

# Evolution of simple multicellular life cycles in dynamic environments

Yuriy Pichugin<sup>1</sup>, Hye Jin Park<sup>1</sup>, and Arne Traulsen<sup>1</sup>

<sup>1</sup>Max Planck Institute for Evolutionary Biology, August-Thienemann-Str. 2, 24306 Plön, Germany

March 12, 2019

## Abstract

The mode of reproduction is a critical characteristic of any species, as it has a strong effect on its evolution. As any other trait, the reproduction mode is subject to natural selection and may adapt to the environment. When the environment varies over time, different reproduction modes could be optimal at different times. The natural response to a dynamic environment seems to be bet hedging, where multiple reproductive strategies are stochastically executed. Here, we develop a framework for the evolution of simple multicellular life cycles in a dynamic environment. We use a matrix population model of undifferentiated multicellular groups undergoing fragmentation and ask which mode maximizes the population growth rate. Counterintuitively, we find that natural selection in dynamic environments generally tends to promote deterministic, not stochastic, reproduction modes.

## 1 Introduction

The ability of organisms to reproduce is a paramount feature of life, and a great diversity of reproduction modes is observed in nature. Even the simplest organisms, such as colonial bacteria and primitive multicellular species reproduce in various ways: by producing unicellular propagules [Koyama et al., 1977], by fragmentation of the colony into two [Keim et al., 2004] or multiple multicellular pieces [Rippka et al., 1979], and by dissolution of the organism into independent cells [Stein, 1958]. A variety of reproduction modes originates from different external and internal conditions [Bonner, 1998, Strassmann et al., 2000, Rainey and Rainey, 2003, Fiegna and Velicer, 2003, Travisano and Velicer, 2004]. The choice of the reproduction mode has a major impact on the later evolution of the species' traits. This aspect is especially important for organisms at the brink of multicellular life: the larger the organism grow, the more complex it can become [Smith et al., 2013]. However, it also means longer developmental time, which might incur additional risks; larger propagules require less protection against unfavourable environmental conditions [Shine, 1978], while smaller propagules can be produced in larger quantities [Macarthur and Wilson, 1967, Pianka, 1970]. Thus, the question of the evolution of reproduction modes of simple multicellular organisms and life cycles in general has a paramount importance for our understanding of the history of life on Earth.

Natural selection favours the life cycle, which utilizes the opportunities and handles challenges faced by species the best. The evolution of life cycles among complex organisms is generally slow and with rare exceptions [Sinervo et al., 2000] occurs unnoticed. At the same time, primitive organisms under selection pressure

35 demonstrate an extraordinary ability to adapt their reproductive strategies. Initially, unicellular *Chlamydomonas*  
36 *reinhardtii* experimentally subjected to selection for fast sedimentation in liquid media has evolved into multi-  
37 cellular clusters reproducing via single cell bottleneck [Ratcliff et al., 2013a]. Similar experiments with budding  
38 yeast *Saccharomyces cerevisiae* show the evolution into snowflake-shaped clusters reproducing by fragmenta-  
39 tion [Ratcliff et al., 2013b]. Selection pressure imposed by another source, e.g., the threat from predators, has  
40 similar effect [Herron et al., 2019]. It was shown that even prokaryotic unicellular life forms are capable to  
41 evolve collective-level traits within a matter of months [Hammerschmidt et al., 2014]. These examples show  
42 that natural selection can drive the adaptation of reproductive strategies.

43 The evolution of life cycles has been investigated from the theoretical perspective as well. Roze and Michod  
44 [2001] have studied the evolution of the propagule size and found that the smaller propagules can be selected  
45 since they are more efficient in elimination of selfish mutants. Tarnita et al. [2013] considered “staying together”  
46 mode of group formation, where cell colonies or organisms grow only by means of division of cells already  
47 comprising them (without immigration). There, Tarnita et al. investigated conditions at which the multicellular  
48 life cycle characterized by stochastic detachment of unicellular propagules outperforms the unicellular life cycle.  
49 Recently, we have performed an extensive investigation for the optimal modes of group reproduction [Pichugin  
50 et al., 2017]. It turns out that only life cycles with a regular schedule of reproduction are favoured by natural  
51 selection. However, all these studies consider only constant environments, where external conditions do not  
52 change with time.

53 It has been shown that the fluctuations of the environment have a significant influence on the life cycles of  
54 many species. Natural examples range from the day-night cycles driving photosynthetic activity [Tamiya et al.,  
55 1953], to the spawning of marine invertebrates synchronised with the lunar cycle [Tessmar-Raible et al., 2011,  
56 Kaiser et al., 2016], to the alteration of seasons affecting availability of food, energy spendings, amount of  
57 daylight, etc [Murphy, 1978, Lenz, 1984, Schierwater and Hauenschild, 1990]. An impact of the environmental  
58 changes on the life cycle has also been investigated in evolutionary experiments [Ratcliff et al., 2012, 2013a,b,  
59 Hammerschmidt et al., 2014]; and even the changes imposed by human interventions into nature [Gross, 1991]  
60 have been reported to have an effect on life cycles.

61 The hallmark phenomenon observed in dynamic environments is bet-hedging, where organisms combine  
62 different reproductive strategies [Philippi and Seger, 1989, Beaumont et al., 2009]. Bet-hedging comes in two  
63 flavours: In “between-clutch” bet-hedging, different organisms of the same species use different reproductive  
64 strategies. An example of this is the blooming of the succulents which must coincide with a hardly predictable  
65 wet season in the desert [Venable, 2007, Gremer and Venable, 2014]. Different plants of the same species  
66 have different time of blooming, so those which catch the wet season will successfully reproduce, while others  
67 perish. Similar processes occur in many other plants including crops [Silvertown, 1984]. In “within-clutch”  
68 bet-hedging, offspring produced together have diverse properties [Einum and Fleming, 2004]. An example is  
69 the diversity among egg size in bird clutches: in mild seasons all eggs are hatched, while in harsh seasons only  
70 the larger eggs with more nutrients can survive [Olofsson et al., 2009].

71 A theory describing demographic dynamics in dynamic environments has been developed in [Tuljapurkar,  
72 1989, Orzack and Tuljapurkar, 1989, Tuljapurkar, 1990, Tuljapurkar et al., 2003] for models with discrete time  
73 and has focused on random fluctuations of environment. The arising method for the population growth rate is  
74 also applicable to continuous time models [Kussel and Leibler, 2005].

75 As shown in experimental evolution studies [Ratcliff et al., 2013a, Hammerschmidt et al., 2014, Herron  
76 et al., 2019], under favourable conditions, the evolution of novel life cycles in microbial populations might occur  
77 within a matter of months. Therefore, the ecological and evolutionary processes in fast reproducing populations  
78 are intertwined with each other. Especially, environmental fluctuations strongly affect smaller groups because  
79 they are more likely to be sensitive to these perturbations in the external environment [Libby and Rainey,

80 2013, van Gestel and Tarnita, 2017]; small changes in the environment might lead to the significant changes  
81 in group behaviour. Yet, the evolution of life cycles of simple multicellularity under dynamic conditions still  
82 remains largely unexplored. To which extent can environmental fluctuations affect the patterns of cell colony  
83 reproduction? What reproduction modes thrive in dynamic environments? Do dynamic environments enrich  
84 the space of life cycles that can evolve, or do they impose additional restrictions? We combine methods from  
85 demographic dynamics in dynamic environments with the general framework of fragmentation mode evolution  
86 to answer these questions.

