- 1 *Title*: Cyclical environments drive variation in life history strategies: a general theory of
- 2 cyclical phenology
- 3

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20 ABSTRACT

21 Cycles, such as seasons or tides, characterize many systems in nature. Overwhelming evidence 22 shows that climate change-driven alterations to environmental cycles—such as longer seasons— 23 are associated with phenological shifts around the world, suggesting a deep link between environmental cycles and life cycles. However, life history evolution in cyclical environments 24 25 remains surprisingly not well understood. Here I build a general demographic framework and ask 26 how life history strategies optimize fitness when the environment perturbs a structured 27 population cyclically, and how strategies should change as cyclicality changes. I show that cycle 28 periodicity alters optimality predictions of classic life history theory because repeated cycles have rippling selective consequences over time and generations. Notably, fitness landscapes that 29 relate environmental cyclicality and life history optimality vary dramatically depending on which 30 trade-offs govern a given species. The model tuned with known life history trade-offs in a marine 31 32 intertidal copepod T. californicus successfully predicted life history variation across natural 33 populations spanning a gradient of tidal periodicities. This framework shows how changes in 34 environmental cycles can drive variation—without complex assumptions of individual responses to cues such as temperature—thus expanding the range of life history diversity explained by 35 theory and providing a basis for adaptive phenology. 36

37 INTRODUCTION

Natural populations in all systems must survive environmental fluctuations. Biologists have long 38 known that a particularly common and powerful mode of fluctuations in nature is cyclical, such 39 40 as seasons. Species around the planet exhibit predictable and sensitive life history transitions that are tightly associated with seasonal cycles, also referred to as phenology. Environmental cycles 41 42 in fact occur beyond just the timescale of seasons, such as daily, tidal, lunar, flood, fire and decadal oscillations, and life histories of species are often associated with cycles at these 43 timescales as well [1-6]. Despite the ubiquity of cycles in nature, and clear empirical evidence of 44 45 the importance of cycles for life histories, we lack a general theory of how life history evolution is shaped by cycles. 46

Over the last few decades perturbations to environmental cycles due to climate change 47 have driven dramatic life history changes such as phenological timing in many species [7-15]. In 48 fact, phenological shifts are widely regarded as the most conspicuous and rapid consequence of 49 50 climate change across marine, freshwater, and terrestrial systems [14]. Notably, different species' 51 phenologies are shifting in different directions, creating phenological mismatches with profound consequences on ecosystem function and health [7,11,16–19]. Disparate case studies of shifts 52 53 that typically invoke individual-level responses to environmental cues such as temperature may be limited in their potential to explain general evolutionary forces due to system-specific 54 idiosyncrasies. On the trailing edge of rapidly accumulating empirical evidence of shifts, 55 56 questions regarding general mechanisms of life history evolution in cyclical environments have 57 emerged to the forefront of theoretical population biology, biodiversity, and climate change science [20–22]. 58

59	A first step in understanding the mechanics of life history evolution in cyclical						
60	environments may be to conceptualize cycles as sequential arrivals of harsh conditions whose						
61	periodicity is not reciprocally affected by local ecological dynamics. An example is the arrival of						
62	winter in seasonal systems. A typical consequence of such cyclical events for a population is						
63	heightened mortality as well as some perturbation to population structure (e.g. seedling mortality						
64	in plants [23]). This consequence not only reduces population size at a given time, but also						
65	impacts the long-term trajectory and fitness of the population [24,25]. It follows that, if periodic						
66	disturbance is an inherent feature of a habitat, fitness is determined by how well a resident						
67	population survives repeated demographic perturbations at regular intervals.						
68	Population ecologists have long been interested in demographic dynamics in variable						
69	environments, including cyclically variable environments [22,26–31]. Life history theorists, on						
70	the other hand, have classically focused on how time-invariant (i.e. constant) perturbations on						
71	age-, size- or stage-classes of populations, mediated by trade-offs between biological processes,						
72	shape life history strategies broadly [32-36]. For example, theory predicts that heightened						
73	juvenile mortality should induce the evolution of reduced reproductive effort. Such predictions						
74	have been widely tested empirically, and effects are often strong, rapid, and heritable [37-42]. So						
75	far, modern models of life history evolution that do incorporate time-variance in the environment						
76	have mainly focused on how optimality predictions are altered by stochasticity (i.e. randomly						
77	variable environments), which yield convenient analytical probabilistic conclusions [22,31,43-						
78	45]. What is not well understood is how life histories are generally shaped by non-random cycles,						
79	despite biological attention to fundamentally cyclical environments such as seasonal systems						
80	[22], and the fact that parametric changes to cycles such as season length are repeatedly						
81	associated with life history changes across systems.						

