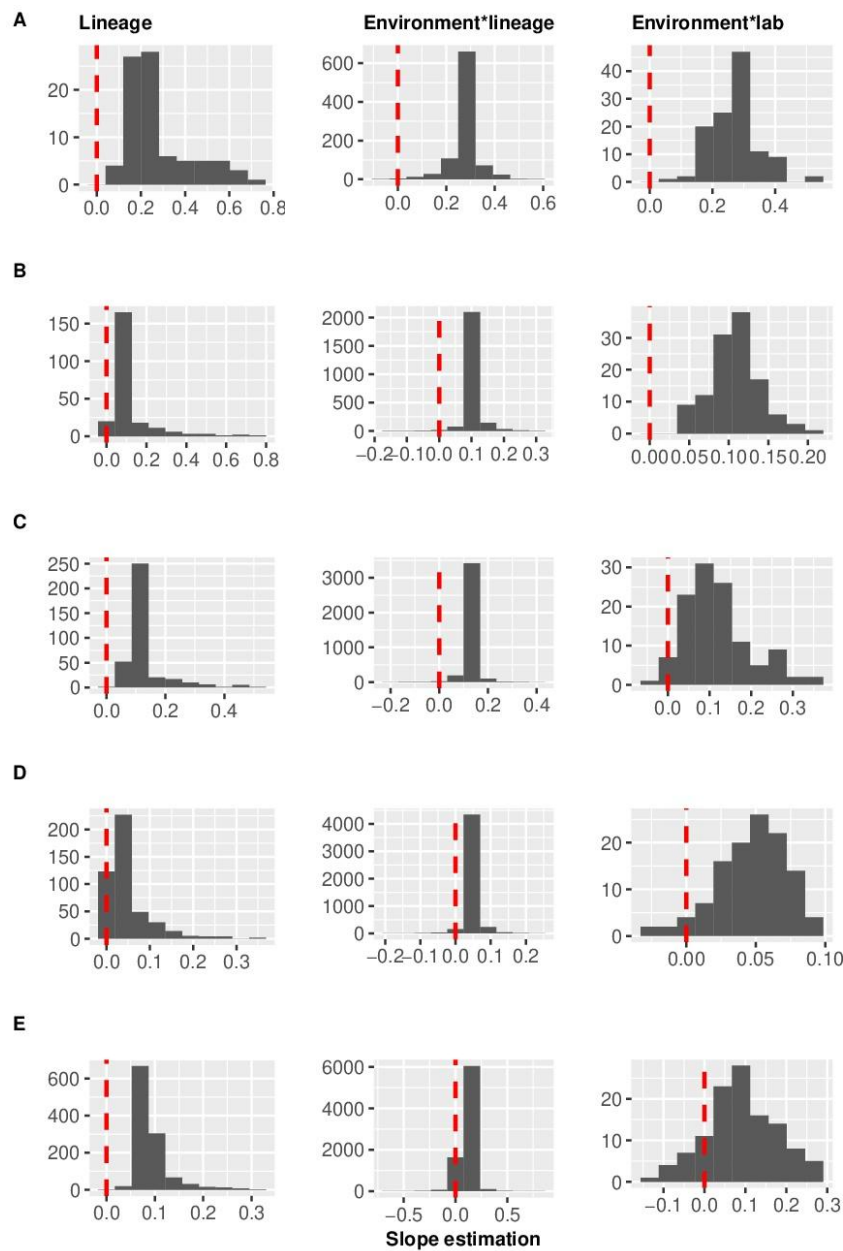
	Diversity	Temp	Lat	pH	Elev	Div *Temp	Div *Lat	Div *pH	Div *Elev
<b>ASV: Genus</b>	<b>0.129***</b> ±0.013	<b>0.044**</b> ±0.016	0.017 ±0.019	0 ±0.018	0 ±0.023	<b>0.043**</b> ±0.014	<b>0.032*</b> ±0.014	0.003 ±0.011	<b>-0.032*</b> ±0.016
<b>Genus: Family</b>	<b>0.094***</b> ±0.009	<b>0.04***</b> ±0.011	-0.009 ±0.01	- <b>0.049**</b> * ±0.009	- 0.003±0. 01	0.019 ±0.01	-0.011 ±0.009	-0.011 ±0.007	-0.005 ±0.009
<b>Family: Order</b>	<b>0.12***</b> ±0.013	0.012 ±0.014	0.002 ±0.021	0 ±0.013	- 0.011±0. 026	0.024 ±0.013	0.01 ±0.013	0.003 ±0.009	-0.015 ±0.014
<b>Order: Class</b>	<b>0.184***</b> ±0.01	0.001 ±0.013	-0.011 ±0.012	-0.002 ±0.012	- 0.008±0. 013	<b>0.036**</b> ±0.012	<b>0.023*</b> ±0.01	-0.003 ±0.01	<b>-0.02</b> <b>±0.01*</b>
<b>Class: Phylum</b>	<b>0.233***</b> ±0.013	-0.025 ±0.015	0.014 ±0.015	0.011 ±0.015	0.032 ±0.019	<b>0.06***</b> ±0.015	<b>0.039**</b> ±0.013	<b>0.029*</b> ±0.013	0.004 ±0.016

623

## 624 Supplementary Figures

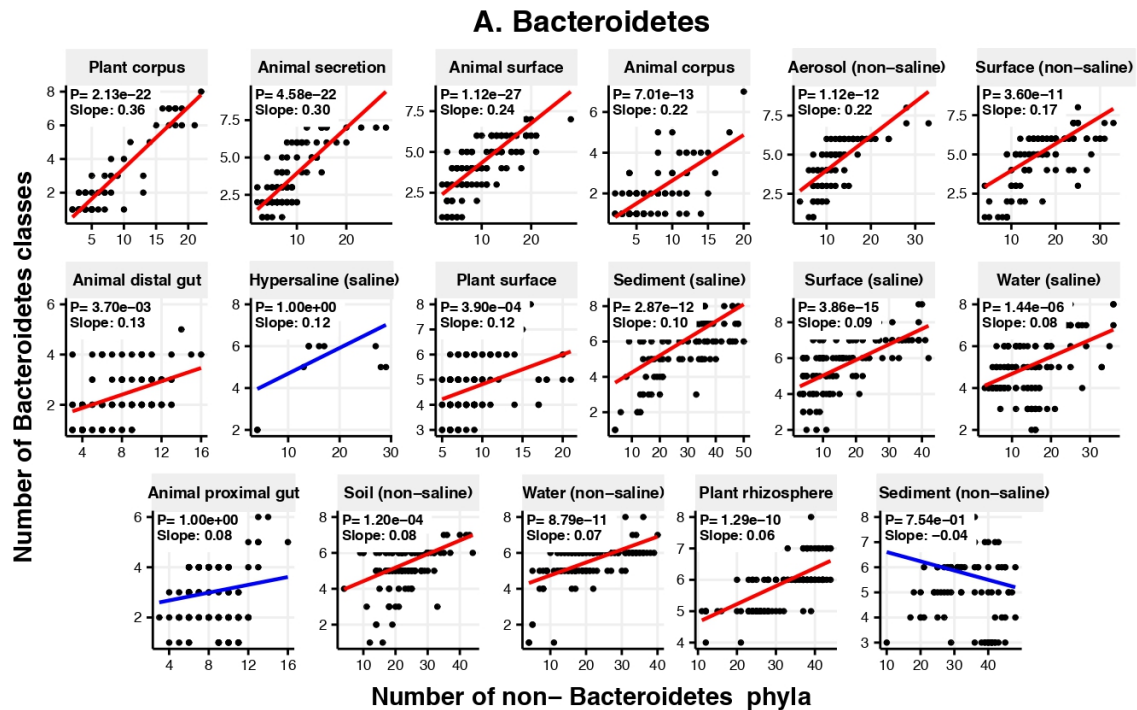
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626 **Figure S1. Distributions of DBD slope estimates across different random effects,**  
627 **from the GLMMs predicting diversification as a function of diversity. (A)**  
628 **Class:Phylum, (B) Order:Class, (C) Family:Order, (D) Genus:Family and (E)**  
629 **ASV:Genus ratios.** Estimation of random effect coefficients from the GLMMs (Table  
630 S1), shows that the effect of diversity on diversification (slope estimates) are generally  
631 positive but could be negative in some lineages or combinations of environment, lineage  
632 (Environment\*Lineage), and the laboratory that submitted the dataset  
633 (Environment\*Lab).  
634



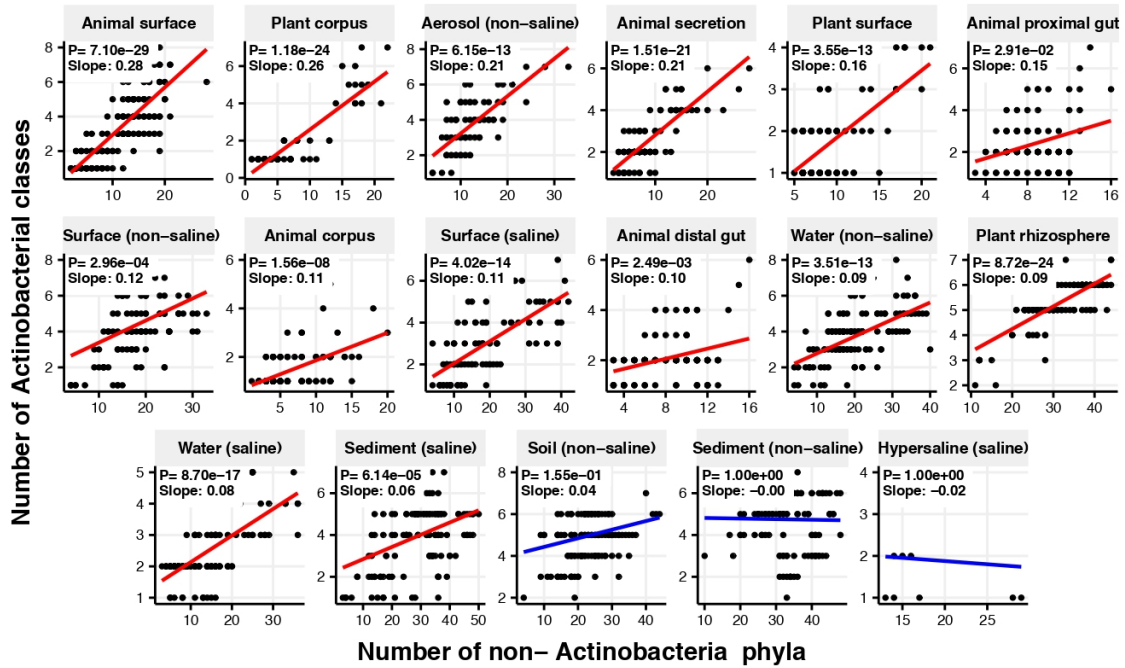
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637 **Figure S2. Diversification as a function of diversity across biomes in the two most**  
 638 **prevalent phyla after Proteobacteria (shown in Figure 2A of the main text). (A)**  
 639 **Bacteroidetes, (B) Actinobacteria.** Linear models are shown for diversification (classes  
 640 number per phylum, y-axis) as a function of diversity (non focal phyla number, x-axis) in  
 641 each of the 17 environments (EMPO3 biomes). P-values are Bonferroni corrected for 17  
 642 tests. Significant ( $P < 0.05$ ) models are shown with red trend lines, non-significant ( $P >$   
 643  $0.05$ ) trends are shown in blue.



