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1 **Rhizosphere microbes and host plant genotype influence the plant metabolome and reduce**  
2 **insect herbivory**

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## 23 Summary

- 24 • Rhizosphere microbes affect plant performance, including plant resistance against insect  
25 herbivores; yet, the relative influence of rhizosphere microbes vs. plant genotype on  
26 herbivory levels and on metabolites related to defense remains unclear.
- 27 • In *Boechnera stricta*, we tested the effects of rhizosphere microbes and plant genotype on  
28 herbivore resistance, the primary metabolome, and select secondary metabolites.
- 29 • Plant populations differed significantly in the concentrations of 6 glucosinolates (GLS),  
30 secondary metabolites known to provide herbivore resistance in the Brassicaceae, and the  
31 population with lower GLS levels experienced ~60% higher levels of aphid (*Aphis spp.*)  
32 attack; no effect was observed of GLS on damage by a second herbivore, flea beetles  
33 (*Altica spp.*). Rhizosphere microbiome (intact vs. disrupted) had no effect on plant GLS  
34 concentrations. However, aphid number and flea beetle damage were respectively ~3-fold  
35 and 7-fold higher among plants grown in the disrupted rhizosphere microbiome  
36 treatment, and distinct (as estimated from 16s rRNA amplicon sequencing) intact native  
37 microbiomes also differed in their effects on herbivore damage. These differences may be  
38 attributable to shifts in primary metabolic pathways.
- 39 • The findings suggest that rhizosphere microbes can play a greater role than plant  
40 genotype in defense against insect herbivores, and act through mechanisms independent  
41 of plant genotype.

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## 49 **Introduction**

50 The rhizosphere microbiome has been referred to as the “extended phenome” of the plant host  
51 because of the significant effects microbes can have on plant performance (Berendsen *et al.*,  
52 2012). Among other mechanisms, rhizosphere microbes can improve fitness in heterogeneous  
53 environments by reducing damage by herbivores (Tétard & Jones *et al.*, 2007; Pineda *et al.*, 2010,  
54 2017; Badri *et al.*, 2013). However, it is not known how these microbes act in combination with  
55 plant genetic pathways to modulate plant defense. For instance, plant pathways regulating the  
56 production of secondary metabolites greatly influence herbivore damage in field settings  
57 (Windsor *et al.*, 2005; Kerwin *et al.*, 2015, 2017, Francisco *et al.*, 2016a, b). Likewise, plants  
58 grown in the presence of an intact rhizosphere microbiome experience reduced herbivory in  
59 controlled settings in comparison to those grown with an experimentally disrupted microbiome,  
60 and responses of the host plant metabolome to the microbiome are one mechanism hypothesized  
61 to underlie differences in herbivory (Badri *et al.*, 2013). Given the pervasive negative effects of  
62 herbivores on plant fitness in natural settings and on crop yield (Mitchell *et al.*, 2016), it would  
63 be valuable to characterize if and by what mechanisms microbes reduce herbivore damage in  
64 agroecologically relevant settings, and in particular if microbes and plant genotypes affect  
65 herbivory by similar or different metabolic mechanisms (Heil, 2008; War *et al.*, 2012; Pineda *et*  
66 *al.*, 2017).

67 Plant genetic variation is known to modulate plant resistance to insect herbivores by  
68 controlling the accumulation of specific primary and secondary metabolites (Kliebenstein, 2004;  
69 Bolton, 2009; Chan *et al.*, 2010; Zhou *et al.*, 2015; Wagner & Mitchell-Olds, 2017). Some  
70 primary metabolites are associated with cell wall composition and thickness and can thereby  
71 affect insect feeding rates (Kärkönen *et al.*, 2005; War *et al.*, 2012; Malinovsky *et al.*, 2014).  
72 Likewise, the concentration in leaves of the primary metabolite, ascorbic acid, is directly  
73 associated with insect feeding behavior (Goggin *et al.*, 2010; Zebelo & Maffei, 2015).  
74 Glucosinolates, secondary metabolites found in the Capparales, combine with myrosinases when  
75 plant tissue is disturbed, creating bioactive products that deter further feeding (Wittstock &  
76 Burow, 2010). In *Boechera stricta*, one QTL was identified that controlled both glucosinolate  
77 concentration and insect feeding (Schranz *et al.*, 2009). Similarly, diverse multi-locus  
78 glucosinolate genotypes that mimic natural variants in *Arabidopsis thaliana* differed significantly

79 in damage by chewing insects (Kerwin *et al.*, 2015), clearly demonstrating that plant genotype  
80 influences the extent of damage by insect herbivores.