Here I explore the general relationship between periodicity of cycles and life history 82 83 optimization. By taking a demographic life history theory approach I address the ultimate causal mechanism behind phenological traits and their shifts, given that phenology is fundamentally a 84 85 study of how life cycle transitions are fit to environmental cycles and proximate causes of phenological expression, such as response to temperature cues, vary widely across systems [20]. 86 87 I hypothesize that rates of life cycle transitions relative to cycle periodicity, balanced by trade-offs between current and future investments considering impending arrivals of predictable 88 population perturbations, determine fitness. Thus I expect the period length of cycles to influence 89 90 the consequences of trade-offs and shape evolutionary predictions of life history rates. I analyze these predictions by calculating which life history strategy in a population confers maximum 91 fitness in a given periodic regime, and then studying how that optimal life history changes as 92 periodicity changes. I explore how various trade-off assumptions impact these optimality curves 93 to understand how different species in nature—whose life histories are in reality shaped by 94 95 different sets of trade-offs—may be differentially affected by the same change in periodic regime. Next I test my theoretical predictions in the copepod *Tigriopus californicus* (Copepoda: 96 Harpacticoida), a crustacean found in rock pools in the supralittoral (upper tidal) zone along the 97 98 North American Pacific coast. Populations are disturbed periodically by wave-wash at high tide, and experience population decline and heightened juvenile mortality periodically. Periodicity of 99 disturbance varies among populations depending on regional tidal patterns and pool height on the 100 101 shore. T. californicus provides an ideal system to study life history variation in cyclical systems 102 across populations due to its short generation time and short disturbance cycles, the rare opportunity to sample from homogenized whole populations, and ease of quick sampling and 103 104 trait measurements yielding large amounts of within- and across-population data. Across 19

natural populations of *T. californicus* in two regions of northern Washington I ask: do

- 106 disturbance cycle periodicity and known trade-offs together predict life history variation across
- 107 populations?
- 108
- 109 METHODS

110 Model construction.

- 111 I consider continuous-time dynamics of a stage-structured population and impose stage-specific
- 112 mortalities at given periodicities (full model description in Supporting Information, section 1).
- 113 First, I describe a population in two stages of broad relevance: juveniles and reproducing adults.
- 114 I express dynamics in abundances of the two stages as a system of ordinary differential equations
- 115 $dJ/dt = -(\mu+d)J + fA$ and $dA/dt = \mu J \gamma A$, that can be expressed as matrix **M**:

$$\mathbf{M} = \begin{bmatrix} -(\mu + d) & f \\ \mu & -\gamma \end{bmatrix}$$
(1)

where J is juveniles, A is adults, μ is the rate at which juveniles mature into reproducing adults, *d* is background mortality of juveniles, *f* is the reproductive rate of adults, and γ is background mortality of adults. Then, via eigendecomposition of **M**, I express the solution at time *t* as:

$$J(t) = \frac{A(0)v_{(2)1} - J(0)v_{(2)2}}{v_{(2)1}v_{(1)2} - v_{(2)2}v_{(1)1}}v_{(1)1}e^{\lambda_{1}t} + \frac{J(0)v_{(1)2} - A(0)v_{(1)1}}{v_{(2)1}v_{(1)2} - v_{(2)2}v_{(1)1}}v_{(2)1}e^{\lambda_{2}t}$$

$$A(t) = \frac{A(0)v_{(2)1} - J(0)v_{(2)2}}{v_{(2)1}v_{(1)2} - v_{(2)2}v_{(1)1}}v_{(1)2}e^{\lambda_{1}t} + \frac{J(0)v_{(1)2} - A(0)v_{(1)1}}{v_{(2)1}v_{(1)2} - v_{(2)2}v_{(1)1}}v_{(2)2}e^{\lambda_{2}t}$$

$$(2)$$

119 where $v_{(i)j}$ is the jth element of the ith eigenvector corresponding to eigenvalue λ_i of **M**. This 120 solution describes simple structured population dynamics in an undisturbed environment, but by 121 eigendecomposing the system I isolate *t* which will eventually allow me to study demographic 122 dynamics as a direct function of period length between disturbances. To make the solutions

explicit with respect to period T, I let t = T, and at time T multiply the structure by S_J and S_A

