# <sup>1</sup> One size does not fit all: Caste and sex differences in

# <sup>2</sup> the response of bumblebees (Bombus impatiens) to

**3** chronic oral neonicotinoid exposure.

4	
5	
6	
7	
8	Authors: Melissa W. Mobley <sup>1</sup> and Robert J. Gegear <sup>1</sup> ,*
9	
10	
11	
12	
13	
14	
15	
16	
17	Affiliation:
18	
19	1. Department of Biology and Biotechnology, Worcester Polytechnic Institute,
20	Worcester, 01609-2280, U.S.A
21	*, Corresponding author. Email: rgegear@wpi.edu
22	
23	

## 24 Abstract

25 Neonicotinoid insecticides have been implicated in the rapid global decline of bumblebees over recent years, particularly in agricultural and urban areas. While there is much known about 26 27 neonicotinoid toxicity effects at the colony stage of the bumblebee annual cycle, far less is known 28 about such effects at other stages critical for the maintenance of wild populations. In the present 29 work, individual-based feeding assays were used to show that chronic consumption of the widely 30 used neonicotinoid clothianidin at a field-realistic average rate of 3.6 and 4.0 ng/g bee/day reduces survival of queen and male bumblebees, respectively, within a 7-day period whereas 31 consumption at a similar rate of 3.9 ng/g·bee/day in workers had no effect on survival. To test 32 33 the hypothesis that males have a lower tolerance for oral clothianidin exposure than workers due to their haploid genetic status, RNAseg analysis was used to compare the transcriptomic 34 35 responses of workers and males to chronic intake of clothianidin at a sub-lethal dose of 0.37ng/day for 5 days. Surprisingly, only 25/100 clothianidin-induced putative detoxification 36 37 genes had expression levels that differed in a sex-specific manner, with 17 genes showing increased expression in workers. Sub-lethal clothianidin intake also induced changes in genes 38 associated with a variety of other major biological functions, including locomotion, reproduction, 39 40 and immunity. Collectively, these results suggest that chronic oral toxicity effects of neonicotinoids are greatest during mating and nest establishment phases of the bumblebee life 41 42 cycle. Chronic oral toxicity tests on males and queens are therefore required in order to fully assess the impact of neonicotinoids on wild bumblebee populations. 43

44

45

# 46 Introduction

47 Bumble bees have rapidly declined in abundance, species richness, and geographic distribution on a global scale over recent years [1, 2]. In North America alone, nearly half of the 48 49 bumblebee species have reached historically low numbers [3-5], including one species, Bombus 50 affinis, which was recently listed as an endangered species by the U.S. Fish and Wildlife Service. 51 From an ecological perspective, these declines pose a significant threat to the function and 52 diversity of temperate ecosystems due to the critical keystone role that bumblebees play as pollinators of native flowering plants. Reports of parallel reductions in bee-pollinated plant species 53 suggest that pollinator decline-mediated effects on wildlife diversity at higher trophic levels may 54 55 already be well underway [1, 6, 7]. It is therefore imperative that all anthropogenic stressors contributing to species decline in bumblebees be identified and mitigated as soon as possible. 56

57 One stressor thought to contribute to wild bumblebee decline in urban and agricultural 58 areas is a newly developed class of pesticides called 'neonicotinoids' [8]. From a pest 59 management perspective, neonicotinoids are a highly effective form of insect control because 60 they specifically target nicotinic acetylcholine receptors in the insect central nervous system, 61 leading to paralysis and death [9]. Unlike most other pesticide classes, neonicotinoids are also 62 systemic, meaning that they are readily taken up by the roots and distributed throughout the entire 63 plant, thereby protecting it from pest damage over the entire growing season [10].

Despite these benefits, neonicotinoids pose a significant threat to beneficial insects such as pollinators because they are also translocated to floral nectar and pollen, presenting a route of oral exposure. Of particular concern to wild pollinator populations is the fact that neonicotinoids can be transported away from the area of application to adjacent natural areas [11, 12] and then contaminate wildflower resources [2, 13, 14]. For example, clothianidin, one of the newest and most potent neonicotinoid formulations, has been detected in numerous wildflower species heavily visited by insect pollinators at concentrations ranging from 4 to 215 ppb [2, 14].

Compounding this issue, neonicotinoid residues can persist in the soil for years after a single application [15-17], and increase in concentration with repeated annual application [18]. Wild bumblebees and other insect pollinators therefore have the potential to be orally exposed to neonicotinoids at many life stages occurring outside of the blooming period and geographic location of the targeted plant species.

76 Yet, the overwhelming majority of neonicotinoid toxicity studies on bumblebees to date have focused only on the colony stage, relying on metrics such as worker survival, larval growth, 77 78 and reproductive output to estimate the potential impact of exposure on wild populations [19]. 79 While such colony-focused risk assessment protocols may be adequate in the context of crop 80 pollination, determining the impact of widespread neonicotinoid use on threatened bumblebee species in an ecological context requires more comprehensive assessment protocols that 81 82 consider potential impacts on queens and males, whose survival during mating, overwintering, 83 and nest establishment stages of the life cycle has a direct effect on population stability (Fig 1).

84

Fig 1. Annual cycle of social bumblebees. (1) Mated gueens emerge from their overwintering 85 site in the spring and forage while search for a suitable nest site. (2) Queens collect floral 86 87 resources and store them in the nest until the first brood of worker bees emerges. (3) Workers forage for the colony so that it can continue to produce more workers and later, gueens and males 88 (reproductives). (4) Virgin queens and males leave the nest to feed and search for mates. Once 89 90 mated, queens, which are the only individuals to survive through the winter months, continue to 91 feed until they find a suitable overwintering site where they reside until the spring, thus completing 92 the cycle.

93

94 Several morphological and physiological characteristics of queen, worker, and male 95 bumblebees suggest that individuals from each group may vary in their capacity to cope with 96 chronic oral neonicotinoid exposure. For example, males and workers have a smaller body size

97 than queens and therefore may experience mortality effects at lower neonicotinoid concentrations, as has been shown at the species level [20]. Males are also haploid whereas 98 queens and workers are diploid, which may limit gene products available for males to detoxify 99 100 Such 'haploid susceptibility' is known to occur in the context of neonicotinoids. 101 immunocompetence [21], but has yet to be studied in the context of metabolic resistance to 102 pesticides. Alternatively, males and queens have greater energetic demands due to their 103 reproductive status [22, 23] and therefore may have fewer resources available for detoxification 104 processes than sterile workers .