81         The microbiome also affects many aspects of plant performance including resistance to  
82 insect feeding, but it is not known if this is via similar mechanisms to those determined by  
83 natural genetic variation in plants (Berendsen *et al.*, 2012; Bulgarelli *et al.*, 2013; Hubbard *et al.*,  
84 2017). Microbes could influence the plant metabolome through their effects on host plant access  
85 to nutrients (Wetzel *et al.*, 2016; Gomez Casati, 2016). The profile of cell wall lignin, an  
86 important physical barrier against herbivore feeding, can be affected by plant access to limiting  
87 nutrients like nitrogen, where plants with greater access to nitrogen have stronger cell walls and  
88 experience less feeding damage (Blodgett *et al.*, 2005; Barros *et al.*, 2015). Likewise, nitrogen  
89 and phosphorus availability can influence glucosinolate content, which can alter plant response  
90 to insect feeding (Ernst, 1998; Del Carmen Martínez-Ballesta *et al.*, 2013). Rhizosphere  
91 microbes are known to improve host plant access to limiting nutrients like nitrogen and  
92 phosphorus, and could thereby indirectly influence physical and chemical plant defenses against  
93 insect herbivores (Richardson *et al.*, 2009; Richardson & Simpson, 2011; Bulgarelli *et al.*, 2013).  
94 Nevertheless, the magnitude of plant genotype *vs.* microbial effects on field herbivory remains  
95 unknown as does the extent to which diverse primary and secondary metabolites are influenced  
96 by rhizosphere microbes.

97         In the current study, we test if rhizosphere microbes and host plant population had similar  
98 (or different) effects on primary metabolites and glucosinolate secondary metabolites, and  
99 ultimately on plant responses to insect herbivores in complex field settings. We first characterize  
100 microbial community structure to test if microbiome treatments yield distinct rhizosphere  
101 microbial communities and if intact native microbiomes differ in community composition, and to  
102 evaluate what microbes may be differentially abundant across treatments. Next, we address how  
103 the rhizosphere microbiome and plant population influence the concentration of primary  
104 metabolites and glucosinolates, and how differences in the plant metabolome attributable to  
105 rhizosphere microbes and to plant population relate to damage by insect herbivores.

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107 **Materials and Methods**

## 108 Plant Material and Growth Conditions

109 To investigate the influence rhizosphere microbes and host plant genotype have on the  
110 metabolome and response to herbivory, we performed 3 related types of experiments: 1) we  
111 characterized differences in rhizosphere microbial communities across distinct inoculation  
112 treatments as well as differences in GLS and primary metabolites among plant populations and  
113 between rhizosphere microbiome treatments, 2) in two independent greenhouse experiments, we  
114 determined the effect of population *vs.* rhizosphere microbiome on the extent of damage by  
115 aphids, and 3) in a field experiment, we tested the effect of population *vs.* rhizosphere  
116 microbiome on damage by flea beetles.

117 We used seeds from five geographically separated (minimum distance between  
118 populations = 4 km) of *Boechnera stricta* populations found within the Medicine Bow, Sherman,  
119 and Sierra Madre mountain ranges of Wyoming: Crow Creek (Sherman), Road 234 (Medicine  
120 Bow), Sandstone (Sierra Madre), South Brush Creek (Medicine Bow) and Webb Springs (Sierra  
121 Madre). We collected soils from 3 of these sites to be used as microbial inoculate: Crow Creek,  
122 Road 234, and Webb Springs (**Table S1**). All seeds and soils were collected under permits from  
123 the National Forest Service (LAR1082).

124 For all experiments, seeds were surface sterilized using a 15% bleach solution and placed  
125 on petri plates to germinate. At the emergence of root radicles, plants were transplanted to 5cm  
126 diameter pots containing a mixture of sterilized Redi-Earth potting mix (Sunagro, Agawam, MA,  
127 USA) and soil inoculate. Soil inoculate was created by mixing 30g of fresh soil with 270ml of  
128 RO H<sub>2</sub>O and filtered through 1,000 $\mu$ m, 500 $\mu$ m, and 212 $\mu$ m sieves to remove large soil  
129 particulates (van de Voorde *et al.*, 2012). For the “intact microbiome” treatments, 2ml of the  
130 resulting solution was added to pots prior to planting. For the “disrupted microbiome” treatment,  
131 we filtered soil inoculate through a 0.2 $\mu$ m mesh to remove all microorganisms. This approach  
132 removes microbes while retaining soil nutrients, thus controlling for the potentially confounding  
133 effect of nutrients on plant performance. All experiments were performed at the Williams  
134 Conservatory at the University of Wyoming and field sites at the Agriculture Experiment Station  
135 in Laramie, WY, USA.

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137 Characterization of microbial communities

138 We harvested rhizospheres from three replicate plants grown in each of the intact (Crow Creek,  
139 Road 234) and disrupted microbiome treatments. To extract rhizosphere bacterial DNA, plant  
140 roots were agitated for 15 minutes in phosphate-buffered saline (PBS) to separate soil particles  
141 from plant roots. After removing plant roots, the soil and PBS solution was centrifuged at 3000  
142 rpm for 15 minutes as described in Hubbard *et al.* (2017). Next, the supernatant was discarded,  
143 and 250 mg of the pellet was transferred to bead tubes from the Mobio Power Soil DNA  
144 Isolation Kit (Mobio Laboratories, Carlsbad, CA, USA). DNA was extracted following the  
145 manufacturer's instructions. A soilless blank was included with each round of extraction to serve  
146 as a negative control. After each round of extractions, we performed PCR to quantify DNA  
147 yields and ensure reagent sterility.