124 which correspond to juvenile- and adult-specific survival associated with disturbance. The

125 combined system can be expressed as the matrix \mathbf{P} (S10):

$$P = \begin{bmatrix} S_{J} \frac{\left[(v_{(1)2} e^{\lambda_{2}T} v_{(1)2} - v_{(1)1} e^{\lambda_{1}T} v_{(2)2}) \right]}{v_{(2)1} v_{(1)2} - v_{(2)2} v_{(1)1}} & S_{J} \frac{\left[(v_{(1)1} e^{\lambda_{1}T} v_{(2)1} - v_{(1)2} e^{\lambda_{2}T} v_{(1)1}) \right]}{v_{(2)1} v_{(1)2} - v_{(2)2} v_{(1)1}} \\ S_{A} \frac{\left[(v_{(2)2} e^{\lambda_{2}T} v_{(1)2} - v_{(1)2} e^{\lambda_{1}T} v_{(2)2}) \right]}{v_{(2)1} v_{(1)2} - v_{(2)2} v_{(1)1}} & S_{A} \frac{\left[(v_{(1)2} e^{\lambda_{1}T} v_{(2)1} - v_{(2)2} e^{\lambda_{2}T} v_{(1)1}) \right]}{v_{(2)1} v_{(1)2} - v_{(2)2} v_{(1)1}} \end{bmatrix}$$
(3)

Matrix-multiplying initial abundances by **P** would thus give stage structure after existing in a constant environment for time T and experiencing a disturbance event that incurs stage-specific mortalities. More interestingly, I use this framework to ask: what are the consequences of different combinations of life history traits on the fitness of a population given that it resides in disturbance regime T?

131

132 Fitness.

133 The dominant eigenvalue (λ) of a population transition matrix is a widely used measure of relative fitness because it represents how well the population will perform in the long run 134 compared to other hypothetical populations with different life history strategies [25,36]. This 135 metric, equivalent to 'r' in demography and life history theory, does not capture consequences of 136 137 short-term transient dynamics [46,47], but has been useful for drawing broad life history evolution predictions and conceptualizing relative fitness that match well with empirical 138 observations [24,25,36]. In stochastic environments fluctuations in instantaneous growth rates 139 140 may lead λ to give inaccurate evolutionary predictions. In systems that can be modelled by 141 periodic switching between environments, however, eigenvalues and eigenvectors of the matrix product of constituent matrices describing the different environmental states can be used for 142 demographic and life history analyses in exactly the same way as they are used in time-invariant 143

144	theory [25,48]. My matrix \mathbf{P} is equivalent to periodic models since the system switches between
145	an undisturbed phase and disturbance, and the switching periodicity and population matrix
146	elements do not fluctuate randomly (see Supporting Information, section 3, Fig. S1 for
147	simulation results). Thus, here I use the dominant eigenvalue of P (hereafter referred to as λ_P) as
148	the measure of relative fitness to compare the theoretical performance of life history strategies in
149	a periodically time-variant framework, and characterize general selective pressures on life
150	history strategies as a function of cycle periodicity.

151

152 Life history trade-offs.

Life history evolution is a matter of optimization because limited resources must be allocated 153 into various biological processes involving trade-offs [36,49]. The exact shapes of trade-off 154 155 functions in organisms are famously difficult to measure, let alone justify in model assumptions 156 [49,50]. Here I take a conservative approach and assume simple linear trade-offs to investigate 157 general patterns in optimality as a function of the environment without making more complex 158 physiological assumptions. To express a trade-off between any two traits in the construction of a fitness landscape, I computationally set the vector of the range of values of one trait in 159 160 decreasing order as the other increases, imposing a negative slope between the two traits. When two traits do not trade off, one of the traits remains at the mean of its range as the other varies 161 through its own range. I varied the combinatory inclusions of trade-offs among the four key 162 parameters to create model variants and investigate their relative fit to the data. 163

164

165 Fitness landscapes and optimal life history strategies.

All realizations of **P**—and thus the construction of fitness landscapes—must be constrained within the space of the interacting life history parameters, μ , *d*, *f*, and γ . In this presentation I constrained the space with known *T. californicus* life history ranges and trade-offs to demonstrate one example of the usage of this framework, but constraints can be set flexibly to represent any given species (see Supporting Information, section 2.5 for descriptions and citations for parameterization).

