Hydrology-informed metapopulation modeling of liver fluke
transmission in the Lawa Lake complex of northeast Thailand
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## 20 Abstract:

A metapopulation model linking local hydrology with transmission of the liver fluke 21 Opisthorchis viverrini in a lake system in northeast Thailand was developed and 22 parameterized using infection data from 2008-2016. A rainfall-runoff model and other 23 24 hydrologic data were used to assess level of connectivity between villages and the influence of upstream communities on parasite distribution in the study area, while the 25 disease transmission model was expanded from a single-village model for O. viverrini 26 transmission. Connectivity between villages and hydrologic variables was assessed 27 28 monthly and showed strong seasonality trends. The metapopulation model improved upon the single-village model in its fit to historical data patterns for the six village 29 30 clusters with the introduction of the new time-variable parameters. Results suggest 31 there are three unique hydrologic-epidemiologic regimes within the Lawa Lake system in response to upstream watersheds and risk of overland flooding that contribute to risk 32 for O. viverrini infection. Similar approaches using a hydrologic submodel to inform a 33 mechanistic disease transmission model could be applied across many water-related 34 disease systems. 35

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#### 37 Author Summary:

While hydrology is intuitively understood to influence transmission dynamics of waterrelated diseases, limited research exists that explicitly links hydrologic and infectious disease data. In this work, we use an approach that leverages a rainfall-runoff model to better understand water movement into, out of, and around Lawa Lake in northeast Thailand and how that affects fate and transport of the multiple waterborne life stages of *Opisthorchis viverrini*. To model disease transmission, we represent six village clusters around the lake using known infection prevalence data of humans, cats and dogs, snails,
and fish to parameterize and fit a metapopulation model. The connectivity between
village clusters and external inputs of parasites are derived from the hydrology data and
the rainfall-runoff model. Results suggest three unique hydrologic regimes that also
reflect unique patterns in disease prevalence among the different hosts. Other waterrelated disease systems can use similar approaches to assess the impacts of water on
pathogen transmission dynamics.

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#### 52 Introduction

53 Opisthorchiasis, infection with liver flukes of the genus *Opisthorchis*, is a disease 54 whose transmission and distribution are largely determined by hydrology. The 55 parasites' egg and cercarial forms require sufficient water and transport to the next 56 intermediate host for the transmission cycle to be sustained. Eggs are excreted in the 57 feces of infected final hosts (humans, and reservoir cats, dogs, and other mammals to a 58 limited extent); if not safely treated or contained, these eggs reach the amphibious 59 habitat of intermediate host snails, which consume the eggs and enable maturation to 60 the cercarial stage. The cercariae are then released back into water, where they swim 61 and seek out the second intermediate host cyprinid fish. They encyst in the fish, which 62 if then consumed raw or undercooked by humans or certain other mammals can 63 migrate to the bile ducts and mature into adult worms. Water is a sustaining force for 64 this parasitic life cycle, and its movement permits viable infection at each successive host stage. 65

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The major liver fluke of interest in Thailand is *Opisthorchis viverrin*i. Given the

known disease burden in Thailand, O. viverrini has been a public health priority there, 68 69 where it is transmitted to humans via the consumption of popular local raw and fermented fish dishes such as koi pla and pla som. The highest prevalence of 70 opisthorchiasis and cholangiocarcinoma (CCA), the fatal bile duct cancer associated 71 72 with O. viverrini infection, are found in northeast Thailand in the region surrounding Khon Kaen [1]. Historical hotspots of opisthorchiasis and CCA were in the villages 73 74 around Lawa Lake. While much research has been conducted on the pathology of 75 opisthorchiasis and CCA, there is limited literature addressing the ecological and 76 hydrological aspects of parasite transmission in the environment [2,3].

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78 Lawa Lake is an approximately 4000-acre body of water that is highly vegetated 79 and subject to significant hydrologic changes caused by seasonal variation in northeast Thailand. A peak in liver fluke infections is seen with lag following the rainy season in 80 Thailand, as flooding facilitates the spread of fecal contamination and coincides with 81 the rapid increase in snail populations [1]. Since several weeks are required for the 82 parasite to mature through its life stages, high infection rates in fish are seen in the late 83 84 rainy season and summer (July-January). Low infection risks occur in the dry season 85 and summer (March-June). A primary industry on the lake is fishing, which contributes 86 to the environment mediating ongoing liver fluke transmission in the local area [4]. The hydrology of the Lawa Lake region is exceedingly complex and disturbed, as significant 87 88 changes have occurred in recent years due to construction of new irrigation canals and 89 ditches, new culverts and spillways that are opened and closed in the flooding season, 90 and fish ponds that have become increasingly popular as a source of food and revenue.