105 To explore these possibilities, the present study uses a novel individual-based feeding assay to directly compare the mortality response of queen, worker, and male Bombus impatiens 106 to chronic consumption of the widely used neonicotinoid clothianidin at field-realistic 107 108 concentrations of 5-10ppb. Assaying bumblebee test populations at the individual level provides 109 a much more robust estimate of chronic lethality level than more traditional assays on small 110 groups (e.g. microcolonies) because drug intake rates can be directly monitored in all test individuals and then easily standardized by adjusting for individual variation in body size [24]. Our 111 feeding experiments revealed that chronic lethality effects of clothianidin are much greater in 112 113 males than workers even when controlling for differences in body size. To test the potential role of male haploidy in driving this sex difference, we then used RNAseg analysis to compare the 114 transcriptomic responses of workers and males to consumption of clothianidin at a sub-lethal daily 115 116 dose over a 5-day period. Following previous work on bumblebees by [25], putative genes related 117 to clothianidin detoxification were identified and classified based on the two well-characterized xenobiotic detoxification phases in insects, with Phase 1 processes metabolizing the xenobiotic 118 119 to less toxic compounds through oxidation, hydrolysis, and reduction reactions, and Phase 2 120 processes breaking down the metabolites produced from Phase 1 processes through conjugation 121 reactions. Based on the findings of [25] that another bumblebee species (B. huntii) has fewer constitutively expressed detoxification genes than workers, it was predicted that male Bombus 122

*impatiens* would have fewer and/or lower expression of inducible detoxification genes comparedto workers.

125

# 126 **Results**

#### 127 Chronic oral toxicity effects of clothianidin differ among queen, worker,

#### 128 and male bumblebees

Chronic consumption of clothianidin at field-realistic concentrations increased mortality rates in 129 130 queen, worker, and male test populations (Fig 2). However, clothianidin concentrations needed 131 to produce such effects varied in a caste- and sex-dependent manner. Table 1 shows intake rates reducing survival in gueens, workers, and males controlling for differences in body size and 132 133 daily clothianidin dose. Compared to 0ppb controls, queens consuming clothianidin at a 134 concentration of 10ppb (mean intake rate  $\pm$  SD = 3.61 $\pm$ 0.71 ng/g bee/day) had reduced survival over the 7-day test period (X<sup>2</sup>=15.6, df=1, P<0.0001). While workers orally exposed to clothianidin 135 at a concentration of 10ppb (mean intake rate  $\pm$  SD = 5.48 $\pm$ 1.95 ng/g·bee/day) also had reduced 136 survival ( $X^2$ =54.7, df=1, P<0.0001), with the test population reaching 50% mortality at day 5. 137 Interestingly, worker consumption mean  $(\pm SD)$  intakes rates of 3.85  $(\pm 1.6)$  ng/g bee/day (7ppb 138 139 solution) had no effect of survival, indicating that workers are better able to cope with chronic oral clothianidin exposure than queens when controlling for consumption rate and body size. 140

141

Fig 2. Kaplan-Meier survival plots for *Bombus impatiens* (A) queens, (B) workers, and (C) males, orally exposed to test solutions containing different concentrations of clothianidin daily for 7 days. Black = 0ppb; red= 5 ppb; green = 7 ppb; blue = 10ppb. A Mantel-Cox Logrank analysis was used to test for differences in survival between each treatment group relative to the 0ppb control group over the 7-day period. \*\*, p<0.0001.

#### 147

148

149Table 1. Response of queen, worker, and male bumblebees (Bombus impatiens) to chronic150clothianidin exposure controlling for differences in body size and daily dose. Workers and151males consumed 75  $\mu$ L of solution per day and due to their larger size, queens consumed 200  $\mu$ L152of each solution per day. Yellow shading = no effect on survival; blue shading = reduced survival.153All values are shown as mean ± SD.

	Concentration of clothianidin solution						
	5ppb(ng/mL)		7ppb(ng/mL)		10ppb(ng/mL)		
	Body mass (g)	Intake rate (ng/g.bee/da y)	Body mass(g)	Intake rate (ng/g.bee/da y)	Body mass (g)	Intake rate (ng/g.bee/da y)	
Queens	0.58±0.1 3	1.79±0.39	0.59±0.1 1	2.45±0.48	0.57±0.1 1	3.61±0.71	
Worker s	0.16±0.0 5	2.60±0.83	0.15±0.0 4	3.85±1.6	0.15±0.0 4	5.48±1.95	
Males	0.14±0.0 4	2.94±0.86	0.14±0.0 4	4.01±1.08	0.13±0.0 3	6.05±1.35	

154

155

Similar to workers and queens, males consuming the 10 ppb clothianidin solution (mean intake rate  $\pm$  SD = 6.05  $\pm$ 1.35 ng/g·bee/day) had reduced survival compared to the 0ppb control (X<sup>2</sup>=38.8, df=1, P<0.0001); however, the test population reached 50% mortality after just 3 days. In addition, males consuming clothianidin at a concentration of 7 ppb (mean intake rate  $\pm$  SD = 4.01  $\pm$ 1.08 ng/g·bee/day) showed reduced survival compared to 0ppb controls (X<sup>2</sup>=17.7, df=1, 161 P<0.0001), with the test population reaching 50% mortality at day 6. The vast majority of queens,

workers, and males fully consumed the clothianidin solutions during testing (Table S1), indicating

that differences in mortality were caused by differences in the toxicity effects of clothianidin rather

than starvation (anti-feeding behavior).

165

#### 166 Sex differences in the transcriptomic response of bumblebees to sub-

#### 167 **lethal doses of clothianidin**

168 RNAseq analysis revealed that chronic oral exposure to clothianidin at a sub-lethal 169 concentration of 5ppb for 5 days affected expression of 147 genes in workers and 200 genes in 170 males. Of the 347 pooled clothianidin-induced genes (147 worker, 200 male), only 15 were 171 present in both sexes, yielding a total of 332 unique clothianidin-induced genes. Of these genes, 172 100 were classified as having a putative role in detoxification and the other 232 were associated 173 with other biological functions (Tables 2, S2, and S3).