172 Using $\lambda_{\rm P}$ I construct fitness landscapes for μ and f simultaneously for each model. Here I focus on μ and f because they are life history traits for which I can collect large amounts of 173 174 paired data in *T. californicus*, but it should be noted that fitness landscapes can be created for any life history trait in the original system of differential equations. For each landscape, I scan across 175 the range of μ or f for a given value of T, while varying all other traits according to trade-off 176 177 relationships included in the given model. Therefore I construct a vertical gradient of relative λ_P per T. To construct a landscape, I calculate gradients of relative λ_P across the horizontal axis of T. 178 179 The optimal trait per T is the trait that maximizes λ_P per T. Finally, to get the curve of optimal 180 trait values across the axis of T I track values associated with maximum λ_P across T.

181

182 **Empirical investigation in** *Tigriopus californicus*.

Tigriopus californicus is a copepod found widely along the North American Pacific coast (see
Supporting Information, section 2.1 for detailed description of natural history). Dense
populations reside in rock pools above the intertidal zone at varying heights [51–53], which
accordingly experience tide cycle disturbance at varying periodicities. When tide levels
cyclically reach pool heights and waves wash through pools, *T. californicus* cling onto the rocky
benthos in order to prevent being flushed down to open water or to the lower intertidal zone [53].

189 If they are washed down, predators that do not occur in *T. californicus* pools feed on them 190 quickly, and re-colonization of *T. californicus* into the pools appear to be low [53,54]. Despite clinging, tidal disturbance was shown to always decrease population size, and in particular, incur 191 192 heightened juvenile mortality (Supporting Information, section 3, Fig. S3). 193 I sampled 19 isolated populations across two sites in northern Washington, USA (Neah 194 Bay, Friday Harbor) in order to capture a wide gradient of disturbance periodicities (see 195 Supporting Information, sections 2.2-2.4 for detailed description of data collection). I quantified the periodicity of tidal disturbance in each pool via timeseries analysis of pool temperature data 196 197 over 4 months at 5-minute intervals, taking abnormal drops in temperature as signals of wave flush (see Supporting Information, section 2.2). I siphoned entire isolated populations out of rock 198 pools, and subsampled individuals after homogenizing them, to get representative population 199 200 samples. I reared 30 mating pairs captured from each population in common garden settings. In these lines I measured rate of maturity (μ in the model) and rate of reproduction (f in the model) 201 202 (see Supporting Information, section 2.4 for detailed description of trait measurements). 203

204 Likelihood and model fitting.

I calculated the log-likelihoods of the optimality curves of the two focal life history traits μ and fproduced by each model variant given the variance and covariance of the μ and f data. Each model is a different trade-off model (Supporting Information, Fig. S2, Table S2). Every model has the same number of estimated parameters because they only differ in how the parameters trade off in the construction of the fitness landscapes, which is included computationally by aligning parameter range sequences in reverse order. Therefore model selection criteria that penalize number of parameters such as AIC were not used. Each model produces optimality

curves of μ and f (dominant eigenvalue of matrix **P**) given trade-off relationships, across values of disturbance period (T). I searched for the maximum log-likelihood of each model given μ and f data simultaneously within the space of $S_A \ge S_J$ and compared maximum log-likelihoods of the 13 model variants.