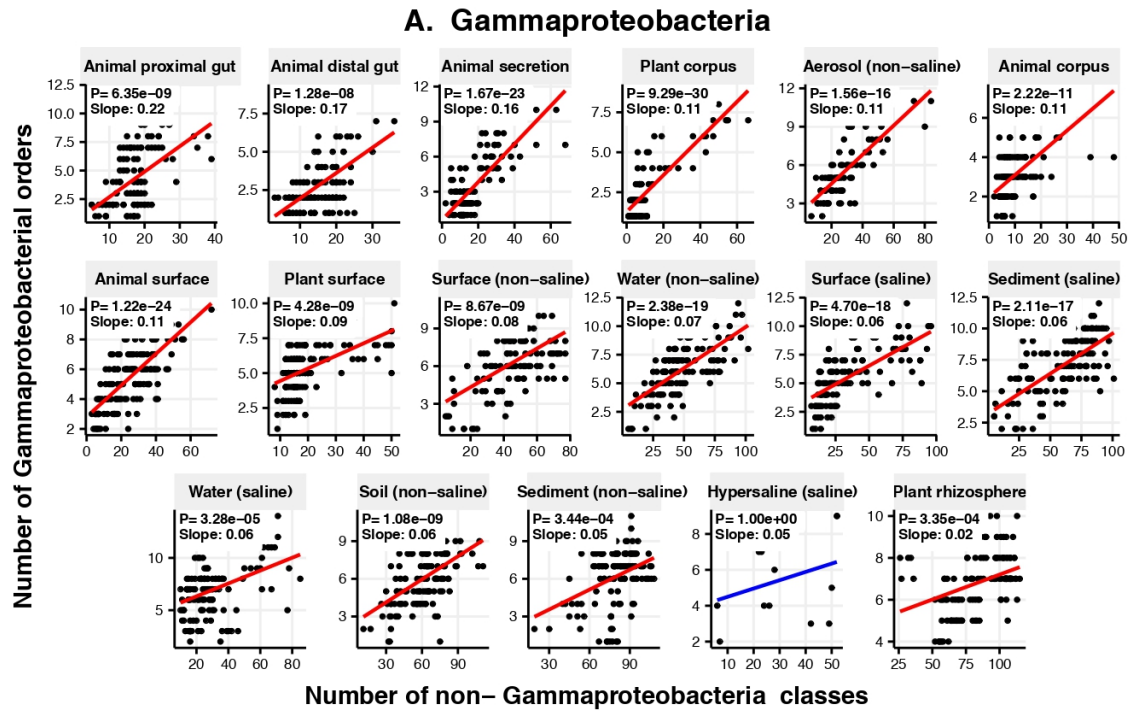
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## B. Actinobacteria



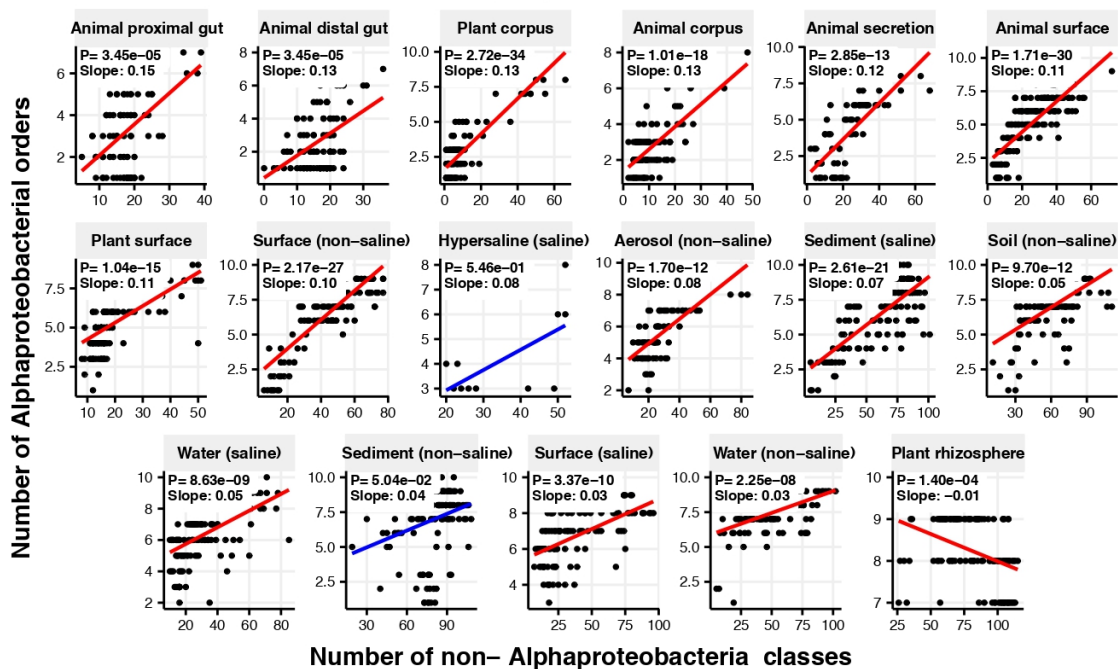
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652 **Figure S3. Diversification as a function of diversity across biomes in the three most**  
653 **prevalent classes.** Linear models are shown for diversification (orders per class, y-axis)  
654 as a function of diversity (non-focal classes, x-axis) in each of the 17 environments  
655 (EMPO3 biomes). P-values are Bonferroni corrected for 17 tests. Significant ( $P < 0.05$ )  
656 models are shown with red trend lines, non-significant ( $P > 0.05$ ) trends are shown in  
657 blue.



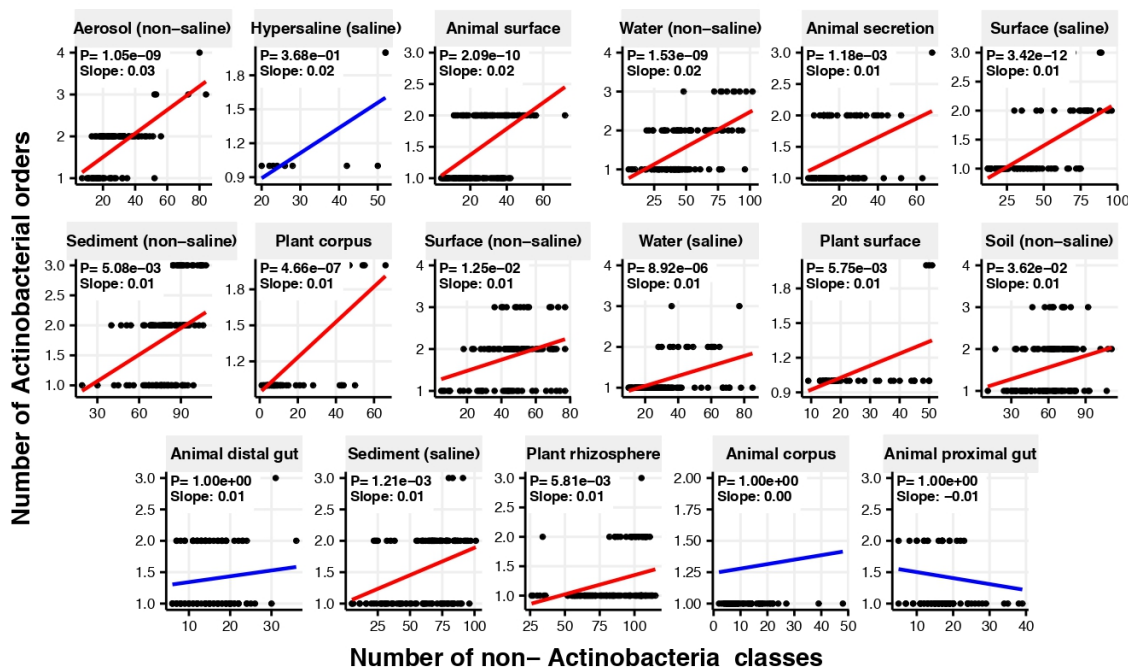
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## B. Alphaproteobacteria