148         Amplicon library preparation of the V4V5 region of the 16s ribosomal subunit gene  
149 (518F and 926R) and amplicon sequencing on the Illumina MiSeq platform (Illumina, San  
150 Diego, CA, USA) was performed at the Marine Biological Laboratories (Woods Hole, MA,  
151 USA) as described in Newton *et al.* (2015). We used the R package *dada2* to filter and trim reads  
152 based on quality, infer error rates, merge paired end reads, remove chimeras, and assign  
153 taxonomy using the Silva reference database (ver. 128) (R Core Team, 2013; Quast *et al.*, 2013;  
154 Callahan *et al.*, 2016). After normalizing to account for differences between samples in read  
155 number, we performed *adonis* (permutational multivariate analysis of variance using  
156 dissimilarity matrices) and Principal Coordinate Analysis (PCoA) on Jaccard (presence-absence)  
157 and Bray-Curtis (abundance) dissimilarities based on non-rarefied reads to test if the microbial  
158 treatments were compositionally distinct using the R packages *Phyloseq* and *Vegan* (McMurdie  
159 & Holmes, 2013; Oksanen, 2015). Additionally, we used *DESeq2* to identify taxa that were  
160 differentially abundant, and may explain differential effects of intact vs. disrupted rhizosphere  
161 microbiomes on the plant metabolome and herbivore defense (McMurdie *et al.*, 2014; Love *et*  
162 *al.*, 2014). All sequences have been deposited into the Short Read Archive (SRA) under project  
163 number PRJNA449164.

164

165 Characterization of the primary and secondary metabolome

166 To characterize the influence of plant genotype and rhizosphere microbiome on *B. stricta*'s  
167 metabolome, replicates of the Road 234 and Crow Creek populations grown in the Crow Creek,  
168 Road 234, and disrupted microbiome treatments were harvested and dried for 7 days at 65°C ( $N$   
169 = 90). From each sample, 20mg of dried leaf tissue was sent to the West Coast Metabolomics  
170 Center (Davis, CA, USA), and samples were used for either high throughput glucosinolate or  
171 primary metabolite characterization (Kliebenstein *et al.*, 2001a,b,c; Fiehn, 2016). Methyl ethyl  
172 and methylpropyl GLS were identified by comparison to previously reported profiles and relative  
173 retention times. Unidentified glucosinolates are named by their retention time (Schranz *et al.*,  
174 2009).

175 For analysis of the glucosinolate data, we first divided the area under peaks by dry leaf  
176 mass to obtain a measurement of glucosinolate concentration. We used Two-Way ANOVAs to  
177 characterize the effects plant population and microbiome status (intact vs. disrupted) had on  
178 glucosinolate concentration; we used False Discovery Rate corrections to correct for multiple  
179 comparisons. Tukey's Honest Significant Differences *post hoc* test was used to determine the  
180 source of significant microbiome differences when observed.

181 For the primary metabolite data, we first corrected for drift in machine sensitivity and  
182 used Two-Way ANOVAs and False Discovery Rate corrections to correct for multiple  
183 comparisons in MetaboAnalyst 3.0 to test for the effect of plant population and microbiome  
184 status on metabolite concentration (Xia *et al.*, 2016). Next, we used hypergeometric tests in  
185 MetaboAnalyst 3.0 using the *Arabidopsis thaliana* pathway library  
186 (KEGG, <http://www.genome.jp/kegg/>) to identify pathways significantly affected by plant  
187 population and rhizosphere microbiome. It was necessary to have separate replicates for  
188 metabolomic evaluation and tests of herbivore damage due to the comparatively small plant size  
189 and large mass of tissue needed for metabolomics.

190

191 Plant response to insect herbivores

192 *Aphids (Aphis spp.)*

193 On August 11 2015, we planted 20 replicates of the Crow Creek and Road 234 populations in  
194 Crow Creek, Road 234, and disrupted microbiome treatments ( $n = 120$ ), using a fully  
195 randomized design. After 6 weeks of growth, there was an unplanned infestation of aphids in the  
196 greenhouse. Aphids on the plant were counted to characterize the association of herbivory with  
197 both plant population and microbiome status. Because the original experiment was not designed  
198 to test for aphid resistance, we re-tested the effect of rhizosphere microbiome by exposing  
199 replicates of Road 234 plants in two microbiome treatments, Road 234 and disrupted, to aphids  
200 ( $N=18$ ). Specifically, seedlings were transplanted from petri plates to soil on September 12 2016  
201 and grown in the William's Conservatory until October 20 before being placed in a fully  
202 randomized array in a greenhouse containing *Zea mays* with aphids. Aphid counts were taken  
203 daily for 10 days.