92 Metapopulation modeling is commonly used to better understand the 93 connectivity and influence of discrete human populations and environmental patches 94 on each other. Metapopulation modeling is especially powerful in understanding 95 pathogen transmission in complex, interrelated environments, where contiguous areas 96 and their inhabitants influence each other. Connectedness between these environmental 97 and host patches can occur in multiple ways, including migration of humans between 98 villages and movement of a waterborne pathogen from a section of river or lake 99 adjacent to one population to another section adjacent to a separate population. The 100 second example demonstrates hydrological influence on a disease transmission system, 101 as waterborne diseases as diverse as cholera and schistosomiasis rely on advective 102 transport to expose new susceptible individuals with pathogens excreted or shed by 103 infected individuals. Hydrologic patterns are time-varying and markedly local in 104 nature. Given this environmental complexity, hydrologic fate and transport of 105 pathogens are difficult to study. In addition, motile waterborne parasitic forms, such as 106 liver flukes and schistosome cercariae, have independent mobility behaviors, making 107 hydrologic flows not entirely representative of how these parasites are dispersed [5,6]. 108

109 Research connecting hydrology with waterborne disease transmission is an 110 emerging field with recent work on cholera and schistosomiasis [7,8,9]. Tracing the 111 spread of pathogens in the environment is challenging, and countervailing forces make 112 it difficult to ascribe an increase or decrease in human infection to trends in 113 meteorology and climate or consequent hydrology. Long-term studies of climatic 114 changes in rainfall patterns or the influence of dams allow more definitive statements 115 about impacts on disease transmission, though these are also complicated by shifts in host and vector habitat and in seasonal patterns that may disrupt or exacerbate host andvector growth and reproduction [10,11].

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119 In this work, a metapopulation disease transmission model is developed and 120 parameterized to assess hydrologic connectivity and O. viverrini parasite movement 121 between six village clusters around Lawa Lake in Khon Kaen Province, Thailand, and 122 how that is reflected by opisthorchiasis prevalence in hosts. Understanding liver fluke 123 transmission in this seasonal, hydrologically connected environment with modeling can 124 help define the scale of transmission processes and thereby optimize environmental 125 control and treatment to have maximum impact on reducing disease transmission in 126 this setting and others.

127

### 128 Methods

The model structure is an extension of the modeling framework presented in 129 130 [12]. The six village clusters studied are now connected in a metapopulation framework 131 to account for exchange of parasites and hosts between village clusters and their 132 associated environments. This enables the model to include the influence of population-133 level factors, spatial heterogeneity, and degrees of connectedness between patches. This 134 metapopulation model leverages information about hydrologic connectivity between 135 village and host clusters to understand the movement of the liver fluke parasite's 136 various forms in the environment as mediated by water. To consider local hydrologic 137 impacts on the liver fluke transmission cycle, five main factors are included: 1) egg 138 inputs into the system from upstream watersheds; 2) egg inputs into the system from 139 overland flooding; 3) snail and fish mobility due to hydrologic connectivity; 4) snail and 140 fish available habitat; and 5) hotspots where infectious snails come into contact with

susceptible fish. These factors are modeled monthly with seasonality to account forchanging patterns throughout the year.

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144 The metapopulation model connects the six villages or geographically proximate 145 village clusters shown in Fig 1 around Lawa Lake in Khon Kaen Province of northeast 146 Thailand. The villages or village clusters described here are CCK, Lawa, BT (cluster of 5 147 villages), NNK, KSR, and DPD (cluster of 2 villages), which were chosen and clustered 148 based on geographical location and how historical human infection survey data was 149 collected. The six clusters further sort into three groups based on impact or lack thereof 150 of flooding and upstream watersheds. The two upstream watersheds (Fig 2) flow into 151 Lawa Lake at locations adjacent to NNK and BT and contribute to egg input there. 152 Villages CCK and Lawa are in proximity to the Chi River and are most susceptible to 153 seasonal overland flooding. KSR and DPD are the villages most "downstream" and are 154 relatively isolated from major flooding or upstream drainage impacts. Villages within 155 the upstream watershed have not had as significant treatment and control programs as 156 the villages around Lawa Lake, and reports suggest that upstream villages still have 157 high infection prevalence values over 30%.

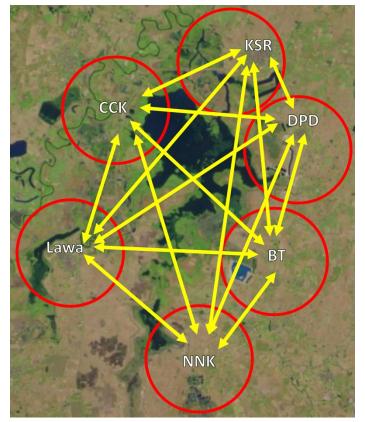


Fig 1: Connectivity between six village clusters around Lawa Lake (Map data: Landsat, USGS).

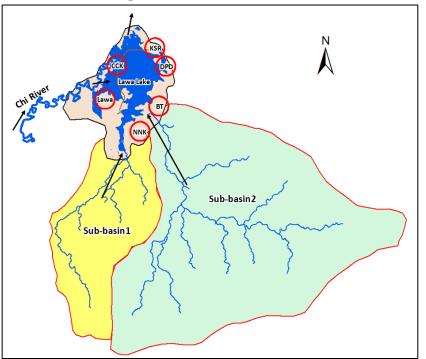


Fig 2: Two upstream sub-basins flow into Lawa Lake near NNK and BT, respectively (QGIS).

160	These hydrologic impacts seem to mirror trends in epidemiological patterns.
161	Previous research highlighted the north-to-south gradient of nitrogen and salinity in the
162	water that predicted higher snail abundance [13]. NNK, the southernmost village
163	cluster, had the highest infection prevalence at the time points when it was studied
164	(Table 1). KSR and DPD, the northernmost villages and the farthest from the Chi River,
165	had the lowest baseline prevalence of O. viverrini infection before the control program
166	started. In the single-village model, varying transmission parameters by village did not
167	fully capture the different patterns that occurred in the six village clusters when
168	connectivity was not included. Therefore, the role of hydrology and connectivity
169	between villages needs to be considered in the disease transmission model to better
170	account for the patterns observed.