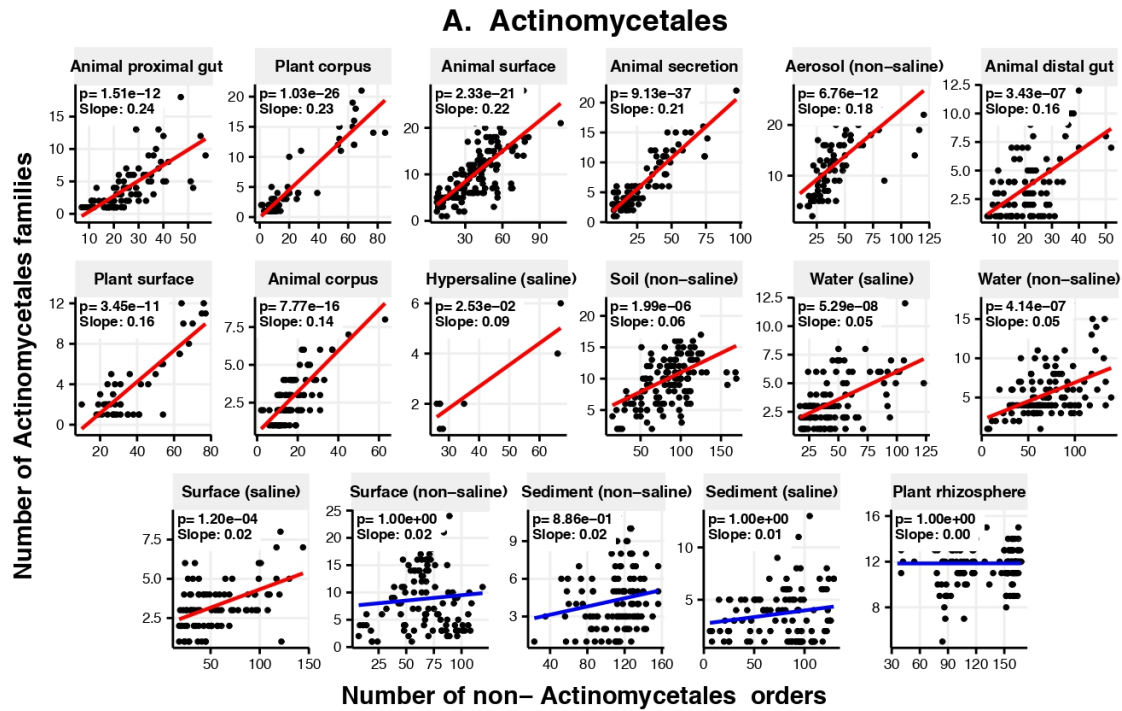
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## C. Actinobacteria



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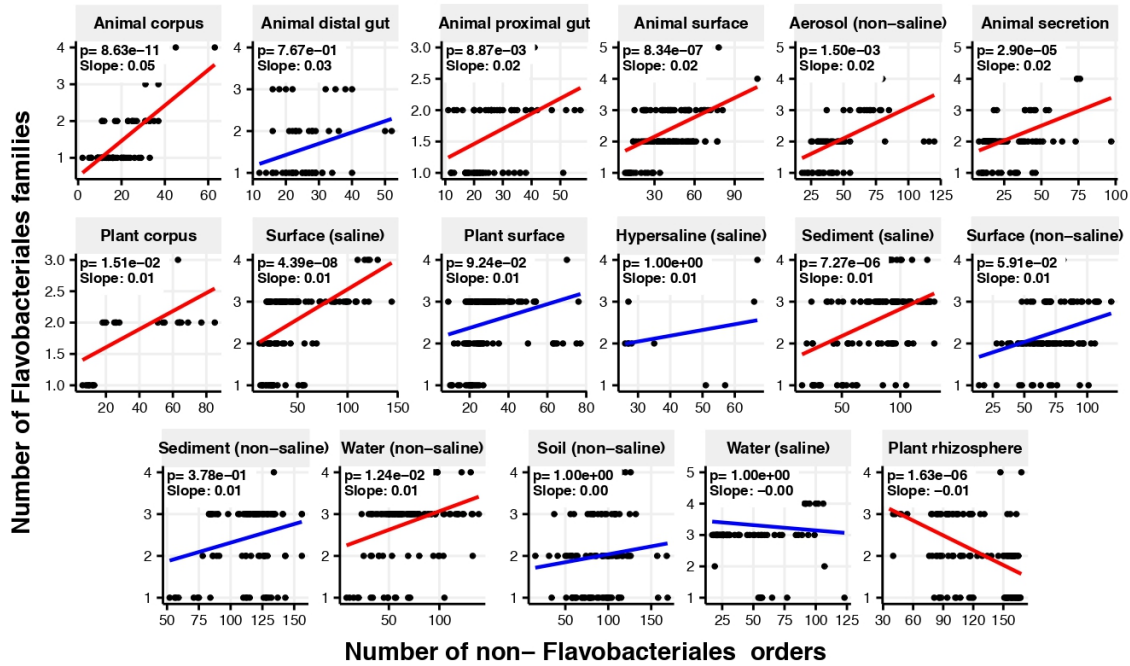
666 **Figure S4. Diversification as a function of diversity across biomes in the three most**  
 667 **prevalent orders.** Linear models are shown for diversification (families per order, y-  
 668 axis) as a function of diversity (non-focal orders, x-axis) in each of the 17 environments  
 669 (EMPO3 biomes). P-values are Bonferroni corrected for 17 tests. Significant ( $P < 0.05$ )  
 670 models are shown with red trend lines, non-significant ( $P > 0.05$ ) trends are shown in  
 671 blue.



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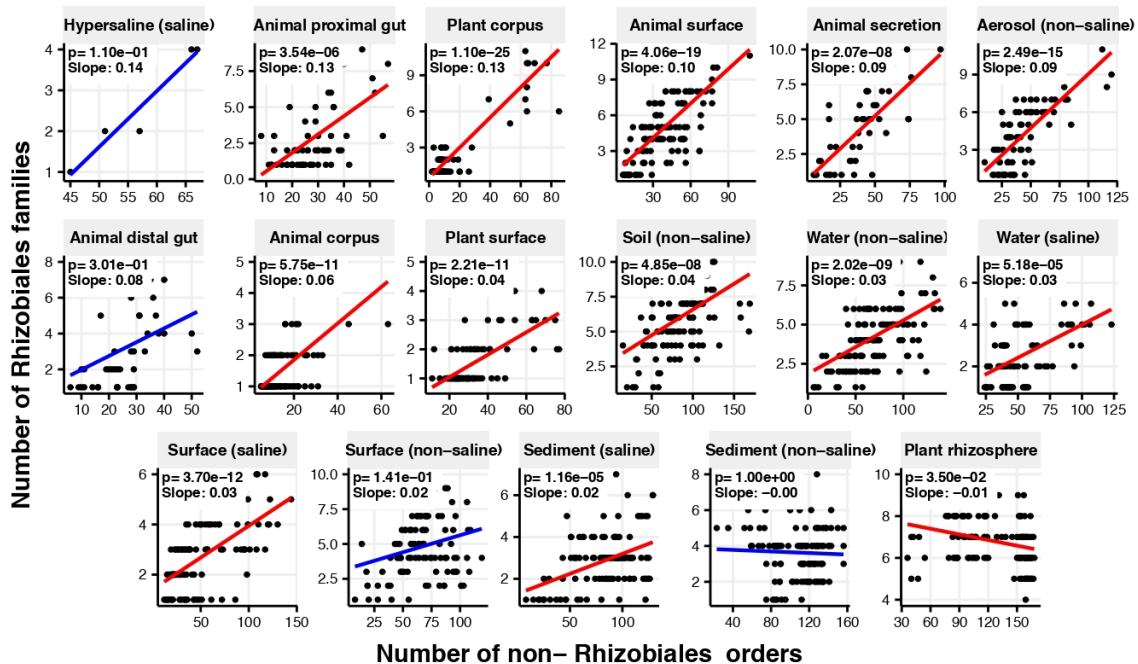


