

# Understanding and predicting extreme wildfires in the contiguous United States

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## Abstract

Wildfires are becoming more frequent in parts of the globe, but predicting where and when extreme events occur remains difficult. To predict wildfire extremes across the contiguous United States, we integrate a 30 year wildfire record with meteorological and housing data in spatiotemporal Bayesian models with spatially varying nonlinear effects. We compared models with different distributions for the number and sizes of large fires. A zero-inflated negative binomial model for fire counts and a lognormal model for burn areas provided the best performance. This model attains 99% interval coverage for the number of fires and 92% coverage for fire sizes over a five-year withheld data set. Dryness and air temperature strongly regulate wildfire risk, with precipitation and housing density playing weaker roles. Statistically, these drivers affect the chance of an extreme wildfire in two ways: by altering fire size distributions, and by altering fire frequency, which influences sampling from the tails of fire size distributions. We conclude that recent extremes that have burned hundreds of thousands of acres should not be surprising, and that the contiguous United States may be on the verge of experiencing even larger (million acre) wildfire extremes.

**Key words:** fire; wildfire; Bayesian; spatiotemporal; extremes; climate.

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## 1 Introduction

2 Wildfire frequency and burned area has increased over the past couple decades in the  
3 United States (Dennison et al. 2014; Westerling 2016), and elsewhere (Krawchuk et al.  
4 2009; Pechony and Shindell 2010). In addition to the ecological and smoke impacts associ-  
5 ated with increased burned area, there has been an increasing interest in extreme wildfires  
6 (Williams 2013) given their impact on human lives and infrastructure (Kochi et al. 2010;  
7 Diaz 2012). While case studies of particular extremes provide insight into what caused  
8 past events (Peterson et al. 2015), predictions of future extremes at a national level could  
9 inform disaster related resource allocation. Here, we consider an extreme wildfire to be  
10 a fire with the largest burned area over a bounded spatiotemporal domain, e.g., within a  
11 spatial region and a temporal interval.

12 Factors driving wildfire extremes vary in space and time (Barbero et al. 2014), but it is  
13 unclear how best to account for this in a predictive model. Previous efforts have used year  
14 or region-specific models, aggregating over space or time (Bermudez et al. 2009), tempo-  
15 rally or spatially explicit models (Mendes et al. 2010), and spatial models with year as a  
16 covariate (Díaz-Avalos, Juan, and Serra-Saurina 2016). Recently, rich spatiotemporal mod-  
17 els have been described with linear, spatially constant covariate effects (Serra, Saez, Juan,  
18 et al. 2014; Serra, Saez, Mateu, et al. 2014). However, linear, spatially constant effects  
19 are suboptimal over large spatial domains with nonlinear drivers (Fosberg 1978, Goodrick  
20 (2002), Preisler et al. (2004); Preisler and Westerling 2007; Balshi et al. 2009; Krawchuk et  
21 al. 2009; Pechony and Shindell 2009; Vilar et al. 2010; Woolford et al. 2011; Woolford et  
22 al. 2014). For example, global wildfire probability shows a hump-shaped relationship with  
23 temperature and moisture (Moritz et al. 2012). Interactions among drivers also impose  
24 nonlinearity, e.g., in hot and dry climates fuel sparsity limits fire spread (McLaughlin and  
25 Bowers 1982), but in cold and wet climates, fires are energy limited (Krawchuk and Moritz  
26 2011).

27 Prediction is also complicated by uncertainty in which distribution(s) to use to assign  
28 probabilities to extreme events. The generalized Pareto distribution (GPD) has frequently

29 been used (Bermudez et al. 2009; Jiang and Zhuang 2011), but the GPD requires a thresh-  
30 old to delineate extreme events (Davison and Smith 1990, Coles (2014)). The utility and  
31 validity of a threshold for extremes in a heterogeneous region is debatable (Tedim et al.  
32 2018). Recently proposed metastatistical extreme value (MEV) approaches do not re-  
33 quire such a threshold (Marani and Ignaccolo 2015; Zorzetto, Botter, and Marani 2016).  
34 In the MEV framework, the occurrence and size of future events, and the parameters of  
35 their distributions are treated as random variables which together imply a distribution  
36 for extremes. This approach has roots in compound distributions (Dubey 1970; Wiitala  
37 1999), doubly stochastic processes (Cox and Isham 1980), superstatistics (Beck and Co-  
38 hen 2003), and the Bayesian posterior predictive distribution (Gelman et al. 2013). The  
39 link to Bayesian inference is particularly useful, as it provides an easy way to propagate  
40 uncertainty forward to to predictions of extremes (Coles, Pericchi, and Sisson 2003).

41 Here, we extend the MEV perspective to account for non-linear, spatially varying wildfire  
42 dynamics with the goal of predicting and explaining extreme wildfire events across the  
43 contiguous United States. We aim to predict occurrence (where and when), and magni-  
44 tude (burn area) of large wildfires at a monthly time scale. Such predictions can be used  
45 to prioritize reactive fire suppression resources or inform proactive wildfire risk mitigation.

## 46 **Methods**

### 47 **Data description**

48 We acquired wildfire event data for the contiguous United States from the Monitoring  
49 Trends in Burn Severity (MTBS, [www.mtbs.gov](http://www.mtbs.gov)) program (Eidenshink et al. 2007), which  
50 includes spatiotemporal information on the occurrence of wildfires in the United States  
51 from 1984 to 2015. The MTBS data contain fires greater than 1000 acres in the western  
52 U.S. and greater than 500 acres in the eastern U.S. For consistency across the U.S., we  
53 discarded all records in the MTBS data less than 1000 acres, retaining 10,315 fire events  
54 (Figure 1A). Each event in the MTBS data has a discovery date, spatial point location,

55 and final size.

56 To explain fire size and occurrence, we used a combination of meteorological variables in-  
57 cluding humidity, air temperature, precipitation, and wind speed. These variables were  
58 selected on the basis of previous work, and also with an aim to capture directly the effects  
59 of easily interpretable meteorological quantities. Meteorological layers were acquired from  
60 the gridMET data (Abatzoglou 2013) that blends monthly high-spatial resolution (~4-km)  
61 climate data from the Parameter-elevation Relationships on Independent Slopes Model  
62 (Daly et al. 2008) with high-temporal resolution (hourly) data from the National Land  
63 Data Assimilation System (NLDAS2) using climatologically aided interpolation. The resul-  
64 tant products are a suite of surface meteorological variables summarized at the daily time  
65 step and at a 4-km pixel resolution. Daily total precipitation, minimum relative humidity,  
66 mean wind speed, and maximum air temperature were averaged at a monthly time step for  
67 each of 84 Environmental Protection Agency level 3 (L3) ecoregions for each month from  
68 1984 to 2015 (Omernik 1987; Omernik and Griffith 2014). We also computed cumulative  
69 monthly precipitation over the previous 12 months for each ecoregion-month combina-  
70 tion. We chose to segment the U.S. with level 3 ecoregions as a compromise between the  
71 more numerous (computationally demanding) level 4 ecoregions, and the coarser level 2  
72 ecoregions.

73 We used publicly available housing density estimates that were generated based on the  
74 U.S. 2000 decennial census as explanatory variables that may relate to human ignition  
75 pressure (Radeloff et al. 2010). These are provided at decadal time steps, and spatially  
76 at the level of census partial block groups. To generate approximate measures of hous-  
77 ing density at monthly time intervals, we used a simple linear interpolation over time for  
78 each block group, then aggregated spatially across block groups to compute mean housing  
79 density for each ecoregion in each month.

## 80 **Model development**

81 We built two types of models: one describing the occurrence of fires within each L3 ecore-  
82 gion over time (i.e., the total number of fires occurring in each ecoregion for each month  
83 from 1984 - 2015), and another describing the size of each wildfire in each ecoregion and  
84 month. For occurrence models, the response variable was a count (number of fires), and  
85 for burn area models, the response was a continuous positive quantity (size of each fire  
86 event). We used the period from 1984 to 2009 for training, withholding the period from  
87 2010 to 2015 to evaluate predictive performance.

### 88 **Fire occurrence**

89 We constructed four models for fire occurrence and compared their predictive performance  
90 based on test-set log likelihood and posterior predictive checks for the proportion of zeros,  
91 maximum count, and total count. The models differed in the distributions used in the like-  
92 lihood, representing counts as a Poisson, negative binomial, zero-inflated Poisson, or zero-  
93 inflated negative binomial random variable. The Poisson distribution is a common choice  
94 for counts, and the negative binomial distribution provides an alternative that can account  
95 for overdispersion. The zero-inflated versions of these distributions include a component  
96 to represent extra zeros, which might be expected to work well if there are independent  
97 processes that determine whether nonzero counts are possible (Lambert 1992).

98 For spatial units (ecoregions)  $s = 1, \dots, S$  and time steps (months)  $t = 1, \dots, T$ , each  
99 model defines a probability mass function for  $n_{s,t}$ : the number of fires over 1000 acres in  
100 ecoregion  $s$  and time step  $t$ . For each of the four count distributions under consideration,  
101 location parameters  $\mu_{s,t}$  and (for zero-inflated models) structural zero inflation parameters  
102  $\pi_{s,t}$  were allowed to vary in space and time. We used a log link function to ensure that  
103  $\mu_{s,t} > 0$ , and a logit link function to ensure that  $\pi_{s,t} \in (0, 1)$ . Concatenating over spatial  
104 and temporal units, so that  $\boldsymbol{\mu} = (\mu_{s=1,t=1}, \mu_{s=2,t=1}, \dots, \mu_{s=S,t=1}, \mu_{s=S,t=2}, \dots, \mu_{s=S,t=T})$ , and  
105 similarly for  $\boldsymbol{\pi}$ , we modeled distributional location and (when applicable) zero inflation  
106 parameters as:

$$\log(\boldsymbol{\mu}) = \alpha^{(\mu)} + \mathbf{X}\boldsymbol{\beta}^{(\mu)} + \boldsymbol{\phi}^{(\mu)} + \log(\mathbf{a}),$$

$$\text{logit}(\boldsymbol{\pi}) = \alpha^{(\pi)} + \mathbf{X}\boldsymbol{\beta}^{(\pi)} + \boldsymbol{\phi}^{(\pi)},$$

107 where  $\alpha^{(\mu)}$  and  $\alpha^{(\pi)}$  are scalar intercept parameters,  $\mathbf{X}$  is a known  $(S \times T) \times p$  design matrix,  
108 where  $p$  is the number of input features,  $\boldsymbol{\beta}^{(\mu)}$  and  $\boldsymbol{\beta}^{(\pi)}$  are column vector parameters of  
109 length  $p$ ,  $\boldsymbol{\phi}^{(\mu)}$  and  $\boldsymbol{\phi}^{(\pi)}$  are column vector parameters of length  $S \times T$  containing spatiotem-  
110 poral adjustments, and  $\mathbf{a}$  is a known offset vector of areas for spatial unit  $s = 1, 2, \dots, S$ ,  
111 repeated  $T$  times.

## 112 **Burn area**

113 We developed five candidate models for burn area, each of which specified a different dis-  
114 tribution for burn areas (Reed and McKelvey 2002; Hernandez et al. 2015), including the  
115 generalized Pareto (Hosking and Wallis 1987), tapered Pareto (Schoenberg, Peng, and  
116 Woods 2003), lognormal, gamma, and Weibull distributions. We evaluated each model  
117 in terms of test set log likelihood, and posterior predictive checks for burn area extremes.  
118 We defined the response  $y_i$  as the number of acres burned over 1000 for the  $i^{\text{th}}$  fire event,  
119 which occurred in spatial unit  $s_i$  and time step  $t_i$ .