Table 1: Infection prevalence (%) and mean intensity in positive individuals (EPG) for six village clusters around Lawa Lake. \* indicates less sensitive diagnostic method (Kato-Katz or Kato thick smear).

Village Cluster	2008	2010	2011	2012	2014	2015
ССК	54.9% (593)		33.0%*	44.3% (139)		
Lawa	67.1% (501)	63.1% (108)	19.0%*	(16)		8.7%*
BT	61.9% (346)	37.2% (131)		35.0% (136)	9.0%*	14.2%*
NNK	74.1% (499)			50.0% (61)		
KSR	16.4% (101)					
DPD	22.1% (112)	36.5% (82)				14.6%*

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To understand the effects of complex hydrologic factors on snail habitat, fish access to these habitats, and the pathways of parasite transmission, a hydrologic model of the Lawa Lake system was utilized to simulate flow patterns and changes in water levels over time. This model is a rainfall/runoff model common in hydrology that 176 considers the transport of water through a system originating from upstream in the 177 catchment basin or from precipitation. It uses the Soil & Water Assessment Tool 178 (SWAT) model to generate runoff for the PCSWMM hydraulic model to determine 179 hydraulic parameters of Lawa Lake such as flood depth and extent, flow velocity, and 180 travel time [14]. The inputs for the SWAT model include meteorological data (rainfall, 181 temperature, relative humidity, and windspeed) from the Thai Meteorological 182 Department, soil type, land use, and a digital elevation model (DEM) generated from 183 satellite imagery. For the hydraulic model, hydrologic structures and key parts of Lawa 184 Lake were surveyed for elevation at 1m x 1m resolution using drone data, and a 2D 185 model integrating runoff, the improved DEM, and meteorological data from 2008 to 186 2016 was developed using PCSWMM. Outputs include water level and flow vectors for 187 the time points modeled between 2008 and 2016. Calibration was conducted with 188 available precipitation and gauge data in the area from the Bureau of Water 189 Management and Hydrology, Royal Irrigation Department, in Thailand. Fig 3 190 demonstrates an example of the variation in flows predicted by the model over the 191 course of a calendar year encompassing the rainy and dry seasons; arrow direction and 192 thickness represent the relative change in flow vectors. While all of the villages are 193 adjacent to Lawa Lake, inflows and outflows as well as the relevant snail and fish 194 populations are not static, homogeneously distributed, or relevant to each village. The 195 rainy season is characterized by high and active flows that generally peak in October 196 with flooding from the Chi River varying from year to year. The dry season has 197 relatively little hydrologic activity establishing connectivity between village clusters. 198

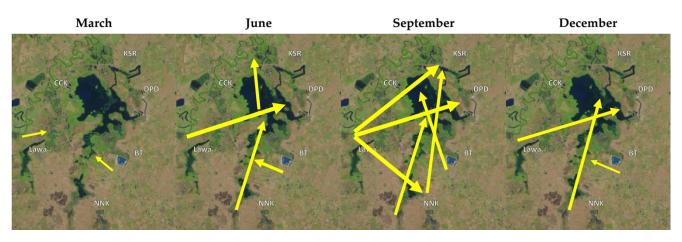


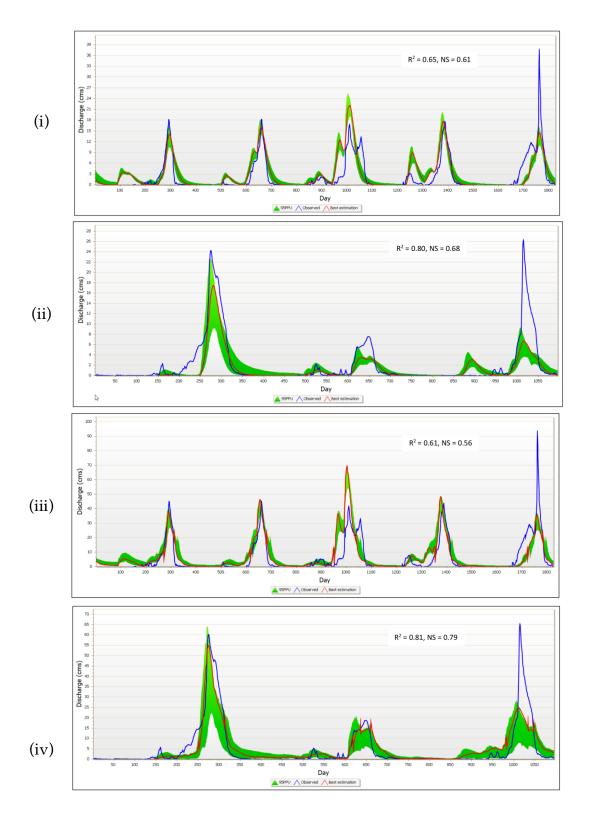
Fig 3: Hydrologic flows change dynamically throughout the year (semiquantitative interpretation of hydrology model results for an annual cycle) (Map data: Landsat, USGS).