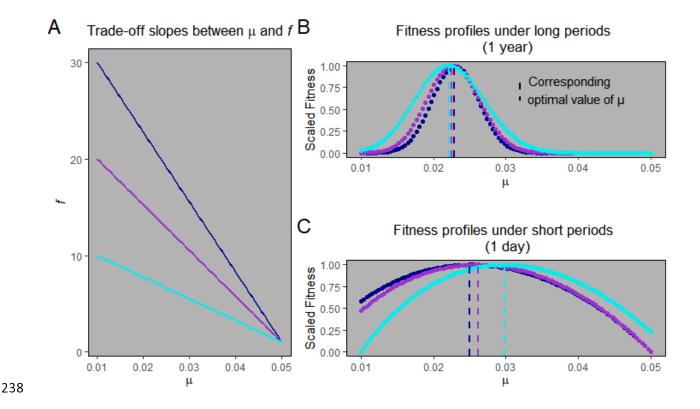
216

217 **RESULTS**

218 Cycle periodicity alters optimal life history predictions.

219 Classic life history theory balances costs and benefits of key biological investments such as 220 development, reproduction, and survival to predict fitness profiles of life history traits [36,55,56]. 221 Here I incorporated these classic balance considerations but imposed cyclical perturbations to 222 population structure and asked if the evolutionary consequences of the balances change as a 223 function of cycle periodicity. Using this framework, I analyzed the role of cost (slope of trade-off, Fig. 1A) on the fitness profile of a life history trait (maturation rate) in two scenarios: one in 224 which period length is long enough (e.g. to fit more than 10 generations in a period) that the 225 226 effect of discrete cycles on the evolution of life history rates should be small (Fig. 1B), and another in which period length is at a relevant timescale to generation time (Fig. 1C). The former 227 approaches classic formulations of optimal life history predictions based on trade-offs alone [55]. 228 229 The latter shows that external periodic perturbations significantly change optimality predictions. 230 In the latter scenario all trade-off cost assumptions predict higher optimal values of maturation 231 rate compared to the former. The spread of fitness profiles is also larger in the latter scenario, 232 which may suggest weaker selection or higher variability of maturation rate within a population. Lastly, and most interestingly, the relationship between trade-off cost and optimality is exactly 233 reversed between the two scenarios: the lowest cost case produces the lowest optimal maturation 234 rate under long periods but the highest optimum under short periods, and vice versa. These 235

- results show that cyclical perturbations significantly alter classic predictions of optimal life
- history that are solely based on trade-offs.



239 **Figure 1.** Three hypothetical cost functions between μ —rate at which juveniles mature into reproducing 240 adults—and f— adult fecundity—are analyzed while keeping linear trade-offs between μ and f with their 241 respective stage-specific background survival rates (d and γ) constant. Colors of cost functions (A) 242 correspond to colors of fitness profiles of μ in (B) and (C). Dashed lines in (B) and (C) show peaks of 243 fitness profiles which correspond to optimal values of μ . Periodicity of cyclical perturbation to population 244 structure is set to be much greater than generation time in (B), and at a relevant time scale (<generation 245 time) in (C). Under short periods (C), all cost functions produce higher optimal μ values, wider fitness 246 profiles, and an exactly reversed relationship between cost and optimality compared to long periods (B).

247

248 Periodicity and trade-offs interact to produce diverse life histories.

249 Optimal life history varies nonlinearly as a function of disturbance cycle period, even with

assumptions of simple linear trade-offs between traits (Fig. 2). For example, with stage-

251 structured mortality under disturbance and trade-offs between maturation and background adult 252 survival, predicted maturation and fecundity increase rapidly with increasing periodicity at low 253 periods, but change little at high periods (Supporting Information, section 3, Fig. S2C). 254 Shapes of optimality curves (optimal life history vs. period) can vary dramatically depending on which trade-offs are included. For example, maturation rate and fecundity are 255 256 expected to evolve in opposite ways as periodicity changes when maturation rate trades off with 257 background juvenile survival and fecundity (Supporting Information, section 3, Fig. S2G) 258 compared to when maturation rate trades off with fecundity and background adult survival 259 (Supporting Information, section 3, Fig. S2I). Similarly, maturation rate and fecundity increase 260 together under maturation-background adult survival trade-off (Supporting Information, section 3, Fig. S2C), but decline together under background juvenile-adult survival trade-off (Supporting 261 262 Information, section 3, Fig. S2E) or fecundity-background adult survival trade-off (Supporting Information, section 3, Fig. S2F), and are insensitive to environmental cycle change under 263 264 maturation-background juvenile survival trade-off (Supporting Information, section 3, Fig. S2A). 265 In the next section I show that the model that includes known trade-offs in *T. californicus* has the 266 highest likelihood given T. californicus-specific life history data; but it is important to note that no one model is necessarily better than another in a general sense because different species in 267 nature will have different levels of complexity and rank order of trade-offs between life history 268 traits [55,49,57]. 269

Overall, these results support the hypothesis that, beyond the long-accepted
understanding that the mean and variance of population structure perturbations shape life history
variation [24,34,35,39], periodicity of such perturbations plays an important role in life history
variation.

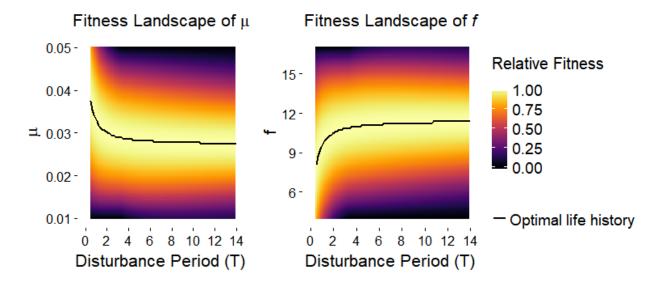




Figure 2. Example fitness landscapes of two focal life history traits, μ (rate of maturity) and f

276 (reproductive rate) which assumes lower juvenile survival with each disturbance event ($S_A = 0.9, S_J =$

277 0.6), and trade-offs between μ and f, between μ and d, and between f and γ . Heat shows normalized

fitness of a life history strategy compared to all other strategies in a disturbance regime (T). Curves track

the optimal (maximum fitness) life history trait across T.