120 Because each burn area distribution has a different parameterization, we included covari-  
121 ate effects in a distribution-specific way. For the generalized Pareto distribution (GPD),  
122 we assumed a positive shape parameter, leading to a Lomax distribution for exceedances  
123 (Bermudez et al. 2009). The GPD and Lomax shape parameters are related by  $\kappa^{(GPD)} =$   
124  $1/\kappa^{(L)}$ , and the GPD scale parameter is related to the Lomax scale and shape parameters  
125 by  $\sigma^{(GPD)} = \sigma^{(L)}/\kappa^{(L)}$ . We introduced covariate dependence via the Lomax scale parame-  
126 ter using a log link. For event  $i$ ,  $\log(\sigma_i^{(L)}) = \alpha + \mathbf{X}_{(s_i, t_i)}\boldsymbol{\beta} + \phi_{s_i, t_i}$ , where  $\alpha$  is an intercept  
127 parameter,  $\boldsymbol{\beta}$  is a length  $p$  vector of coefficients,  $\mathbf{X}_{(s_i, t_i)}$  is a row vector from  $\mathbf{X}$ , and  $\phi_{s_i, t_i}$   
128 is a spatiotemporal adjustment for  $s_i$  and  $t_i$ . For the tapered Pareto model, we modeled  
129 the shape parameter as  $\log(\kappa_i) = \alpha + \mathbf{X}_{(s_i, t_i)}\boldsymbol{\beta} + \phi_{s_i, t_i}$ . The lognormal model included

130 covariate dependence via the location parameter:  $\mu_i = \alpha + \mathbf{X}_{(s_i, t_i)}\boldsymbol{\beta} + \phi_{s_i, t_i}$ . The gamma  
131 model used a log link for the expected value:  $\log(E(y_i)) = \alpha + \mathbf{X}_{(s_i, t_i)}\boldsymbol{\beta} + \phi_{s_i, t_i}$ . Last, we  
132 modeled the Weibull scale parameter as  $\log(\sigma_i) = \alpha + \mathbf{X}_{(s_i, t_i)}\boldsymbol{\beta} + \phi_{s_i, t_i}$ . More detail on the  
133 parameterization of each burn area distribution is provided in the Appendices.

### 134 **Accounting for nonlinear forcing**

135 The design matrix  $\mathbf{X}$  was constructed to allow for spatially varying nonlinear effects of  
136 housing density and meteorological drivers. We used B-splines to account for nonlinearity  
137 (Figure 2) and allowed the coefficients for each basis vector to vary spatially (Wood 2017).  
138 First, we constructed univariate B-splines for log housing density, wind speed, same month  
139 precipitation, previous 12 month precipitation, air temperature, and humidity, with five  
140 degrees of freedom (including an intercept) for each variable. This step generated 30 basis  
141 vectors (five for each of six variables).

142 To allow for spatial variation in these nonlinear effects, we added interaction effects be-  
143 tween each of the basis vectors and ecoregions (Brezger and Lang 2006; Kneib, Hothorn,  
144 and Tutz 2009). The hierarchical nesting of ecoregion designations (Figure 1B-D) lends  
145 itself to such interactions. Conceptually, coefficients in a level 3 ecoregion may be related  
146 to coefficients in the level 2 ecoregion containing the level 3 region, the level 1 ecoregion  
147 containing the level 2 region, and a global effect. The coefficient associated with a basis  
148 vector for any level 3 ecoregion is treated as a sum of a global effect, a level 1 ecoregion  
149 adjustment, a level 2 ecoregion adjustment, and a level 3 ecoregion adjustment. Thus,  
150 for every univariate basis vector, we included interaction effects with ecoregion at each of  
151 the three ecoregion levels. This allows borrowing of information across space (level 3 ecore-  
152 gions in a level 2 ecoregion are often adjacent), and for regions that are ecologically similar.  
153 We also included adjustments on the global intercept for each level 1, 2, and 3 ecoregion to  
154 account for spatial variation that is unrelated to climate or housing density. This specifica-  
155 tion induces sparsity in  $\mathbf{X}$  that we exploit to increase the efficiency of computing  $\boldsymbol{\mu}$  and  $\boldsymbol{\pi}$ .  
156 In total,  $\mathbf{X}$  has  $p = 3,472$  columns, with 97% zero entries.

## 157 **Prior specification**

158 To avoid overfitting, we used a regularized horseshoe prior on the coefficients associated  
159 with the spatially varying nonlinear effects described above (Piironen, Vehtari, and others  
160 2017). This prior places high probability close to zero, while retaining heavy enough tails  
161 that nonzero coefficients are not shrunk too strongly toward zero. This is consistent with  
162 our prior expectation that most of the coefficients associated with the columns in  $\mathbf{X}$  were  
163 close to zero. For the zero inflated count models, we used a multivariate horseshoe to allow  
164 information sharing between the zero inflated and distribution specific location parameters  
165 (Peltola et al. 2014). For the remaining count models and all burn area models, this was  
166 a univariate horseshoe prior. Spatiotemporal random effects were constructed using a  
167 temporally autoregressive, spatially intrinsically autoregressive formulation (Besag and  
168 Kooperberg 1995; Banerjee, Carlin, and Gelfand 2014). Details of these priors and the  
169 resulting joint distributions are provided in the Appendices.

## 170 **Posterior predictive inference and extremes**

171 We used the posterior predictive distribution to check each model and make inference on  
172 extremes. The posterior predictive distribution provides a distribution for replications  
173 of observed data ( $y^{\text{rep}}$ ), and predictions of future data (Gelman et al. 2013). Conceptu-  
174 ally, for a “good” model,  $y^{\text{rep}}$  should be similar to observed training data  $y$ , and future  
175 predictions should be similar to future data. Distributions over both quantities can be  
176 obtained by conditioning  $y$  and marginalizing over model parameters  $\theta$ , e.g.,  $[y^{\text{rep}}|y] =$   
177  $\int [y^{\text{rep}}|\theta][\theta|y]d\theta$ .

178 Posterior predictive distributions facilitate model checks that compare predicted and ob-  
179 served test statistics (Gelman, Meng, and Stern 1996). To evaluate whether a model cap-  
180 tures tail behavior, one can compare an empirical maximum ( $T(y) = \max(y)$ ) to the pre-  
181 dicted distribution of maxima  $T(y^{\text{rep}})$ . We also include predictive checks for the proportion  
182 of zero counts, and totals for count and burn area models. Posterior predictive inference  
183 for maxima is similar in spirit to the MEV approach. Both obtain a distribution over max-



184 ima by marginalizing over unknowns including the number of events, size of each event,  
185 and parameters of their distributions (Marani and Ignaccolo 2015). However, a Bayesian  
186 approach explicitly conditions on the observed data to obtain a posterior distribution of  
187 parameters.

188 Seeing this connection is useful in the context of including priors and propagating uncer-  
189 tainty to derived parameters. Concretely, for any ecoregion  $s$  and timestep  $t$ , if we define  
190 a particular maximum fire size conditional on a fire having occurred as  $z_{s,t}$ , and let  $Z_{s,t}$   
191 represent the random variable of maximum fire size, then the cumulative distribution func-  
192 tion (CDF) for  $z_{s,t}$  is given by  $\Pr(Z_{s,t} \leq z_{s,t}) = F(y_{(s,t)})^{n_{s,t}}$ , where  $F(y_{(s,t)})$  is the CDF  
193 of fire size, and  $n_{s,t}$  is the number of wildfire events. The CDF for  $z_{s,t}$  can be inverted to  
194 produce a quantile function that permits computation of prediction intervals for maxi-  
195 mum fire sizes, conditional on fires having occurred. Given a collection of posterior draws  
196 from a burn area model that parameterize  $F(y_{(s,t)})$ , and a collection of posterior draws of  
197  $n_{s,t}^{\text{rep}}$  from the posterior predictive distribution of a wildfire count model, a posterior dis-  
198 tribution for the CDF or quantile function of maximum fire size can be generated which  
199 combines the two models to facilitate inference on the distribution of extremes.

## 200 **Parameter estimation**

201 We used a combination of variational approximations and Hamiltonian Monte Carlo meth-  
202 ods to sample from the posterior distributions of count and burn area models. A varia-  
203 tional approximation (Kucukelbir et al. 2015) was used for count models to quickly iden-  
204 tify a preferred model and avoid excessive multi-day model runs. Models were fit in the  
205 Stan probabilistic programming language using the `rstan` package (Carpenter et al. 2016;  
206 Stan Development Team 2018). The best performing count model and all burn area mod-  
207 els were fit using the No-U-Turn Sampler (Hoffman and Gelman 2014). We ran four chains  
208 for 1000 iterations each, and discarded the first 500 iterations as warmup. Convergence  
209 was assessed using visual inspection of trace plots, with potential scale reduction statistic  
210 values  $\hat{R} \geq 1.1$  as an indicator convergence failure (Brooks and Gelman 1998).

## 211 **Implementation**

212 All data processing, model fitting, and visualization were implemented with open source  
213 software, primarily in the R programming language (R Core Team 2017), and wrapped in  
214 a reproducible workflow via GNU Make and Docker (Stallman, McGrath, and Smith 2004;  
215 Boettiger 2015). Data cleaning and transformation required the R packages `assertthat`  
216 (Wickham 2017a), `lubridate` (Grolemund and Wickham 2011), `Matrix` (Bates and Maech-  
217 ler 2018), `pbapply` (Solymos and Zawadzki 2018), `splines` (R Core Team 2018), `tidyverse`  
218 (Wickham 2017b), and `zoo` (Zeileis and Grothendieck 2005). Spatial data were processed  
219 with `raster` (Hijmans 2017), `rgdal` (Bivand, Keitt, and Rowlingson 2018), `sf` (Pebesma  
220 2018), and `spdep` (Bivand and Piras 2015). Finally, we used `cowplot` (Wilke 2017), `ggre-`  
221 `pel` (Slowikowski 2018), `ggthemes` (Arnold 2018), `patchwork` (Pedersen 2017), and `RCol-`  
222 `orBrewer` (Neuwirth 2014) for visualization. The manuscript was written in R Mark-  
223 down (Allaire et al. 2018). Analyses were run on an Amazon Web Services `m5.2xlarge`  
224 EC2 instance with four physical cores and 32 GB of RAM, and the whole workflow re-  
225 quires  $\approx 72$  hours. All code to reproduce the analysis is available on GitHub at [https:](https://github.com/mbjoseph/wildfire-extremes)  
226 [//github.com/mbjoseph/wildfire-extremes](https://github.com/mbjoseph/wildfire-extremes) (Joseph 2018).

## 227 **Results**

### 228 **Wildfire occurrence**

229 The zero-inflated negative binomial distribution performed best on the held-out test set  
230 (Table 1), and was able to recover the proportion of zeros, count maxima, and count to-  
231 tals in posterior predictive checks for both the training and test data (Figure 3). All of  
232 the other count models that we considered exhibited lack of fit to at least one of these  
233 statistics in posterior predictive checks. Hereafter, we report results from the zero-inflated  
234 negative binomial model.

235 Minimum relative humidity and maximum air temperature had the strongest effects on

236 both the zero-inflation component and the expected value of the negative binomial com-  
237 ponent (Figure 4, posterior median for  $\rho$ : 0.663, 95% credible interval (CI): 0.361 - 0.86).  
238 The model uncovered unique effects of meteorological variables at level 1, 2, and 3 ecore-  
239 gions (Figure 5). For example, a positive interaction effect between the second air temper-  
240 ature basis vector and the L1 Great Plains ecoregions indicates that the expected number  
241 of wildfires in plains ecoregions with cold conditions is high relative to other ecoregions.  
242 The Ozark/Ouachita-Appalachian forest and Ozark Highlands were also identified as hav-  
243 ing region-specific temperature effects (Figure 5). Twelve month total precipitation also  
244 had region specific effects in the Mississippi Alluvial and Southeast Coastal Plains ecore-  
245 gion, where it tended to reduce fire risk (Figure 5). In contrast, increasing cumulative  
246 twelve month precipitation tended to increase risk in desert ecoregions (Figure 4). Housing  
247 density showed a unimodal relationship to expected count (Figure 4), with lower expected  
248 counts in sparsely populated ecoregions, and higher expected counts with moderately pop-  
249 ulated ecoregions.

250 Posterior 95% credible interval coverage for the number of fires over 1000 acres in the test  
251 set was 98.9%. The lowest test set interval coverage was 91.7%, in the Cross Timbers L3  
252 ecoregion. When observed counts fell outside the 95% prediction interval, counts were  
253 larger than predicted 100% of the time. The largest difference between observed numbers  
254 and predicted 97.5% posterior quantiles (the upper limit for the 95% credible interval) oc-  
255 curred for the Columbia Mountains/Northern Rockies L3 ecoregion in August 2015, when  
256 36 fires over 1000 acres occurred and at most 23.025 were predicted. For nearly half of the  
257 level 3 ecoregions (44 of 85), accounting for 40.2% of the land area of the contiguous U.S.,  
258 the zero-inflated negative binomial model had 100% test set prediction interval coverage.

## 259 **Wildfire burned areas**

260 The lognormal distribution performed best on the test set (Table 2), and captured tail-  
261 behavior better than other burn area distributions (Figure 6). The GPD model was too  
262 heavy-tailed to adequately capture the pattern in the empirical data, predicting fires far  
263 larger than those observed in the training and test sets (Figure 6). The tapered Pareto

264 distribution was too light-tailed (Figure 6). The gamma and Weibull models performed  
265 very poorly overall on the test set (Table 2), apparently due to a lack of congruence be-  
266 tween the shapes of these distributions and the actual burn area distribution. Despite a  
267 poor fit to the bulk of the wildfire burn area distribution, both performed adequately in  
268 the upper tails (Figure 6). Hereafter we present results for the lognormal model, which  
269 had the highest test set log likelihood and captured tail behavior of the empirical fire size  
270 distribution.