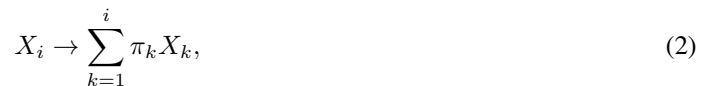
## 87 2 Methods

### 88 2.1 Life cycle of a group-structured population in a static environment

89 We consider a population model, where cells are nested into groups. Reproduction of cells leads to growth of  
90 groups, but external cells are never integrated into groups (no “coming together” in the sense of Tarnita et al.  
91 [2013]). The dynamics of the population is driven by a number of biological reactions representing cell growth  
92 and group fragmentation. After each cell division, cells either stay together as a group or fragment [Tarnita  
93 et al., 2013, Pichugin et al., 2017]. If they stay together, the group size increases, while the number of groups  
94 in the population is unchanged. Such events are given by the reactions



95 where  $X_i$  denotes the group of size  $i$ . Each cell in a group of  $i$  cells has the same birth rate  $b_i$ , and thus the  
96 growth rate of the group is  $ib_i$ . In our model, birth rates  $b_i$  represent the processes influencing the cell growth:  
97 they summarise the benefits and costs of a group living within a certain environment. The birth rate of a solitary  
98 cell can be set to one ( $b_1 = 1$ ) without loss of generality. On the other hand, after division, if a group of cells  
99 fragments instead of staying together, both the total number of cells and the number of groups in the population  
100 increase. The fragmentation results in reactions



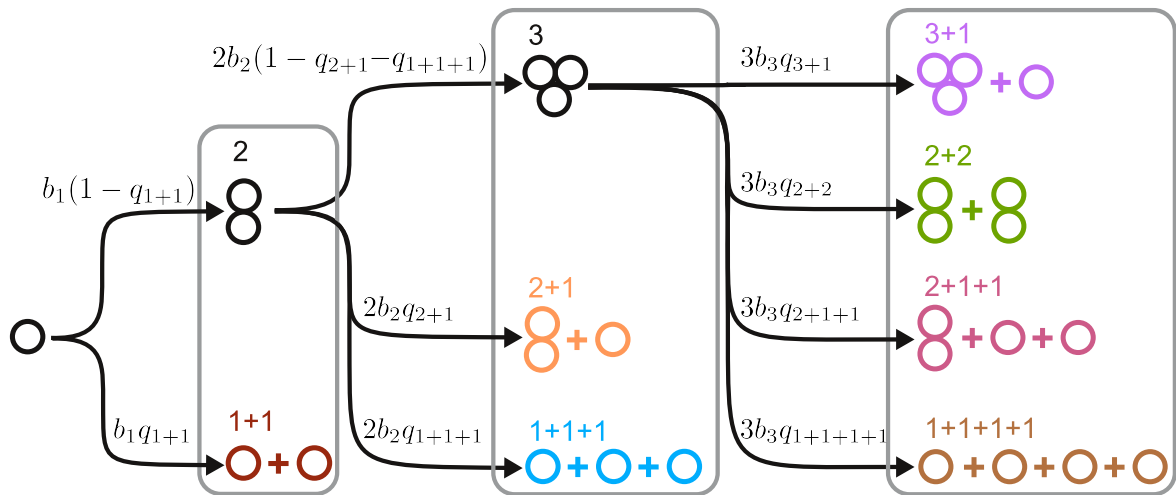
101 where  $\pi_k$  is the number of produced fragments of size  $k$ . Since fragmentation event conserves the total number  
102 of cell,  $\sum_{k=1}^i k\pi_k = i + 1$ .

103 For example, upon reaching size two after a division from a solitary cell, a group may split into two inde-  
104 pendent cells, i.e. execute *fragmentation pattern* 1+1, or cells may stay together increasing the group size from  
105 1 to 2. Upon reaching size three, they may fragment into either a bi-cellular group and an independent cell  
106 (fragmentation pattern 2+1), three independent cells (fragmentation pattern 1+1+1), or cells may stay together  
107 making the group size 3. Upon reaching size four, a group may fragment to one of four fragmentation patterns:  
108 3+1, 2+2, 2+1+1, or 1+1+1+1 (see Fig. 1A), or stay as the group of size 4 and so on. For the sake of calculation  
109 efficiency and illustrative purposes, we limit the maximal group size  $n$  to 3 in our numerical simulations. How-  
110 ever, the approach we developed and our analytical results are not constrained by this limit and are applicable  
111 to populations with any group sizes.

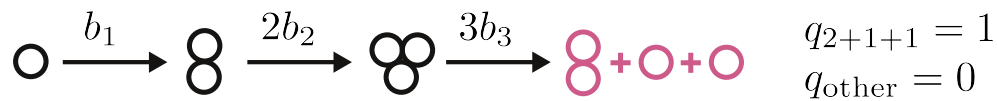
112 For an arbitrary life cycle, the rates of reaction are proportional to the probability of the fragmentation  
113 pattern  $\kappa$  to occur, denoted as  $q_\kappa$ . Thus the set of fragmentation probabilities,

114  $\mathbf{q} = (q_{1+1}; q_{2+1}, q_{1+1+1}; q_{3+1}, q_{2+2}, q_{2+1+1}, q_{1+1+1+1}; \dots)$ , defines the *fragmentation mode* of a population.  
115 In the special case where only a single fragmentation reaction occurs, i.e. all fragmentation probabilities except  
116 one are zero, the life cycle represents a regular schedule of the group development and fragmentation. We refer

## A mixed life cycle



## B pure life cycle



**Figure 1: Structure of “staying together” life cycles for a population with the maximal group size  $n = 3$ .**

**A.** Schematic figure for the life cycles of a population with  $n = 3$ . Birth rates  $b_i$  are identical for all cells in the same group size  $i$ . Fragmentation probability set  $\mathbf{q} = (q_{1+1}; q_{2+1}, q_{1+1+1}; q_{3+1}, q_{2+2}, q_{2+1+1}, q_{1+1+1+1})$  determines the life cycle of the group structure. **B.** Pure life cycles are obtained in a special case where a single fragmentation probability is equal to one, while all others are zero. An example of a pure life cycle using the fragmentation mode  $\mathbf{q} = (0; 0, 0; 0, 0, 1, 0)$  is presented. In a pure life cycle, all groups follow a regular schedule of development and fragmentation.

117 to such cases as *pure life cycles*, see Fig. 1B. In other cases, commonly referred to as *mixed life cycles*, the sum  
 118 of fragmentation probabilities at each size cannot exceed one:  $q_{1+1} \leq 1$  and  $q_{2+1} + q_{1+1+1} \leq 1$ ; at the maximal  
 119 size it has to be one, so in our simulations we used  $q_{3+1} + q_{2+2} + q_{2+1+1} + q_{1+1+1+1} = 1$  (see Fig. 1A).

120 The rates of reactions are density independent and therefore, lead to a set of linear differential equations  
 121 describing the population dynamics, see Appendix A and [Pichugin et al., 2017] for details,

$$\frac{d\mathbf{x}}{dt} = A\mathbf{x}, \quad (3)$$

122 where  $\mathbf{x}$  is the vector for abundances  $x_i$  of groups size  $i$ ,  $\tilde{t}$  is time, and  $A$  is the projection matrix. In a static  
 123 environment, the projection matrix  $A$  does not change over time. Therefore, the population dynamics converges  
 124 to a stationary regime where

$$\lim_{\tilde{t} \rightarrow \infty} \mathbf{x}(\tilde{t}) = e^{\lambda \tilde{t}} \mathbf{w}. \quad (4)$$

125 Here,  $\lambda$  is the leading eigenvalue of the projection matrix  $A$ , and  $\mathbf{w}$  is the right eigenvector associated with  $\lambda$ .  
 126 The growth rate of the population is determined by  $\lambda$  expressed in terms of birth rates  $b_i$  and the fragmentation  
 127 mode  $\mathbf{q}$ . Hence, evolution favours the life cycle which gives the largest  $\lambda$ . For a static environment, it has  
 128 been shown that an evolutionarily optimal life cycle must be a pure binary life cycle, where a parental group  
 129 fragments into exactly two offspring groups [Pichugin et al., 2017].

## 130 2.2 Growth of the group-structured population in a dynamic environment

131 Next, we investigate the evolution of life cycles under dynamic environmental conditions, where growth rates  
 132  $b_i$  do not remain the same through time. Here, we consider a dynamic environment in a form of a regular switch  
 133 between two seasons  $\mathcal{S}_1$  and  $\mathcal{S}_2$ . Each season is characterized by its own set of birth rates:  $\mathcal{S}_1 = (1, b_2^1, b_3^1, \dots)$   
 134 and  $\mathcal{S}_2 = (1, b_2^2, b_3^2, \dots)$ , respectively. Consequently,  $\mathcal{S}_1$  and  $\mathcal{S}_2$  may favour different life cycles. In this setting,  
 135  $\mathcal{S}_1$  lasts for time duration  $\tau_1$  and then switches to  $\mathcal{S}_2$ , which lasts for  $\tau_2$ . Hence, the dynamic environment is  
 136 determined by the two sets of birth rates and two season lengths  $\mathcal{D} = \{\mathcal{S}_1, \tau_1; \mathcal{S}_2, \tau_2\}$ .