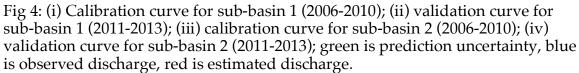
Because gauge data was not available for the sub-basins upstream of Lawa Lake, 200 201 a Soil and Water Assessment Tool - Calibration and Uncertainty Procedures (SWAT-202 CUP) model was used to calibrate and validate runoff into the Lawa Lake system from 203 these sources generated by the SWAT model. Inputs for the upstream sub-basins 204 included a 30m x 30m DEM, land use data, and soil type data from the Land 205 Development Department of Thailand. To map the hydrologic features of Lawa Lake 206 and finer scale structures, a drone was used to chart these areas in greater detail and 207 determine elevations where water was flowing into or out of Lawa Lake. Sub-basin 208 calibration and validation graphs are shown in Fig 4; 2006-2010 data was used for 209 calibration, and 2011-2013 data was used for validation. The R<sup>2</sup> values range from 0.61 to 210 0.81; both the calibration and validation models miss late peaks in their runs. In the case 211 of calibration, fitting the less extreme peaks may have disadvantaged the model from predicting the major discharge in 2010. In the case of validation, the 2011 peak 212 213 discharge was better modeled, but 2013 was missed by a large margin for reasons that 214 are not entirely clear but may be related to the different timing of precipitation-driven 215 flooding in 2013 compared with other years.

Based on these data and tools, the presence/absence of connections between 217 218 parcels of water associated with different village clusters were assessed. For example, 219 year-round one would not expect KSR to contribute to transmission in Lawa or NNK 220 but would expect the reverse to be true. In March, when less rainfall and flooding occur, 221 each village cluster is relatively isolated, with the exception of the relatively weak 222 influence of near neighbors (Lawa to CCK or BT to DPD). The origin of flows is also 223 subject to a differential dilution effect; contaminated waste from the upstream 224 watershed (Fig 2) would more strongly impact NNK than DPD or KSR and would 225 experience the effects sooner.

226

227 The mathematical model of disease transmission (which incorporates data from 228 the hydrologic model as variables and parameters) is an expansion of the single-village 229 model described in [12] to connect the six village clusters and uses as its state variables 230 the infection prevalence in humans, reservoir hosts (cats and dogs), snails, and fish. Base 231 equations are included below in Equations (1) - (5). Initial values were set from the 232 baseline surveys in 2008, and base transmission parameters are carried over from the 233 single-village model found using Markov chain Monte Carlo (MCMC) methods to fit the 234 model onto known infection prevalence data as described below. The infection 235 prevalence data includes surveys using two different methods: formalin-ether 236 concentration technique (FECT) and Kato-Katz (or Kato thick smear). In Thailand, FECT 237 has been regarded as the significantly more sensitive method to detect O. viverrini 238 infection based on available data and because the protocol is intended to make





microscopic examination easier [15,16]. Within the fish state variable is a fish demography model that captures the small window of time in the first few months of a fish's life when it is susceptible to cercarial infection (before its scales harden and it becomes more resistant). This model assumes a maximum fish lifespan of 4 years before either being caught or natural death.

244

245 Using this approach, the fish (*F*) differential equations (4) and (5) solve for the 246 number of infected fish in a cluster, whereas the human (*H*, equation 1), reservoir host 247 (*R*, equation 2), and snail (*S*, equation 3) differential equations solve for infection 248 prevalence as in the single-village model. The egg inputs come from the Chi River and 249 the two upstream watersheds and affect the snail state variable, contributing to the 250 force of infection in that linkage; these time-varying parameters are derived on a 251 monthly basis from the rainfall-runoff model. Other egg inputs from open defecation 252 and disposal of septic tank sludge are not modeled due to lack of information about where and when they occur. The egg inputs from overland flooding of the Chi River 253 254 were assumed to affect CCK and Lawa villages equally and were calculated by using 255 flow measurements from the river and multiplying by a scalar to relate the impact of 256 that water source with the upstream sub-basins. The first and second upstream sub-257 basins' outflow were modeled to contribute eggs to the systems in BT and NNK 258 exclusively and multiplied by their own scalars to translate those flows into 259 contributions to human and reservoir host infection.

261

$$\frac{dI_{H,i}}{dt} = \beta_{FH,i} \left( 1 - I_{H,i}(t) \right) \sum_{j=1}^{n} r_{FH_{i,j}}(t) * \gamma \sum_{k=1}^{3} I_{F_{j,k}}(t) - (\mu_H + \alpha_{PZQ}(t)) I_{H,i}(t)$$
(1)

$$\frac{dI_{R,i}}{dt} = \beta_{FR,i} \left( 1 - I_{R,i}(t) \right) \sum_{j=1}^{n} r_{FR_{i,j}}(t) * \gamma \sum_{k=1}^{3} I_{F_{j,k}}(t) - (\mu_R + \alpha_{PZQ,R}(t)) I_{R,i}(t)$$
(2)

$$\frac{dI_{S,i}}{dt} = \beta_{HS,i} \left( 1 - I_{S,i}(t) \right) \left( \sum_{j=1}^{n} r_{HS_{i,j}} I_{H,j}(t) N_{H,j}(t) + \sum_{j=1}^{n} r_{RS_{i,j}} I_{R,j}(t) N_{R,j}(t) \right) - \mu_{S} I_{S,i}(t)$$
(3)

$$\frac{dI_{F,i}^{0}}{dt} = \beta_{SF,i} \left( N_{F,i}^{0}(t) - I_{F,i}^{0}(t) \right) \lambda(t) \sum_{j=1}^{n} r_{SF_{i,j}} I_{S,j}(t) N_{S,j}(t) - \mu_{F} I_{F,i}^{0}(t)$$
(4)

$$\frac{dI_{F,i}^{1-3}}{dt} = -(\gamma + \mu_F)I_{F,i}^{1-3}(t)$$
(5)