271 Relative humidity was the primary driver of expected burn area for a fire event (Figure  
272 7A). The first basis vector for mean daily minimum relative humidity was the only coeffi-  
273 cient with a 95% credible interval that did not include zero (posterior median: 1.66, 95%  
274 CI: (0.57 - 2.28)). This nonlinear effect can be observed in Figure 7B as an increase in the  
275 expected burn area below 20% mean daily minimum humidity. This leads to a seasonality  
276 gradient among ecoregions of expected fire sizes, with little or no seasonal signal in typi-  
277 cally humid ecoregions such as Marine West Coast Forests of the Pacific Northwest, and  
278 seasonal oscillations in ecoregions that have periodic fluctuations between dry and humid  
279 conditions such as the Temperate Sierras (Figure 7C). There was not strong evidence that  
280 meteorological variables had spatially variable effects on expected wildfire burn area.

281 Overall, 95% posterior predictive interval coverage in the test set for burn areas was 92.3%.  
282 The lowest test set coverage was 0%, for the Eastern Great Lakes Lowlands L3 ecoregion,  
283 followed by 50%, for the Central California Valley L3 ecoregion, though these ecoregions  
284 had just 1 and 2 wildfire events in the test set. When observed fire sizes fell outside the  
285 95% prediction interval, 23.3% of wildfires were smaller than predicted, and 76.7% of wild-  
286 fires were larger than predicted. The largest discrepancy between the actual size of a wild-  
287 fire and the predicted 97.5% posterior quantile was observed with the Wallow Fire in 2011  
288 which burned 563,655 acres, but the predicted upper limit for size was 50,136. We inves-  
289 tigate this discrepancy further in the case study below. The lognormal burn area model  
290 achieved 100% interval coverage in 26 of 66 ecoregions that had wildfire events in the test  
291 set, accounting for 29% of the land area of the contiguous U.S.

## 292 Inference on extremes

293 By combining the output of the event count and burn area models, we derived posterior  
294 prediction intervals for the size of the largest fire in a month for each region (the “burn  
295 area maximum”), integrating over uncertainty in the number of fires, as well as the log-  
296 normal mean and standard deviation for burn area. We evaluated the posterior distri-  
297 bution for the quantile function of the  $n$ -sample maximum of a lognormal distribution  
298 ( $\exp(\mu + \sigma\sqrt{2}\text{erf}^{-1}(2P^{1/n} - 1))$ ), where  $\text{erf}^{-1}$  is the inverse error function, and  $P$  is a prob-  
299 ability) to generate prediction intervals for maximum fire sizes by month and ecoregion,  
300 conditional on one or more fires having occurred. In the holdout period from 2010 to 2015,  
301 a 99% prediction interval achieved 77.6% interval coverage, with 13.6% of the burn area  
302 maxima (109 fire events) being larger than predicted (Figure 8). The model predicted  
303 the total area burned over the entire contiguous United States in test period from 2010  
304 to 2015 to be 28,401,914 (95% CI: (18,037,280 - 50,996,560) and the actual value was  
305 26,639,835.

306 While fires over a million acres in size have happened historically in the contiguous U.S.  
307 (Pernin 1971), all fires in the training and test sets were below one million acres. If we  
308 extrapolate, the probability that at least one fire exceeded one million acres in the period  
309 from 2010 to 2015 was estimated to be between 0.171 and 0.661 (95% CI), with a posterior  
310 median of 0.33. The highest estimated probability for a million acre event was 0.014 (pos-  
311 terior median), with a 95% CI of (0, 0.272) seen for the Southwestern Tablelands ecoregion  
312 in June 2011. The second highest probability for a million acre event was 0.003 (posterior  
313 median), with a 95% CI of (0, 0.048) seen for the Arizona/New Mexico Mountains ecore-  
314 gion in June 2011. Aggregating spatially, we estimated monthly probabilities of a million  
315 acre wildfire. These probabilities show seasonal signals corresponding to peak fire seasons,  
316 with a shift toward higher and broader peaks beginning in the 21st century (Figure 9).

## 317 **Error analysis case study: the 2011 Wallow Fire**

318 To better understand how well the model could or could not anticipate notable extreme  
319 events, and why, we used the largest fire in the test set as a case study. The Wallow Fire  
320 was accidentally ignited on May 29, 2011 by two campers in the L3 Arizona/New Mex-  
321 ico Mountains ecoregion. It burned through the month of June and into early July. The  
322 model underpredicted the total burn area of the Wallow Fire. Integrating over uncer-  
323 tainty in the predicted number of fires and expected fire size, the 99% credible interval  
324 for the maximum fire size for May 2011 was (1,820 - 270,304) acres, but the Wallow Fire is  
325 recorded as 563,655 acres in the MTBS data.

326 We evaluated the contribution of each covariate to the linear predictor functions of the  
327 three model components (lognormal mean for burn areas, negative binomial mean for  
328 counts, and the logit probability of the zero-inflation component) to understand why these  
329 predictions differed. We defined the contribution of a variable as the dot product of the  
330 elements in the design matrix  $\mathbf{X}$  corresponding to a particular driver variable (e.g., hu-  
331 midity), and the estimated coefficients in  $\beta$  corresponding to that variable. This provides  
332 a quantitative measure of how each input variable contributes to the linear predictor for  
333 an ecoregion, and incorporates the overall, level 1, level 2, and level 3 ecoregion adjust-  
334 ments on these effects. Humidity is the primary driver of variation in the model's pre-  
335 dictions overall, and June 2011 - the month after ignition - favored more large fires, with  
336 drier, hotter conditions (Figure 10). The 99% credible interval for June 2011 was (10,053  
337 - 1,005,827) acres, which contains the true value. Evidently, conditions in May that drove  
338 (under)predictions of maximum burn area were not representative of the conditions over  
339 most of the Wallow Fire's duration. The failure of the model to correctly predict the size  
340 of the Wallow fire suggests potential avenues for improvement, discussed below.

## 341 Discussion

342 Extreme wildfires are often devastating, but perhaps they need not be surprising. By  
343 allowing the non-linear effects of weather and housing density to vary across space, we  
344 were able to achieve good predictive accuracy for fire extremes over a five-year prediction  
345 window. We estimate a moderate chance of wildfires larger than what has been observed  
346 in recent decades, perhaps even over one million acres. Such predictions can support short  
347 to medium term wildfire management and probabilistic hazard assessment.

348 Driving a model with meteorological features raises challenges related to predictive un-  
349 certainty and covariate shift - a change in the underlying distribution of forcing variables,  
350 potentially outside of the historic range. Ideally, this uncertainty would be propagated for-  
351 ward in a predictive model, possibly through stacking of predictive distributions that are  
352 generated from multiple models of future climate dynamics (Yao et al. 2017). But, even  
353 if one had a perfect forecast, novel conditions present a challenge for predictive modeling  
354 (Quionero-Candela et al. 2009). For example, the High Plains ecoregion had its highest  
355 monthly precipitation, lowest 12 month running precipitation, driest, hottest, and windi-  
356 est conditions in the test set period, so that the range of environmental conditions in the  
357 training data did not encompass the range of future conditions. Extrapolating beyond the  
358 range of training inputs is generally difficult, but the hierarchical spatial effect specifica-  
359 tion used here allows partial pooling among climatically similar ecoregions that can inform  
360 such predictions, unlike models fit separately to disjoint spatial regions.

361 Human-caused climate change is expected to increase fire activity in the western U.S.  
362 (Rogers et al. 2011; Westerling et al. 2011; Moritz et al. 2012; Abatzoglou and Williams  
363 2016) and elsewhere (Flannigan et al. 2009), which when coupled with the nonlinear ef-  
364 fect of human-density provides a key inferential wrinkle. While most U.S. ecoregions are  
365 increasing in human density over time, some of these ecoregions are in the range of val-  
366 ues in which this increases the expected number of large fires, while others are so pop-  
367 ulated that further increases would reduce the chance of a large fire. The hump-shaped  
368 effect of human density on the expected number of large fires is likely driven by ignition

369 pressure and fire suppression (Balch et al. 2017). As human density increases from zero,  
370 ignition pressure increases, but eventually landscapes become so urbanized, fragmented,  
371 and/or fire-suppressed that wildfire risk decreases (Syphard et al. 2007; Bowman et al.  
372 2011; Bistinas et al. 2013; Knorr et al. 2013; Mcwethy et al. 2013; Syphard et al. 2017;  
373 Nagy et al. 2018). At intermediate density, wildfire dynamics respond to human ignition  
374 and altered fuel distributions (Guyette, Muzika, and Dey 2002), but these responses de-  
375 pend on environmental context and characteristics of the human population (Marlon et  
376 al. 2008; Li et al. 2009). This model indicates that the combination of moderate to high  
377 human density and dry conditions would nonlinearly increase the chance of an extreme fire  
378 event. Both human density and dryness are expected to increase in the future across large  
379 swaths of the U.S. (Lloyd, Sorichetta, and Tatem 2017; Stavros et al. 2014, Radeloff et  
380 al. (2010)), with potential implications for human mortality, health risks from smoke and  
381 particulate emission, and the financial burden of wildfire management (Reid et al. 2016;  
382 Radeloff et al. 2018).

383 This work points to promising directions for future predictive efforts. Default choices such  
384 as Poisson and GPD distributions should be checked against alternative distributions. Fur-  
385 ther, the predictive skill of this model seems to suggest that ordinary events provide infor-  
386 mation on extremes, which would not be the case if the generative distribution of extremes  
387 was completely unique. Previous case studies have identified that extremes or anomalies  
388 in climatological drivers play a role in the evolution of extreme wildfires (Peterson et al.  
389 2015), but for this work, monthly averages of climatological drivers over fairly large spatial  
390 regions were used, which may smooth over anomalous or extreme conditions. Enhancing  
391 the spatiotemporal resolution of predictive models could better represent climatic and  
392 social drivers of fire dynamics and provide localized insights into fire dynamics to inform  
393 decision-making. This raises computational challenges, but recent advances in distributed  
394 probabilistic computing (Tran et al. 2017), efficient construction of spatiotemporal point  
395 processes (Shirota and Banerjee 2018), and compact representations of nonlinear spatial  
396 interactions (Lee and Durbán 2011) may provide solutions.

397 The Wallow Fire case study reveals at least one limitation of increasing the spatiotempo-



398 ral resolution. When the model predictions are driven by covariates that are summarized  
399 in space and time (e.g. a mean across an ecoregion in a month), summary values may  
400 not represent conditions that are most relevant to an event. With a discrete space-time  
401 segmentation, events can occur at the boundary of a spatiotemporal unit, e.g., if a fire  
402 spreads into an adjacent ecoregion or ignites on the last day of the month. Large wild-  
403 fires can span months, and a model that only uses conditions upon ignition to predict  
404 total burn area can fail to account for conditions that change over the course of the event.  
405 Modeling ignitions as a point process in continuous space and time (Brillinger, Preisler,  
406 and Benoit 2003), and explicitly modeling subsequent fire duration and spatial dynamics  
407 could better separate conditions that ignite fires from those that affect spread. Such an  
408 approach might be amenable to including information on fuel continuity, which is likely  
409 to limit the size of extremely large fires and did not factor into the current models predic-  
410 tions (Rollins, Morgan, and Swetnam 2002; Hargrove et al. 2000). To the extent that a  
411 model reflects the generative process for extreme events, the decomposition of contribu-  
412 tions to the model's predictions may provide insight into attribution for meteorological  
413 and anthropogenic drivers of extremes.

414 This paper presents and evaluates a statistical approach to explain and predict extreme  
415 wildfires that incorporates spatially varying non-linear dynamics. The model reveals con-  
416 siderable differences in fire dynamics among ecoregions spanning the mountain west to the  
417 great plains, deserts, and eastern forests, and suggests a decent chance of very large fires  
418 exceeding one million acres in the contiguous U.S. Predictive approaches such as this can  
419 inform decision-making by placing probabilistic bounds on the number of wildfires and  
420 their sizes, while provide deeper insights into wildfire ecology.