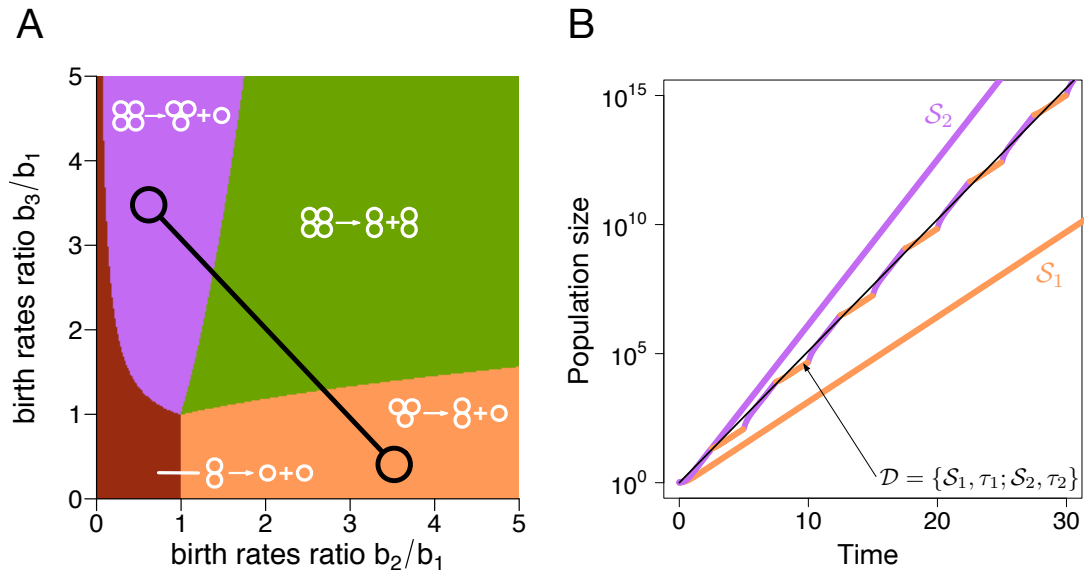
137 In the dynamic environment, the growth rate of the population cannot be characterized by a single projection  
 138 matrix. However, the demographical dynamics within a single season is still described by a single projection  
 139 matrix. Therefore, we numerically simulated the population growth in a dynamic environment using the cor-  
 140 responding projection matrix during each season. We follow the method used in [Kussel and Leibler, 2005] to  
 141 compute the average population growth rate ( $\Lambda$ ) in a dynamic environment ( $\mathcal{D}$ ) over a whole sequence of sea-  
 142 sons, as a slope in the logarithm of the populations size against time, see Fig. 2 and Appendix B. The average  
 143 growth rate  $\Lambda$  in a dynamic environment plays the same role as the leading eigenvalue  $\lambda$  of the projection matrix  
 144 in a static environment: the life cycle with higher  $\Lambda$  will eventually outgrow others with lower  $\Lambda$ .

145 For each studied dynamic environment  $\mathcal{D}$ , we numerically find evolutionarily optimal life cycles by max-  
 146 imizing  $\Lambda(\mathbf{q}, \mathcal{D})$  with respect to the vector of fragmentation probabilities  $\mathbf{q}$ , see Appendix C. Note that we  
 147 perform optimization on a multi-dimensional lattice of  $\mathbf{q}$ , so an accuracy of the optimal life cycle  $\mathbf{q}$  is limited  
 148 by the lattice spacing, which we set to 0.05. We repeat the optimization for different initial conditions to take  
 149 into account the possibility of multiple local optima.

## 150 3 Results

### 151 3.1 Limit regimes of dynamic environments

152 For a given pair of seasons  $\{\mathcal{S}_1, \mathcal{S}_2\}$ , we screened a wide range of seasons lengths combinations  $\{\tau_1, \tau_2\}$  and  
 153 found a set of locally optimal life cycles for each dynamic environment  $\mathcal{D} = \{\mathcal{S}_1, \tau_1; \mathcal{S}_2, \tau_2\}$ . We present



**Figure 2: Unconstrained population growth in dynamic environment can be approximated by exponential growth.** **A.** The map of evolutionarily optimal life cycles in static environments. Only four pure binary life cycles are evolutionarily optimal: 1+1, 2+1, 3+1, or 2+2. The dynamic environment with seasons  $S_1$  and  $S_2$  is represented by a pair of interconnected circles. **B.** The growth of population executing a pure life cycle 3+1 ( $\mathbf{q}_{3+1} = (0; 0, 0; 1, 0, 0, 0)$ ) in dynamic and static environments. Each line shows the temporal growth of the population size. Coloured lines correspond to the population growth in static environments  $S_1 = (1, 3, 0.5)$  and  $S_2 = (1, 0.5, 3)$ , respectively. The two-colored line indicates the population growth in the dynamic environment, alternating between two seasons  $S_1$  and  $S_2$  with  $\tau_1 = \tau_2 = 2.5$ . While the growth in the dynamic environment is complicated in general, it can be approximated very well by exponential growth (thin black line).

154 the results of this screening in the form of optimality maps, which indicate optimal life cycles at a given set of  
 155 season lengths. Each pixel on a map represents a set of season lengths  $\{\tau_1, \tau_2\}$ , and the colour of a pixel is given  
 156 by the all optimal life cycles found in many optimizations from different initial conditions, see Appendix D.  
 157 For convenience, we convert parameters  $\tau_1$  and  $\tau_2$  into the season turnover period  $T \equiv \tau_1 + \tau_2$  and the ratio  
 158 of season lengths  $t \equiv \tau_1/\tau_2$  and present the obtained map using  $T$  and  $t$ . Examples of optimality maps are  
 159 presented in Fig. 3.

160 We focus our analysis on extreme regimes: prevalence of the the single season ( $t \ll 1$  or  $t \gg 1$ ), short  
 161 seasons ( $T \ll 1$ ), and long seasons ( $T \gg 1$ ). The numerical simulations show that at intermediate lengths of  
 162 seasons ( $t \sim 1$  and  $T \sim 1$ ), the behaviour of the system is intermediate between these extremes, see Fig. 3.

163 The simplest behaviour occurs in the prevalence regime ( $t \ll 1$  or  $t \gg 1$ ). In this case, the influence of the  
 164 shorter season on the population growth is negligible. The growth rate in such dynamic environment is close to  
 165 the growth rate in the static environment given by the long season, i.e.  $\Lambda(\mathbf{q}, \mathcal{D}) \approx \lambda(\mathbf{q}, \mathcal{S}_1)$  when the first season  
 166 is much longer ( $t \gg 1$ ). As a consequence, a pure binary life cycle, which is optimal in the static environment  
 167 provided by the first season  $\mathcal{S}_1$ , is evolutionarily optimal in the dynamic environment  $\mathcal{D}$ , where the first season  
 168 prevails ( $t \gg 1$ ). Similarly, the life cycle optimal in the static environment provided by the second season  $\mathcal{S}_2$   
 169 is evolutionarily optimal in dynamic environments, where  $t \ll 1$ . All numerically obtained optimality maps  
 170 confirm this.