Connectivity rates between patches were varied on a monthly basis between 262 263 2008 and 2016 based on the hydrology model results to account for snail and fish 264 mobility due to hydrologic connectivity. These connectivity parameters were derived 265 by assessing the fate and transport of parcels of water in a village cluster's area and what proportions reached other village clusters in the Lawa Lake system. These r266 parameters are unitless and vary from 0 to 1, describing the proportion of each village 267 268 cluster's force of infection for that host stage that affects each other village. Fig 5 shows 269 examples of how these connectivity parameters varied by village pairing, month, and 270 year (Figs 5-8 were produced in R using the ggplot2 package [17,18]).

The parameters carried over from the single-village model are listed in Table S1. The  $\beta$  values are transmission parameters and are village cluster-specific (Table S2). The r connectivity parameters are specific to each patch-to-patch relationship (Fig 1),  $\gamma$  is the fish catch rate describing the fraction of the total fish population caught at each time step,  $\lambda(t)$  is a gating function to control fish birth, death, and aging processes,  $\mu$  are mortality rates, and  $\alpha_{PZQ}$  are PZQ treatment events.



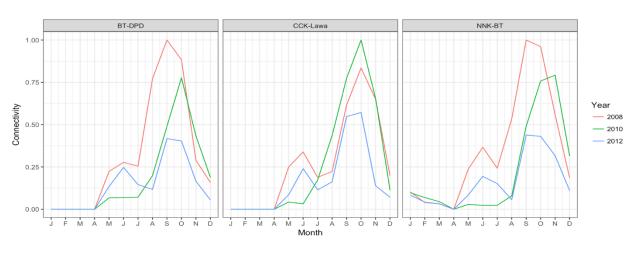


Fig 5: Village connectivity for BT-DPD, CCK-Lawa, and NNK-BT in 2008, 2012, and 2012

279 While a daily time scale could be preferable for assessing hydrologic impact, 280 historical data only captured month-to-month variability. Because human infection 281 survey data only measures prevalence and not incidence, this time scale is reasonable 282 for this study. From the hydrology model, the extent of water surface area at a suitable 283 depth (under 0.3m) for the contact events between juvenile cyprinid fish and the aquatic 284 snail intermediate hosts ("hotspots" for infection transfer) were used and to estimate N<sub>s</sub> 285 and N<sub>F</sub>. The transport time for a parcel of water between each village cluster was 286 employed to estimate the time scale of movement between the locations, but these 287 interactions happened on the order of days and not months and therefore the model's

288	time scale was not granular enough to introduce a time lag component. There was
289	insufficient data to track fish mobility between patches, so fishermen's movement data
290	from [19] was used to inform the exchange of fish in inter-village commerce as
291	contributing to human infection from other village clusters. Table 2 summarizes these
292	linkages and use of hydrology-related data in the disease transmission model. Fig 5
293	shows examples of the variability between village connectivity across months and years
294	for the human/reservoir-to-snail and snail-to-fish transmission processes (the
295	remaining connectivity parameter graphs are shown in the Appendix). The general
296	trends persist from year to year, but the timing of peaks differ and affect village
297	connections differently. The year 2008 produced stronger connectivity for BT-DPD and
298	NNK-BT, while 2010 had stronger connectivity for CCK-Lawa.

299

Hydrology-related transmission impact	Quantification method	Retained in model?
Egg inputs from overland flooding	Rainfall-runoff model output from Chi River summarized on monthly basis	Yes
Egg inputs from upstream water basins	Sub-basin model output summarized on monthly basis	Yes
Snail and fish mobility	Snail: Patch connectivity from rainfall- runoff model; Fish: fishermen catch data	Yes
Snail and fish available habitat	Snail: Rainfall-runoff model output and GIS analysis; Fish: N/A	Yes
Hotspots for snail-to- fish contact	Rainfall-runoff model output and GIS analysis	No

Table 2: Description of linkages between hydrology model and disease transmission model

300

#### 302 **Results**

303 Fig 6 shows the metapopulation model results for the six village clusters in the 304 base scenario with the metapopulation model establishing relative connectivity between 305 villages. Prevalence data points are included, distinguishing between the more sensitive 306 FECT surveys and the less sensitive Kato method surveys. The heterogeneity of 307 outcomes reflects the data: some villages saw reductions in infection prevalence to less 308 than 10% (KSR, DPD), yet a few villages continued to have predicted prevalence values 309 greater than 20% (BT and NNK). The steep drops in the graph were treatment events, 310 when a subgroup of villagers was tested for infection and given praziquantel if they 311 tested positive (the model assumes 100% drug efficacy). Model simulations were run for 312 eight years between 2008 and 2016.