174

175 Table 2. Sex-specific expression of inducible detoxification genes in bumblebee workers 176 and males orally exposed to clothianidin at a sub-lethal daily dose over 5 consecutive days. Phase 1 detoxification processes include oxidation, hydrolysis, and reduction reactions 177 178 while Phase 2 processes include conjugation reactions. Expression levels were determined through RNAseq analysis. Upper portion = genes with increased expression in workers; lower 179 portion = genes with increased expression in males. Clothianidin-inducible genes were initially 180 181 identified by comparing transcriptomic expression patterns between individuals fed either 5ppb clothianidin solution or the same volume of solution containing 1.4% DMSO (vehicle control) 182 183 solution separately for each sex and then pooling genes with significantly altered expression 184 together.

DETOXIFICATION GENES: INCREASED EXPRESSION IN WORKERS VERSUS MALES					
Gene ID	Gene ID Value increase			Phase	
	value	Increase			
BIMP10198	0.001	6.07	cytochrome P450 6k1-like	1	
BIMP23510	0.021	2.45	cytochrome P450 306a1	1	
BIMP20393	0.013	1.85	dipeptidase 1	1	
BIMP19106	0.004	1.84	fatty-acid amide hydrolase 2-like	1	
BIMP24166	0.011	1.71	U6 snRNA phosphodiesterase	1	
BIMP18394	0.019	1.69	N-sulphoglucosamine sulphohydrolase	1	
BIMP11124	0.021	1.63	probable cation-transporting ATPase 13A3	1	
BIMP13907	0.048	1.46	ribonuclease P protein subunit p25-like protein	1	
BIMP18346	0.027	1.45	cholinesterase transcript variant X3	1	
			short-chain dehydrogenase family 16C		
BIMP18541	0.020	1.40	member 6	1	
BIMP10339	0.005	3.25	probable serine/threonine-protein kinase yakA	2	
BIMP12444	0.023	3.12	hydroxymethylglutaryl-CoA synthase 1	2	
BIMP18546	0.011	2.12	putative glycerol kinase 5	2	
BIMP13921	0.029	1.82	lymphokine-activated killer T-cell protein kinase	2	
BIMP18798	0.033	1.67	probable deoxyhypusine synthase	2	
BIMP21529	0.027	1.48	death-associated protein kinase 1-like	2	
BIMP17254	0.046	1.44	glycerol kinase	2	
DETOXIFIC		GENES: INC	REASED EXPRESSION IN MALES VERSUS WO	ORKERS	
P- Fold Annotation		Annotation	Phase		
	value	increase	Annotation	i nase	
BIMP23965	0.003	5.66	lysozyme 2-like	1	

BIMP20243	0.030	5.61	alkaline phosphatase 4-like	1
BIMP21241	0.001	4.86	phospholipase A2-like	1
BIMP16635	0.019	3.97	probable cytochrome P450 304a1	1
BIMP17003	0.021	3.47	cytochrome P450 6k1-like	1
			inosine-5'-monophosphate dehydrogenase 1b-	
BIMP22299	0.024	1.83	like	1
BIMP23899	0.008	1.62	sarcosine dehydrogenase mitochondrial	1
BIMP14887	0.015	2.39	glutathione S-transferase 1-1	2

185

186

187

188 Only 77 of the 332 clothianidin-sensitive genes were differentially expressed between workers and males: 48 genes had higher expression in workers and 29 genes had higher 189 expression in males. Of these 77 genes, only 25 were associated with detoxification processes 190 191 17 (10 Phase 1 and 7 Phase 2) had higher expression in workers and 8 (7 Phase 1 and 1 Phase 192 2) had higher expression in males (Table 2; Fig 3). Of the remaining 52 genes associated with 193 biological functions other than detoxification, 31 genes had higher expression in workers and 21 194 genes had higher expression in males. Biological functions, fold changes, and accession 195 numbers for these 52 genes are shown in Table S2. Although many of these genes may play some as yet unknown functional role in detoxification, currently known biological functions include 196 197 immunity, learning and memory, reproduction, and signal transduction. We also found eight 198 genes whose function we could not identify based on current databases. Information for the 199 255/332 clothianidin-inducible genes with conserved expression between workers and males are 200 provided in Table S3.

```
201
```

Fig 3. Heatmap showing the 25 putative detoxification genes that were differentially expressed between workers and males after 5 consecutive days of oral exposure to clothianidin. Phase 1 detoxification processes include oxidation, hydrolysis, and reduction reactions while Phase 2 processes include conjugation reactions. Legend on the right shows relative fold change corresponding to each color.

- 207
- 208

# 209 **Discussion**

210 The present study shows that chronic consumption of field-realistic doses of the widely 211 used neonicotinoid pesticide clothianidin has differential effects on the survival of queen, worker, 212 and male bumblebees. Controlling for body size, queens and males consuming clothianidin at an 213 average rate of 3.6 and 4.0 ng clothianidin/g bee/day, respectively, had reduced survival over a 214 7-day period whereas a similar average consumption rate in workers (3.9 ng/g bee/day) had no Although clothianidin reduced queen and male survival at similar daily 215 mortality effect. 216 consumption rate, only male test populations reached 50% mortality over the testing period, 217 indicating that chronic oral lethality effects for clothianidin are slightly greater in males than 218 queens. The observed mortality response of all test groups to chronic clothianidin consumption 219 cannot be attributed to a starvation effect as food avoidance and vomiting behaviors were not 220 observed during the testing period. Collectively, these findings demonstrate that gueens and 221 males are much more vulnerable to clothianidin toxicity effects than workers, highlighting the 222 importance of incorporating queen and male survival assays into current colony-focused 'higher tier' protocols for assessing the potential risk of various neonicotinoid formulations to wild 223 pollinator populations [19]. In addition, the observed reductions in survival in *B. impatiens* workers 224 225 chronically consuming clothianidin at a concentration of 10ppb is notably less than the 20ppb

chronic lethality threshold reported for worker honeybees [26] as well as the 25ppb threshold reported for workers of the European bumblebee species *B. terrestris* [27], indicating that chronic oral lethality effects of clothianidin are relatively high in *B. impatiens*. Our findings thus add to the rapidly growing body of evidence that neonicotinoid oral toxicity effects can vary considerably among insect pollinator species [28-31].