## B. Flavobacteriales



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## C. Rhizobiales



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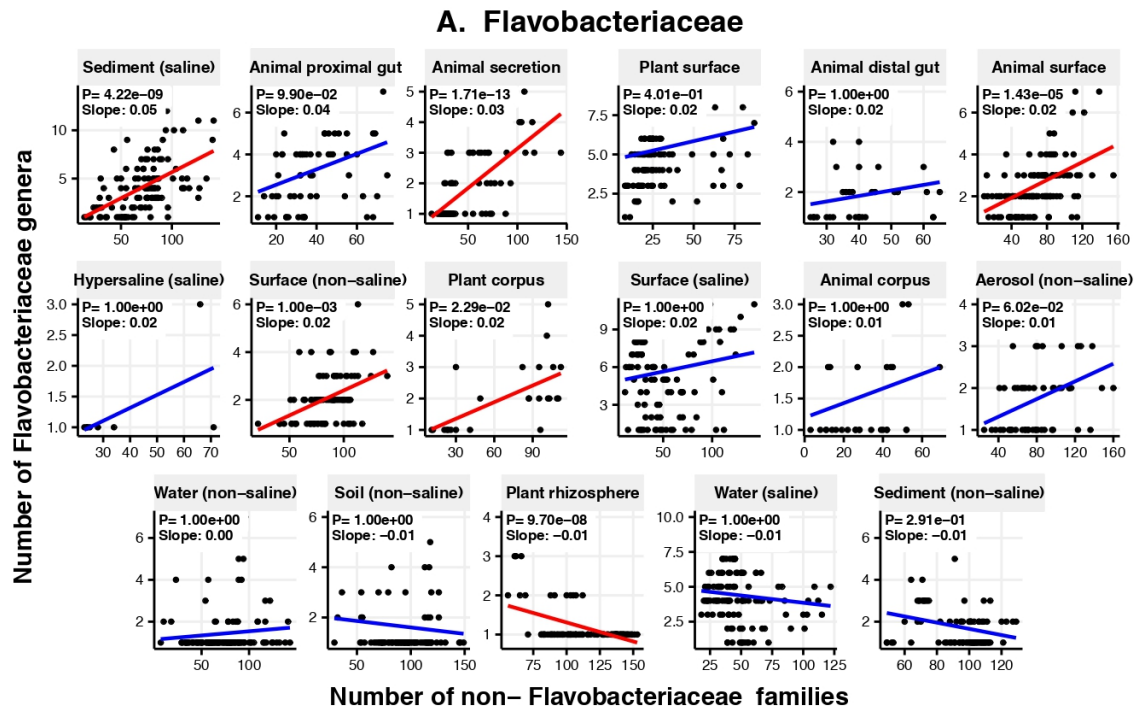
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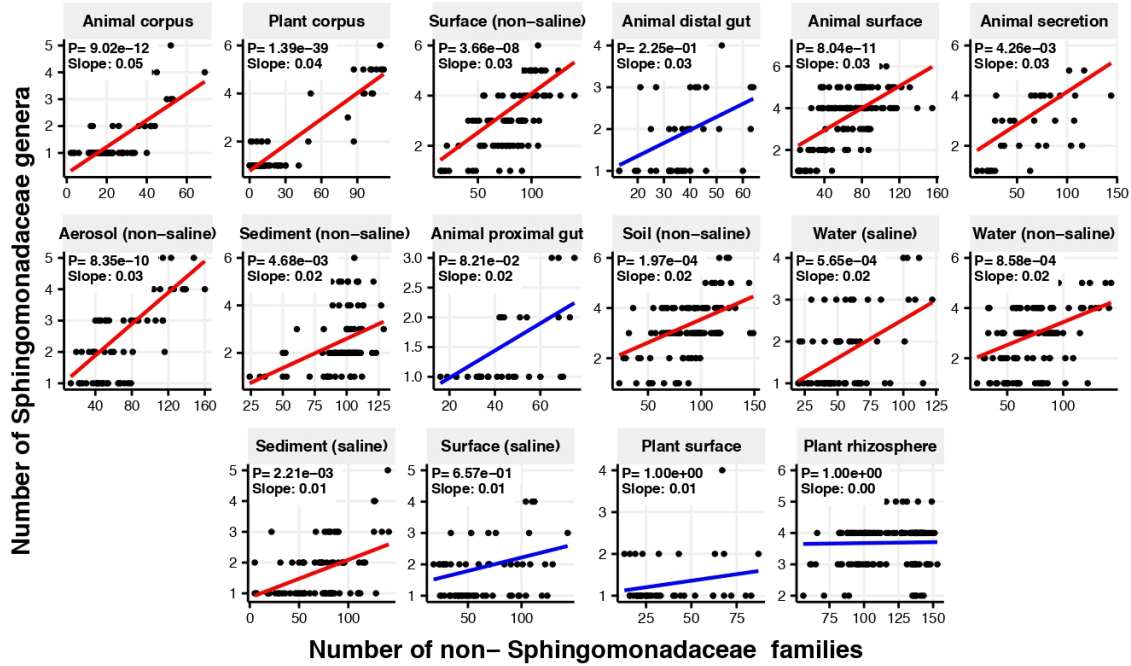
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681 **Figure S5. Diversification as a function of diversity across biomes in the three most**  
682 **prevalent families.** Linear models are shown for diversification (genera per family, y-  
683 axis) as a function of diversity (non-focal families, x-axis) in each of the 17 environments  
684 (EMPO3 biomes). P-values are Bonferroni corrected. Significant ( $P < 0.05$ ) models are  
685 shown with red trend lines, non-significant ( $P > 0.05$ ) trends are shown in blue.



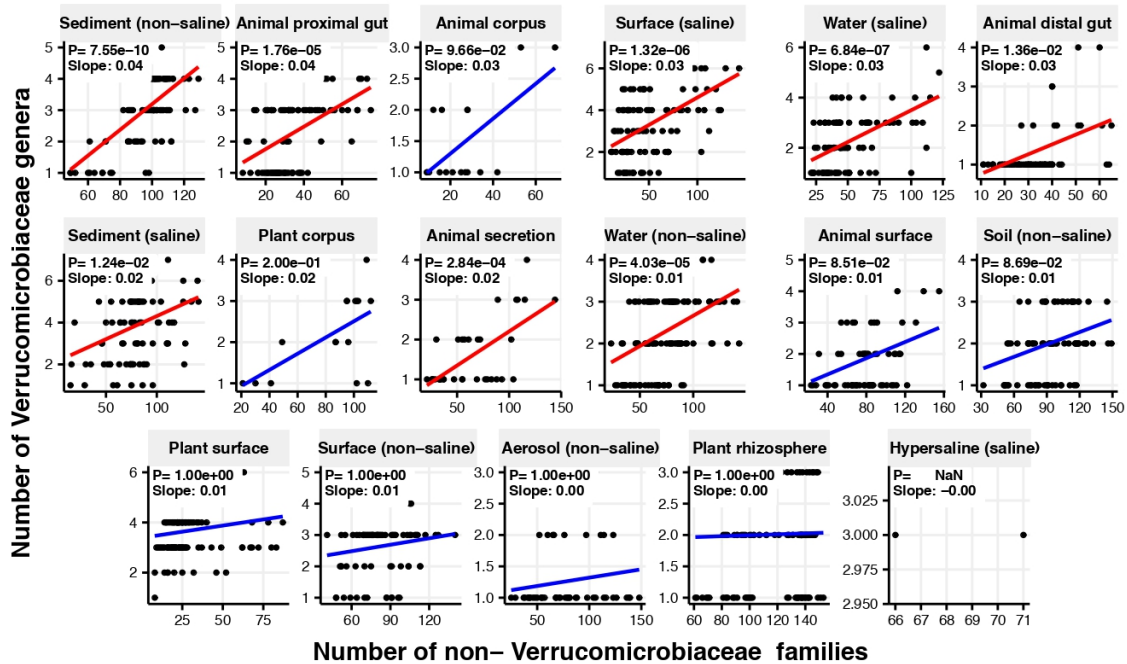
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## B. Spingomonadaceae



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## C. Verrucomicrobiaceae



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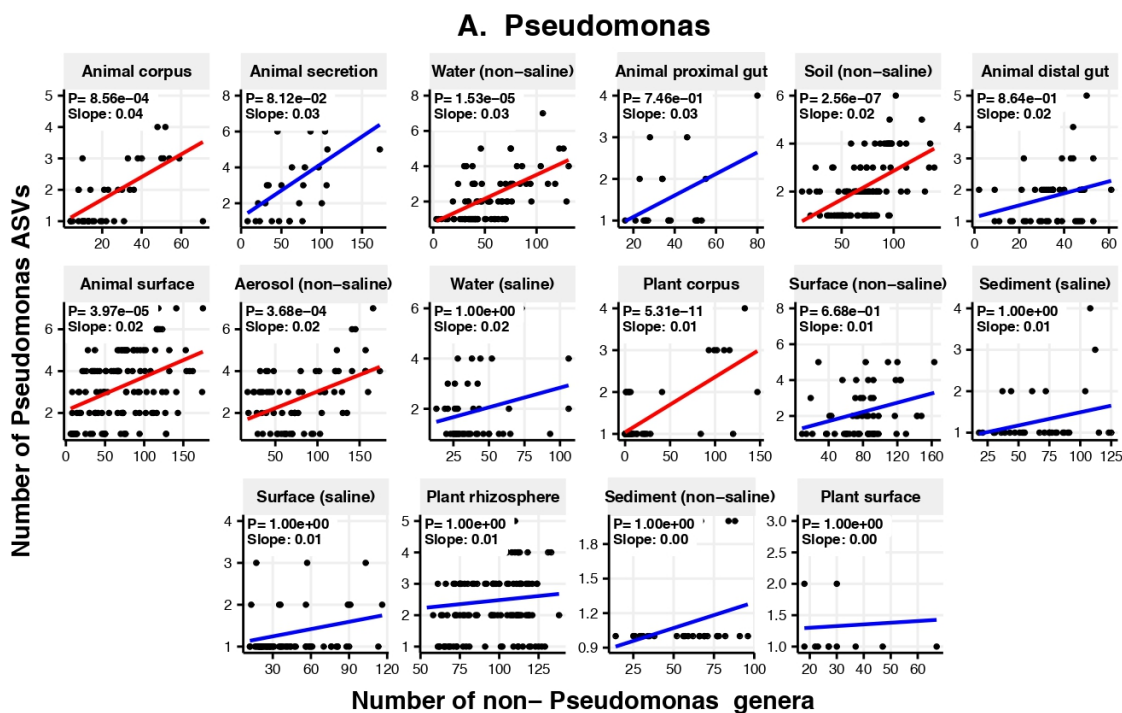
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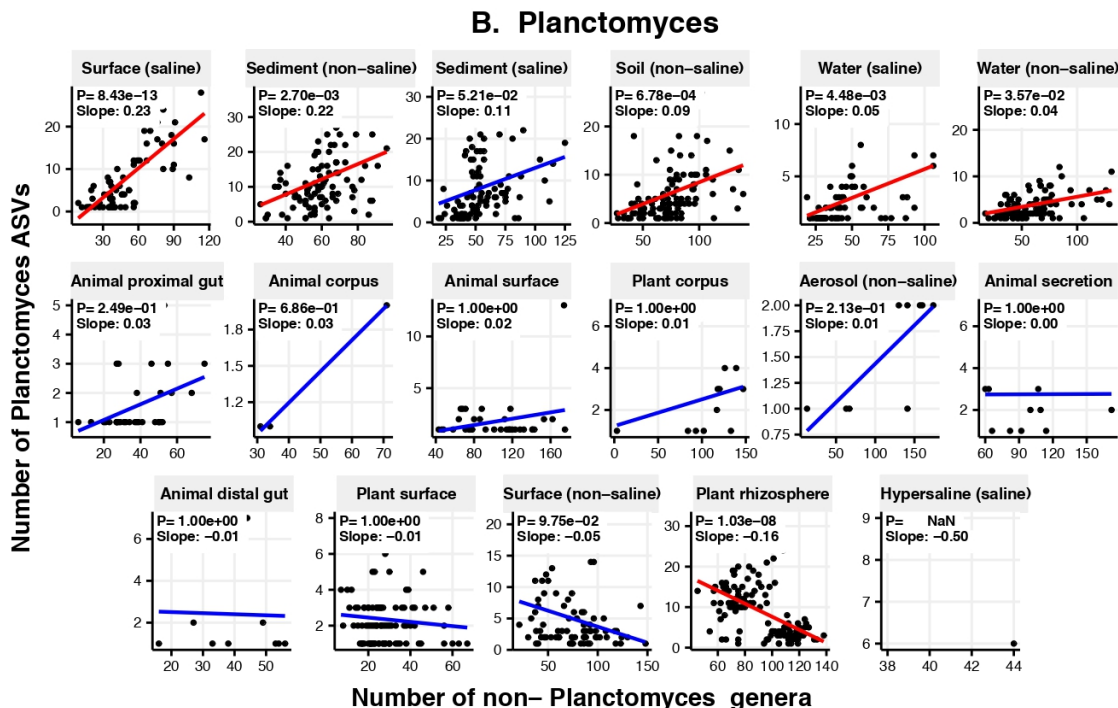
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694 **Figure S6. Diversification as a function of diversity across biomes in the three most**  
 695 **prevalent genera.** Linear models are shown for diversification (ASVs per genus, y-axis)  
 696 as a function of diversity (non-focal genera, x-axis) in each of the 17 environments  
 697 (EMPO3 biomes). P-values are Bonferroni corrected. Significant ( $P < 0.05$ ) models are  
 698 shown with red trend lines, non-significant ( $P > 0.05$ ) trends are shown in blue.