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## 727 Tables

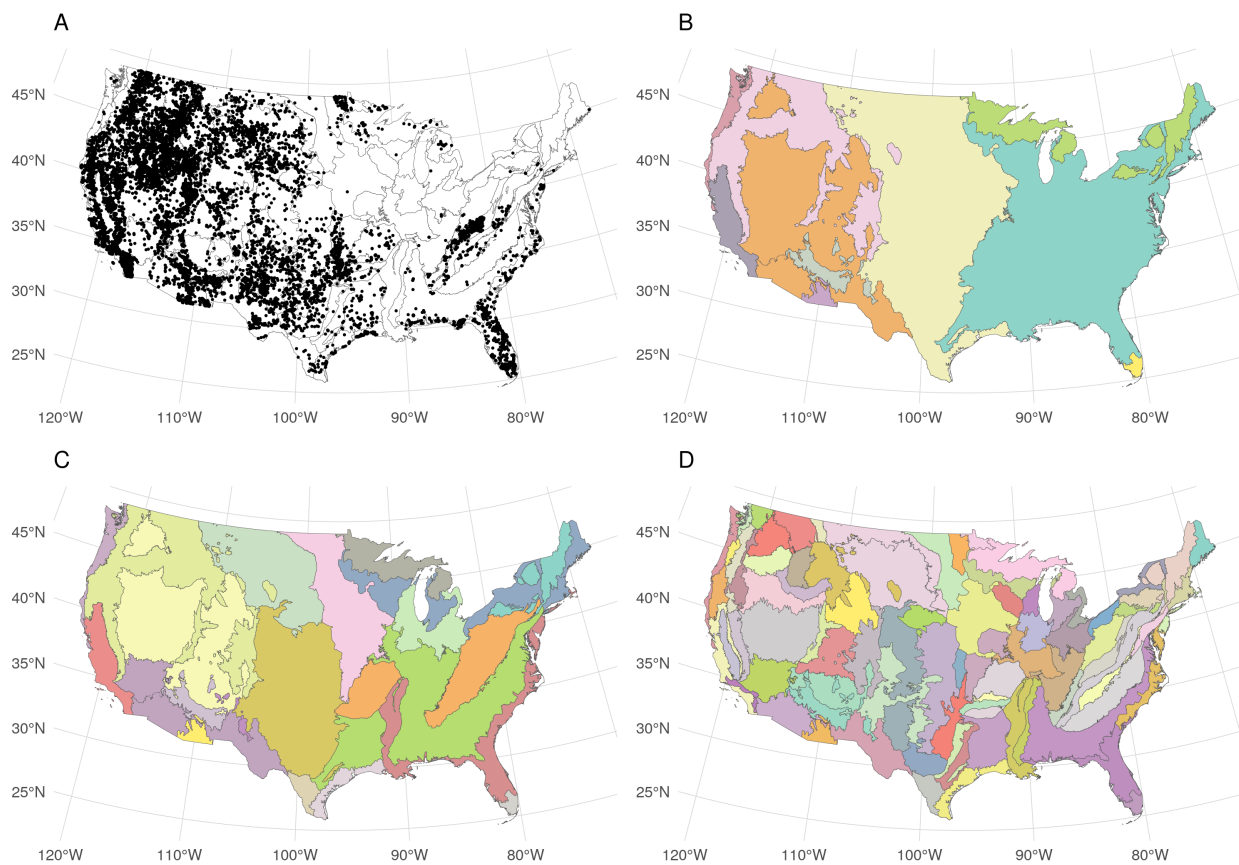
**Table 1.** Performance of count models on the test set in descending order. Posterior means are provided with standard deviations in parentheses.

Model	Holdout log likelihood
ZI Negative binomial	-3068 (55)
Negative binomial	-3090 (60)
ZI Poisson	-3526 (71)
Poisson	-4194 (124)

**Table 2.** Performance of burn area models on the test set in descending order. Posterior means are provided with standard deviations in parentheses.

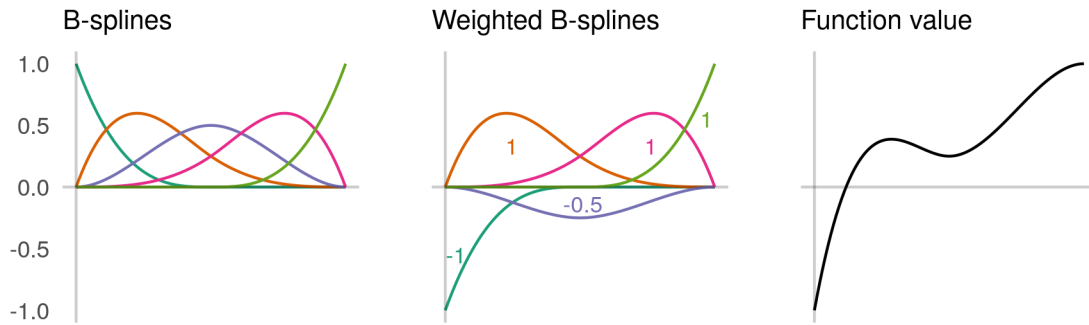
Model	Holdout log likelihood
Lognormal	-22341 (39)
Generalized Pareto	-22375 (42)
Tapered Pareto	-22388 (48)
Weibull	-23499 (230)
Gamma	-26191 (908)

728 **Figures**

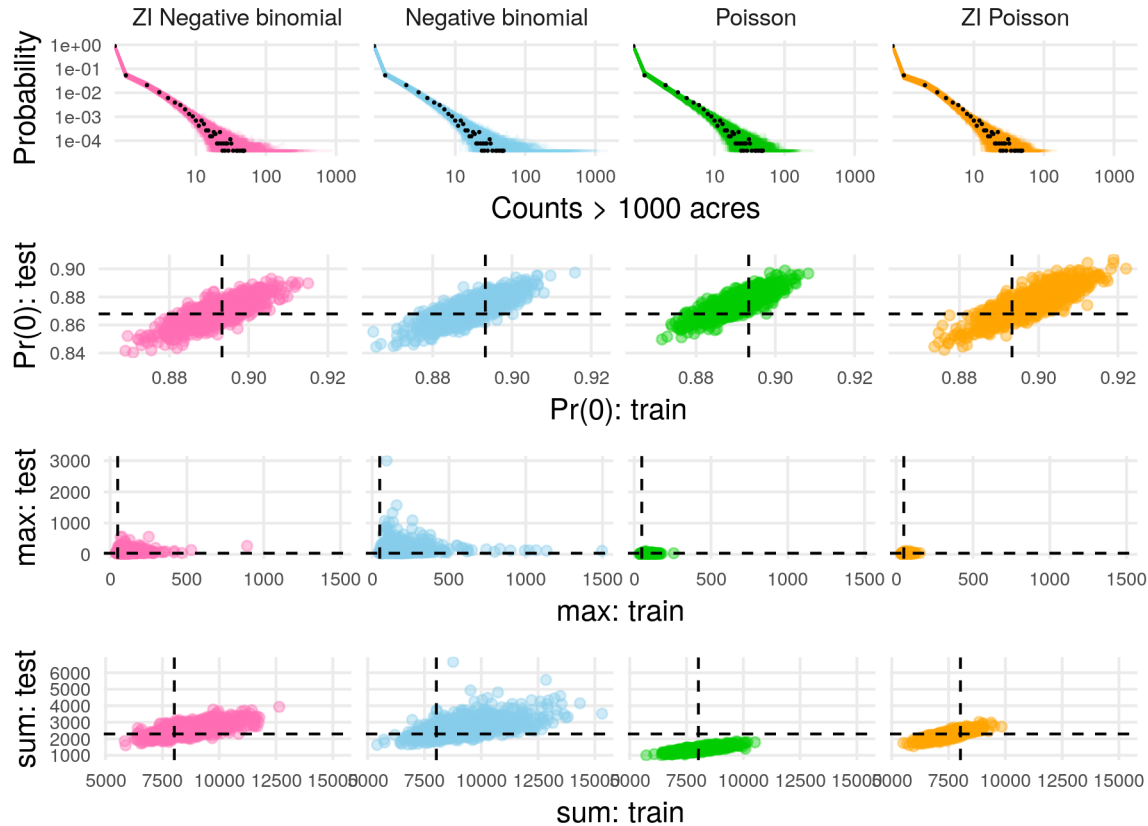


**Figure 1.** A. Fire ignition locations are shown as points across the study region. Colors in panels B, C, and D show level 1, 2, and 3 ecoregions respectively.

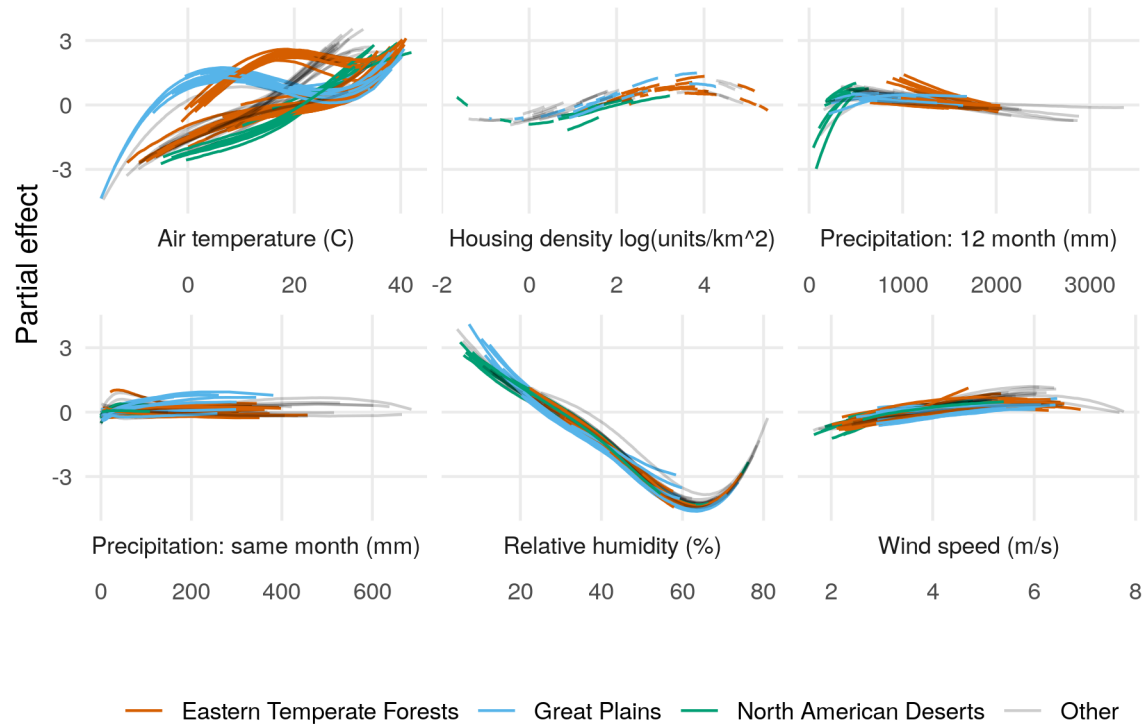




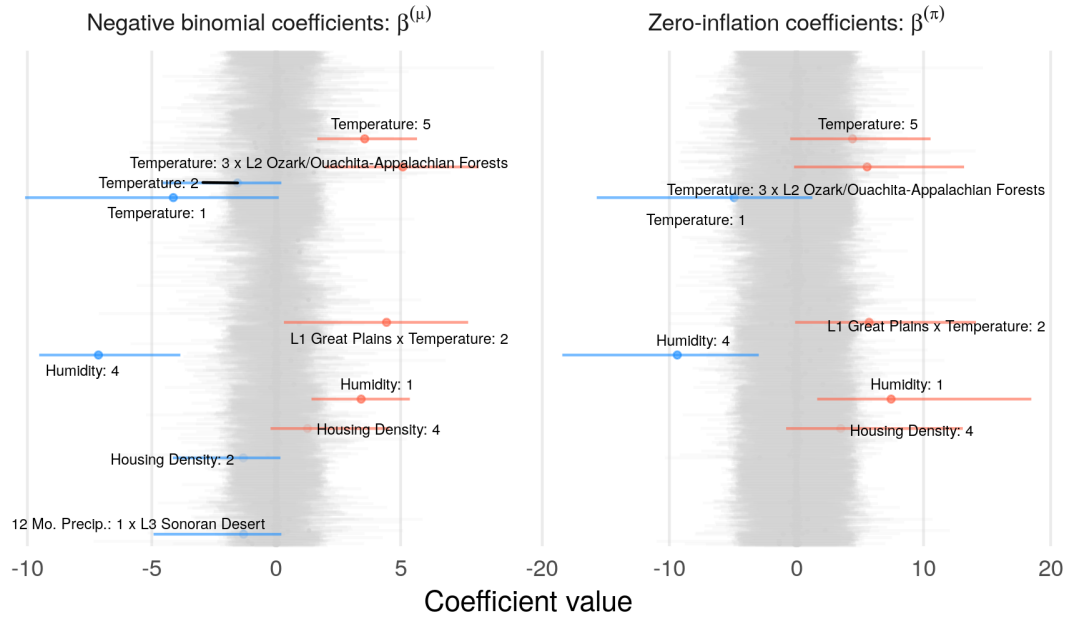
**Figure 2.** Conceptual figure to illustrate the use of B-splines to construct nonlinear functions. In the left panel, five B-spline vectors are shown, which map values of an input variable (on the x-axis) to a value on the y-axis. The middle panel shows the same B-spline vectors, but weighted (multiplied) by real numbers, with the weights illustrated as annotations. These weighted B-spline vectors are summed to produce the values of a nonlinear function (right panel).