171 For short seasons ( $T \ll 1$ ), the population composition changes little within a single period of seasons  
 172 change. Thus, demographical changes in the population occur at much slower time scale than changes of the  
 173 environment. As such, the system effectively experiences the average environment with season lengths being  
 174 the weights of each component [Gokhale and Hauert, 2016]. Thus, in the *short seasons approximation* — the  
 175 population growth rate is given by the growth rate in the averaged static environment  $\bar{\mathcal{S}}$

$$\Lambda_{\text{SSA}}(\mathbf{q}, \mathcal{D}) \approx \lambda(\mathbf{q}, \bar{\mathcal{S}}), \quad (5)$$

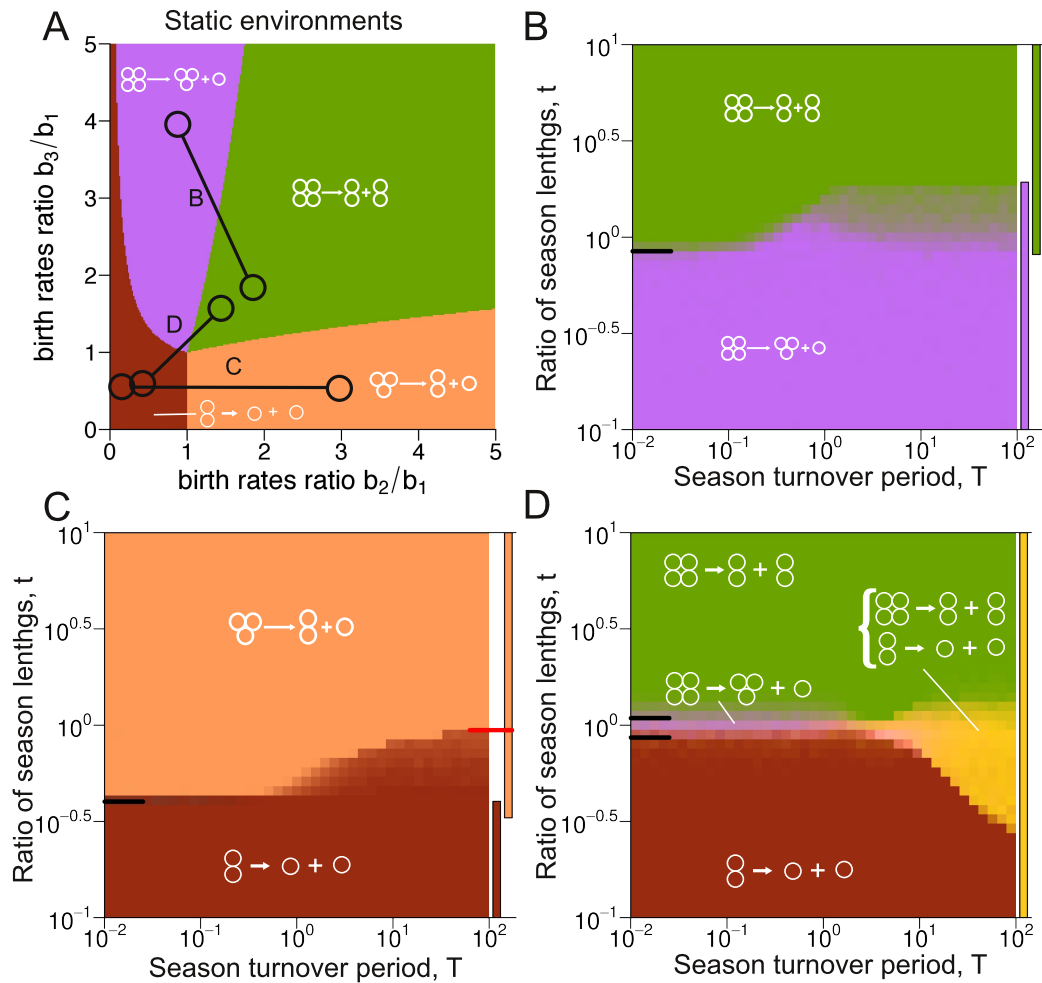
176 where

$$\begin{aligned} \bar{\mathcal{S}} &= \frac{\tau_1}{\tau_1 + \tau_2} \mathcal{S}_1 + \frac{\tau_2}{\tau_1 + \tau_2} \mathcal{S}_2 \\ &= \frac{t}{1+t} \mathcal{S}_1 + \frac{1}{1+t} \mathcal{S}_2. \end{aligned} \quad (6)$$

177 This approximation also allows us to use the results of the optimal life cycles in a static environment  
 178 [Pichugin et al., 2017]. These imply that for any dynamic environment with short seasons, there is only a  
 179 single evolutionarily optimal life cycle, which is a pure binary life cycle. In addition, the short seasons ap-  
 180 proximation allows us to explicitly find the border value of  $t$  separating the area of optimality of favoured life  
 181 cycles in the static environment given by  $\mathcal{S}_1$  and  $\mathcal{S}_2$ , respectively. The border is determined by the ratios of  
 182 season lengths  $t$  at which the average environment lies at the border between areas of optimality in static envi-  
 183 ronments. Numerical simulations reproduce this border well, see the predicted border marked at the left sides  
 184 of the optimality maps in Fig. 3.

185 For the long season length ( $T \gg 1$ ), the population reaches the stationary regime within each season, and  
 186 the transient growth regime between two adjacent seasons can be negligible. This suggest the *long seasons*  
 187 *approximation* — the population growth rate is given by the weighted average of growth rates in static environ-  
 188 ments

$$\begin{aligned} \Lambda_{\text{LSA}}(\mathbf{q}, \mathcal{D}) &\approx \frac{\tau_1}{\tau_1 + \tau_2} \lambda(\mathbf{q}, \mathcal{S}_1) + \frac{\tau_2}{\tau_1 + \tau_2} \lambda(\mathbf{q}, \mathcal{S}_2) \\ &= \frac{t}{1+t} \lambda(\mathbf{q}, \mathcal{S}_1) + \frac{1}{1+t} \lambda(\mathbf{q}, \mathcal{S}_2). \end{aligned} \quad (7)$$



**Figure 3: Evolutionary optimality of life cycles in extreme regimes can be described by analytical approximations.**

**A** The dynamic environment used in panels B, C, and D are represented by connected circles. **B** The optimality map featuring only pure life cycles. When season 1 is much longer than season 2,  $t \gg 1$ , regardless of the overall seasons turnover length, the optimal life cycle is always the pure life cycle 2+2. For the opposite case,  $t \ll 1$ , 3+1 is always optimal. For short seasons,  $T \ll 1$ , the border between two life cycles is located at the position predicted by the short seasons approximation (black tick line on the left side of the map). For long seasons  $T \gg 1$ , for this dynamic environment, the transition between two life cycles is performed through the bi-stability between pure life cycles, as suggested by long seasons approximation (see main text). Coloured bars on the right side of the map show the areas of stability of pure life cycles according to the long seasons approximation. **C** The optimality map featuring a mixed life cycle at the long seasons regime. For  $t > 0.287$ , pure life cycle 1+1 is not evolutionarily optimal, yet the mixed life cycle featuring pattern 1+1 persists up to  $t = 0.485$ , where it disappears in a saddle-node bifurcation indicated by red mark (see main text). **D** The pure life cycles 1+1 and 2+2 can be executed within the same mixed life cycle at different seasons (see main text), and such a combination (yellow) might be evolutionarily optimal in long seasons regime. Exact parameters used in calculations and simulations are presented in Appendix L.



189 Under the long seasons approximation at intermediate values of  $t$ , the optimal life cycle is not necessarily pure,  
 190 and there might be more than one locally optimal life cycle, see examples on Fig. 3C. The transition in  $t$  from  
 191 the prevalence of one season to another is non trivial in this case. These changes in optimal life cycles along  $t$   
 192 in the long seasons regime ( $T \gg 1$ ) are our focus in the remaining part of this study.

### 193 3.2 Borders of the prevalence regimes

194 In the limits  $t \rightarrow \infty$  and  $t \rightarrow 0$ , the optimal life cycles are determined by the prevalent season. However, as  
 195 we increase the other season length, the optimal life cycle may change. In this section, we examine the border  
 196 of these prevalence regimes. We begin by considering extremely small  $t$  and measure how long the optimal life  
 197 cycle persists against the increase of  $t$ .

198 We sampled 40,000 pairs of seasons and investigated at which  $t$  the prevalence regime is violated, see  
 199 Appendix E for details of sampling. We found that the border of the prevalence regime significantly varies  
 200 between season combinations. For some pairs of seasons, the prevalence regime is extremely robust while some  
 201 others show that the prevalence regime is extremely fragile; even at  $t = 10^{-4}$  the prevalence was violated.

202 We found that the main factor determining the robustness of the prevalence regime is the environment in the  
 203 prevalent season. The most fragile prevalence regimes were observed for environments at the border between  
 204 two different types of life cycles and environments promoting unicellular life cycle, see Fig. 4A. At the same  
 205 time, the most robust prevalence regimes were observed for environments far from optimality borders, see  
 206 Fig. 4B.

### 207 3.3 Stability of life cycles in the long seasons regime

208 In this section, we investigate what kind of life cycles emerge to be evolutionary optimal at intermediate  $t$  in  
 209 the long seasons regime ( $T \gg 1$ ). In the long seasons regime, the population growth in a dynamic environment  
 210 can be inferred from the growth rates in two stationary environments, see Eq. (7). Therefore, the analysis of  
 211 evolutionary optimality of life cycles can be performed with relatively simple expressions.