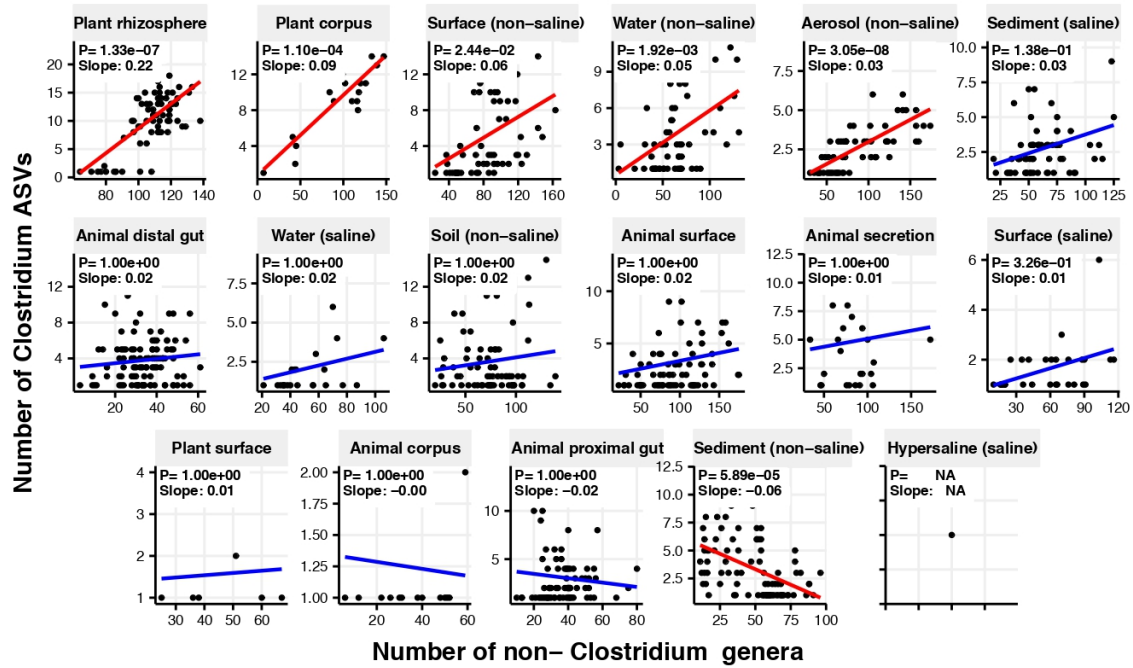


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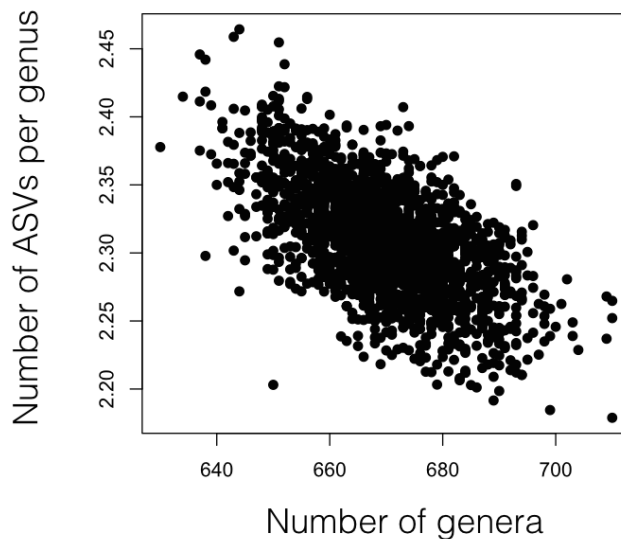
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### C. Clostridium



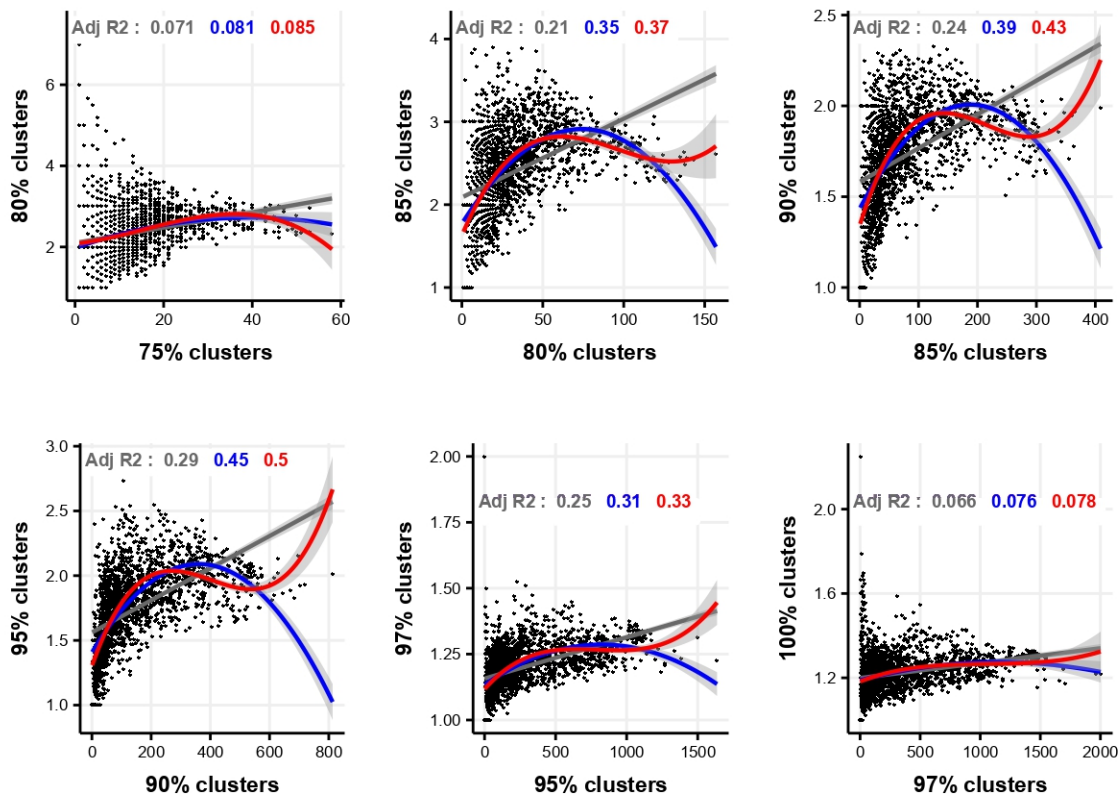
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707 **Figure S7. Permuted EMP data is biased toward a negative diversity-diversification**  
708 **relationship.** We permuted the EMP dataset of 2,000 samples each rarefied to 5,000  
709 sequences/sample and took 2,000 simulated samples, by picking from the overall  
710 distribution of 155,002 unique ASVs across all samples, weighted by their total number  
711 of observations. Thus, the 'true' patterns of co-occurrence between ASVs in the same  
712 sample (and thus any 'biologically true' pattern of either DBD or EC models) is removed  
713 from the data. The permutations yield a negative relationship between diversity (number  
714 of genera) and diversification (number of ASVs per genus): slope = -0.002; Pearson  
715 correlation = -0.61;  $P < 2.2 \cdot 10^{-16}$ .  
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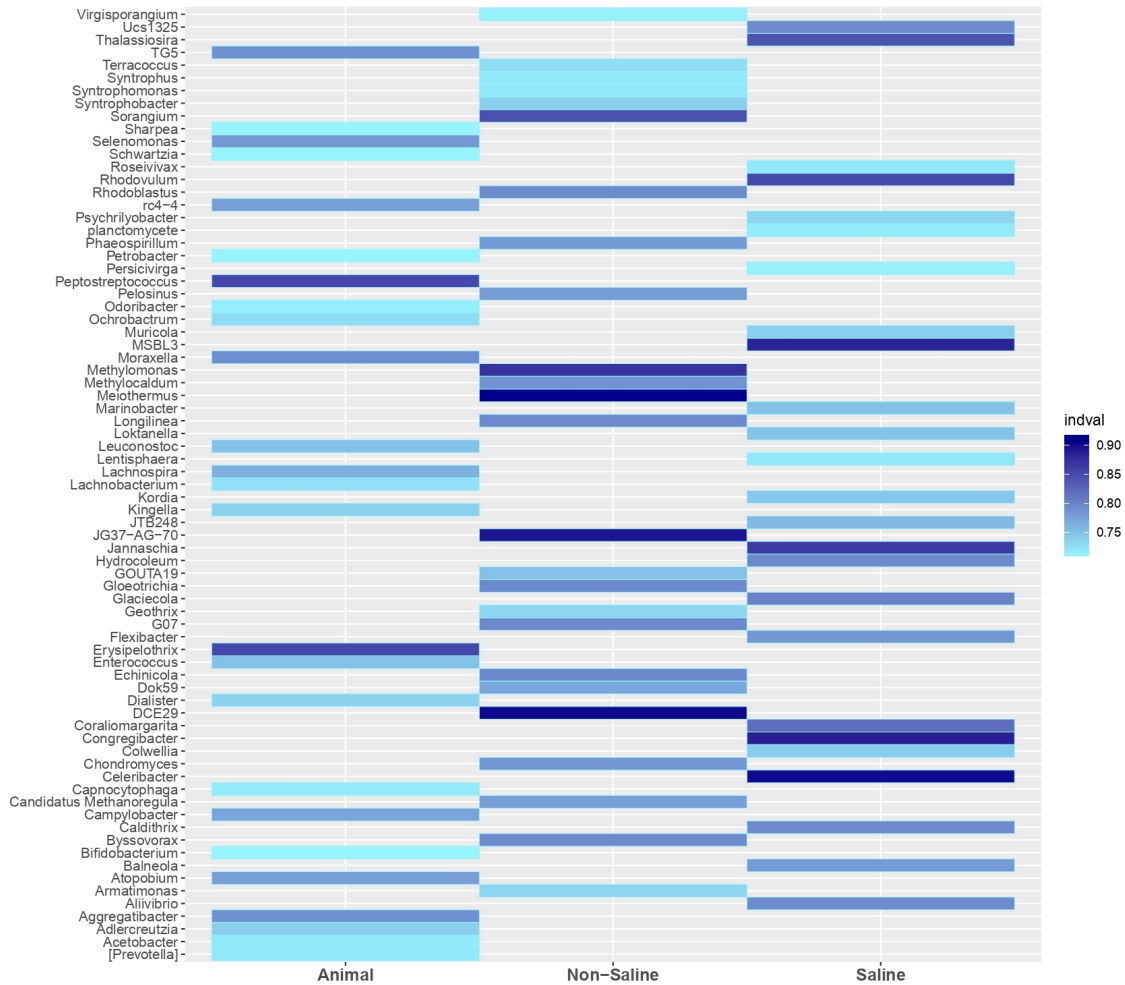
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719 **Figure S8. Linear, quadratic and cubic models for the relationship between**  
720 **diversification and diversity for varying levels of % nucleotide identity.** Diversity  
721 was estimated as the number of clusters at a focal level ( $d_i$ ) and diversification as the  
722 mean of the clusters at the rank above ( $d_{i+1}/d_i$ ). All  $P$ -values are  $< 0.001$ . Linear fit  
723 (grey); quadratic fit (blue), cubic fit (red); same colors for the associated adjusted  $R^2$ . The  
724 x-axis (diversity) shows the number of clusters at the focal percent-identity level ( $d_i$ ), and  
725 the y-axis (diversification) is the mean of the clusters at the rank above ( $d_{i+1}/d_i$ ).