204

205 *Flea beetles (Altica spp.)*

206 In a fully randomized design, we grew up 20 replicates of each of the Crow Creek, Road 234,  
207 Sandstone, South Brush Creek, and Webb Springs populations in the Crow Creek, Road 234,  
208 Webb Springs, and disrupted microbiome treatments ( $N = 400$ ). Seedlings were transferred to  
209 pots on June 6 2016 and grown in a greenhouse until July 6. On July 6, 8 replicates of each  
210 population  $\times$  microbiome treatment combination were transplanted into a field with a history of  
211 flea beetle occurrence (CH, CW personal observations). Plants were checked daily for flea  
212 beetles and scored qualitatively for whole plant damage on a 0-4 scale (0 = 0-20%, 1 = 21-40%,  
213 2 = 41-60%, 3 = 61-80%, 4 = 81-100% damage). On July 18, we harvested the 2<sup>nd</sup> true leaf to  
214 quantify herbivore damage. Leaves were scanned using an Epson V700 scanner (Epson, Long  
215 Beach, CA, USA), and we used ImageJ to quantify leaf damage (Schneider *et al.*, 2012).

216

217 *Statistical Analyses*

218 We used Two-Way ANOVAs and Tukey's Honest Significant Differences *post hoc* tests to  
219 characterize the effect of plant population and rhizosphere microbiome status on susceptibility to



220 aphids or flea beetles (de Mendiburu, 2016). All plots were made using the package *ggplot2*  
221 (Wickham, 2009).

222

## 223 **Results**

### 224 Characterization of microbial communities

225 After processing, we retained 1 763 803 high quality reads from a total of 2 466 906 raw reads,  
226 where read count per sample ranged from 119 042 to 313 038 reads (**Table S2**). From a planned  
227 comparison within a Jaccard presence-absence analysis, the intact (Crow Creek and Road 234)  
228 treatment differed significantly from the disrupted microbiome treatment ( $P = 0.014$ ; **Fig. 1a**);  
229 further, the Crow Creek microbiome differed from that of Road 234 ( $P = 0.043$ ). Likewise, from  
230 a planned comparison within a Bray-Curtis abundance analysis, the intact (Crow Creek and Road  
231 234) treatment differed significantly from the disrupted microbiome treatment ( $P = 0.011$ ; **Fig.**  
232 **1b**), and Crow creek differed from Road 234 ( $P = 0.048$ ). Further, differential abundance  
233 analysis (DESeq2) revealed 36 genera from 14 phyla differentially abundant between the intact  
234 vs. disrupted microbiome treatments (**Fig. 2**). Of the 27 genera associated with the intact  
235 microbiome treatment, 15 have been previously reported as beneficial to plant performance  
236 (Preston, 2004; Khan *et al.*, 2017).

237

### 238 Characterization of primary and secondary metabolites and pathways

239 Plant population, rhizosphere microbiome, and their interaction significantly affected *B. stricta*'s  
240 metabolome. Six out of the nine measured glucosinolates differed significantly between  
241 populations, where five of the six were in higher concentration in plants from the Road 234  
242 population (**Table 1**). Further, population significantly influenced the concentration of 33  
243 primary metabolites (**Table S3**) corresponding to three metabolic pathways, all of which were at  
244 higher concentrations in Crow Creek plants (**Table 2**).

245 In contrast to plant population, rhizosphere microbiome status influenced the  
246 concentration of only one of a total of nine glucosinolates, and for this one glucosinolate, the  
247 concentration was higher in plants from the disrupted microbiome treatment (**Table 1**).

248 Additionally, microbiome status affected the concentrations of 22 metabolites and four metabolic  
249 pathways (**Table S3**). Metabolites in three of the four pathways were in higher concentrations in  
250 the disrupted treatment, while metabolites in the pentose and glucuronate interconversion  
251 pathway were in higher concentrations in the intact microbiome treatment (**Table 2**). 23  
252 metabolites were significantly affected by interactions between microbiome status and  
253 population (**Table S3**). Further, one pathway, glyoxylate and dicarboxylate metabolism, was  
254 significantly affected by both population and microbiome status, where metabolites were in  
255 higher concentrations in Crow Creek plants and the disrupted treatment (**Table 2**).

256

## 257 Plant response to insect herbivores

### 258 *Aphids*

259 In the first experiment in fall 2015, plants from the Crow Creek population had ~3 (60%) more  
260 aphids per plant than Road 234 plants (**Fig. 3a; Table 3**). Further, in both aphid experiments,  
261 microbiome status significantly influenced the number of aphids per plant (**Table 3**). In the first  
262 experiment, plants in the disrupted microbiome treatment had ~9 (~225%) more aphids than  
263 plants in the intact microbiome treatments on average (**Fig. 3b**). Plants grown in the intact Road  
264 234 microbiome had ~3 (113%) more aphids than those grown in the intact Crow Creek  
265 microbiome. Likewise, in the second experiment, plants in the disrupted treatment had more than  
266 330% more aphids (~3) than plants in the intact treatment (**Fig. 3c**). Thus, the magnitude of the  
267 rhizosphere microbiome effect was larger than that of plant genotype.

268

### 269 *Flea beetles*

270 We did not observe a significant effect of plant population on leaf or whole plant damage by flea  
271 beetles (**Fig. S1; Table 3**). Plants in the disrupted microbiome treatment experienced  
272 significantly greater leaf and whole plant damage than plants in the intact treatment (**Fig. 4a, b;**  
273 **Table 3**). For instance, plants in the disrupted treatment, on average, experienced ~700% more  
274 leaf damage and more than 200% greater whole plant damage compared to plants in the intact

275 treatment (**Fig. 4a, b**). The three intact microbiomes did not differ significantly in their effect on  
276 flea beetle damage (**Fig. 4a, b**).