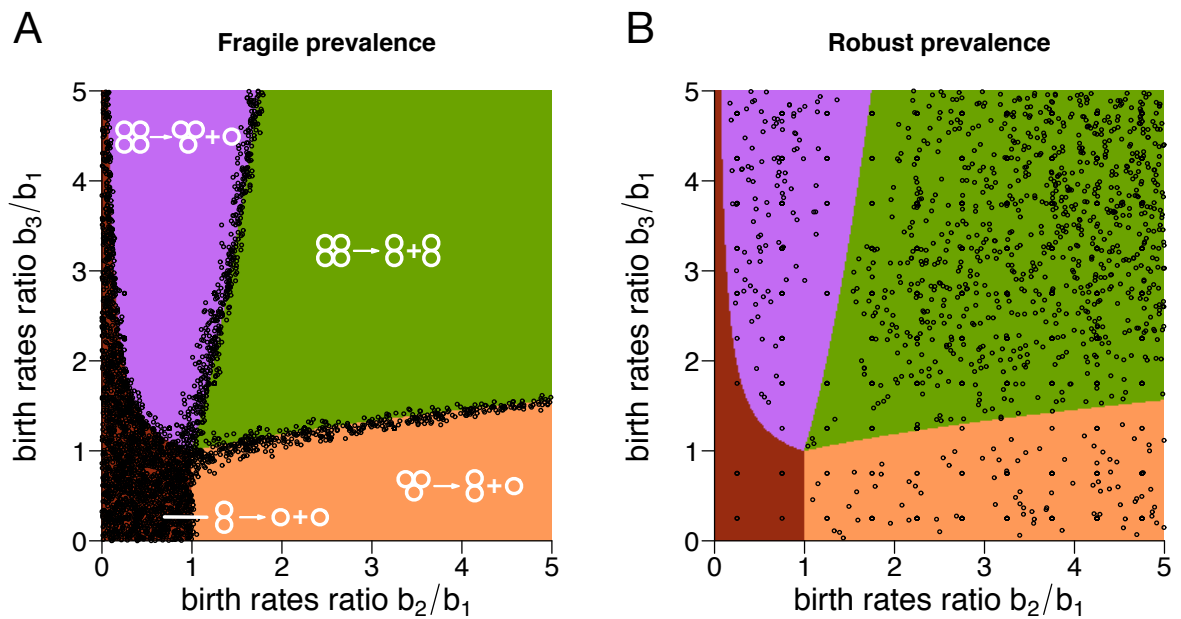
212 An arbitrary fragmentation mode ( $\mathbf{q}$ ) is a local optimum of the growth rate  $\Lambda$ , when any small change in  
 213 the probabilities set  $\mathbf{q}$  leads to a decrease in the population growth rate. For this to happen, all fragmentation  
 214 patterns  $\kappa$  must fulfil the conditions

$$\begin{cases} \left. \frac{\partial \Lambda}{\partial q_\kappa} \right|_{\mathbf{q}} < 0 & \text{if } \kappa \text{ is not executed } (q_\kappa = 0), \\ \left. \frac{\partial \Lambda}{\partial q_\kappa} \right|_{\mathbf{q}} = 0 \text{ and } \left. \frac{\partial^2 \Lambda}{\partial q_\kappa^2} \right|_{\mathbf{q}} < 0 & \text{if } \kappa \text{ is mixed with other patterns of the same size } (0 < q_\kappa < 1), \\ \left. \frac{\partial \Lambda}{\partial q_\kappa} \right|_{\mathbf{q}} > 0 & \text{if } \kappa \text{ is the only executed pattern of its group size } (q_\kappa = 1). \end{cases} \quad (8)$$

215 First, we consider the evolutionary optimality of pure life cycles, where only one fragmentation pattern occurs,  
 216 see Fig. 1B. These life cycles establish a regular schedule of group growth and reproduction, which is commonly  
 217 observed in nature. From the perspective of our model, in pure life cycles  $q_\kappa = 0$  or  $q_\kappa = 1$ , the investigation  
 218 of evolutionary optimality invokes only the first order derivatives, see Eq. (8). A pure life cycle  $\mathbf{q}$  becomes  
 219 evolutionary unstable when an admixture of at least one of absent fragmentation patterns (with  $q_\kappa = 0$ ) no  
 220 longer decreases the growth rate:  $\frac{\partial \Lambda}{\partial q_\kappa} = 0$ . For a given pair of the pure life cycle  $\mathbf{q}_1$  and admixture life cycle  
 221  $\mathbf{q}_2$ , this is achieved at the ratio of the season length

$$t_s(\kappa_1, \kappa_2) = - \left. \frac{\lambda'_x(\mathbf{q}(x), \mathcal{S}_2)}{\lambda'_x(\mathbf{q}(x), \mathcal{S}_1)} \right|_{x=1}, \quad (9)$$

222 where  $\mathbf{q}(x)$  is the mixed life cycle in which fragmentation occurs by  $\kappa_1$  with probability  $x$  and occurs by  $\kappa_2$   
 223 with probability  $1 - x$ , see Appendix F for details. Knowing the values of  $t_s$  for all pairs  $(\kappa_1, \kappa_2)$  makes it



**Figure 4: The robustness of the prevalence regime is mainly determined by the position of the prevalent environment.**  
**A.** Circles show prevalent seasons, for which the violation of the prevalence regime occurred already when the fraction of time  $t$  spend in the second, randomly chosen environment is small,  $t < 0.1$ . These non-robust seasons either promote unicellularity or are located at the borders between areas of optimality - where at least two life cycles have similar growth rate. **B.** Circles show prevalent seasons, for which the violation of the prevalence regime occurred only when the fraction of time spend in the other environment, is very large,  $t > 10$ . The distribution of these seasons has low density in the border regions.

224 possible to outline the ranges of  $t$  where each pure life cycle is optimal. We denote this optimality areas as  
 225 coloured bars to the right of each optimality map, see Fig 3.

226 Next, we consider mixed life cycles, which can emerge as evolutionary optimal in the long seasons regime.  
 227 Above, we have shown that only pure life cycles are evolutionary optimal if there is effectively a single season  
 228 ( $t \ll 1$  or  $t \gg 1$ ). Therefore, as  $t$  approaching these extreme values, evolutionary optimal mixed life cycles  
 229 cease to exist. This happens by one of two scenarios: either a mixed life cycle transforms into a pure one, or  
 230 it merges with the local minimum of  $\Lambda$  and disappears in a saddle-node bifurcation. The majority of mixed  
 231 life cycles observed in our simulations feature only two fragmentation patterns. For these life cycles, transition  
 232 from a mixed optimal life cycle into a pure one occurs at values  $t$  given by Eq. (9). The saddle-node bifurcation  
 233 (if exists) occurs at  $x = x^*$  and  $t = t^*$  satisfying

$$\frac{\lambda''_{xx}(\mathbf{q}(x), \mathcal{S}_1)}{\lambda'_x(\mathbf{q}(x), \mathcal{S}_1)} \Big|_{x=x^*} = \frac{\lambda''_{xx}(\mathbf{q}(x), \mathcal{S}_2)}{\lambda'_x(\mathbf{q}(x), \mathcal{S}_2)} \Big|_{x=x^*}, \quad (10)$$

234 and

$$t^*(\kappa_1, \kappa_2) = - \frac{\lambda'_x(\mathbf{q}(x), \mathcal{S}_2)}{\lambda'_x(\mathbf{q}(x), \mathcal{S}_1)} \Big|_{x=x^*}, \quad (11)$$

235 see Appendix G for detailed analysis of mixed life cycles optimality. We highlight the positions of saddle-node  
 236 bifurcations  $t^*$  by red marks on the right hand side of our optimality maps, see Fig 3C.