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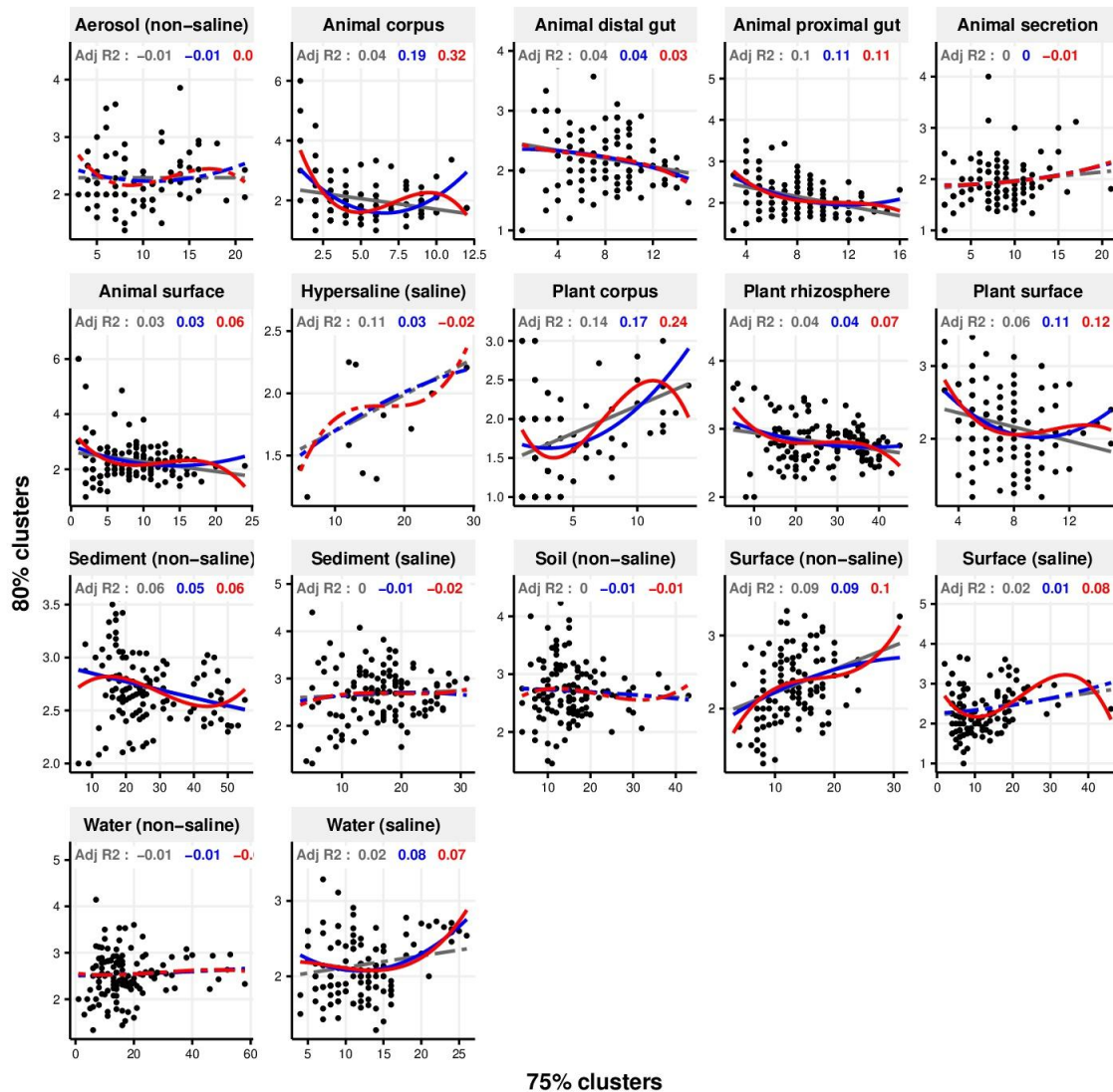
732 **Figure S9. Resident genera of environment clusters.** Results from indicator species  
 733 analysis illustrated as a heatmap. Only the 25 resident genera with the highest indval  
 734 indices and  $P < 0.05$  (permutation test) are shown for every environment cluster (animal-  
 735 associated, non-saline and saline free). For the full results see **Supplementary Data file**  
 736 **2.**  
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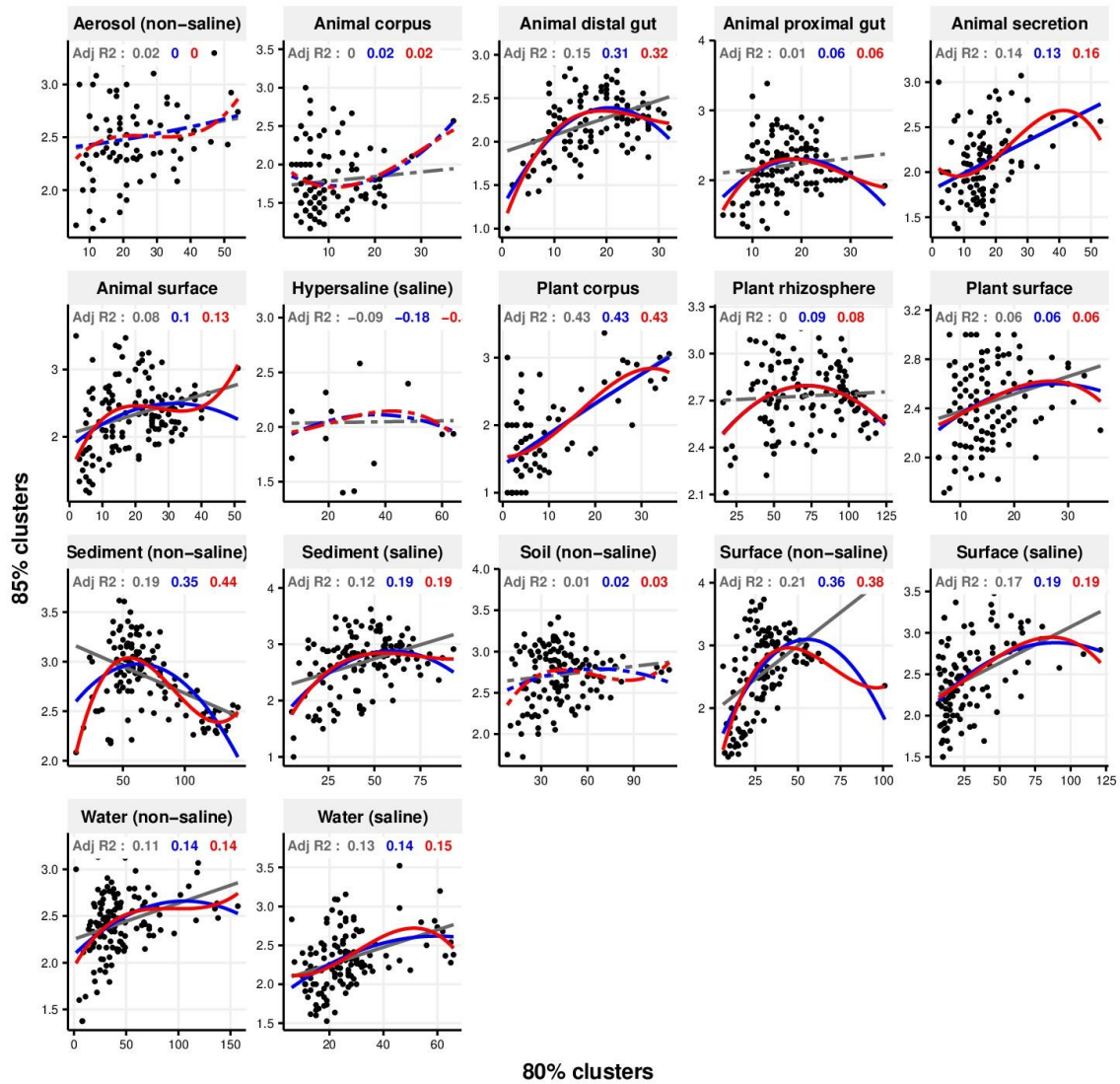
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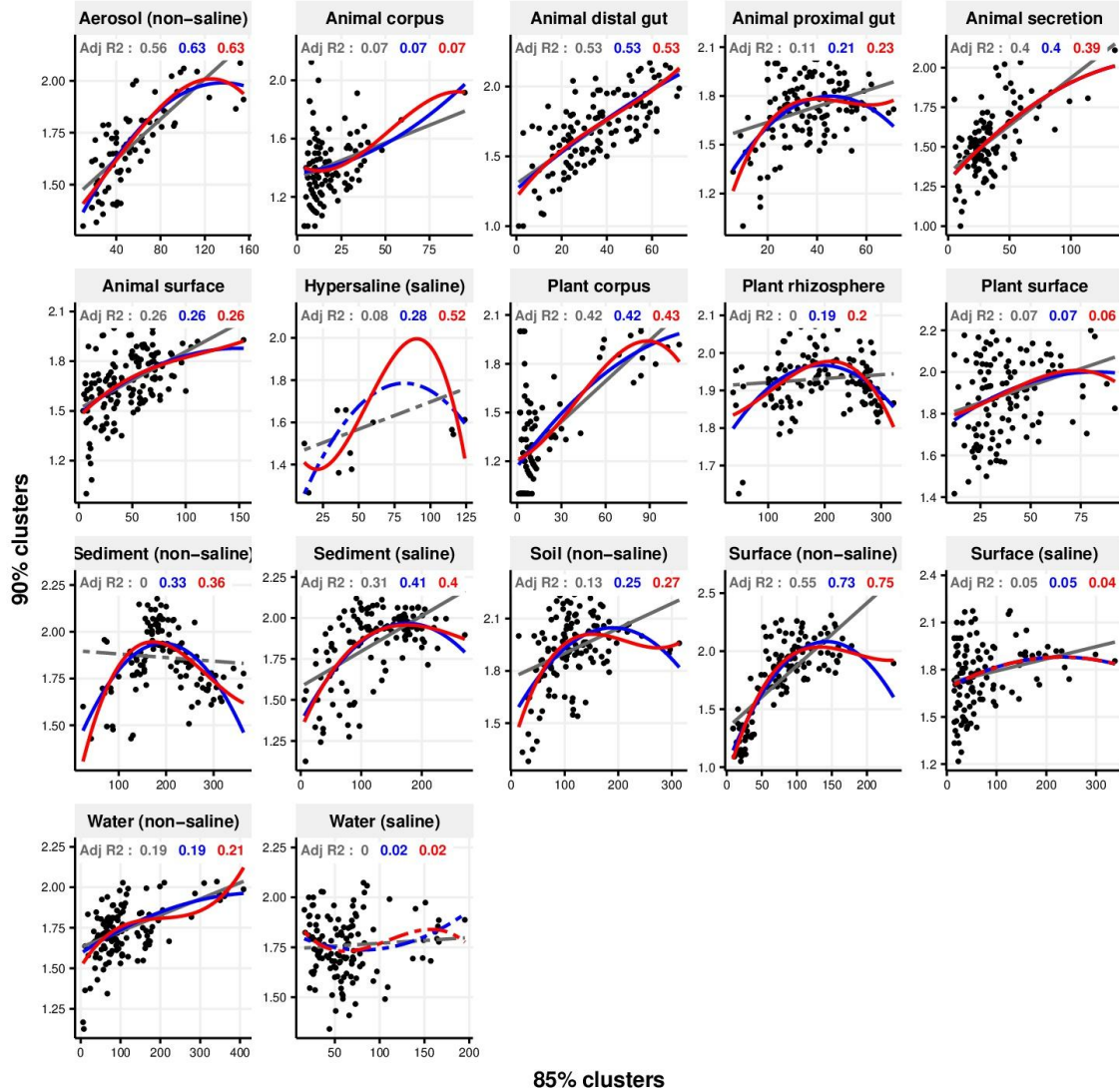
741 **Figure S10. Linear, quadratic and cubic models for diversification-diversity**  
 742 **relationship for each environment type for varying levels of % nucleotide identity.**  
 743 Diversity was estimated as the number of clusters at a focal level ( $d_i$ ) and diversification  
 744 as the mean of the clusters at the rank above ( $d_{i+1}/d_i$ ). Linear (grey), quadratic (blue) and  
 745 cubic (red), with corresponding adjusted R-squared values in the same color. *P*-values are  
 746 Bonferroni corrected for 17 tests. Significant,  $P < 0.05$  (solid lines), non-significant  
 747 (dashed lines). The x-axis shows the number of clusters at the focal percent-identity level  
 748 ( $d_i$ ), and the y-axis is the mean of the clusters at the rank above ( $d_{i+1}/d_i$ ).