280

281 *Tigriopus* trade-offs predict life history variation across a periodicity gradient.

Temperature time series analyses confirmed that there is a broad range of disturbance cycle 282 periodicities across T. californicus pools across the two regions (Supporting Information, section 283 284 2.1; section 3, Fig. S3A, B; section 4, Table S1). These sampled pools provided a gradient of periodic regimes against which I tested optimal life history predictions. Daily temperature 285 regimes, which may contribute to life history differences [58,59], were not significantly different 286 287 among pools of varying periodicity regimes across the two regions (Supporting Information, section 3, Fig. S4). Disturbance always caused higher juvenile mortality than adult mortality in 288 289 subsampled disturbance events, with mean juvenile mortality of 41% and mean adult mortality of 290 6% (Supporting Information, section 3, Fig. S3C).

291

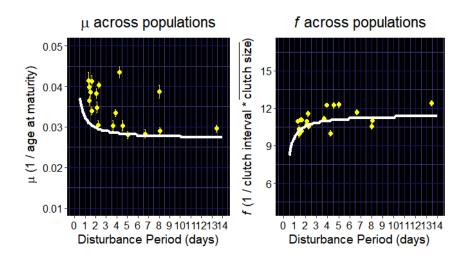


Figure 3. Mean (±se) values of the two focal life history traits µ and f across 19 T. californicus
populations, against mean disturbance period determined by timeseries analysis of wave disturbance
signals in each pool. Curves are optimal life history functions across periodicity (T) fit simultaneously to
µ and f.

297

292

298 Life history traits shift as disturbance period changes across T. californicus populations (Fig. 3), following model predictions (Fig. 2). The best model (likelihood maximizing when μ 299 and f are fit simultaneously) was the one that assumed trade-offs between maturation rate and 300 301 fecundity, between maturation and juvenile survival, and between fecundity and adult survival (Table 1), consistent with known trade-offs in *T. californicus* (Supporting Information, section 302 2.1). While the fit of the model to the data does not elucidate the selective strength of cyclical 303 disturbance on life histories, the statistical support demonstrates that the hypothesis based on 304 305 optimization in cyclical environments is consistent with the data and cannot at this time be rejected. Finally, model variants with double or tertiary trade-off assumptions generally fit better 306 than ones with only single trade-offs (Supporting Information, section 3, Fig. S2; section 4, 307 308 Table S2). These comparisons among model variants suggest that multidimensional trade-off 309 relationships—which are typically avoided in empirical measurements or model assumptions of 310 life history evolution [49,50], but gaining some attention [60,61]—may actually be important in

- 311 predicting life history optimization in cyclical environments because trade-off consequences
- 312 change as a function of cycle period.
- 313
- Table 1. Rank-ordered top five model variants with their life history trade-off assumptions

Model (Trade-off)	Log-likelihood		
$d \leftrightarrow \mu \leftrightarrow f \leftrightarrow \gamma$	441.771		
$\mu \leftrightarrow f \leftrightarrow \gamma$	441.709		
$d \leftrightarrow \mu \leftrightarrow f$	441.527		
$d \leftrightarrow f \leftrightarrow \gamma$	441.120		
$\mu \leftrightarrow f$	431.774		

315

316 Arrows denote trade-offs between life history traits. Model variants have different trade-off inclusions,

317 but have the same number of estimated parameters because linear trade-offs were included

318 *computationally by setting parameter ranges of those traits in opposing (increasing vs. decreasing) order.*

319 *Therefore model comparison criteria that penalize number of parameters were not needed. Likelihoods*

are based on simultaneous fitting of μ and f. See Supporting Information, section 4, Table S2 for the full

321 *list of models*.