237 Finally, we found a distinct solution, when the maximal size of groups produced in one fragmentation pattern  
 238 is smaller than the minimal size of offspring produced by another pattern. With the maximal group size 3, there  
 239 is a single such pair:  $\kappa_1 = 1+1$  and  $\kappa_2 = 2+2$ . When the two seasons in the dynamic environment favour  
 240 the pure fragmentation modes 1+1 and 2+2, the optimal life cycle in a dynamic environment in a long seasons  
 241 regime is always  $\mathbf{q}_C = \mathbf{q}_{1+1} + \mathbf{q}_{2+2} = (1; 0, 0; 0, 1, 0, 0)$ , see Appendix I for details. In the long seasons  
 242 regime, a population employing  $\mathbf{q}_C$  is capable to execute pure life cycle 1+1 during seasons favouring 1+1  
 243 over 2+2; and pure life cycle 2+2 during seasons favouring 2+2 over 1+1. We call such scenario a coexisting  
 244 fragmentation mode  $\mathbf{q}_C$  and the optimality map in Fig. 3D presents this. Except in this special case, numerical  
 245 simulations confirm our analytical results.

### 246 3.4 The spectrum of evolutionary optimal life cycles in dynamic environments is di- 247 verse

248 In this section, we consider, which fragmentation patterns can contribute to evolutionarily optimal life cycles.  
 249 We begin our analysis from a specific scenario, where all birth rates are similar to each other. From a technical  
 250 point of view, the situation where all birth rates are equal, constitutes a neutral environment at which all growth  
 251 rates of any mixed or pure life cycles are equal. We consider near neutral environments, where cell birth rates in  
 252 both seasons slightly deviate from one  $b_i = 1 + \varepsilon\beta_i$ ,  $\varepsilon \ll 1$ , where  $\beta_i$  and thus  $b_i$  is different in the two seasons.  
 253 As a consequence, in the vicinity of this neutrality point ( $\varepsilon \ll 1$ ), the growth rate of any life cycle is close to one  
 254 and can be represented in a form  $\Lambda(\mathbf{q}, \mathcal{D}) = 1 + \varepsilon\Lambda_1 + O(\varepsilon^2)$ , where  $\Lambda_1$  is associated with the first derivatives,  
 255 see Appendix J. Biologically, this corresponds to a scenario, where living in a group has only minimal impact  
 256 on the cell growth. This scenario is seems to be relevant in the early stages of the evolution of multicellularity,  
 257 where benefits of the group formation are minimal due to the absence of adaptations to collective life.

258 There, the case where the maximal group size is limited to two is analytically trackable, see Appendix J. For  
 259 groups not exceeding size two, only three fragmentation patterns are available: 1+1 (unicellularity), 2+1 (binary  
 260 fragmentation) and 1+1+1 (multiple fragmentation). We proved that only three types of life cycles can emerge in  
 261 near neutral environments: pure unicellularity and binary fragmentation, as well as the mixed life cycle utilizing

262 both of them, see Fig. 5. The multiple fragmentation pattern is unable to contribute to evolutionarily optimal  
263 life cycles here.

264 Releasing the size constraints, we found that in near neutral environments, the evolutionary optimality of  
265 life cycles is independent on the seasons turnover period  $T$ . Life cycles evolve similarly in both the short and the  
266 long season regime, see Fig. 6A and Appendix K for a proof. As a consequence, in near neutral environments,  
267 the optimal pure life cycles can be inferred from the short seasons approximation. Since the areas of optimality  
268 map are separated by narrow borders in the order of  $\varepsilon$  in the short seasons regime, the pure binary fragmentation  
269 modes are evolutionarily optimal for the majority of dynamic environments with any season turnover period.

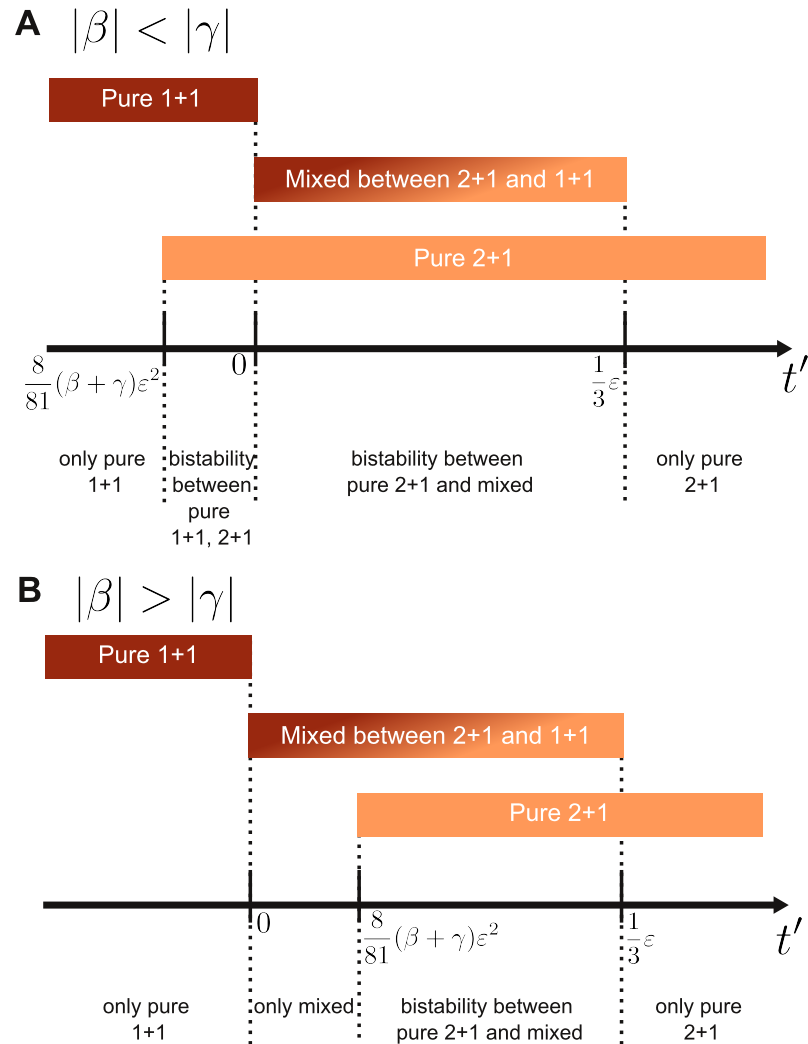
270 If the birth rates are not in the vicinity of the neutral point  $b_i = 1$ , the set of optimal life cycles violates this  
271 scheme. Beyond the near neutral environment, we find more complex life cycles. We find that the fragmentation  
272 pattern presented in the short seasons regime may be absent in the long seasons regime, see Fig. 6B. The  
273 opposite is also possible, a fragmentation pattern absent in the short seasons may appear in the long seasons (as  
274 a component of a mixed life cycle, though), see Fig. 6C. Moreover, the multiple fragmentation, which cannot be  
275 evolutionarily optimal in any static environment, can evolve in a dynamic environment (again, as a component  
276 of the mixed life cycle), see Fig. 6D.

### 277 **3.5 Robustness of pure binary fragmentation**

278 While the potential diversity of the life cycles in dynamic environment is huge, an exotic behaviour is rare,  
279 and requires a fine balance of cell birth rates profiles  $\{\mathcal{S}_1, \mathcal{S}_2\}$  and seasons lengths  $\{\tau_1, \tau_2\}$ . In the data set  
280 used in section 3.2, for each of 40,000 pairs of seasons, we screened 41 different season length ratios (in  
281 total 1,640,000 dynamic environments), see Appendix E for details. For each environment, we found and  
282 characterized the set of evolutionarily optimal life cycles, see Fig. 7. The majority of dynamic environments  
283 (87%) promoted a unique pure life cycle. A smaller fraction (13%) featured coexistence of multiple local  
284 optima. A much smaller fraction of dynamic environments (1.9%) exhibited mixed life cycles, the majority of  
285 which was composed of two fragmentation patterns. Finally, multiple fragmentation was observed in a tiny set  
286 of environments (0.04%). Therefore, we conclude that pure life cycles should be a widespread evolutionary  
287 strategy in changing environment.