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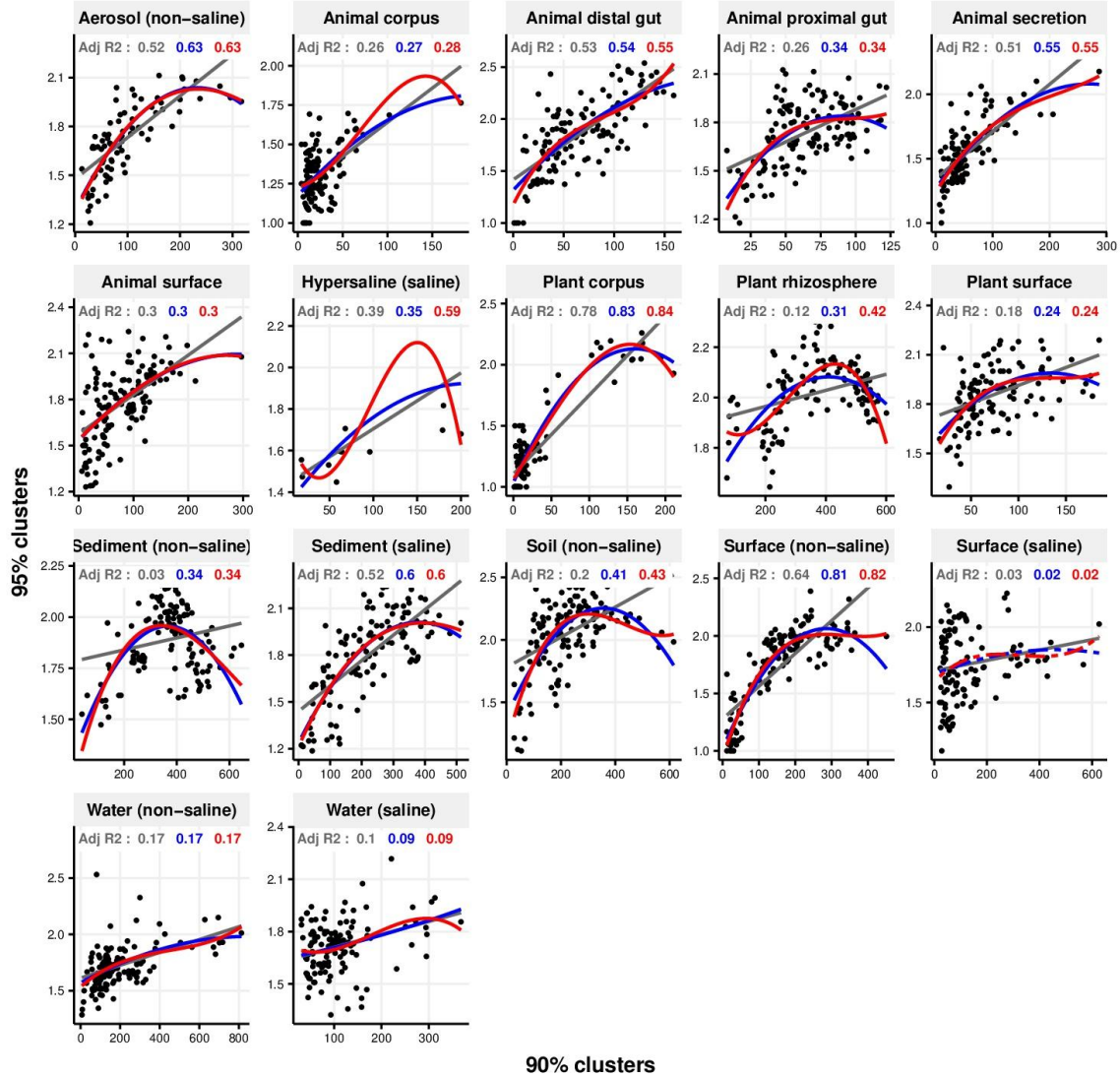