322

323 **DISCUSSION**

Ecologists have long assumed that environmental cycles are important for life cycle-324 325 related traits. But growing knowledge of phenological shifts have generated confusion regarding 326 how environmental cycles shape life history strategies and thus transition rates of life cycle 327 phases. Various species in the same community undergoing the same change in seasonal cycles often exhibit phenological shifts in opposite directions, suggesting an interaction between 328 external cycles and internal mechanisms of life history optimization. My framework addresses 329 330 how the consequence of trade-offs, a fundamental driver of life history evolution, is influenced by cycles. Results show that environmental cycles can significantly alter traditional predictions 331 of life history evolution that are based on assumptions of population structure perturbations and 332 trade-offs alone. A version of the model tuned with known T. californicus trade-offs successfully 333

predicted the shape of life history variation across natural periodicity regimes, demonstrating thepower of this mechanistic framework.

Varying trends in phenological shifts among species in fact offer current, global 336 337 examples of the creation of life history variation. They provide a hidden opportunity to address a fundamental puzzle in ecology and evolution: why are life histories diverse in nature? Here I 338 339 show that the interaction between environmental cycles and life history trade-offs is a simple 340 mechanism that can account for large variations in life histories. First, due to the non-linear 341 relationships between cycle period and optimal traits, the same magnitude of period change can 342 induce different magnitudes of life history evolution between two populations of a species that are in different cyclical regimes (Fig. 2). Second, different trade-offs produce varying shapes of 343 optimality curves (Supporting Information, section 3, Fig. S2), and thus the same change in 344 345 period can induce an increase, decrease, or no change in a life history trait for different species in the same system depending on what trade-offs are biologically important for those species. 346 347 Environmental cycle periodicity is diverse across systems (such as growing season lengths across 348 a latitudinal gradient), and trade-off architectures among populations and species vary widely 349 due to physiological constraints, environmental conditions, and reaction norms [49]. Combined, cycles and trade-offs can produce a wide array of predicted life history strategies. Testing this 350 351 mechanism in species that are controlled by different trade-offs, either across populations in 352 different cyclical regimes or within a single population through time in a habitat undergoing a 353 change in cycle periodicity—for instance due to climate change—will provide fruitful avenues 354 for further exploring this perspective.

355

356 Stochasticity, ESS models, and gene flow.

357 Cycles in nature, of course, are not perfectly periodic. The present study focuses on the 358 consideration of period, or interval length between autocorrelated events. The mechanistic 359 influence of fundamentally cyclical environments on life history evolution is noticeably 360 understudied compared to probabilistic expectations in stochastic environments [22], even though regular cycles on various time scales are common in nature. Periodic models can be used 361 362 to address a real aspect of nature that is difficult or impossible to address explicitly with stochastic models: cyclicality. Here, I take advantage of the fact that periodic models allow the 363 use of matrix properties such as the dominant eigenvalue to infer relative fitness within a 364 365 fluctuating system [25,48] and analyze conditions for optimization. By doing so I uncover a 366 novel mechanistic relationship between cyclicality and life history evolution. However, cyclicality and stochasticity are both important aspects of nature. For instance, stochastic 367 368 fluctuations in instantaneous population growth rate can significantly modify evolutionary trajectories predicted by time-invariant or periodic theoretical assumptions [46,47,62]. Studying 369 370 the relative influences of periodicity and stochasticity on optimal strategy, and on how quickly a 371 population evolves to its predicted optimal strategy, are the obvious next steps that will add more richness to the perspective offered here. 372

Optimality curves in my model framework are attractors that represent evolutionary stable strategies (ESS) because I take the long-run growth rate of populations (dominant eigenvalue of **P**) as the measure of fitness as is commonly done in demography and life history theory. ESS models are useful for the purpose of predicting general directions of selection over a long term.ESS models take a non-genetic perspective on broad selective forces, although a genetical justification for optimization of a quantitative trait is given by the fact that a mutation can invade the population if it confers a higher *r* on its carriers [24]. Optimization models and

380 quantitative genetics models are approximately equal for constrained multivariate systems [63]. 381 In *T. californicus*, selection on optimal life histories may be obscured if high gene flow among nearby populations exists due to wave transport. However, colonization rates and genetic 382 383 exchange have been repeatedly observed to be low in this system [64–66], and demographic 384 dynamics given high mortality rates caused by tidal disturbance likely out-scale population 385 genetic dynamics on the time scale of tide cycles. In this study I deliberately chose populations 386 that were deemed to be well isolated given field observations. But the level of gene flow may vary depending on locality due to habitat characteristics, and may contribute to some of the 387 388 variance within populations and deviations of population means from ESS predictions. Nonetheless, my model fitting results suggest that ESS assumptions predict T. californicus life 389 histories reasonably well given a population's periodic regime. 390 391

001

392 Trade-off functions.

Trade-offs between traits can be nonlinear, and multidimensional architectures of trade-offs can 393 394 be extremely difficult to measure [49,57,50]. Here I have taken the conservative approach of assuming linear trade-offs among modeled life history variables to focus on the demonstration 395 that consequent optimality curves across periodicity are nonlinear, and that a diverse set of 396 optimality curves can be produced with different trade-offs. The simple linear assumption still 397 performs well, at least with T. californicus life history data from my sample populations. 398 399 However to test this framework further in different species, different functions can and should be 400 used if the relationship between two traits is known to be nonlinear.