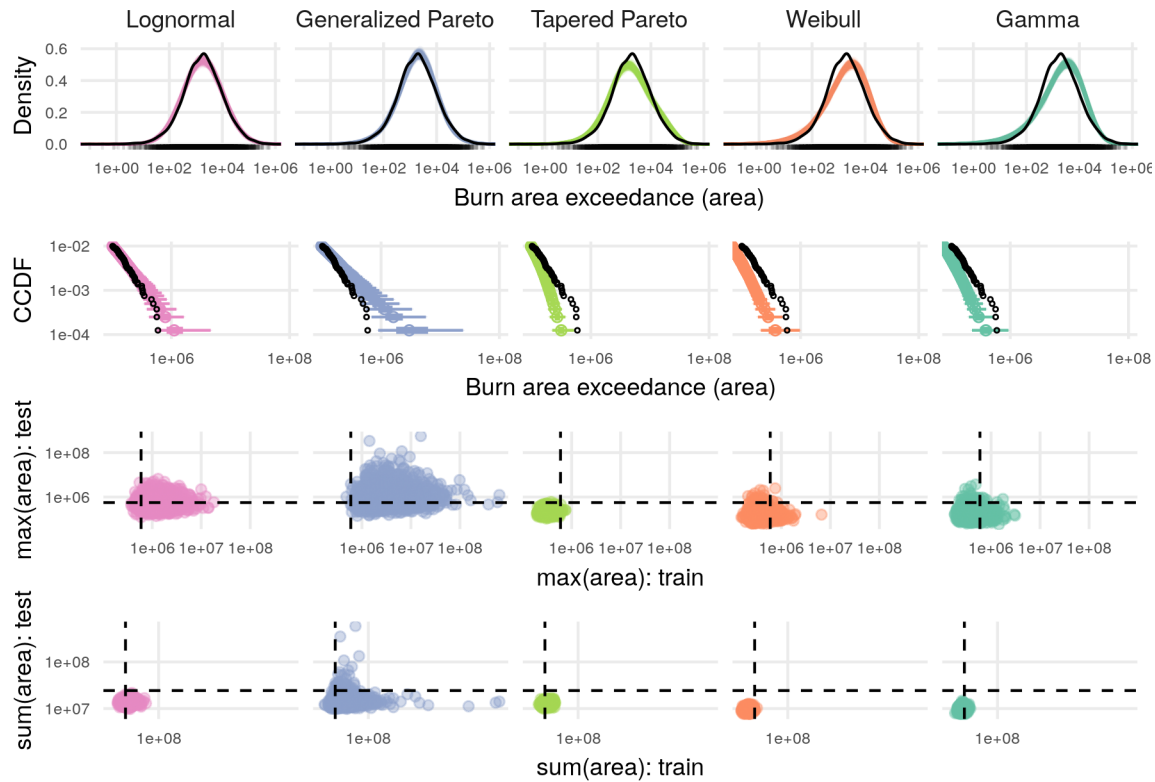
**Figure 3.** Count predictive checks. Row one shows observed count frequencies as black points and predicted frequencies as lines. Rows two, three, and four show predicted proportions of zeros, maxima, and sums (respectively) in the training and test data, with empirical values as dashed lines. Rows two through four facilitate comparison of performance on training and test sets. Ideally, model predictions cluster around the dashed lines for both the training (x-axis direction) and test (y-axis direction) sets, leading to a tight cluster of points at the intersection of the dashed lines.



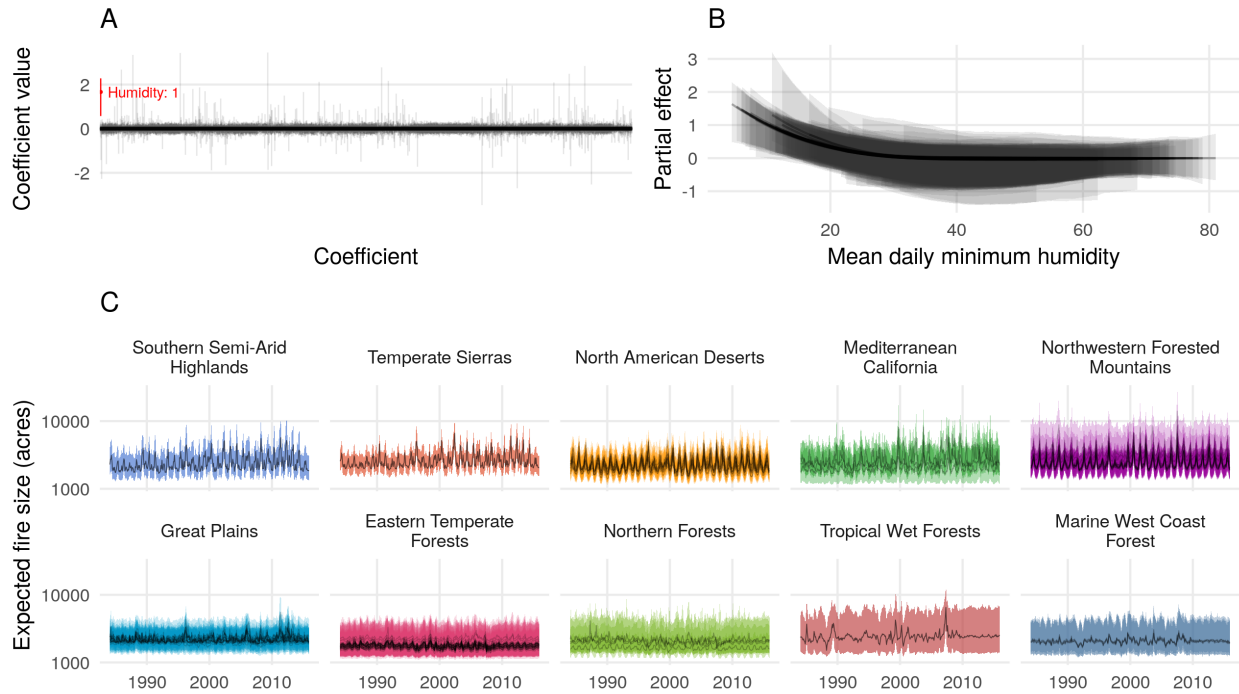
**Figure 4.** Partial effects on the log-transformed negative binomial mean component of the zero-inflated negative binomial model for each level 3 ecoregion, colored by level 1 ecoregion. Lines are posterior medians. Results are similar for the zero-inflation component.



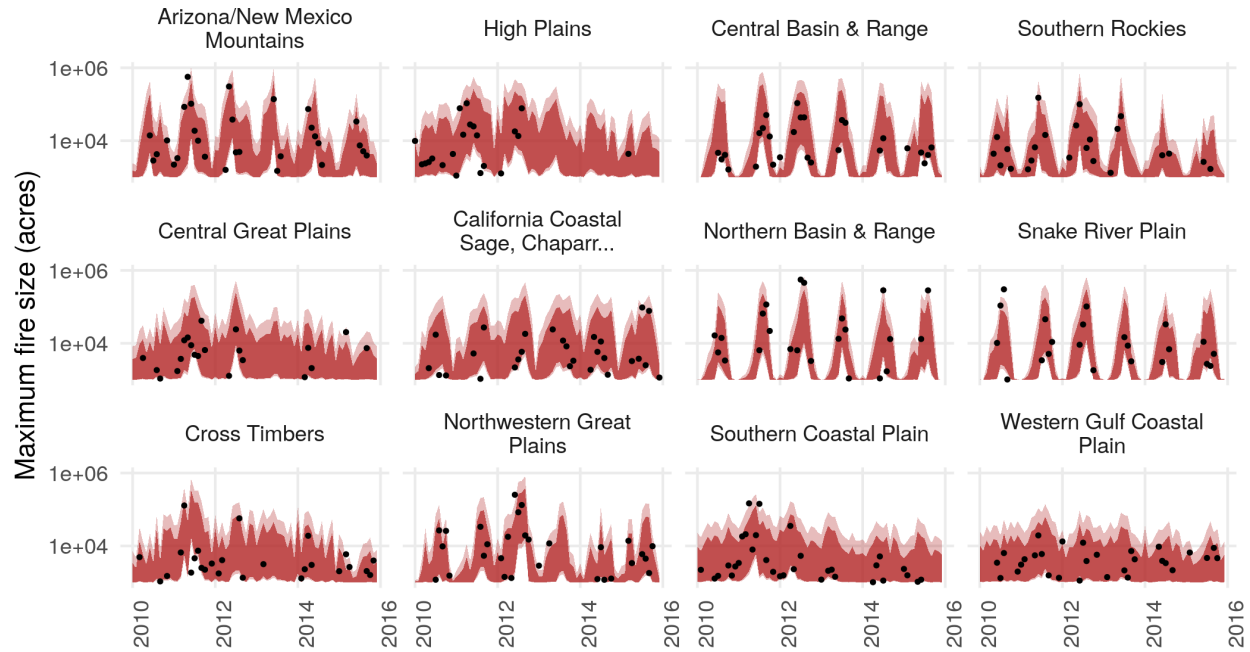
**Figure 5.** Caterpillar plots of zero inflated negative binomial model coefficients,  $\beta^{(\mu)}$  (left) and  $\beta^{(\pi)}$  (right). Horizontal line segments denote 95% credible intervals. Grey segments indicate coefficients with a less than 87% posterior probability of being positive or negative, and colored segments indicate coefficients that are probably positive (red) or negative (blue). B-spline vectors are indicated by colons, e.g., Humidity:1 indicates the first basis vector corresponding to humidity. Interactions between variables a and b are represented as a x b. Level 1 ecoregions are represented by L1 ecoregion name, and L2 and L3 indicate level 2 and 3 ecoregions.



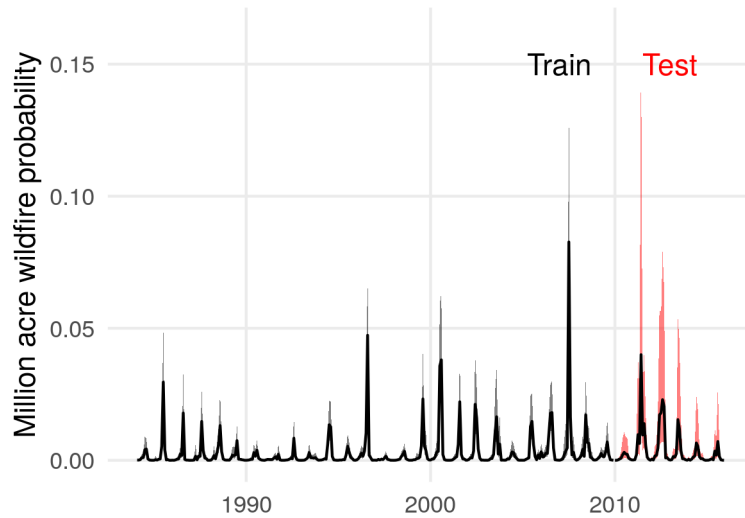
**Figure 6.** Predictive checks for burn area models. The top row shows predicted density in color and empirical density for the training set in black, which reveals overall lack of fit for the gamma and Weibull models. Row two shows the complementary cumulative distribution function (CCDF) at the tails, with 95% and 50% prediction intervals shown in color and observed data as black points, which shows that the Generalized Pareto distribution predicts values that are too extreme. The third and fourth rows show checks for maximum and total burn areas in the training and test set, with observed values as dashed lines and posterior draws as colored points. These final two rows facilitate checks for summary statistics on both the training and test set, with the ideal model generating predictions (colored points) clustered close to where the dashed lines intersect.



**Figure 7.** **A.** Estimated posterior medians and 95% credible intervals for each of the 3,473 coefficients associated with expected burn area. Only one coefficient - the first basis vector for humidity - had a 95% credible interval that excluded zero, shown in red. This effect is visualized in **B**. Partial effects of mean daily minimum humidity for each level 3 ecoregion, with posterior medians drawn as lines, and the 95% credible intervals as ribbons. **C.** Monthly time series of expected fire sizes for every level 3 ecoregion, faceted and colored by level 1 ecoregions sorted by mean humidity. Lines are posterior medians and ribbons are 95% credible intervals.