277

## 278 **Discussion**

279 The past several decades have witnessed a ground-breaking description of plant genetic controls  
280 underlying defense (Coley *et al.*, 1985; Bennett *et al.*, 1994; Kroymann *et al.*, 2003). More  
281 recently, studies have considered the effects of the rhizosphere microbiome on plant resistance to  
282 insect herbivores (Tétard & Jones *et al.*, 2007; Pineda *et al.*, 2010, 2017; Badri *et al.*, 2013).

283 Although the effect of plant genotype on many metabolic defense pathways is well-  
284 characterized, the effect of microbes on the metabolome and the relative contribution of plant  
285 genotype *vs.* microbes to herbivore defense remain unclear. Here, we compared the effects of  
286 host plant population and rhizosphere microbes on metabolites and on plant responses to insect  
287 herbivores. We observed that host plant populations differed by 60% or non-significantly in  
288 susceptibility to aphids and flea beetles, respectively. Notably, intact *vs.* disrupted rhizosphere  
289 microbiome status led to 3- and 7-fold differences in defense against aphids and flea beetles, and  
290 different intact native microbiomes differed 2-fold in aphid susceptibility. Thus, the magnitude  
291 of rhizosphere microbial effect was much greater than that of plant genotype, a result consistent  
292 with the recent observation that some human phenotypes are better-predicted from the  
293 microbiome than the host genome (Sze & Schloss, 2016). The underlying mechanism of defense  
294 was seemingly distinct, because plant populations differed significantly in GLS while intact *vs.*  
295 disrupted rhizosphere microbiomes elicited differences in the primary metabolome but not GLS  
296 concentrations of the host plant.

297 Plant genotype has been shown to affect diverse primary and secondary metabolites linked to  
298 defense (Mauricio & Rausher, 1997; Mauricio, 1998; Kerwin *et al.*, 2015; Francisco *et al.*,  
299 2016b). Here, plant population influenced several primary metabolic pathways as well as  
300 glucosinolate secondary metabolites and host plant defense against one herbivore. The three  
301 primary metabolic pathways affected by population were all in higher concentrations among  
302 plants from Crow Creek relative to Road 234 population, while 6 out of 9 measured GLS were  
303 affected by population and the majority (5) of these were found at higher concentration in plants

304 from the Road 234 population (**Table 1, 2**). Despite the metabolomic differences between Road  
305 234 and Crow Creek plants, we did not find significant differences in flea beetle damage  
306 between populations (**Fig. S1**). This result is consistent with previous studies of flea beetle-  
307 *Boechera stricta* interactions. In a study examining the distribution of natural *B. stricta*  
308 populations, there was an inverse relationship between *B. stricta* occurrence and flea beetle  
309 frequency despite natural variation in glucosinolate content between lines (Naithani *et al.*, 2014).  
310 *B. stricta*'s inability to adapt to flea beetles may be explained by flea beetles' ability to utilize  
311 glucosinolates as a nutrient source (Beran *et al.*, 2014). By contrast, plants from the Road 234  
312 population had significantly fewer aphids per plant than Crow Creek plants (**Fig. 3a**). It is  
313 possible that both primary metabolites and GLS contribute to reduced damage by this second  
314 insect herbivore. For example, Goggin *et al.* (2010) found that plants with higher concentrations  
315 of ascorbate were more attractive to many insect herbivores than plants with lower ascorbate  
316 concentrations, and higher concentrations of ascorbate and aldarate metabolism could  
317 correspondingly be contributing to higher rates of aphids on Crow Creek plants. Further, studies  
318 of *A. thaliana* have shown that glucosinolate concentration affects plant response to aphid  
319 feeding (Kliebenstein *et al.*, 2005; de Vos *et al.*, 2007; Wittstock & Burow, 2010; Zhou *et al.*,  
320 2015). Together, the metabolomic patterns are consistent with the view that Crow Creek plants  
321 may be attracting more aphids and have less protection against aphid feeding than Road 234  
322 plants.

323 Our microbial treatments differed significantly in the presence-absence and abundance of  
324 bacterial taxa (**Fig. 1**), and these microbiome differences contributed could have affected defense  
325 by improving resource availability. Differential abundance analysis identified 36 bacterial genera  
326 that differed in abundance between intact *vs.* disrupted microbiome treatments (**Fig. 2**). 64 of the  
327 84 bacterial taxa were in higher abundance in the rhizosphere of plants grown with an intact  
328 microbiome and may explain performance differences between plants across treatments. For  
329 instance, the genus *nitrobacter* was in higher abundance in the intact microbiome treatment than  
330 the disrupted microbiome treatment. In soils, *nitrobacter* is known to convert nitrite ( $\text{NO}_2^-$ ) to  
331 nitrate ( $\text{NO}_3^-$ ), making nitrogen more readily available to the host plant (Kumar *et al.*, 1983;  
332 Richardson *et al.*, 2009). As a result, plants grown in the intact microbiome treatments may have  
333 greater access to nitrogen and thus potentially higher concentrations of primary or secondary  
334 metabolites relevant to defense (Ernst, 1998; Blodgett *et al.*, 2005; Del Carmen Martínez-

335 Ballesta *et al.*, 2013; Barros *et al.*, 2015). Additional studies, using microbial cultures and  
336 controlled manipulations of microbial inoculates, will clarify the causal effect that different  
337 microbial taxa have on plant performance.