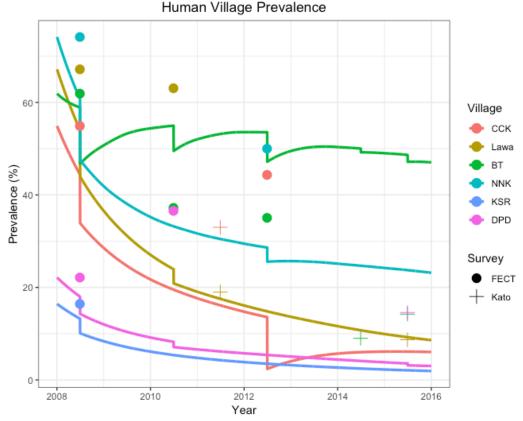


Fig 6: Metapopulation model run for human infection prevalence in villages around Lawa Lake

Figs 7 and 8 show infection prevalence values for intermediate snail and fish 313 314 hosts. In snails, the prevalence cycles seasonally with most patch populations decreasing to 0.1% from initial values of 0.2% (with the exception of Lawa where 315 316 prevalence approaches 0%). NNK has the highest final snail prevalence value at 0.18%, which is well within the range of what would be expected in this type of environment. 317 318 For fish prevalence, because the initial conditions are disparate and based on baseline 319 survey data, the model behavior is quite different. There is a seasonal aspect to their 320 dynamics though this is dampened for most villages except NNK, where it is readily 321 apparent. The end prevalence values range from 8-41%, with CCK, Lawa, BT, and DPD 322 having the lowest values and NNK having the highest.

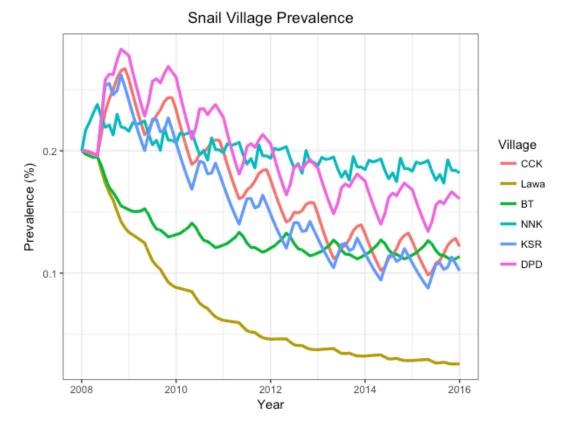


Fig 7: Snail prevalence values for metapopulation model

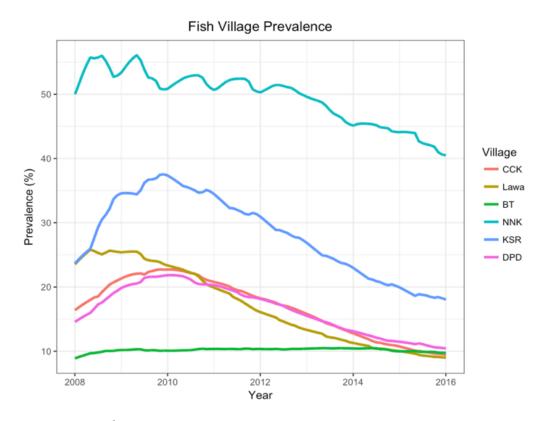


Fig 8: Fish prevalence values for metapopulation model

#### 324

## 325 Discussion

Compared with the single-village model, the metapopulation model no longer 326 327 over-predicts final infection prevalence for the villages in 2016. Instead, the model now 328 occasionally under-predicts prevalence for some data points, notably, CCK in 2012. This 329 specific issue can be partially accounted for by the number of treatment-naïve 330 individuals included in the 2012 infection surveys. Interpretation of the data leads to 331 some specualtion about the meaning of the surveys and their different diagnostic 332 methods. BT stands out as the modeled village with the least improvement (though it 333 overpredicts the 2012 and 2014 data). According to control program managers, BT villagers were the least participatory in the Lawa Project and may therefore have reaped 334 335 the least benefit from the control program. Given that this mathematical model is

primarily concerned with infectious indviduals actively shedding parasites into the
environment rather than asymptomatic cases, the Kato-based surveys from 2011, 2014,
and 2015 may actually capture the most infectious and relevant individuals in the
system and therefore be useful for thinking about infection prevalence patterns.
However, fitting all of the data points accurately would be nearly impossible since most
villages show non-monotonic patterns, and infection burden builds up over time.

342

343 Two major questions are, given the discrepancy between model fits and the data, 344 whether the data accurately reflects the reality of the disease transmission situation and 345 should the model by tuned with yet more parameters to get a close fit to the data. While 346 this dataset is relatively complete and informative by the standards of NTDs, it still 347 lacks enough time points, consistency in collection methods, and large enough sample 348 proportions of the populations to give a detailed picture of the *O. viverrini* infection 349 situation around Lawa Lake. The data (Table 1) show major swings across time points 350 and discrepancies between the survey methods. Therefore, while the FECT data was 351 used to fit via MCMC the transmission parameters in the single-village model, no 352 parameters were fit for the metapopulation model because of the increase in model 353 complexity and number of parameters (eighteen transmission parameters alone 354 compared with three per single-village model, plus seventy-two time-varying 355 connectivity parameters between the six village clusters). For this reason, the 356 connectivity parameters were derived exclusively from hydrology submodel results. 357 The metapopulation model is more believable than the single-village model in part 358 because of its inclusion of external hydrologic influences and connecitivity and in part 359 because the modeled behavior is more nuanced in the clustered patterns of village 360 prevalence change it exhibits. The lack of parameters fitted to this model structure and