401

402 Links to evolution of seasonal phenologies.

In seasonal environments cyclical arrival of harsh meteorological conditions (e.g. winter) can
incur large demographic perturbations and thus strongly influence population dynamics [67,68].
Here I show that if periodic arrivals of disturbance incur significant demographic perturbations,
the costs of having non-optimal life history strategies ripple across generations and cycle periods;
thus, cyclical perturbations play an important role in driving the evolution of life history
transition rates.

Period is not the only parameter of cycles, however. Particularly for seasons, cycle 409 amplitude may also shape phenologies in important ways, and is shifting with climate change in 410 411 many natural systems (e.g. seasonal CO_2 cycle amplitude [69,70]). Amplitude of seasonal cycles 412 may play two roles for evolution. First, amplitude is associated with intensity of disturbance, which can be explored with survivorship functions in my theoretical framework. If the pattern of 413 414 stage-specific mortality associated with cyclical disturbance is clear, such as in T. californicus 415 and many seasonal species, then heightened intensity of cyclical disturbance will likely increase 416 strength of selection. Second, amplitude reflects the rate of environmental change within cycle 417 phases. Rate of change may be important for cue-detection and plastic responses. For example 418 many plants in seasonal environments are well known for tracking growing degree-days as a way 419 of taking cues on the passing of the seasons [71]. In my theoretical framework, cyclical 420 disturbances arrive without warning and simply incur repeated penalties on individuals and cohorts that had non-optimal life history strategies for the given regime. In reality there may be a 421 422 number of continuously changing environmental variables in T. californicus pools such as 423 salinity, and I cannot exclude the possibility that, like plants, birds, or many aquatic invertebrates, 424 T. californicus possess biological mechanisms to use cues from continuously changing 425 parameters to plastically alter their phenotypes. Nonetheless, it is notable that a simple

426 framework that does not include the possibility of plastic responses predicts variation in life 427 histories as a function of periodicity, in what are likely continuously changing habitats. Future phenological work should consider the relative roles of demographic influences such as those 428 429 discussed here and plastic response to cues that can be tracked along continuous cycles. 430 When considering phenological evolution in cyclical environments, the relative scaling of 431 life cycles and environmental cycles becomes important. For instance, a perennial species must 432 endure multiple seasonal cycle periods per generation. An annual species' generation on the other hand fits within a single cycle period. In both cases, consequences of fitness-related 433 434 phenotypes in one generation carry over to subsequent generations via intergenerational tradeoffs in life histories [36], but the trajectory of evolution may differ between the two because of 435 the number of cycle periods a generation experiences. Further, the model framework presented 436 here assumes overlapping generations but many annual organisms have non-overlapping 437 generations and synchronous phenologies. The evolutionary consequences of non-overlapping 438 439 generations and synchronization in a population in cyclical environments should be explored 440 further.

Phenology is the study of how life cycle schedules are fit to environmental cycles. A 441 442 phenological trait is a manifestation of the aggregate life history strategy of a species [16], and expression timings of traits are ultimately controlled by transition rates between life history 443 stages [20]. Phenological studies typically measure one representative phenotype such as 444 445 flowering time in association with proximate drivers such as temperature or precipitation. But 446 phenotypes covary and therefore one must consider trade-offs and competing selective forces with a whole-life perspective in order to understand the evolution of cyclical phenological traits. 447 448 Here I placed such connections in the general context of environmental cycles, of which the

449	annual seasonal	cycle is one example	e. and tested	mechanistic	predictions on	the relatively	v short
772	annual scasonal	cycle is one chample	e, and tested	moonamotic	predictions on	the relatively	y 01

- timescale of tide cycles which yielded large amounts of data across many cycle periods and
- 451 generations quickly. This framework provides a basis for analyzing, comparing, and predicting
- 452 adaptive phenological shifts in changing seasonal environments.
- 453

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