338 As in previous studies of plant-microbe interactions, we found that the disruption of rhizosphere  
339 microbiomes alters plant performance (Lau & Lennon, 2011; Badri *et al.*, 2013; Chialva *et al.*,  
340 2018) and here specifically defense and several metabolites. As noted above, we found that the  
341 differences in the rhizosphere microbiome led to much greater differences in defense than did  
342 host plant population, and specifically plants grown in the disrupted microbiome treatment  
343 experienced significantly higher aphid prevalence (**Fig. 3bc**) and flea beetle damage (**Fig. 4**).  
344 While there were no differences in GLS between microbiome treatments, plants grown in the  
345 disrupted microbiome treatment had higher concentrations of primary metabolites in three  
346 pathways, and plants grown in the intact microbiome treatments had higher concentrations of  
347 metabolites in one pathway that may have altered defense (**Table 2**). Higher concentrations of  
348 metabolites in the citrate cycle among plants grown in the disrupted microbiome treatment and  
349 higher concentrations of metabolites pentose and glucuronate interconversions in plants grown in  
350 the intact microbiome treatments could contribute to the observed differences in plant response  
351 to herbivory. Upregulation of the citrate cycle is an indicator of plant stress, which can reduce  
352 plant defenses (Obata & Fernie, 2012; Bauerfeind & Fischer, 2013) and may be a factor among  
353 plants grown in the disrupted microbiome treatment. By contrast, higher concentrations of  
354 metabolites of the pentose and glucuronate interconversion pathway can increase the strength of  
355 cell walls, which could make feeding more difficult for insects (Urbanczyk-Wochniak & Fernie,  
356 2005; Silveira *et al.*, 2013) and could be important among plants in the intact microbiome  
357 treatment. It is unlikely that differences in herbivore susceptibility between the microbiomes  
358 simply reflect differences in plant vigor and poor plant growth; although plants in disrupted  
359 microbiome treatments performed poorly in response to insect herbivores in comparison to all  
360 intact microbiomes, plants from the disrupted treatment were consistently significantly larger  
361 than plants grown in the intact Crow Creek microbiome treatment in another set of experiments  
362 (CH & CW unpublished data). Together, these results suggest that plants in the disrupted  
363 microbiome treatments are more vulnerable and less well-defended than plants in the intact  
364 microbiome treatment, likely making them more susceptible and attractive to insect herbivores.

365 Although many metabolites and pathways were differentially influenced by either rhizosphere  
366 microbes or plant population, there was one point of overlap (**Table S3**). Metabolites in the  
367 glyoxylate and dicarboxylate metabolism pathway were in higher concentrations in plants in the  
368 disrupted treatment and among Crow Creek plants (**Table 2**). Interestingly, both experienced  
369 significantly higher rates of aphid prevalence than their counterparts. However, the role of  
370 glyoxylate and dicarboxylate metabolism in plant defense is not clear. While more work needs to  
371 be done to elucidate the role of this pathway in defense, the results suggest this is one case where  
372 plant genotype acted in a manner parallel to rhizosphere microbiome with regard to the host  
373 plant metabolome.

374 In sum, we observed that plant population and rhizosphere microbiome differentially affected the  
375 plant metabolome and plant response to insect herbivory in complex field settings. Plant  
376 genotype was associated with differences in glucosinolate levels as well as infestation by aphids,  
377 consistent with the well-described effects of this secondary metabolite on damage (Schranz *et*  
378 *al.*, 2009; War *et al.*, 2012; Zhou *et al.*, 2015). However, the rhizosphere microbiome had much  
379 larger effects on defense than did plant genotype, and seemingly influenced defense by a  
380 different metabolomic mechanism than plant genotype, specifically a mechanism other than  
381 GLS. While we cannot exclude unmeasured metabolites, the effect of the rhizosphere  
382 microbiome may be explained by differences in primary metabolites previously implicated in  
383 plant defense or by metabolites with as yet uncharacterized effects on defense. Given the  
384 pronounced effects on herbivory, further characterizing the metabolomic intermediates between  
385 the rhizosphere microbiome and host plant defense comprises a promising avenue for future  
386 research as does functional characterization of the causal microbes.

387

### 388 **Conflicts of Interest**

389 The authors do not report any conflicts of interest.

390

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396

### 397 **Author Contributions**

398 C.J.H., D.K., and C.W. designed the experiments, C.J.H., B.L., R.M., and M.T.B performed the  
399 research, C.J.H, B.L., L.M., B.E.E., D.K., and C.W. analyzed the data, and all authors  
400 contributed to the writing of the manuscript.