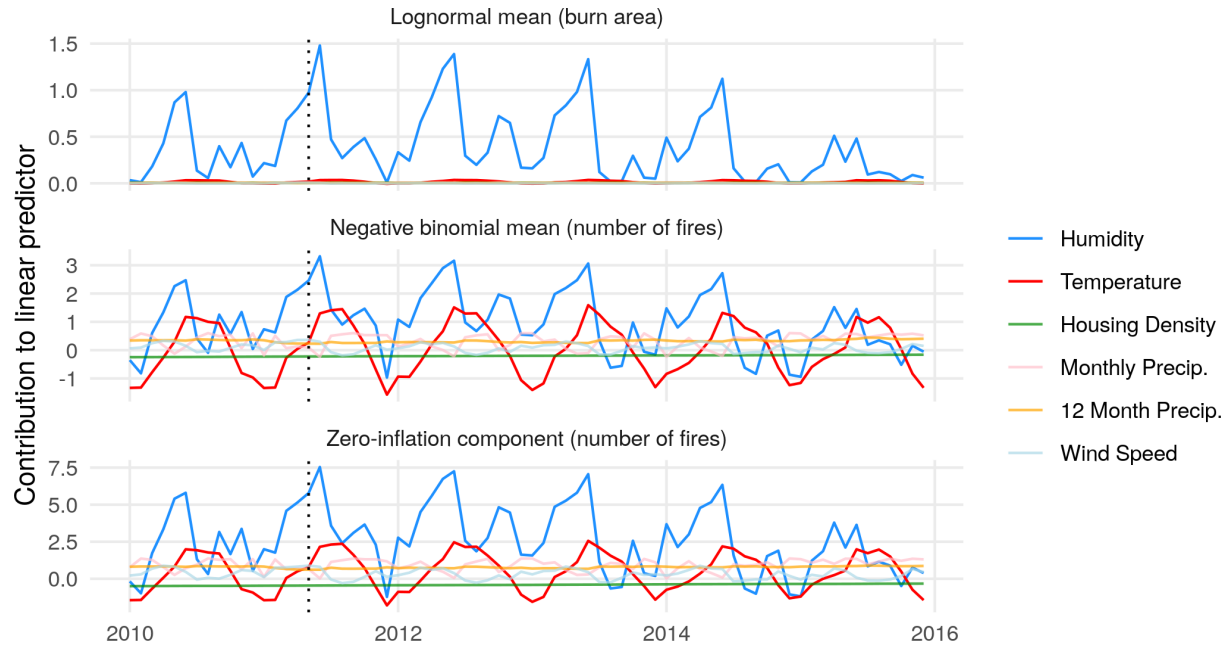


**Figure 8.** Posterior 99% (light red) and 95% (dark red) prediction intervals for the burn area of the largest fire event by month and level 3 ecoregion in the test set, shown for ecoregions with wildfires in more than 20 months. Empirical maxima are shown as black dots.



**Figure 9.** Estimated monthly posterior probabilities that one or more fire events exceed one million acres. The line represents the posterior median, and shaded region represents an 80% credible interval. The training period up to 2010 is shown in black, and the test period for which data were withheld during parameter estimation is shown in red.





**Figure 10.** Posterior median contribution of each input variable to the linear predictor function of model components for the Arizona/New Mexico Mountains level 3 ecoregion from 2010-2016. A dotted vertical line marks May 2011, when the Wallow Fire ignited. Vertical positions of colored lines show contributions to the linear predictor function of each model component.

## 729 Appendices

### 730 Prior specifications

731 Prior distributions were chosen to regularize coefficients on the distribution specific means  
732  $\beta^{(\mu)}$  and structural zero parameters  $\beta^{(\pi)}$ . We used a regularized horseshoe prior on these  
733 coefficients, which shrinks irrelevant coefficients towards zero, while regularizing nonzero  
734 coefficients (Piironen, Vehtari, and others 2017). For zero-inflated models, we used a multi-  
735 variate version of the regularized horseshoe (Peltola et al. 2014):

$$\begin{pmatrix} \beta_j^{(\mu)} \\ \beta_j^{(\pi)} \end{pmatrix} \sim N\left(\mathbf{0}, \begin{pmatrix} \tau_1^2 \tilde{\lambda}_{1,j}^2 & \rho \tau_1 \tau_2 \tilde{\lambda}_{1,j} \tilde{\lambda}_{2,j} \\ \rho \tau_1 \tau_2 \tilde{\lambda}_{1,j} \tilde{\lambda}_{2,j} & \tau_2^2 \tilde{\lambda}_{2,j}^2 \end{pmatrix}\right),$$

$$\tilde{\lambda}_{m,j}^2 = \frac{c_m^2 \lambda_j^2}{c_m^2 + \tau_m^2 \lambda_j^2},$$

736 for each response dimension  $m = 1, 2$  and coefficient  $j = 1, \dots, p$ . Here  $\rho$  is a correlation  
737 parameter,  $\tau_1$  and  $\tau_2$  are global variance hyperparameters,  $c_1$  and  $c_2$  are hyperparameters  
738 that determine the amount of shrinkage on the largest coefficients, and  $\lambda_j$  is a local scale  
739 parameter drawn from a half-Cauchy distribution that control the amount of shrinkage  
740 applied to coefficient  $j$  (Piironen, Vehtari, and others 2017). With this prior specification,  
741 information can be shared across the two response dimensions through the correlation pa-  
742 rameter  $\rho$ , and/or through the local scale parameters  $\lambda_j$ . For count models without struc-  
743 tural zeros (the Poisson and negative binomial models), this multivariate prior simplifies to  
744 a univariate regularized horseshoe prior.

745 Spatiotemporal random effects were constructed using a temporally autoregressive, spa-  
746 tially intrinsically autoregressive formulation (Besag and Kooperberg 1995; Banerjee, Car-  
747 lin, and Gelfand 2014). Temporarily suppressing the superscript that indicates whether  
748 the effects are on  $\mu$  or  $\pi$ , and denoting column  $t$  from an  $S \times T$   $\Phi$  as  $\phi_t$  we have:

$$\phi_{t=1} \sim N(\mathbf{0}, (\tau^{(\phi)}(\mathbf{D} - \mathbf{W}))^{-1})$$

$$\phi_t \sim N(\eta\phi_{t-1}, (\tau^{(\phi)}(\mathbf{D} - \mathbf{W}))^{-1}), \quad t = 2, \dots, T$$

749 where  $\eta$  is a temporal dependence parameter,  $\tau^{(\phi)}$  is a precision parameter,  $\mathbf{D}$  is an  $S \times S$   
750 diagonal matrix with entries corresponding to the number of spatial neighbors for each  
751 spatial unit, and  $\mathbf{W}$  is an  $S \times S$  spatial adjacency matrix with nonzero elements only when  
752 spatial unit  $i$  is a neighbor of spatial unit  $j$  ( $w_{i,j} = 1$  if  $i$  is a neighbor of  $j$ , and  $w_{i,j} = 0$   
753 otherwise, including  $w_{i,i} = 0$  for all  $i$ ).  $\tau^{(\phi)}$  is a precision parameter. We imposed a soft  
754 identifiability constraint that places high prior mass near  $\sum_{s=1}^S \phi_{t,s}^* = 0$  for all  $t$ .

755 We applied a univariate regularized horseshoe prior to all  $\beta$  coefficients in burn area mod-  
756 els (Piiironen, Vehtari, and others 2017):

$$\beta_j \sim N\left(0, \tau^2 \tilde{\lambda}_j^2\right), \quad \tilde{\lambda}_j^2 = \frac{c^2 \lambda_j^2}{c^2 + \tau^2 \lambda_j^2},$$

757 Spatiotemporal random effects were constructed in the same way as for the count models.

## 758 Joint distributions

759 Here we provide the unnormalized posterior densities for each model. Square brackets rep-  
760 resent a probability mass or density function. Parameterizations for model likelihoods are  
761 provided first, followed by the factorization of the joint distribution, with explicit priors.

### 762 Poisson wildfire count model

763 We used the following parameterization of the Poisson distribution:

$$[n|\mu] = \frac{\mu^n e^{-\mu}}{n!},$$

764 where  $\mu$  is the mean and variance.

765 The unnormalized posterior density of this model is:

$$\begin{aligned} & [\boldsymbol{\beta}^{(\mu)}, \alpha^{(\mu)}, \boldsymbol{\phi}, \sigma^{(\phi)}, \eta, \boldsymbol{\lambda}, c, \tau | \mathbf{N}] \propto \\ & \prod_{s=1}^S \prod_{t=1}^T [n_{s,t} | \boldsymbol{\beta}^{(\mu)}, \alpha^{(\mu)}, \phi_{s,t}] \times \\ & [\phi_1 | \sigma^{(\phi)}] \prod_{t=2}^T [\phi_t | \phi_{t-1}, \sigma^{(\phi)}, \eta] \times \\ & \prod_{j=1}^p [\beta_j^{(\mu)} | \lambda_j, c, \tau] [\lambda_j] \times \\ & [\sigma^{(\phi)}] [\eta] [c] [\tau] [\alpha^{(\mu)}] \end{aligned}$$

$$\begin{aligned}
 &= \prod_{s=1}^S \prod_{t=1}^T \text{Poisson}(n_{s,t} | \exp(\alpha^{(\mu)} + \mathbf{X}_{(s,t)} \boldsymbol{\beta}^{(\mu)} + \phi_{s,t})) \times \\
 &\quad \text{Normal}(\boldsymbol{\phi}_1 | \mathbf{0}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
 &\quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t | \eta \boldsymbol{\phi}_{t-1}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
 &\quad \prod_{j=1}^p \text{Normal}\left(\beta_j^{(\mu)} | 0, \frac{\tau^2 c^2 \lambda_j^2}{c^2 + \tau^2 \lambda_j^2}\right) \times \text{Cauchy}^+(\lambda_j | 0, 1) \times \\
 &\text{Normal}^+(\sigma^{(\phi)} | 0, 1^2) \times \text{Beta}(\eta | 1, 1) \times \text{Inv-Gamma}(c^2 | 2.5, 10) \times \\
 &\quad \text{Normal}^+(\tau | 0, 5^2) \times \text{Normal}(\alpha^{(\mu)} | 0, 5^2).
 \end{aligned}$$

766 **Negative binomial wildfire count model**

767 We used the following parameterization of the negative binomial distribution:

$$[n|\mu, \delta] = \binom{n + \delta - 1}{n} \left(\frac{\mu}{\mu + \delta}\right)^n \left(\frac{\delta}{\mu + \delta}\right)^\delta,$$

768 where  $\mu$  is the mean, and  $\delta$  is a dispersion parameter.

769 The unnormalized posterior density of this model is:

$$\begin{aligned} & [\boldsymbol{\beta}^{(\mu)}, \boldsymbol{\alpha}^{(\mu)}, \boldsymbol{\phi}, \sigma^{(\phi)}, \eta, \boldsymbol{\lambda}, c, \tau, \delta \mid \mathbf{N}] \propto \\ & \prod_{s=1}^S \prod_{t=1}^T [n_{s,t} | \boldsymbol{\beta}^{(\mu)}, \boldsymbol{\alpha}^{(\mu)}, \boldsymbol{\phi}_{s,t}, \delta] \times \\ & [\boldsymbol{\phi}_1 | \sigma^{(\phi)}] \prod_{t=2}^T [\boldsymbol{\phi}_t | \boldsymbol{\phi}_{t-1}, \sigma^{(\phi)}, \eta] \times \\ & \prod_{j=1}^p [\beta_j^{(\mu)} | \lambda_j, c, \tau] [\lambda_j] \times \\ & [\sigma^{(\phi)}] [\eta] [c] [\tau] [\boldsymbol{\alpha}^{(\mu)}] [\delta] \end{aligned}$$

$$\begin{aligned} & = \prod_{s=1}^S \prod_{t=1}^T \text{Negative Binomial}(n_{s,t} | \exp(\boldsymbol{\alpha}^{(\mu)} + \mathbf{X}_{(s,t)} \boldsymbol{\beta}^{(\mu)} + \boldsymbol{\phi}_{s,t}), \delta) \times \\ & \text{Normal}(\boldsymbol{\phi}_1 | \mathbf{0}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\ & \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t | \eta \boldsymbol{\phi}_{t-1}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\ & \prod_{j=1}^p \text{Normal}\left(\beta_j^{(\mu)} | 0, \frac{\tau^2 c^2 \lambda_j^2}{c^2 + \tau^2 \lambda_j^2}\right) \times \text{Cauchy}^+(\lambda_j | 0, 1) \times \\ & \text{Normal}^+(\sigma^{(\phi)} | 0, 1^2) \times \text{Beta}(\eta | 1, 1) \times \text{Inv-Gamma}(c^2 | 2.5, 10) \times \\ & \text{Normal}^+(\tau | 0, 5^2) \times \text{Normal}(\boldsymbol{\alpha}^{(\mu)} | 0, 5^2) \times \text{Normal}^+(\delta | 0, 5^2). \end{aligned}$$

770 **Zero-inflated Poisson wildfire count model**

771 We used the following parameterization of the zero-inflated Poisson distribution:

$$[n|\mu, \pi] = I_{n=0}(1 - \pi + \pi e^{-\mu}) + I_{n>0}\pi \frac{\mu^n e^{-\mu}}{n!},$$

772 where  $\mu$  is the Poisson mean, and  $1 - \pi$  is the probability of an extra zero.