231 In addition to increasing mortality, chronic oral exposure to clothianidin over a short time 232 period (5 days) was found to have profound sub-lethal effects on workers and males at the 233 genomic level, altering expression of a total of 332 genes associated with a wide variety of 234 biological functions, including immunity, neuronal signal transduction, locomotion, reproduction, and several fundamental cellular processes (Tables 2, S2, and S3). These results are consistent 235 with a growing number of studies at the organismal level showing that neonicotinoids can have 236 237 substantial sub-lethal effects on pollinators [32-36]. For example, clothianidin consumption induced changes to genes associated with reproductive output in males such as outer dense fiber 238 protein 3-like and lutropin-choriogonadotropic hormone receptor-like (Table 2), which is 239 240 consistent with previous work showing that oral exposure to imidacloprid reduced sperm viability 241 in males bees by 50% [37]. Under natural conditions, chronic exposure to low doses of 242 clothianidin in wildflower nectar thus has the potential to influence the dynamics of wild bumblebee 243 populations by reducing male and queen reproductive capacity and at slightly higher doses, reducing the number of mating individuals in the fall and the number of gueens establishing nests 244 245 in the spring.

Somewhat surprisingly, the genomic response of workers and males to sub-lethal clothianidin doses did not provide unequivocal evidence that the haploid genetic status of males renders them less able to cope with oral clothianidin exposure. Of the 100 putative detoxification genes that were affected by clothianidin, 75 had a similar expression level in workers and males. These similarly expressed genes included many 'classic' detoxification genes such as cytochrome P450s, glutathione S-transferases, and carboxylesterases [38]. There was, however, a strong

252 sex bias in direction predicted by the haploid susceptibility hypothesis for the 25 remaining 253 clothianidin-induced detoxification genes that differed between the sexes. Compared to males, 254 workers had higher expression in more than twice genes associated with Phase 1 hydrolysis reactions and remarkably, higher expression in seven times as many genes associated with 255 256 Phase 2 conjugation reactions. The fact that workers and males had a similar number of highly 257 expressed Phase 1 genes related to oxidation-reduction reactions suggests that are were equally capable of initially breaking down clothianidin into intermediate metabolites, but differed in the 258 259 capacity to subsequently transform such metabolites into non-toxic compounds (Phase 2 260 reactions). These results are similar to those reported for constitutively expressed detoxification genes in B. huntii [25], which is closely related to B. impatiens. However, it must be noted that Xu 261 et al. identified over 250 constitutively expressed genes for conjugation compared to the 18 262 inducible genes found in the present study. Thus, despite the fact that there were substantial 263 264 differences in the mortality response to oral clothianidin exposure, they were not strongly 265 correlated with the number and expression of inducible detoxification genes. Of course it is possible that haploidy in males results in the reduced expression of few key genes in the 266 267 clothianidin detoxification pathway, but further genetic manipulation experiments are required to 268 determine whether or not this is the case.

269 The present work demonstrates the potential conservation benefits of using a genomic approach to quantify exposure of individual pollinators to extremely low doses of neonicotinoids, 270 an idea that has been recently advocated by others [39]. Currently, one of the major problems 271 272 with studying neonicotinoid effects on wild bee populations is that individual tissue samples have concentrations that are too low to be quantified with a high level of accuracy and reliability using 273 conventional methodology and instrumentation. Indeed, the main laboratory at the United States 274 275 Department of Agriculture responsible for measuring neonicotinoid levels in tissue samples was 276 unable to detect clothianidin in our 5ppb in test solution, yet it produced a robust genetic response in both males and workers. To address this problem, the current study has identified a set of 277

278 clothianidin-induced genes that could potentially be used to develop highly sensitive 'eco-tools' 279 for assessing neonicotinoid exposure wild bumblebee populations. Such tools could also be used 280 to gauge sub-lethal effects on behavior (e.g. reduced cognitive function) and physiology (e.g. reduced sperm production), which are often difficult to assess in the field at the organismal level. 281 282 Comparative genomic approaches would also provide unique insight into why chronic oral toxicity levels for neonicotinoids vary considerably among (and within) wild bees species [31], greatly 283 accelerating efforts to determine the impact of neonicotinoid use on wild pollinators and the global 284 285 biodiversity that they ultimately support.

286

# 287 Materials and Methods

### 288 Experimental Design

**Bumble bees:** Virgin gueens, workers, males were obtained from commercial *Bombus* 289 impatiens colonies (Biobest Biological Systems, Learnington, Canada). 290 Colonies initially 291 contained approximately 100 workers and were subsequently supplied with 30% (stock) sugar solution (Grade A pure honey mixed with distilled water to the desired concentration, measured 292 with a Bellingham & Stanley hand-held refractometer (Suwanee, USA)) and wildflower pollen ad 293 294 *libitum* to facilitate continued production of workers and later production of gueens and males. 295 Multiple colonies were used across experimental trials to control for potential inter-colony differences in clothianidin sensitivity. 296

For chronic oral toxicity tests, queen, worker and male bees were collected and placed in a 4°C refrigerator. Once immobile, individuals were weighed to the nearest milligram, marked with acrylic paint for identification, and then placed in a 16oz plastic housing container with a screen lid for ventilation (Fig S1). Each lid had a circular opening with a removable plastic plug that was positioned directly above a plastic 'feeding bowl' fastened to the bottom of the container.

In this way, test solutions could be dispensed into the bowls with minimal disturbance to the bee.
Each container was also supplied with 1-1.5 grams of a pollen paste (a dough-like mixture of and
stock sugar solution), which was replaced as necessary.

All housing containers were kept in Percival Scientific (Perry, USA) environmental chambers set to 22-25°C, 50 % humidity, and a 12 hour light/dark cycle. Bees were fed 75  $\mu$ L (workers and males) or 200  $\mu$ L (queens) of untreated stock sugar solution once daily within 3 hours of the start of the light cycle. Preliminary experiments with stock sugar solution determined that these volumes are completely consumed by individuals over a 24-hour period while having no effect on survival or activity level. All handling of housing containers was conducted under red light in order to minimize additional stress on test bees.

312

**Clothianidin Solutions:** Analytical-grade clothianidin (Sigma-Aldrich, USA) was dissolved in dimethyl sulfoxide (DMSO), as has been done in other neonicotinoid toxicity studies [40], to a concentration of  $5 \times 10^6$  ng/mL or ppb (parts per billion) and serially diluted down to  $5 \times 10^2$  or  $5 \times 10^1$  ppb. Diluted clothianidin solution was added to stock sugar solution to create feeding solutions of 10, 7, and 5 ppb (ng/mL). Concentrations of clothianidin selected were well within the range present under field conditions [41]. Drug mixtures were prepared within 2 days of the start of a trial, stored in conical tubes in the dark at 4°C and only taken out for immediate use.