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415 **References**

- 416 **Badri D V., Zolla G, Bakker MG, Manter DK, Vivanco JM. 2013.** Potential impact of soil  
417 microbiomes on the leaf metabolome and on herbivore feeding behavior. *New Phytologist* **198**:  
418 264–273.
- 419 **Barros J, Serk H, Granlund I, Pesquet E. 2015.** The cell biology of lignification in higher  
420 plants. *Annals of Botany* **115**: 1053–1074.
- 421 **Bauerfeind SS, Fischer K. 2013.** Testing the plant stress hypothesis: stressed plants offer better  
422 food to an insect herbivore. *Entomologia Experimentalis et Applicata* **149**: 148–158.
- 423 **Bennett RN, Wallsgrove R, M. 1994.** Secondary metabolites in plant defence mechanisms. *New*  
424 *Phytologist* **127**: 617–633.
- 425 **Beran F, Pauchet Y, Kunert G, Reichelt M, Wielsch N, Vogel H, Reinecke A, Svato A, Mewis I, Schmid D,**  
426 **et al. 2014.** Phyllotreta striolata flea beetles use host plant defense compounds to create their own  
427 glucosinolate-myrosinase system. *Proceedings of the National Academy of Sciences* **111**: 7349–7354.
- 428 **Berendsen RL, Pieterse CMJ, Bakker PAHM. 2012.** The rhizosphere microbiome and plant  
429 health. *Trends in Plant Science* **17**: 478–486.
- 430 **Blodgett JT, Herms DA, Bonello P. 2005.** Effects of fertilization on red pine defense chemistry  
431 and resistance to *Sphaeropsis sapinea*. *Forest Ecology and Management* **208**: 373–382.
- 432 **Bolton MD. 2009.** Primary Metabolism and Plant Defense—Fuel for the Fire. *Molecular Plant-*  
433 *Microbe Interactions* **22**: 487–497.
- 434 **Bulgarelli D, Schlaeppi K, Spaepen S, van Themaat EVL, Schulze-Lefert P. 2013.** Structure  
435 and Functions of the Bacterial Microbiota of Plants. *Annual Review of Plant Biology* **64**: 807–  
436 838.
- 437 **Bundy JG, Davey MP, Viant MR. 2009.** Environmental metabolomics: a critical review and  
438 future perspectives. *Metabolomics* **5**: 3–21.
- 439 **Callahan BJ, McMurdie PJ, Rosen MJ, Han AW, Johnson AJA, Holmes SP. 2016.** DADA2:  
440 High-resolution sample inference from Illumina amplicon data. *Nature methods* **13**: 581–3.



- 441 **Del Carmen Martínez-Ballesta M, Moreno DA, Carvajal M. 2013.** The physiological  
442 importance of glucosinolates on plant response to abiotic stress in Brassica. *International journal*  
443 *of molecular sciences* **14**: 11607–25.
- 444 **Chan EKF, Rowe HC, Kliebenstein DJ. 2010.** Understanding the evolution of defense  
445 metabolites in *Arabidopsis thaliana* using genome-wide association mapping. *Genetics* **185**: 991–  
446 1007.
- 447 **Chialva M, Salvioli di Fossalunga A, Daghino S, Ghignone S, Bagnaresi P, Chiapello M,**  
448 **Novero M, Spadaro D, Perotto S, Bonfante P. 2018.** Native soils with their microbiotas elicit a  
449 state of alert in tomato plants. *New Phytologist*.
- 450 **Coley PD, Bryant JP, Chapin FS. 1985.** Resource Availability and Plant Antiherbivore  
451 Defense. *Science* **230**: 895–899.
- 452 **Ernst WHO. 1998.** Sulfur metabolism in higher plants: potential for phytoremediation.  
453 *Biodegradation* **9**: 311–318.
- 454 **Fiehn O. 2016.** Metabolomics by Gas Chromatography-Mass Spectrometry: Combined Targeted  
455 and Untargeted Profiling. *Current Protocols in Molecular Biology*. Hoboken, NJ, USA: John  
456 Wiley & Sons, Inc., 30.4.1-30.4.32.
- 457 **Francisco M, Joseph B, Caligagan H, Li B, Corwin JA, Lin C, Kerwin R, Burow M,**  
458 **Kliebenstein DJ. 2016a.** The Defense Metabolite, Allyl Glucosinolate, Modulates *Arabidopsis*  
459 *thaliana* Biomass Dependent upon the Endogenous Glucosinolate Pathway. *Frontiers in Plant*  
460 *Science* **7**: 774.
- 461 **Francisco M, Joseph B, Caligagan H, Li B, Corwin JA, Lin C, Kerwin RE, Burow M,**  
462 **Kliebenstein DJ. 2016b.** Genome Wide Association Mapping in *Arabidopsis thaliana* Identifies  
463 Novel Genes Involved in Linking Allyl Glucosinolate to Altered Biomass and Defense.  
464 *Frontiers in Plant Science* **7**: 1010.
- 465 **Goggin FL, Avila CA, Lorence A. 2010.** Vitamin C content in plants is modified by insects and  
466 influences susceptibility to herbivory. *BioEssays* **32**: 777–790.
- 467 **Gomez Casati DF. 2016.** Metabolomics Applications in Plant Biotechnology. *Journal of*