773 The unnormalized posterior density of this model is:

$$\begin{aligned} & [\boldsymbol{\beta}^{(\mu)}, \alpha^{(\mu)}, \boldsymbol{\beta}^{(\pi)}, \alpha^{(\pi)}, \boldsymbol{\phi}^{(\mu)}, \boldsymbol{\sigma}^{(\phi, \mu)}, \boldsymbol{\eta}^{(\mu)}, \boldsymbol{\phi}^{(\pi)}, \boldsymbol{\sigma}^{(\phi, \pi)}, \boldsymbol{\eta}^{(\pi)}, \boldsymbol{\lambda}, c, \tau, \rho | \mathbf{N}] \propto \\ & \prod_{s=1}^S \prod_{t=1}^T [n_{s,t} | \boldsymbol{\beta}^{(\mu)}, \alpha^{(\mu)}, \boldsymbol{\beta}^{(\pi)}, \alpha^{(\pi)}, \phi_{s,t}^{(\mu)}, \phi_{s,t}^{(\pi)}] \times \\ & [\phi_1^{(\mu)} | \boldsymbol{\sigma}^{(\phi, \mu)}] \prod_{t=2}^T [\phi_t^{(\mu)} | \phi_{t-1}^{(\mu)}, \boldsymbol{\sigma}^{(\phi, \mu)}, \boldsymbol{\eta}^{(\mu)}] \times \\ & [\phi_1^{(\pi)} | \boldsymbol{\sigma}^{(\phi, \pi)}] \prod_{t=2}^T [\phi_t^{(\pi)} | \phi_{t-1}^{(\pi)}, \boldsymbol{\sigma}^{(\phi, \pi)}, \boldsymbol{\eta}^{(\pi)}] \times \\ & \prod_{j=1}^p [\beta_j^{(\mu)}, \beta_j^{(\pi)} | \lambda_j, c, \tau, \rho] [\lambda_j] \times \\ & [\boldsymbol{\sigma}^{(\phi, \mu)}] [\boldsymbol{\sigma}^{(\phi, \pi)}] [\boldsymbol{\eta}^{(\mu)}] [\boldsymbol{\eta}^{(\pi)}] [\alpha^{(\mu)}] [\alpha^{(\pi)}] [\rho] \prod_{m=1}^2 [c_m] [\tau_m] \end{aligned}$$

$$\begin{aligned}
&= \prod_{s=1}^S \prod_{t=1}^T \text{ZIP}(n_{s,t} | e^{\alpha^{(\mu)} + \mathbf{X}_{(s,t)} \boldsymbol{\beta}^{(\mu)} + \phi_{s,t}^{(\mu)}}, \text{logit}^{-1}(\alpha^{(\pi)} + \mathbf{X}_{(s,t)} \boldsymbol{\beta}^{(\pi)} + \phi_{s,t}^{(\pi)})) \times \\
&\quad \text{Normal}(\boldsymbol{\phi}_1^{(\mu)} | \mathbf{0}, ((\sigma^{(\phi, \mu)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
&\quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t^{(\mu)} | \eta^{(\mu)} \boldsymbol{\phi}_{t-1}^{(\mu)}, ((\sigma^{(\phi, \mu)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
&\quad \text{Normal}(\boldsymbol{\phi}_1^{(\pi)} | \mathbf{0}, ((\sigma^{(\phi, \pi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
&\quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t^{(\pi)} | \eta^{(\pi)} \boldsymbol{\phi}_{t-1}^{(\pi)}, ((\sigma^{(\phi, \pi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
&\quad \prod_{j=1}^p \text{N} \left( \begin{pmatrix} \beta_j^{(\mu)} \\ \beta_j^{(\pi)} \end{pmatrix} \middle| \mathbf{0}, \begin{pmatrix} \tau_1^2 \frac{c_1^2 \lambda_j^2}{c_1^2 + \tau_1^2 \lambda_j^2} & \rho \tau_1 \tau_2 \sqrt{\frac{c_1^2 \lambda_j^2}{c_1^2 + \tau_1^2 \lambda_j^2}} \sqrt{\frac{c_2^2 \lambda_j^2}{c_2^2 + \tau_2^2 \lambda_j^2}} \\ \rho \tau_1 \tau_2 \sqrt{\frac{c_1^2 \lambda_j^2}{c_1^2 + \tau_1^2 \lambda_j^2}} \sqrt{\frac{c_2^2 \lambda_j^2}{c_2^2 + \tau_2^2 \lambda_j^2}} & \tau_2^2 \frac{c_2^2 \lambda_j^2}{c_2^2 + \tau_2^2 \lambda_j^2} \end{pmatrix} \right) \times \\
&\quad \prod_{j=1}^p \text{Cauchy}^+(\lambda_j | 0, 1) \times \\
&\quad \text{Normal}^+(\sigma^{(\phi, \mu)} | 0, 1^2) \times \text{Normal}^+(\sigma^{(\phi, \pi)} | 0, 1^2) \times \\
&\quad \text{Beta}(\eta^{(\mu)} | 1, 1) \times \text{Beta}(\eta^{(\pi)} | 1, 1) \times \\
&\quad \text{Normal}(\alpha^{(\mu)} | 0, 5^2) \times \text{Normal}(\alpha^{(\pi)} | 0, 5^2) \times \text{LKJ}(\rho | 3) \times \\
&\quad \prod_{m=1}^2 \text{Inv-Gamma}(c_m^2 | 2.5, 10) \times \text{Normal}^+(\tau_m | 0, 5^2).
\end{aligned}$$



774 **Zero-inflated negative binomial wildfire count model**

775 We used the following parameterization of the zero-inflated negative binomial distribution:

$$[n|\mu, \delta, \pi] = I_{n=0}(1 - \pi + \pi\left(\frac{\delta}{\mu + \delta}\right)^\delta) + I_{n>0}\binom{n + \delta - 1}{n}\left(\frac{\mu}{\mu + \delta}\right)^n\left(\frac{\delta}{\mu + \delta}\right)^\delta,$$

776 where  $\mu$  is the negative binomial mean,  $\delta$  is the negative binomial dispersion, and  $\pi$  is the probability of an extra zero.

777  $1 - \pi$  is the probability of an extra zero.

778 The unnormalized posterior density of this model is:

$$\begin{aligned} & [\boldsymbol{\beta}^{(\mu)}, \alpha^{(\mu)}, \boldsymbol{\beta}^{(\pi)}, \alpha^{(\pi)}, \boldsymbol{\phi}^{(\mu)}, \boldsymbol{\sigma}^{(\phi, \mu)}, \boldsymbol{\eta}^{(\mu)}, \boldsymbol{\phi}^{(\pi)}, \boldsymbol{\sigma}^{(\phi, \pi)}, \boldsymbol{\eta}^{(\pi)}, \boldsymbol{\lambda}, c, \tau, \rho, \delta | \mathbf{N}] \propto \\ & \prod_{s=1}^S \prod_{t=1}^T [n_{s,t} | \boldsymbol{\beta}^{(\mu)}, \alpha^{(\mu)}, \boldsymbol{\beta}^{(\pi)}, \alpha^{(\pi)}, \phi_{s,t}^{(\mu)}, \phi_{s,t}^{(\pi)}, \delta] \times \\ & [\phi_1^{(\mu)} | \boldsymbol{\sigma}^{(\phi, \mu)}] \prod_{t=2}^T [\phi_t^{(\mu)} | \phi_{t-1}^{(\mu)}, \boldsymbol{\sigma}^{(\phi, \mu)}, \boldsymbol{\eta}^{(\mu)}] \times \\ & [\phi_1^{(\pi)} | \boldsymbol{\sigma}^{(\phi, \pi)}] \prod_{t=2}^T [\phi_t^{(\pi)} | \phi_{t-1}^{(\pi)}, \boldsymbol{\sigma}^{(\phi, \pi)}, \boldsymbol{\eta}^{(\pi)}] \times \\ & \prod_{j=1}^p [\beta_j^{(\mu)}, \beta_j^{(\pi)} | \lambda_j, c, \tau, \rho] [\lambda_j] \times \\ & [\boldsymbol{\sigma}^{(\phi, \mu)}][\boldsymbol{\sigma}^{(\phi, \pi)}][\boldsymbol{\eta}^{(\mu)}][\boldsymbol{\eta}^{(\pi)}][\alpha^{(\mu)}][\alpha^{(\pi)}][\rho][\delta] \prod_{m=1}^2 [c_m][\tau_m]. \end{aligned}$$

$$\begin{aligned}
&= \prod_{s=1}^S \prod_{t=1}^T \text{ZINB}(n_{s,t} | e^{\alpha^{(\mu)} + \mathbf{X}_{(s,t)} \boldsymbol{\beta}^{(\mu)} + \phi_{s,t}^{(\mu)}}, \delta, \text{logit}^{-1}(\alpha^{(\pi)} + \mathbf{X}_{(s,t)} \boldsymbol{\beta}^{(\pi)} + \phi_{s,t}^{(\pi)})) \times \\
&\quad \text{Normal}(\boldsymbol{\phi}_1^{(\mu)} | \mathbf{0}, ((\sigma^{(\phi, \mu)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
&\quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t^{(\mu)} | \eta^{(\mu)} \boldsymbol{\phi}_{t-1}^{(\mu)}, ((\sigma^{(\phi, \mu)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
&\quad \text{Normal}(\boldsymbol{\phi}_1^{(\pi)} | \mathbf{0}, ((\sigma^{(\phi, \pi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
&\quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t^{(\pi)} | \eta^{(\pi)} \boldsymbol{\phi}_{t-1}^{(\pi)}, ((\sigma^{(\phi, \pi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\
&\quad \prod_{j=1}^p \text{N} \left( \begin{pmatrix} \beta_j^{(\mu)} \\ \beta_j^{(\pi)} \end{pmatrix} \middle| \mathbf{0}, \begin{pmatrix} \tau_1^2 \frac{c_1^2 \lambda_j^2}{c_1^2 + \tau_1^2 \lambda_j^2} & \rho \tau_1 \tau_2 \sqrt{\frac{c_1^2 \lambda_j^2}{c_1^2 + \tau_1^2 \lambda_j^2}} \sqrt{\frac{c_2^2 \lambda_j^2}{c_2^2 + \tau_2^2 \lambda_j^2}} \\ \rho \tau_1 \tau_2 \sqrt{\frac{c_1^2 \lambda_j^2}{c_1^2 + \tau_1^2 \lambda_j^2}} \sqrt{\frac{c_2^2 \lambda_j^2}{c_2^2 + \tau_2^2 \lambda_j^2}} & \tau_2^2 \frac{c_2^2 \lambda_j^2}{c_2^2 + \tau_2^2 \lambda_j^2} \end{pmatrix} \right) \times \\
&\quad \prod_{j=1}^p \text{Cauchy}^+(\lambda_j | 0, 1) \times \\
&\quad \text{Normal}^+(\sigma^{(\phi, \mu)} | 0, 1^2) \times \text{Normal}^+(\sigma^{(\phi, \pi)} | 0, 1^2) \times \\
&\quad \text{Beta}(\eta^{(\mu)} | 1, 1) \times \text{Beta}(\eta^{(\pi)} | 1, 1) \times \\
&\quad \text{Normal}(\alpha^{(\mu)} | 0, 5^2) \times \text{Normal}(\alpha^{(\pi)} | 0, 5^2) \times \text{LKJ}(\rho | 3) \times \text{Normal}^+(\delta | 0, 5^2) \times \\
&\quad \prod_{m=1}^2 \text{Inv-Gamma}(c_m^2 | 2.5, 10) \times \text{Normal}^+(\tau_m | 0, 5^2).
\end{aligned}$$

779 **Generalized Pareto/Lomax burn area model**

780 We used the following parameterization of the GPD/Lomax distribution:

$$[y|\sigma, \kappa] = \frac{1}{\sigma} \left( \frac{\kappa y}{\sigma} + 1 \right)^{-(\kappa+1)\kappa^{-1}},$$

781 where  $\kappa$  is a shape parameter and  $\sigma$  is a scale parameter.