To confirm that test solutions contained the correct amount of clothianidin, fresh samples of 0, 5, 7, and 10 ppb test solutions (were sent to the U.S.D.A. Agricultural Marketing Service's National Science Laboratories Testing Division (Gastonia, USA). Clothianidin residues were measured using Mass Spectrometry with a LOQ of 5 +/- 1 ppb. Although the 5ppb test solution was below the minimal level of detection by the U.S.D.A. equipment, 7 and 10 ppb test solutions (3 technical replicates each) yielded mean values (+/-SE) of 4.1 +/- 3.5 and 7.3 +/-1.4 ppb, respectively.

#### 327

## **Test Procedure for feeding experiments**

Prior to each trial, all bees were fed stock sugar solution at the previously described volumes daily until there were no observed deaths over a 48-hour period. Individuals not consuming all of the sugar solution within a 24 hour period were also removed. This process ensured that bees were in good health prior to testing.

After the acclimation period, individuals were assigned to either the 0 ppb (control), 5 ppb, 333 334 7ppb, or 10ppb treatment group, with colony origin and bee size balanced among groups. 335 Location of housing containers within and between shelves inside the environmental chamber was randomized in order to control for potential positional effects on survival. Total number of 336 337 bees tested at each clothianidin concentration was as follows: 208 at 0ppb (88 workers, 58 queens, 62 males), 178 at 5ppb (70 workers, 58 queens, 50 males), 178 at 7 ppb (70 workers, 58 queens, 338 339 50 males), and 264 at 10 ppb (108 workers, 84 queens, 72 males). These numbers were obtained 340 through 11 independent worker, 8 gueen, and 12 male trials, with bees in each trial obtained from 341 at least two colonies. Given gueens required a greater volume of test solution to remain active 342 than workers and males due to their larger body size, we expressed exposure level for each group 343 as ng clothianidin consumed per gram of bee per day (ng/g·bee/day). Mean (±SD) mass in grams for queens =  $0.58\pm0.12$  g; workers =  $0.15\pm0.05$  g; males =  $0.14\pm0.04$  g. In this way, sex and 344 caste differences in clothianidin sensitivity could be established while controlling for differences 345 346 in body size and consumption rate.

Each day over a 7-day testing period, individual housing containers were removed from the environmental chamber within three hours after to the beginning of the light cycle, and the number of dead bees was recorded. In the rare case that a test individual was not dead but showed inverted orientation, extremity spasms, and unresponsiveness, it was recorded as dead and euthanized. Presence/absence of test solutions was also checked daily and feeding bowls

were replenished. Immediately following data collection, bees were returned to the environmentalchamber.

To confirm that consumption of dimethyl sulfoxide (DMSO) at the concentration used in 354 experiments had no effect on survival of queens, workers and males, additional feeding trials 355 356 were conducted with stock sugar solution containing 1.4% DMSO. As in the clothianidin test trials, 357 workers and males were fed 75 µL of solution per day and queens were fed 200 µL/day over a 7day period. Survival in the control (0ppb) group (see above for numbers in each group) was then 358 359 compared to survival in the 1.4% DMSO group for queens (24 treated individuals), workers (55 360 treated individuals), and males (87 treated individuals). Results of Mantel-Cox Log-rank analyses revealed that consuming 1.4% DMSO for seven days had no effect on survival of gueens ( $X^2=0.61$ . 361 df=1, P=0.43), workers (X<sup>2</sup>=0.26, df=1, P= 0.60) and males (X<sup>2</sup>=0.08, df=1, P=0.77). 362

363

**Statistical Analysis:** Survival data from each caste was compiled over the full 7-day clothianidin trial period. All remaining survivors were censored. For each clothianidin concentration, Kaplan-Meier survival curves [42] were generated for queens, workers, and males in GraphPad Prism 6 (La Jolla, USA) and a Mantel-Cox Log-rank analysis was used to test for differences in survival rates across all treatment groups and when significant, survival between a single clothianidin concentration and the 0ppb control.

370

### **Test procedure for RNAseq analysis**

To test the haploid susceptibility hypothesis that males have fewer gene products available to detoxify clothianidin than workers, RNAseq analysis was used to conduct genome-wide transcriptome analysis of workers and males after consumption 75uL of either the 5ppb test solution (0.374ng/bee; 3 workers and 3 males) or 1.4% DMSO in stock sugar solution (vehicle control; 3 workers and 3 males) once a day for 5 consecutive days. This clothianidin dose was

377 selected because it did not reduce worker and male survival in our initial series of feeding378 experiments.

379

Treatment and sample preparation: Individuals were euthanized at -20°C 24 hours 380 after the last dose (day 6), immediately dissected on ice by removing legs, wings, and proboscis, 381 and transferred to -80°C until ready for extraction. (This entire process was performed as quickly 382 383 as possible to minimize RNA degradation, usually completed within 5 minutes.) Individual sample 384 extractions were completed in an RNA-specified biosafety cabinet by crushing the bee body through pestle homogenization in Trizol (ThermoFischer Scientific, USA) following the suppliers 385 recommendations for RNA isolation followed by chloroform phase separation. The aqueous 386 phase containing RNA was collected, then purified using a PureLink RNA Mini Kit (ThermoFischer 387 388 Scientific, USA), ending with a 30 uL elution step. Concentration and purity (A260/280 and A260/230) were measured using Nanodrop. Purified RNA was stored at -80°C until use. 389

390

**RNA Sequencing:** Purified RNA samples were diluted to 100ng/uL in sterile, DNase-free, 391 RNase-free water and then shipped to Quick Biology (Pasadena, USA) for processing. Libraries 392 for RNA-Seq were prepared with a KAPA Stranded RNA-Seq Kit. The workflow consisted of 393 394 mRNA enrichment, cDNA generation, and end repair to generate blunt ends, A-tailing, adaptor ligation and PCR amplification. Different adaptors were used for multiplexing samples in one lane. 395 Sequencing was performed on Illumina Hiseg3000/4000 for a pair end 150 run. Data quality check 396 397 was done on an Illumina SAV (San Diego, USA). De-multiplexing was performed with Illumina 398 Bcl2fastg2 v 2.17 program. Reads were first mapped to the latest UCSC transcript set using STAR 399 version 2.4.1d and the gene expression level was quantified using an E/M algorithm in Partek 400 Genomics Suite (Partek, Inc. USA). Gene expression levels were normalized to total counts.