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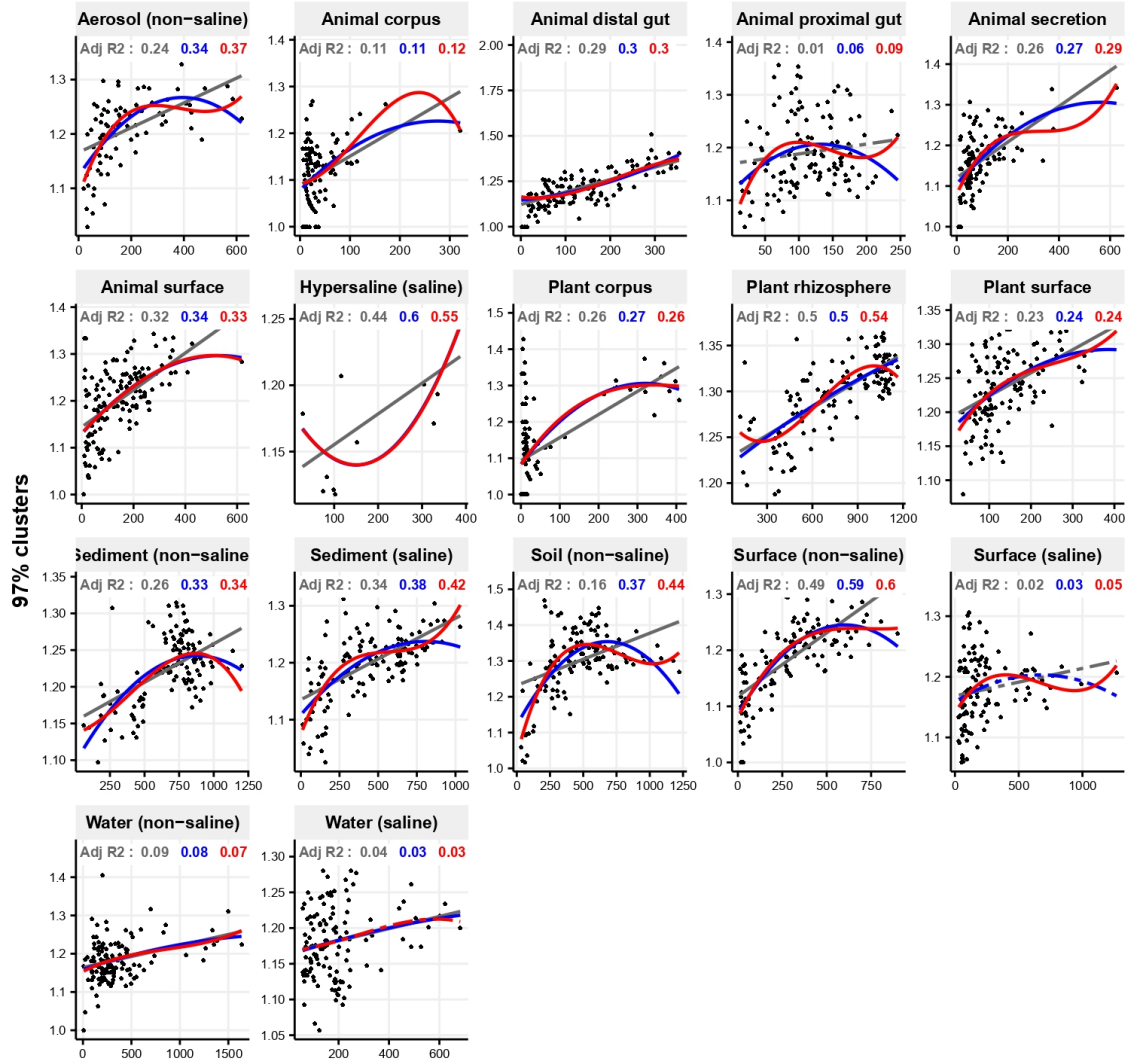


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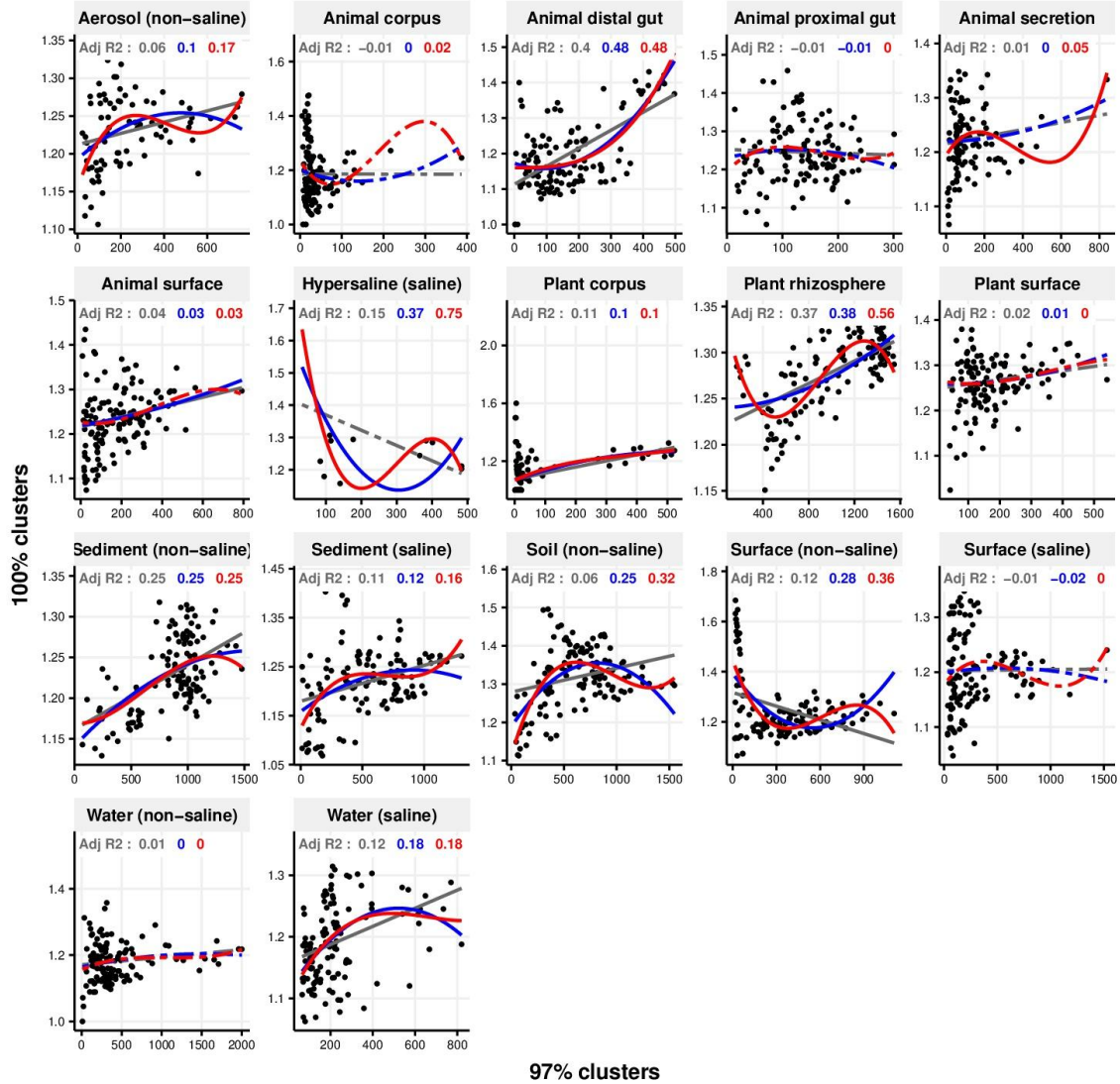
85% clusters



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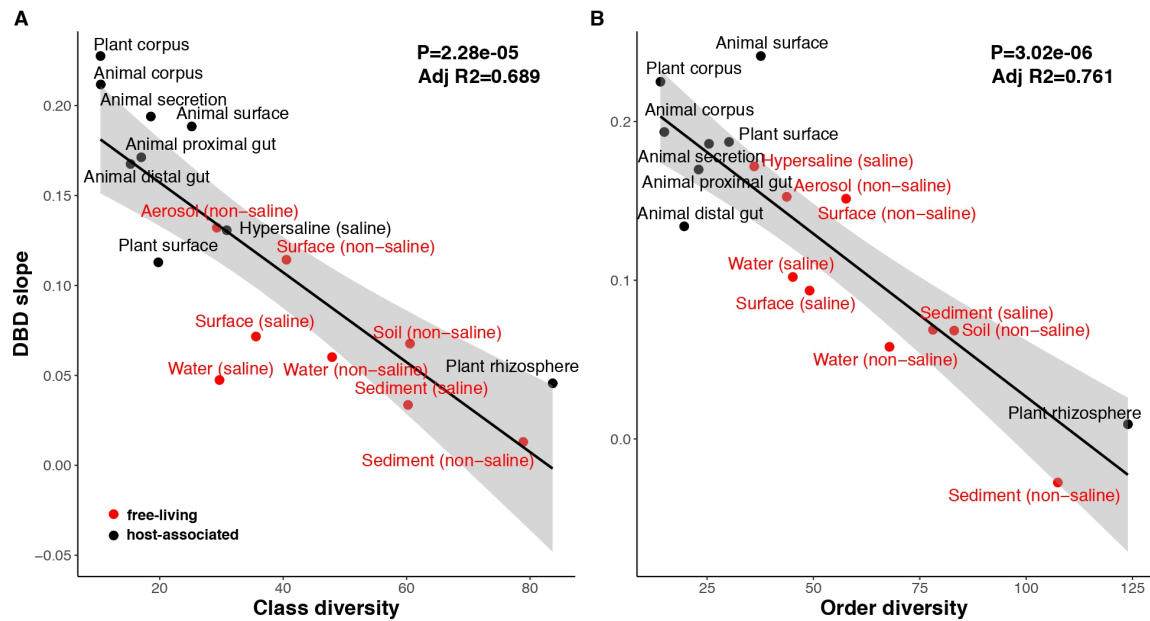
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97% clusters

758 **Figure S11. DBD slope is higher in low-diversity (often host-associated)**  
759 **environments.** The x-axis shows the mean number of (A) classes and (B) orders in each  
760 biome; on the y-axis, DBD slope is the result from the GLMMs predicting diversification  
761 as a function of the interaction between diversity and environment type at (A)  
762 Order:Class and (B) Family:Order ratio (**Supplementary Data file 1 Section 4**). The  
763 line represents a regression line, shaded areas depict 95% confidence limits of the fitted  
764 values.  
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