361 the higher quality data that informed the hydrology model lend a realism to the 362 underlying mechanics of the metapopulation model that improves upon the relatively 363 straightforward transmission framework of the single-village model. The 364 metapopulation model is meant to capture patterns of transmission rather than exact 365 fits to the data. Nonetheless, the lack of information about differences in snail infection 366 and raw fish eating patterns between villages remains a weakness, as they could not be 367 incorporated into the model. Better data on these aspects of the transmission cycle 368 would further strengthen the modeling framework and bring it into greater alignment 369 with reality.

370

371 The patterns observed in these results support the sorting of the six village 372 clusters into three geographical clusters that exhibit different patterns and trends based 373 on human prevalence values. The first cluster, consisting of CCK and Lawa, is in close 374 proximity to the Chi River and is most susceptible to overland flooding experienced 375 during the rainy season. Its villages had high human prevalence values at the start of 376 the control program, which decreased sharply during the period of treatment and 377 control activity. These villages were the headquarters and major focal area of Lawa 378 Project activities, suggesting that they benefited the most from health education and 379 health volunteer engagement. The model is able to accurately account for the decrease 380 in prevalence without making any assumptions about reinfection. Additionally, these 381 villages are in close proximity to heavily fished waters in the lake, as supported by GPS 382 evidence and interviews described in [19].

383

384 The second cluster consists of BT and NNK, the villages to the south and385 southeast of the lake. These village also had high prevalence values at the start of the

386 treatment and control program but experienced much more modest decreases when 387 comparing data from later infection surveys. These villages are most impacted by 388 upstream watersheds draining into Lawa Lake (Fig 2.2), where some villages still have 389 over 50% *O. viverrini* infection prevalence (B. Sripa, unpublished data). Consequently, if 390 open defecation or unsafe disposal of human or reservoir waste is occurring in these 391 watersheds, the runoff will disperse parasite eggs into the canals, ponds, and sections of 392 the lake in close proximity to the second cluster's villages. These villages are also 393 adjacent to the highest concentration of fish ponds in the system and were not a focus of 394 major emphasis for Lawa Project activities.

395

396 The third cluster is KSR and DPD, which are located to the northeast of Lawa 397 Lake. At the start of the control program, these villages had significantly lower 398 prevalence values, which may be attributable to a lower degree of baseline 399 environmental contamination. These villages were significantly affected by neither the 400 Chi River nor the upstream watersheds, so they experienced fewer external inputs of 401 infectious individuals or waterborne forms of the parasite into their local systems. 402 These villages were not adjacent to high concentrations of fish ponds or fishing activity 403 in their local waters and received less emphasis by the Lawa Project compared to the 404 first cluster.

405

Because of the low prevalence of snail infection, the lack of field survey data, and the difficulty measuring snail prevalence precisely, strong claims cannot be made about the model results for the snail intermediate host. However, most field surveys indicate that snail prevalence in this region ranges between 0 and 0.2%, agreeing with the model results [13]. With current diagnostic methods, differentiating between the clusters

would require surveys of tens of thousands of snails at least. In recent years, the 411 412 highest snail prevalence values found are still less than 10% [20]. Much like other snail-413 borne diseases such as schistosomiasis though, only a few snails are required to 414 maintain transmission in an area because of the high number of cercariae they shed into 415 water bodies. Further understanding of where snails are most likely to be infected will 416 help with environmental measurement and control. Bottlenecks of water flow, such as 417 certain ditches and culverts, would concentrate fecal waste and parasite eggs and be 418 zones of likely contact with susceptible snail hosts. Eliminating or protecting these areas 419 could be an effective method of environmental control subject to proper coverage in the 420 environment of interest and patch/cluster connectivity.

421

422 Considering fish, infection prevalence is much higher and therefore it is easier to 423 discern differences between clusters. Because transmission is foodborne, we are 424 interested in the supply chain of food to consumers, which is not necessarily related to 425 proximity between where fish lived and where they are eaten. Fishermen, middlemen, 426 and merchants are all mobile and may choose to sell and distribute fish to other villages 427 to expand their market. In the model simulation, the relative ordering of low to high 428 fish prevalence values in fish hosts differ considerably from the results in humans, with 429 NNK and KSR having the highest infection prevalence. NNK's is driven largely by the 430 initial value, but KSR's is driven by dynamics, as its outcome is quite different from 431 Lawa's, which started with a similar prevalence level. Fish are infected by having 432 infectious snails releasing cercariae into areas with juvenile fish, so KSR is the cluster 433 with the greatest magnitude of this process taking place. Paying attention to fish prevalence results and how they interact with patterns of fish commerce can help 434 435 identify where to target health education interventions related to cooking.