401 All four experimental groups produced a similar number of mean total reads: worker

402 controls  $4.951*10^6 \pm 1.845*10^5$ , worker clothianidin  $5.147*10^6 \pm 4.004*10^5$ , male controls 403  $5.520*10^6 \pm 8.654*10^4$ , and male clothianidin  $5.889*10^6 \pm 1.168*10^5$ . Reads were aligned to the 404 BIMP\_2.0 genome [43], resulting a high mean percentage of alignment in all four groups: worker 405 control 73.20%  $\pm 3.35$ , worker clothianidin  $80.41\% \pm 1.573$ , male control 79.15%  $\pm 1.511$ , male 406 clothianidin 77.58%  $\pm 1.938$ , corresponding with 15,896 identified genes. RNA sequencing data 407 quality and alignment for sex and treatment are shown in Table S4.

408

409 **Statistical Analysis:** Normalized gene expression levels were compared between groups 410 using the differential gene expression (GSA) algorithm in Partek Genomics Suite (Partek, Inc., 411 USA). To initially isolate all putative detoxification genes induced by clothianidin consumption, 412 test and vehicle control gene expression patterns were compared separately for each sex. A 413 gene was considered to be differentially expressed if comparisons yielded both a p-value less 414 than 0.05 and a fold change greater than 1.5. All differentially expressed genes in workers and 415 males were then pooled to create a list of genes with clothianidin-induced expression.

To test for sex differences in the expression level of pooled clothianidin-induced genes, 416 417 expression levels were then compared between workers and males in the test group (bees 418 chronically exposed to clothianidin at concentration of 5ppb). For this comparison, a gene was 419 considered to have sex-specific expression if P<0.05. Genes with this significance threshold were 420 then categorized as being involved in either Phase I (oxidation, hydrolysis and reduction) or Phase 421 II (conjugation with sugars, glutathione, amino acids, etc.) detoxification or assigned to some other biological process. This method of identifying and categorizing putative detoxification genes has 422 423 been used previously in bumblebees and other insects [25] and [44]. Genes were assigned a 424 particular biological processes based on information from the UniProt online gene database 425 (http://www.uniprot.org/).

426

#### 427 Acknowledgements:

428 **General:** We thank C. Emery and J. Letourneau for assistance with feeding and data collection. 429

430

### 431 **References**

- 432 1. Goulson D, Lye GC, Darvill B. Decline and conservation of bumble bees. Annu Rev Entomol.
- 433 2008;53:191-208.
- 434 2. David A, Botías C, Abdul-Sada A, Nicholls E, Rotheray EL, Hill EM, et al. Widespread
- 435 contamination of wildflower and bee-collected pollen with complex mixtures of neonicotinoids and

436 fungicides commonly applied to crops. Environment international. 2016;88:169-78.

- 437 3. Cameron SA, Lozier JD, Strange JP, Koch JB, Cordes N, Solter LF, et al. Patterns of widespread
- 438 decline in North American bumble bees. Proceedings of the National Academy of Sciences.
- 439 2011;108(2):662-7.
- 440 4. Colla SR, Packer L. Evidence for decline in eastern North American bumblebees (Hymenoptera :
- 441 Apidae), with special focus on Bombus affinis Cresson. Biodiversity and Conservation. 2008;17(6):1379-
- 442 91. doi: 10.1007/s10531-008-9340-5. PubMed PMID: WOS:000256754200006.

443 5. Grixti JC, Wong LT, Cameron SA, Favret C. Decline of bumble bees (Bombus) in the North

- 444 American Midwest. Biological Conservation. 2009;142(1):75-84. doi:
- 445 <u>https://doi.org/10.1016/j.biocon.2008.09.027</u>.
- 446 6. Biesmeijer J, Roberts S, Reemer M, Ohlemüller R, Edwards M, Peeters T, et al. Parallel declines in
- pollinators and insect-pollinated plants in Britain and the Netherlands. Science. 2006;313(5785):351-4.
- 448 7. Memmott J, Waser NM, Price MV. Tolerance of pollination networks to species extinctions. Proc
- 449 R Soc B-Biol Sci. 2004;271(1557):2605-11. doi: 10.1098/rspb.2004.2909. PubMed PMID:

450 WOS:000226419100011.

- 451 8. Wood TJ, Goulson D. The environmental risks of neonicotinoid pesticides: a review of the
- 452 evidence post 2013. Environmental Science and Pollution Research. 2017;24(21):17285-325. doi:
- 453 10.1007/s11356-017-9240-x. PubMed PMID: WOS:000406479200003.
- 454 9. Sheets LP, Li AA, Minnema DJ, Collier RH, Creek MR, Peffer RC. A critical review of neonicotinoid
- insecticides for developmental neurotoxicity. Critical reviews in toxicology. 2016;46(2):153-90.
- 456 10. Dively GP, Kamel A. Insecticide residues in pollen and nectar of a cucurbit crop and their
- 457 potential exposure to pollinators. Journal of agricultural and food chemistry. 2012;60(18):4449-56.
- 458 11. Whiting SA, Strain KE, Campbell LA, Young BG, Lydy MJ. A multi-year field study to evaluate the
- 459 environmental fate and agronomic effects of insecticide mixtures. Science of The Total Environment.

460 2014;497–498:534-42. doi: <u>http://dx.doi.org/10.1016/j.scitotenv.2014.07.115</u>.

- 461 12. Main AR, Michel NL, Cavallaro MC, Headley JV, Peru KM, Morrissey CA. Snowmelt transport of
- 462 neonicotinoid insecticides to Canadian Prairie wetlands. Agriculture, Ecosystems & Environment.

463 2016;215:76-84. doi: <u>http://dx.doi.org/10.1016/j.agee.2015.09.011</u>.