782 The unnormalized posterior density of this model is:

$$\begin{aligned} & [\boldsymbol{\beta}, \alpha, \boldsymbol{\phi}, \sigma^{(\phi)}, \eta, \kappa^{(L)}, \boldsymbol{\lambda}, c, \tau | \mathbf{y}] \propto \\ & \prod_{i=1}^{n_{\text{tot}}} [y_i | \boldsymbol{\beta}, \alpha, \phi_{s_i, t_i}, \kappa^{(L)}] \times \\ & [\boldsymbol{\phi}_1 | \sigma^{(\phi)}] \prod_{t=2}^T [\boldsymbol{\phi}_t | \boldsymbol{\phi}_{t-1}, \sigma^{(\phi)}, \eta] \times \\ & \prod_{j=1}^p [\beta_j | \lambda_j, c, \tau] [\lambda_j] \times \\ & [\alpha][c][\tau][\kappa^{(L)}][\eta][\sigma^{(\phi)}] \\ & = \prod_{i=1}^{n_{\text{tot}}} \text{Lomax}(y_i | \kappa^{(L)}, e^{\alpha + \mathbf{X}_{(s_i, t_i)} \boldsymbol{\beta} + \phi_{s_i, t_i}}) \times \\ & \text{Normal}(\boldsymbol{\phi}_1 | \mathbf{0}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\ & \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t | \eta \boldsymbol{\phi}_{t-1}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\ & \prod_{j=1}^p \text{Normal}\left(\beta_j | 0, \frac{\tau^2 c^2 \lambda_j^2}{c^2 + \tau^2 \lambda_j^2}\right) \times \text{Cauchy}^+(\lambda_j | 0, 1) \times \\ & \text{Normal}(\alpha | 0, 5^2) \times \text{Inv-Gamma}(c^2 | 2.5, 10) \times \text{Normal}^+(\tau | 0, 5^2) \\ & \text{Normal}^+(\kappa^{(L)} | 0, 5^2) \times \text{Beta}(\eta | 1, 1) \times \text{Normal}^+(\sigma^{(\phi)} | 0, 1^2). \end{aligned}$$

### 783 Tapered Pareto burn area model

784 We used the following parameterization of the tapered Pareto distribution:

$$[y|\kappa, \nu] = \left(\frac{\kappa}{y} + \frac{1}{\nu}\right) \exp(-x/\nu),$$

785 where  $\kappa$  is a shape parameter and  $\nu$  a taper parameter.

786 The unnormalized posterior density of this model is:

$$\begin{aligned} & [\boldsymbol{\beta}, \alpha, \boldsymbol{\phi}, \sigma^{(\phi)}, \eta, \nu, \boldsymbol{\lambda}, c, \tau | \mathbf{y}] \propto \\ & \prod_{i=1}^{n_{\text{tot}}} [y_i | \boldsymbol{\beta}, \alpha, \phi_{s_i, t_i}, \nu] \times \\ & [\boldsymbol{\phi}_1 | \sigma^{(\phi)}] \prod_{t=2}^T [\boldsymbol{\phi}_t | \boldsymbol{\phi}_{t-1}, \sigma^{(\phi)}, \eta] \times \\ & \prod_{j=1}^p [\beta_j | \lambda_j, c, \tau] [\lambda_j] \times \\ & [\alpha] [c] [\tau] [\nu] [\eta] [\sigma^{(\phi)}] \\ & = \prod_{i=1}^{n_{\text{tot}}} \text{Tapered Pareto}(y_i | e^{\alpha + \mathbf{X}_{(s_i, t_i)} \boldsymbol{\beta} + \phi_{s_i, t_i}}, \nu) \times \\ & \quad \text{Normal}(\boldsymbol{\phi}_1 | \mathbf{0}, ((\sigma^{(\phi)})^{-2} (\mathbf{D} - \mathbf{W}))^{-1}) \times \\ & \quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t | \eta \boldsymbol{\phi}_{t-1}, ((\sigma^{(\phi)})^{-2} (\mathbf{D} - \mathbf{W}))^{-1}) \times \\ & \quad \prod_{j=1}^p \text{Normal}\left(\beta_j | 0, \frac{\tau^2 c^2 \lambda_j^2}{c^2 + \tau^2 \lambda_j^2}\right) \times \text{Cauchy}^+(\lambda_j | 0, 1) \times \\ & \quad \text{Normal}(\alpha | 0, 5^2) \times \text{Inv-Gamma}(c^2 | 2.5, 10) \times \text{Normal}^+(\tau | 0, 5^2) \times \\ & \quad \text{Cauchy}^+(\nu | 0, 1) \times \text{Beta}(\eta | 1, 1) \times \text{Normal}^+(\sigma^{(\phi)} | 0, 1^2). \end{aligned}$$

787 **Lognormal burn area model**

788 We used the following parameterization of the lognormal distribution:

$$[y|\mu, \sigma] = \frac{1}{y} \frac{1}{\sigma\sqrt{2\pi}} \exp\left(-\frac{(\log(y) - \mu)^2}{2\sigma^2}\right),$$

789 where  $\mu$  and  $\sigma$  are location and scale parameters, respectively.

790 The unnormalized posterior density of this model is:

$$\begin{aligned} [\boldsymbol{\beta}, \alpha, \boldsymbol{\phi}, \sigma^{(\phi)}, \eta, \sigma, \boldsymbol{\lambda}, c, \tau | \mathbf{y}] &\propto \\ &\prod_{i=1}^{n_{\text{tot}}} [y_i | \beta, \alpha, \phi_{s_i, t_i}, \sigma] \times \\ &[\boldsymbol{\phi}_1 | \sigma^{(\phi)}] \prod_{t=2}^T [\boldsymbol{\phi}_t | \boldsymbol{\phi}_{t-1}, \sigma^{(\phi)}, \eta] \times \\ &\prod_{j=1}^p [\beta_j | \lambda_j, c, \tau] [\lambda_j] \times \\ &[\alpha][c][\tau][\sigma][\eta][\sigma^{(\phi)}] \\ &= \prod_{i=1}^{n_{\text{tot}}} \text{Lognormal}(y_i | \alpha + \mathbf{X}_{(s_i, t_i)} \boldsymbol{\beta} + \phi_{s_i, t_i}, \sigma) \times \\ &\quad \text{Normal}(\boldsymbol{\phi}_1 | \mathbf{0}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\ &\quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t | \eta \boldsymbol{\phi}_{t-1}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\ &\quad \prod_{j=1}^p \text{Normal}\left(\beta_j | 0, \frac{\tau^2 c^2 \lambda_j^2}{c^2 + \tau^2 \lambda_j^2}\right) \times \text{Cauchy}^+(\lambda_j | 0, 1) \times \\ &\quad \text{Normal}(\alpha | 0, 5^2) \times \text{Inv-Gamma}(c^2 | 2.5, 10) \times \text{Normal}^+(\tau | 0, 5^2) \times \\ &\quad \text{Normal}^+(\sigma | 0, 5^2) \times \text{Beta}(\eta | 1, 1) \times \text{Normal}^+(\sigma^{(\phi)} | 0, 1^2). \end{aligned}$$

791 **Gamma burn area model**

792 We used the following parameterization of the gamma distribution:

$$[y|\kappa, \sigma] = \frac{1}{\Gamma(\kappa)\sigma^\kappa} y^{\kappa-1} \exp(-y/\sigma),$$

793 where  $\kappa$  is a shape parameter and  $\sigma$  a scale parameter.

794 The unnormalized posterior density of this model is:

$$\begin{aligned} & [\boldsymbol{\beta}, \alpha, \boldsymbol{\phi}, \sigma^{(\phi)}, \eta, \kappa, \boldsymbol{\lambda}, c, \tau | \mathbf{y}] \propto \\ & \prod_{i=1}^{n_{\text{tot}}} [y_i | \beta, \alpha, \phi_{s_i, t_i}, \kappa] \times \\ & [\boldsymbol{\phi}_1 | \sigma^{(\phi)}] \prod_{t=2}^T [\boldsymbol{\phi}_t | \boldsymbol{\phi}_{t-1}, \sigma^{(\phi)}, \eta] \times \\ & \prod_{j=1}^p [\beta_j | \lambda_j, c, \tau] [\lambda_j] \times \\ & [\alpha] [c] [\tau] [\kappa] [\eta] [\sigma^{(\phi)}] \\ & = \prod_{i=1}^{n_{\text{tot}}} \text{Gamma}(y_i | \kappa, \kappa / \exp(\alpha + \mathbf{X}_{(s_i, t_i)} \boldsymbol{\beta} + \phi_{s_i, t_i})) \times \\ & \quad \text{Normal}(\boldsymbol{\phi}_1 | \mathbf{0}, ((\sigma^{(\phi)})^{-2} (\mathbf{D} - \mathbf{W}))^{-1}) \times \\ & \quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t | \eta \boldsymbol{\phi}_{t-1}, ((\sigma^{(\phi)})^{-2} (\mathbf{D} - \mathbf{W}))^{-1}) \times \\ & \quad \prod_{j=1}^p \text{Normal}\left(\beta_j | 0, \frac{\tau^2 c^2 \lambda_j^2}{c^2 + \tau^2 \lambda_j^2}\right) \times \text{Cauchy}^+(\lambda_j | 0, 1) \times \\ & \quad \text{Normal}(\alpha | 0, 5^2) \times \text{Inv-Gamma}(c^2 | 2.5, 10) \times \text{Normal}^+(\tau | 0, 5^2) \times \\ & \quad \text{Normal}^+(\kappa | 0, 5^2) \times \text{Beta}(\eta | 1, 1) \times \text{Normal}^+(\sigma^{(\phi)} | 0, 1^2). \end{aligned}$$

795 **Weibull burn area model**

796 We used the following parameterization of the Weibull distribution:

$$[y|\kappa, \sigma] = \frac{\kappa}{\sigma} \left(\frac{y}{\sigma}\right)^{\kappa-1} \exp\left(-\left(\frac{y}{\sigma}\right)^\kappa\right),$$

797 where  $\kappa$  is a shape parameter and  $\sigma$  is a scale parameter.

798 The unnormalized posterior density of this model is:

$$\begin{aligned} [\boldsymbol{\beta}, \alpha, \boldsymbol{\phi}, \sigma^{(\phi)}, \eta, \kappa, \lambda, c, \tau | \mathbf{y}] &\propto \\ &\prod_{i=1}^{n_{\text{tot}}} [y_i | \beta, \alpha, \phi_{s_i, t_i}, \kappa] \times \\ &[\boldsymbol{\phi}_1 | \sigma^{(\phi)}] \prod_{t=2}^T [\boldsymbol{\phi}_t | \boldsymbol{\phi}_{t-1}, \sigma^{(\phi)}, \eta] \times \\ &\prod_{j=1}^p [\beta_j | \lambda_j, c, \tau] [\lambda_j] \times \\ &[\alpha][c][\tau][\kappa][\eta][\sigma^{(\phi)}] \\ &= \prod_{i=1}^{n_{\text{tot}}} \text{Weibull}(y_i | \kappa, \exp(\alpha + \mathbf{X}_{(s_i, t_i)} \boldsymbol{\beta} + \phi_{s_i, t_i})) \times \\ &\quad \text{Normal}(\boldsymbol{\phi}_1 | \mathbf{0}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\ &\quad \prod_{t=2}^T \text{Normal}(\boldsymbol{\phi}_t | \eta \boldsymbol{\phi}_{t-1}, ((\sigma^{(\phi)})^{-2}(\mathbf{D} - \mathbf{W}))^{-1}) \times \\ &\quad \prod_{j=1}^p \text{Normal}\left(\beta_j | 0, \frac{\tau^2 c^2 \lambda_j^2}{c^2 + \tau^2 \lambda_j^2}\right) \times \text{Cauchy}^+(\lambda_j | 0, 1) \times \\ &\quad \text{Normal}(\alpha | 0, 5^2) \times \text{Inv-Gamma}(c^2 | 2.5, 10) \times \text{Normal}^+(\tau | 0, 5^2) \times \\ &\quad \text{Normal}^+(\kappa | 0, 5^2) \times \text{Beta}(\eta | 1, 1) \times \text{Normal}^+(\sigma^{(\phi)} | 0, 1^2). \end{aligned}$$