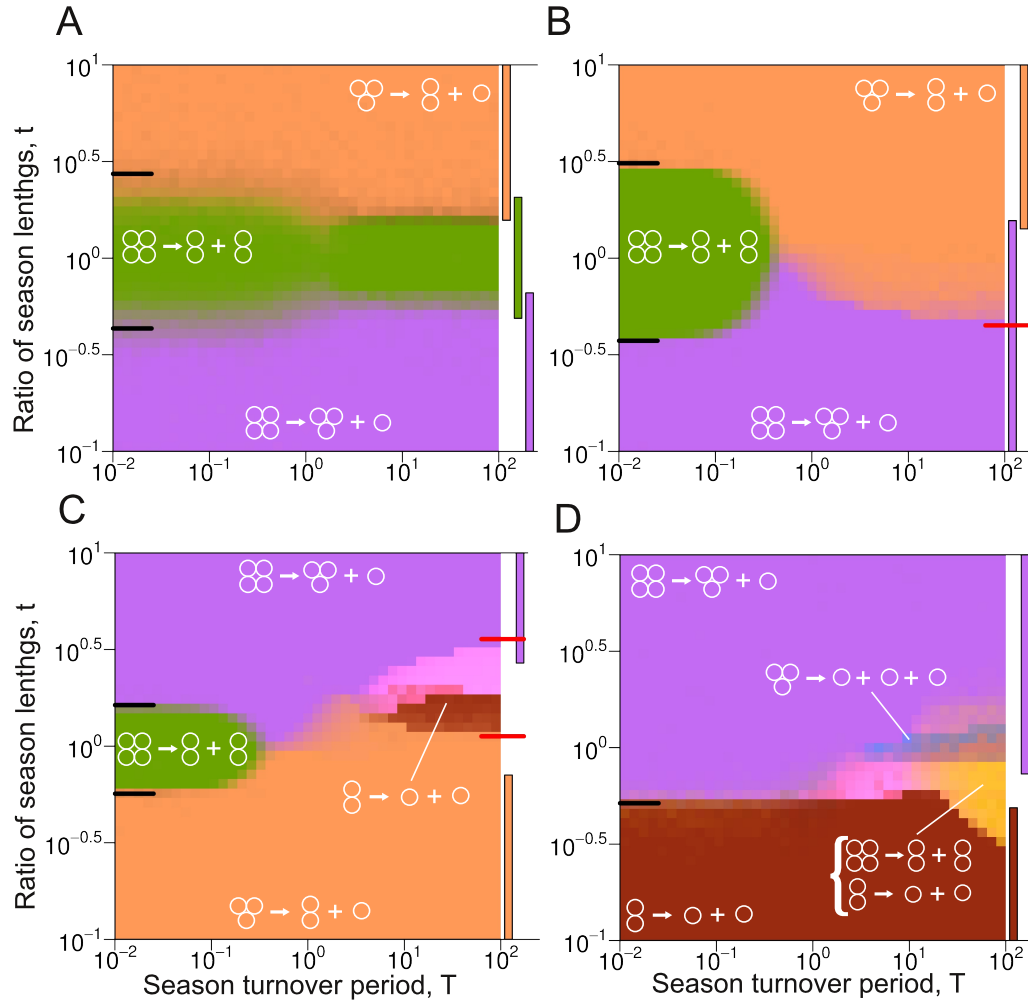
288 To support our result, we investigate the evolution of life cycles of larger colonial organisms. We consider  
289 groups growing up to size  $n = 15$ , so fragmentation must happen upon the birth of 16-th cell in a group. This  
290 size limit is comparable to the size of some volvocales algae, such as *Gonium pectorale* - one of the model  
291 organisms used to study the evolution of multicellularity.

292 We use the cell birth rate profiles  $b_i = 1 + M \left( \frac{i-1}{n-1} \right)^\alpha$ . Investigation of these profiles in static environments  
293 [Pichugin et al., 2017] revealed that  $\alpha \ll 1$  promotes an equal split (8 + 8), and  $\alpha \gg 1$  promotes production  
294 of unicellular propagules (15 + 1). Note that the value of  $M$  has a relatively small influence. With these  
295 profiles, any increase in size is always beneficial to the group. Therefore, we restricted the optimization of life  
296 cycles to only fragmentation patterns of 16-cell groups. We obtain the optimality map, and the result supports  
297 our conclusion, see Fig. 8. Out of 1681 dynamic environments investigated, 1673 (99.5%) promoted a single  
298 unique optimal life cycle in a form of binary fragmentation. Only 8 environments (0.5%) had a mixed optimal  
299 life cycle, and no environment exhibited a coexistence of several local optima, or fragmentation into multiple  
300 pieces.



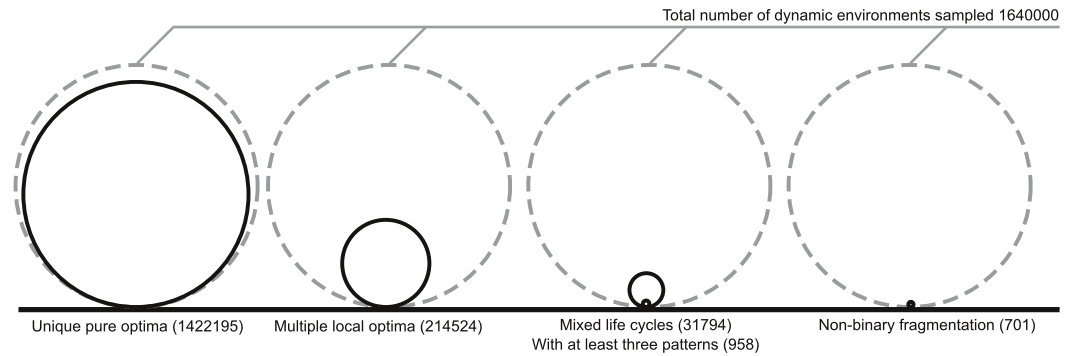
**Figure 5: Only three kinds of life cycles can be evolutionarily optimal in near neutral environments if the maximal group size is  $n = 2$ .**

Populations, in which group size do not exceed two, have an access to three fragmentation patterns: 1+1, 2+1 and 1+1+1. In near neutral environment constructed by  $\mathcal{S}_1 = \{1, 1 + \varepsilon\beta\}$  and  $\mathcal{S}_2 = \{1, 1 + \varepsilon\gamma\}$ , only three fragmentation modes can be evolutionarily optimal: pure 1+1, pure 2+1, and a mixed life cycle simultaneously utilizing fragmentation modes 1+1 and 2+1. The fragmentation mode 1+1+1 does not contribute to an evolutionary optimum under any near neutral dynamic environment. Since the same signs for both  $\beta$  and  $\gamma$  give the same optimal life cycle in both seasons, we focus on different signs:  $\beta > 0$  and  $\gamma < 0$ . **A** For  $|\beta| < |\gamma|$ , there is the range of  $t$  where population exhibits a bi-stability between two pure life cycles. **B** For  $|\beta| > |\gamma|$ , there is the range of  $t$  where only a mixed life cycle is evolutionarily optimal. On both panels, we use rescaled variable for x-axis,  $t' \equiv \frac{t + \gamma/\beta}{|\gamma/\beta|(\beta - \gamma)}$ .



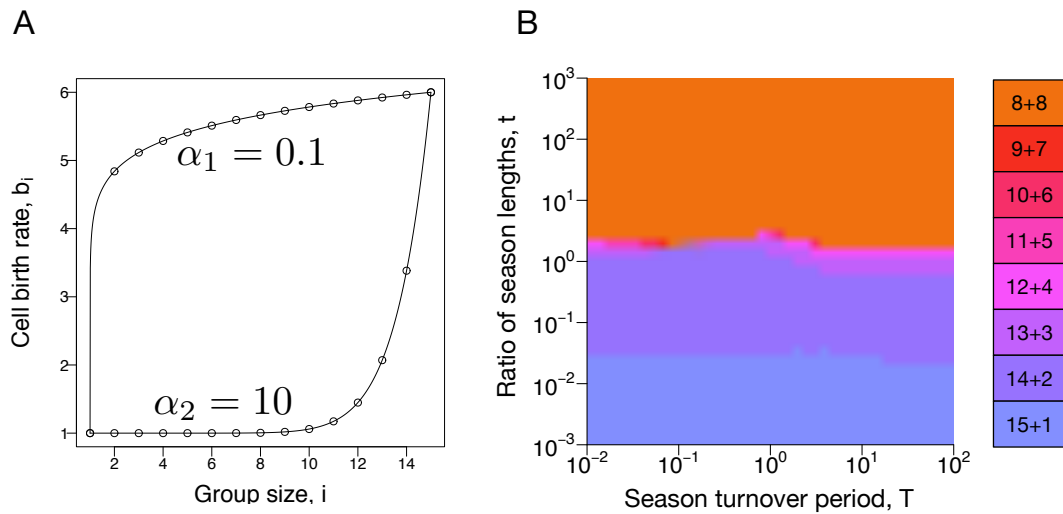
**Figure 6: In the near-neutral dynamic environments, the long and the short seasons regimes promote similar life cycles, while in far-from-neutral environments, these can be very different.**

**A.** In the near neutral environment, the behaviour in the short and long seasons regime are similar. For environments far from neutral, this pattern is violated. **B.** A life cycle present in the short seasons regime (2+2) can be absent from the long seasons regime. **C.** Example in which a fragmentation pattern absent at the short seasons regime (1+1) contribute to the optimal life cycle in the long seasons regime. **D.** Example of a multiple fragmentation pattern (1+1+1, light blue) contributing to the optimal life cycle in the long seasons regime. Black lines on the left hand side of maps indicate borders predicted by the short seasons approximation, coloured bars to the right from maps indicate the locally stable areas of pure life cycles in the long seasons approximation. Red ticks on the right hand side of maps indicate the saddle-node bifurcations  $t^*$  in the long seasons approximation. Exact parameters of all presented calculation are presented in Appendix L.