- 464 13. Pecenka JR, Lundgren JG. Non-target effects of clothianidin on monarch butterflies. Science of
- 465 Nature. 2015;102(3-4). doi: 10.1007/s00114-015-1270-y. PubMed PMID: WOS:000355942200007.
- 466 14. Gels JA, Held DW, Potter DA. Hazards of insecticides to the bumble bees Bombus impatiens
- 467 (Hymenoptera : Apidae) foraging on flowering white clover in turf. Journal of Economic Entomology.
- 468 2002;95(4):722-8. doi: 10.1603/0022-0493-95.4.722. PubMed PMID: WOS:000177448000010.
- 15. Blacquiere T, Smagghe G, Van Gestel CA, Mommaerts V. Neonicotinoids in bees: a review on
- 470 concentrations, side-effects and risk assessment. Ecotoxicology. 2012;21(4):973-92.
- 471 16. Krupke CH, Hunt GJ, Eitzer BD, Andino G, Given K. Multiple routes of pesticide exposure for
- 472 honey bees living near agricultural fields. PLoS One. 2012;7(1):e29268.
- 473 17. Bonmatin J-M, Giorio C, Girolami V, Goulson D, Kreutzweiser DP, Krupke C, et al. Environmental

474 fate and exposure; neonicotinoids and fipronil. Environmental Science and Pollution Research.

- 475 2015;22(1):35-67. doi: 10.1007/s11356-014-3332-7.
- 476 18. de Perre C, Murphy TM, Lydy MJ. Fate and effects of clothianidin in fields using conservation
- 477 practices. Environmental toxicology and chemistry / SETAC. 2015;34(2):258-65. Epub 2014/11/08. doi:
- 478 10.1002/etc.2800. PubMed PMID: 25376402.
- 479 19. Cabrera AR, Almanza MT, Cutler GC, Fischer DL, Hinarejos S, Lewis G, et al. Initial
- 480 recommendations for higher-tier risk assessment protocols for bumble bees, Bombus spp.
- 481 (Hymenoptera: Apidae). Integrated Environmental Assessment and Management. 2016;12(2):222-9. doi:
- 482 doi:10.1002/ieam.1675.
- 483 20. Devillers J, Decourtye A, Budzinski H, Pham-Delègue MH, Cluzeau S, Maurin G. Comparative
- 484 toxicity and hazards of pesticides to Apis and non-Apis bees. A chemometrical study. SAR and QSAR in
- 485 Environmental Research. 2003;14(5-6):389-403. doi: 10.1080/10629360310001623980.
- 486 21. O'Donnell S. The role of male disease susceptibility in the evolution of haplodiploid insect
- 487 societies Proceedings of the Royal Society B: Biological Sciences. 2004;271(1542):979-83. doi:
- 488 10.1098/rspb.2004.2685.
- 489 22. Heinrich B. Bumblebee Economics: Harvard University Press; 1979.
- 490 23. Heinrich B, Heinrich MJE. Size and Caste in Temperature Regulation by Bumblebees.
- 491 Physiological Zoology. 1983;56(4):552-62.
- 492 24. Thompson HM, Wilkins S, Harkin S, Milner S, Walters KFA. Neonicotinoids and bumblebees
- 493 (Bombus terrestris): effects on nectar consumption in individual workers. Pest Management Science.
- 494 2015;71(7):946-50. doi: 10.1002/ps.3868.
- 495 25. Xu J, Strange JP, Welker DL, James RR. Detoxification and stress response genes expressed in a
- 496 western North American bumble bee, Bombus huntii (Hymenoptera: Apidae). BMC Genomics.
- 497 2013;14:874. Epub 2013/12/18. doi: 10.1186/1471-2164-14-874. PubMed PMID: 24330608; PubMed

498 Central PMCID: PMCPMC3878831.

- 499 26. Alkassab AT, Kirchner WH. Impacts of chronic sublethal exposure to clothianidin on winter
- 500 honeybees. Ecotoxicology. 2016;25(5):1000-10. doi: 10.1007/s10646-016-1657-3.
- 501 27. Spurgeon D, Hesketh H, Lahive E, Svendsen C, Baas J, Robinson A, et al. Chronic oral lethal and
- sub-lethal toxicities of different binary mixtures of pesticides and contaminants in bees (Apis mellifera,
- 503 Osmia bicornis and Bombus terrestris). EFSA Supporting Publications. 2016;13(9):1076E-n/a. doi:
- 504 10.2903/sp.efsa.2016.EN-1076.
- 505 28. Abbott V, Nadeau J, Higo H, Winston M. Lethal and sublethal effects of imidacloprid on Osmia
- 506 lignaria and clothianidin on Megachile rotundata (Hymenoptera: Megachilidae). Journal of Economic
- 507 Entomology. 2008;101(3):784-96.
- 508 29. Scott-Dupree CD, Conroy L, Harris C. Impact of currently used or potentially useful insecticides
- 509 for canola agroecosystems on Bombus impatiens (Hymenoptera: Apidae), Megachile rotundata
- 510 (Hymentoptera: Megachilidae), and Osmia lignaria (Hymenoptera: Megachilidae). Journal of Economic
- 511 Entomology. 2009;102(1):177-82.
- 512 30. Cresswell JE, Desneux N, vanEngelsdorp D. Dietary traces of neonicotinoid pesticides as a cause
- of population declines in honey bees: an evaluation by Hill's epidemiological criteria. Pest management
  science. 2012;68(6):819-27.
- 515 31. Arena M, Sgolastra F. A meta-analysis comparing the sensitivity of bees to pesticides.
- 516 Ecotoxicology. 2014;23(3):324-34.
- 517 32. Di Prisco G, Cavaliere V, Annoscia D, Varricchio P, Caprio E, Nazzi F, et al. Neonicotinoid
- 518 clothianidin adversely affects insect immunity and promotes replication of a viral pathogen in honey
- 519 bees. Proceedings of the National Academy of Sciences of the United States of America.
- 520 2013;110(46):18466-71. doi: 10.1073/pnas.1314923110. PubMed PMID: WOS:000326830900045.
- 521 33. Fauser-Misslin A, Sadd BM, Neumann P, Sandrock C. Influence of combined pesticide and