436

437	The model and the data that inform the model have limitations due to their
438	fragmentary nature. Other model assumptions ignore the impact of different parasite
439	burden levels in hosts and the age structure of human populations. The hydrology
440	model was calibrated and validated against available data, leading to a plausible
441	interpretation of the observed hydrologic behavior of the system. However, it could not
442	account for very local effects that eluded its time and spatial scale and may have
443	relevance for the points of contact between stages of the life cycle. The infection survey
444	data may not be representative of the host populations because of sampling bias, but its
445	overall spatial and temporal patterns align with local understanding and experience.
446	

## 447 Conclusion

This work highlights a major driver of persistent O. viverrini infection in 448 449 northeast Thailand: a disturbed and dynamic hydrologic environment that mediates 450 parasite transmission between connected village clusters and environments. This was 451 accomplished by discussing and modeling five different means of hydrology 452 influencing parasite transmission and showing that its effects are significant and 453 complex, acting heterogeneously across the Lawa Lake system. While local infection 454 and contamination may be the main factor driving transmission at higher prevalence 455 levels, as prevalence declines and villages move toward elimination connectivity will play a bigger role in maintaining the transmission cycle and preventing local 456 457 elimination of the parasite.

458

459 The connectivity between water bodies and villages ensures that elimination of460 local infection is not possible without addressing upstream and adjacent environments.

If infectious fecal waste from human and reservoir hosts is continually flushed 461 462 downstream and the source is not treated, transmission will be restarted even if 463 previously disrupted. This argument supports an approach that first targets villages and populations near headwaters and then proceeds further downstream while still 464 465 accounting for human mobility and migration that could reintroduce infection into a 466 previously cleared environment. Flood pulses and overland flooding also risk moving 467 parasites into isolated and disconnected environmental patches on an annual basis, 468 which requires constant treatment and attention to infection status of individuals in 469 those patches. Snail and fish mobility remain little understood but have major relevance 470 for *O. viverrini's* life cycle, specifically how infection propagates in the environment. 471 Targeting the locations where snails and juvenile fish come into close proximity with 472 each other could be a promising environmental control technique but requires greater 473 knowledge about the conditions that enable this transmission process.

474

475 A linked disease transmission-hydrologic modeling approach was employed 476 here that uses hydrology model outputs as time-varying inputs in the disease 477 transmission model to account for seasonal effects of flooding and water movement 478 relevant to the intermediate hosts and waterborne forms of *O. viverrini*. Based on model 479 results, village clusters were grouped into three disease prevalence curve patterns based 480 on presence / absence of upstream and flooding impacts and history of control program 481 intensiveness. Considering these findings, we argue for the use of this modeling 482 approach and its results to inform environmental control of *O. viverrini* and for the need 483 for environmental surveillance. While the specifics of the hydrology, population 484 structure, and pathogen transmission cycle are local and specific in nature, this

- 485 approach can be replicated across a variety of disease systems that are impacted by
- 486 seasonally dynamic water movement.

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# 557 Supporting information

Parameter	Value	Units	Source	Symbol
Natural mortality of snails	1.37E-03	per day	Kruatrachue et al. 1982	$\mu_S$
Parasite dependent mortality of snails	1.37E-03	per day	Chanawong & Waikagul 1991	$\alpha_S$
Mortality of fish	6.85E-04	per day	Suvarnaraksha et al. 2011	$\mu_F$
Parasite dependent mortality of fish	0	per day	Assumption (unstudied)	$\alpha_F$
Natural mortality of humans	3.69E-05	per day	CIA 2015 (Factbook)	$\mu_H$
Human infection clearance by praziquantel	Variable	Episodic	Treatment data from clinics	$\alpha_{PZQ}$
Mortality of reservoir host	2.74E-04	per day	Local interview data	$\mu_R$
Reservoir infection clearance by praziquantel	Variable	Episodic	Cat/dog treatment data from veterinarians	$\alpha_{PZQ,R}$
Transmission parameters (fish-to- human, fish-to- reservoir, human-to- snail, snail-to-fish)	See Table 3.5.1	per day per infectious host/worm	Equilibrium conditions and MCMC	$eta_{FH}eta_{FR},\ eta_{HS},eta_{SF}$
Fish population	3000	fish	Estimate	$N_F$
Snail population	30000	snails	Estimate	N <sub>S</sub>
Human population	Variable	humans	Village censuses	N <sub>H</sub>
Cat and dog population	100	reservoir hosts	Estimate from village censuses	N <sub>R</sub>
Mean human worm count	Variable	worms	Infection survey data	W <sub>H</sub>
Mean reservoir worm count	Variable	worms	Infection survey data	W <sub>R</sub>

Table S1: Parameter values for single-village model

### 558

Equilibrium	ССК	Lawa	ВТ	NNK	KSR	DPD
Fish to Human	2.50E-10	7.24E-11	2.53E-10	1.81E-11	3.32E-07	2.68E-07
Human to Snail	8.90E-09	1.01E-08	4.92E-09	8.92E-09	6.56E-12	1.18E-11
Snail to Fish	7.34E-06	7.10E-06	3.40E-06	8.96E-06	1.67E-05	1.35E-05
МСМС	ССК	Lawa	ВТ	NNK	KSR	DPD
Fish to Human	2.95E-07	3.45E-07	7.26E-07	2.28E-07	3.30E-08	7.75E-08
Human to Snail	1.25E-08	2.53E-09	3.45E-09	2.26E-08	1.28E-08	1.48E-08
Snail to Fish	4.48E-06	7.03E-06	2.23E-06	2.28E-05	7.08E-06	3.89E-06

# Table S2: Beta transmission parameters for single village model