**Figure 7: The majority of dynamic environments promotes a unique evolutionary optimal pure life cycle in a form of binary fragmentation.**

We sampled  $1.64 \cdot 10^6$  different long seasons dynamic environments, see Appendix E. More than one locally optimal fragmentation mode was found only in 13% of them (second circle). In 1.9% of dynamic environments, a set of locally optimal life cycles contained a mixed life cycle (third circle). Majority (97%) of found mixed life cycles executed just two fragmentation patterns. Only 3% of evolutionarily optimal mixed life cycles had three patterns or more. Finally, the number of dynamic environments promoted the evolution of non-binary fragmentation (1+1+1, 2+1+1, or 1+1+1+1) was extremely tiny - 0.04% (last circle). So, the most common result of life cycle optimization was a single local optimum, which happened to be a pure life cycle with binary fragmentation (first circle).



**Figure 8: Monotonic dynamic environments promote binary fragmentation.**

**A.** To investigate the evolution of larger life cycles, we constructed a monotonic dynamic environment, where the birth rates  $b_i$  follow the function  $b_i = 1 + M \left( \frac{i-1}{n-1} \right)^\alpha$  with different exponents  $\alpha$  at different seasons. **B.** The majority of dynamic environments promoted unique pure life cycle with binary fragmentation. Even the environment is far from neutral, short and long seasons regimes exhibit similar evolutionary optimal life cycles. In our investigation, we used  $M = 5$  in both seasons,  $\alpha = 0.1$  for the first season, and  $\alpha = 10$  for the second season. Due to significant computational load, we performed only 20 independent optimizations at each pixel of the map.

## 301 4 Discussion

302 Environmental fluctuations are commonly observed in nature. Under changing conditions, a trait beneficial in  
303 one season may become detrimental in another. Thus, adaptation to a dynamic environment may lead to totally  
304 different phenotypes than those that evolve in a static environment. In this manuscript, we investigated the  
305 influence of a changing environment on the evolution of life cycles in the context of primitive multicellularity.  
306 In our model, unstructured groups grow and eventually reproduce by fragmentation. The growth competition  
307 between different reproduction modes determines which life cycle will spread in population.

308 Our present model uses the minimal set of processes necessary for the multicellular life cycle: birth, growth,  
309 and reproduction of cell colonies. A number of other factors might influence the evolution of life cycles as well:  
310 aggregation of cells [Garcia and De Monte, 2013, Amado et al., 2018], group death [Pichugin et al., 2017], cell  
311 death [Amado et al., 2018, Pichugin and Traulsen, 2018], interactions between different cell types [Garcia et al.,  
312 2014, 2015, Gao et al., 2019], the geometry of groups [Libby et al., 2014], and so forth. However, given the  
313 current state of the field, the understanding of evolutionary dynamics of life cycles even in the minimal setups  
314 is missing. Our study has shown that even with basic processes, the spectrum of evolutionary outcomes is rich  
315 and deserves a dedicated investigation.

316 Previous findings reveal that a static environment puts strong constraints on evolutionarily optimal life cycles  
317 [Pichugin et al., 2017]. There, only pure life cycles can evolve. Moreover, among these, only binary fragmenta-  
318 tion life cycles, featuring fragmentation into two groups, can become evolutionarily optimal. Interestingly, we  
319 found that evolution in dynamic environments can release both constraints.

320 Not only pure, but also mixed life cycles are able to evolve in our model. Being unable to perform well dur-  
321 ing both seasons, groups may employ a stochastic life cycle, where different groups randomly execute different  
322 fragmentation patterns. Thus, mixed life cycles are manifestation of between-clutch bet-hedging within the  
323 scope of our model. We found that in some dynamic environments, a mixed life cycle is the only evolutionarily  
324 optimal strategy, see Fig 5B. Our model also predicts that mixed life cycles may employ fragmentation patterns  
325 which would not contribute to the optimal life cycles under any of static seasonal components alone, see Fig. 6C  
326 and D. Yet, the most abundant scenario was the existence of only one locally optimal life cycle, which utilize  
327 a pure binary fragmentation mode, see Fig. 7. In other words, our model predicts that for simple multicellular-  
328 ity, between-clutch bet-hedging is possible to evolve, but is rarely an evolutionarily optimal strategy – even in  
329 changing environments.

330 Another form of bet-hedging observed among complex multicellular organisms is within-clutch bet-hedging,  
331 where offspring with diverse properties are produced in a single act of reproduction [Einum and Fleming, 2004].  
332 From the perspective of simple multicellular life cycles, an act of reproduction is the distribution of the parental  
333 biomass among offspring. Hence, it is impossible to distinguish between an organism releasing a propagule and  
334 an organism producing two offspring of different size. In both cases, the result of reproduction is a collection  
335 of organisms of different sizes. The traditional point of view on these events is to consider them as asymmetric  
336 division, or propagule formation, and not as a bet-hedging scenario. In other words, while we can distinguish,  
337 who is the parent and who is the offspring in the case of chicken laying eggs, it is hardly possible to do so  
338 for a broken cyanobacteria filament. Thus, the very idea of within-clutch bet-hedging implies more developed  
339 multicellularity than one considered in our study.

340 In the light of the evolution of simple multicellularity, pure life cycles deserve special attention. One of the  
341 conceptual barriers for the species transitioning from unicellular to multicellular existence is the necessity to  
342 develop a predictable life cycle. While, the formal grouping of cells into clusters can give some advantages to  
343 the population [Rainey and Travisano, 1998, Ratcliff et al., 2012], the real strength of multicellularity eventu-  
344 ally comes from beneficial interactions within the groups, such as cooperation or division of labour [Bell and



345 Koufopanou, 1991, Kirk, 1997, Flores and Herrero, 2010, Hammerschmidt et al., 2014]. The ability of cells  
346 to participate in such interactions is not guaranteed beforehand and therefore, must evolve. For this, a regular  
347 schedule of a group growth and reproduction provides a proper basis. In our model, this regularity is obtained  
348 by pure life cycles. Thus, it is an interesting question, to which extent the changing environment can violate the  
349 evolutionary stability of pure life cycles.

350 Despite the mixed life cycles observed in simulations, pure life cycles remain prevalent: less than 2% of  
351 our dynamic environments promoted mixed cycles, see Fig. 7. Among the four analysed limiting regimes,  
352 three favour pure life cycles. If one season occupies a large enough proportion of the seasonal cycle, the  
353 evolutionary optimal life cycle is the same as if the second season does not happen at all. In other words, short  
354 disruptions of environmental conditions are unable to affect the evolutionary optimality of life cycles. The actual  
355 threshold below which the short season is unable to influence the life cycle evolution is a complex function of the  
356 environments and the turnover rate. Nevertheless, this value can be either inferred from numerical simulations  
357 of our model, or estimated from approximations. The short seasons regime only promotes pure life cycles. In  
358 near neutral environments, the behaviour at any seasons turnover time becomes similar to the one expressed at  
359 short seasons – the transitional area between pure life cycles become narrow. Only the long seasons regime can  
360 explicitly promote mixed life cycles. However, these emerge only at intermediary values of season length.

361 These findings are surprising because in dynamic environments, intuitively, mixed life cycles, which com-  
362 bine the best of both worlds, are expected to be optimal. Countering that intuition, we found that pure life cycles  
363 emerge for a wide range of dynamic environments.

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