- 522 parasite exposure on bumblebee colony traits in the laboratory. J Appl Ecol. 2014;51(2):450-9. doi:
- 523 10.1111/1365-2664.12188. PubMed PMID: WOS:000332835600018.
- 524 34. Wu JY, Anelli CM, Sheppard WS. Sub-lethal effects of pesticide residues in brood comb on
- 525 worker honey bee (Apis mellifera) development and longevity. PloS one. 2011;6(2):e14720.
- 526 35. Rabhi KK, Esancy K, Voisin A, Crespin L, Le Corre J, Tricoire-Leignel H, et al. Unexpected effects of
- 527 low doses of a neonicotinoid insecticide on behavioral responses to sex pheromone in a pest insect.
- 528 PLoS One. 2014;9(12):e114411. Epub 2014/12/18. doi: 10.1371/journal.pone.0114411. PubMed PMID:
- 529 25517118; PubMed Central PMCID: PMCPMC4269385.
- 530 36. Decourtye A, Devillers J, Genecque E, Le Menach K, Budzinski H, Cluzeau S, et al. Comparative
- sublethal toxicity of nine pesticides on olfactory learning performances of the honeybee Apis mellifera.
- 532 Archives of environmental contamination and toxicology. 2005;48(2):242-50.
- 533 37. Chaimanee V, Evans JD, Chen Y, Jackson C, Pettis JS. Sperm viability and gene expression in
- 534 honey bee queens (Apis mellifera) following exposure to the neonicotinoid insecticide imidacloprid and
- the organophosphate acaricide coumaphos. Journal of Insect Physiology. 2016;89:1-8. doi:
- 536 <u>http://dx.doi.org/10.1016/j.jinsphys.2016.03.004</u>.
- 537 38. Li X, Schuler MA, Berenbaum MR. Molecular Mechanisms of Metabolic Resistance to Synthetic
- and Natural Xenobiotics. Annual Review of Entomology. 2007;52(1):231-53. doi:
- 539 10.1146/annurev.ento.51.110104.151104. PubMed PMID: 16925478.
- 540 39. Woodard SH, Lozier JD, Goulson D, Williams PH, Strange JP, Jha S. Molecular tools and bumble
- 541 bees: revealing hidden details of ecology and evolution in a model system. Molecular Ecology.
- 542 2015;24(12):2916-36. doi: 10.1111/mec.13198.
- 40. de Oliveira IM, Nunes BVF, Barbosa DR, Pallares AM, Faro LRF. Effects of the neonicotinoids
- 544 thiametoxam and clothianidin on in vivo dopamine release in rat striatum. Toxicology Letters.
- 545 2010;192(3):294-7. doi: <u>http://dx.doi.org/10.1016/j.toxlet.2009.11.005</u>.

	546	41.	Arce AN, David TI, Rand	dall EL, Rodrigues AF	R, Colgan TJ, Wurm	Y, et al. Imp	act of contro	ollec
--	-----	-----	-------------------------	-----------------------	--------------------	---------------	---------------	-------

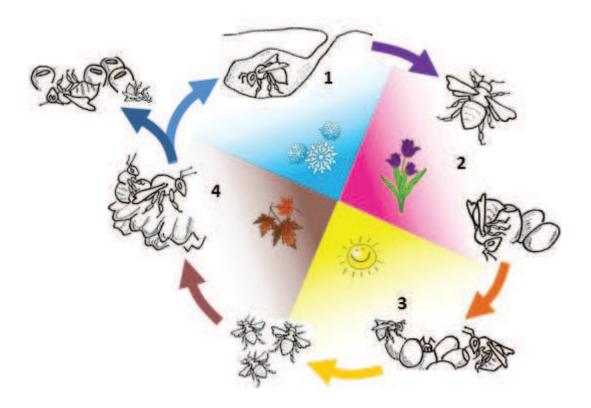
- neonicotinoid exposure on bumblebees in a realistic field setting. J Appl Ecol. 2017;54(4):1199-208. doi:
- 548 10.1111/1365-2664.12792. PubMed PMID: WOS:000405534500020.
- 549 42. Kaplan EL, Meier P. Nonparametric Estimation from Incomplete Observations. Journal of the
- 550 American Statistical Association. 1958;53(282):457-81. doi: 10.2307/2281868.
- 43. Sadd BM, Barribeau SM, Bloch G, de Graaf DC, Dearden P, Elsik CG, et al. The genomes of two
- key bumblebee species with primitive eusocial organization. Genome Biol. 2015;16. doi: ARTN 76
- 553 10.1186/s13059-015-0623-3. PubMed PMID: WOS:000353676700001.
- 554 44. Schröder P, Collins C. Conjugating Enzymes Involved in Xenobiotic Metabolism of Organic
- 555 Xenobiotics in Plants. International Journal of Phytoremediation. 2002;4(4):247-65. doi:
- 556 10.1080/15226510208500086.
- 557
- 558
- 559
- 560 Supporting Information

561

- Fig S1. Schematic of setup for individual feeding assay. Bees were isolated from the colony
  and housed in a 16oz plastic container with a feeding cup positioned directly under an open vent
  in the screen lid. Test sugar solutions were delivered through a hole in the lid with a micropipette.
  Individuals were supplied with pollen *ad libitum*.
  Table S1. Percent of alive queens, workers and males consuming test solutions
- 568 (clothianidin concentration given in ppb) over the 7-day testing period.

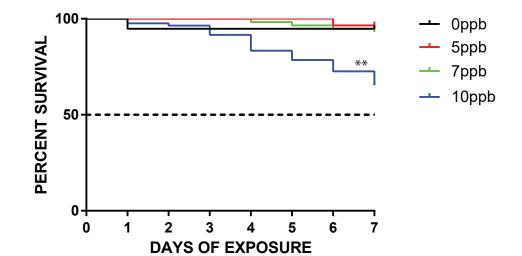
- **Table S2. List of clothianidin-induced non-detoxification genes with increased expression**
- 570 in worker (W) and male (M) Bombus impatiens. Biological process known to be associated
- 571 with each gene is also shown.
- 573 Table S3. List of clothianidin-induced genes showing a similar expression level in
- 574 worker (W) and male (M) Bombus impatiens.
- **Table S4. RNA sequencing results and quality control.**

#### Figure 1

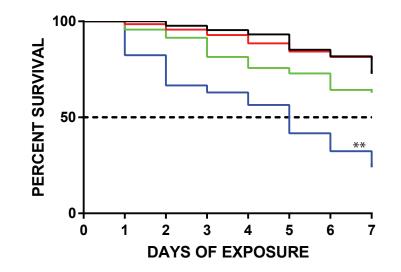


#### FIGURE 3

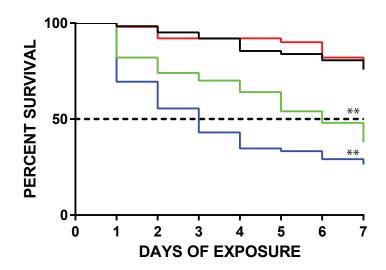
#### A) QUEENS



**B) WORKERS** 



C) MALES



#### Figure 3