- 468 *Postgenomics Drug & Biomarker Development* **6**.
- 469 **Heil M. 2008.** Indirect defence via tritrophic interactions. *New Phytologist* **178**: 41–61.
- 470 **Hubbard CJ, Brock MT, Diepen LT van, Maignien L, Ewers BE, Weinig C. 2017.** The plant  
471 circadian clock influences rhizosphere community structure and function. *The ISME Journal*,  
472 *Published online: 20 October 2017; | doi:10.1038/ismej.2017.172.*
- 473 **Kärkönen A, Murigneux A, Martinant J-P, Pepey E, Tatout C, Dudley BJ, Fry SC. 2005.**  
474 UDP-glucose dehydrogenases of maize: a role in cell wall pentose biosynthesis. *The Biochemical*  
475 *journal* **391**: 409–15.
- 476 **Kerwin R, Feusier J, Corwin J, Rubin M, Lin C, Muok A, Larson B, Li B, Joseph B,**  
477 **Francisco M, et al. 2015.** Natural genetic variation in *Arabidopsis thaliana* defense metabolism  
478 genes modulates field fitness. *eLife* **4**: e05604.
- 479 **Kerwin RE, Feusier J, Muok A, Lin C, Larson B, Copeland D, Corwin JA, Rubin MJ,**  
480 **Francisco M, Li B, et al. 2017.** Epistasis × environment interactions among *Arabidopsis*  
481 *thaliana* glucosinolate genes impact complex traits and fitness in the field. *New Phytologist* **215**:  
482 1249–1263.
- 483 **Khan AL, Waqas M, Asaf S, Kamran M, Shahzad R, Bilal S, Khan MA, Kang S-M, Kim**  
484 **Y-H, Yun B-W, et al. 2017.** Plant growth-promoting endophyte *Sphingomonas* sp. LK11  
485 alleviates salinity stress in *Solanum pimpinellifolium*. *Environmental and Experimental Botany*  
486 **133**: 58–69.
- 487 **Kliebenstein DJ. 2004.** Secondary metabolites and plant/environment interactions: a view  
488 through *Arabidopsis thaliana* tinged glasses. *Plant, Cell and Environment* **27**: 675–684.
- 489 **Kliebenstein DJ, Kroymann J, Brown P, Figuth A, Pedersen D, Gershenzon J, Mitchell-**  
490 **Olds T. 2001a.** Genetic control of natural variation in *Arabidopsis* glucosinolate accumulation.  
491 *Plant physiology* **126**: 811–25.
- 492 **Kliebenstein DJ, Kroymann J, Mitchell-Olds T. 2005.** The glucosinolate–myrosinase system  
493 in an ecological and evolutionary context. *Current Opinion in Plant Biology* **8**: 264–271.

- 494 **Kliebenstein DJ, Lambrix VM, Reichelt M, Gershenzon J, Mitchell-Olds T. 2001b.** Gene  
495 duplication in the diversification of secondary metabolism: tandem 2-oxoglutarate-dependent  
496 dioxygenases control glucosinolate biosynthesis in Arabidopsis. *The Plant cell* **13**: 681–93.
- 497 **Kroymann J, Donnerhacke S, Schnabelrauch D, Mitchell-Olds T. 2003.** Evolutionary  
498 dynamics of an Arabidopsis insect resistance quantitative trait locus. *Proceedings of the National*  
499 *Academy of Sciences* **100**: 14587–14592.
- 500 **Kumar S, Nicholas DJD, Williams EH. 1983.** Definitive <sup>15</sup>N NMR evidence that water serves  
501 as a source of ‘O’ during nitrite oxidation by *Nitrobacter agilis*. *FEBS Letters* **152**: 71–74.
- 502 **Lau JA, Lennon JT. 2011.** Evolutionary ecology of plant-microbe interactions: soil microbial  
503 structure alters selection on plant traits. *New Phytologist* **192**: 215–224.
- 504 **Love MI, Huber W, Anders S. 2014.** Moderated estimation of fold change and dispersion for  
505 RNA-seq data with DESeq2. *Genome Biology* **15**: 550.
- 506 **Malinovsky FG, Fangel JU, Willats WGT. 2014.** The role of the cell wall in plant immunity.  
507 *Frontiers in Plant Science* **5**: 178.
- 508 **Mauricio R. 1998.** Costs of Resistance to Natural Enemies in Field Populations of the Annual  
509 Plant *Arabidopsis thaliana*. *The American Naturalist* **151**: 20–28.
- 510 **Mauricio R, Rausher MD. 1997.** Experimental Manipulation of Putative Selective Agents  
511 Provides Evidence for the Role of Natural Enemies in the Evolution of Plant Defense. *Evolution*  
512 **51**: 1435.
- 513 **McMurdie PJ, Holmes S. 2013.** phyloseq: An R Package for Reproducible Interactive Analysis  
514 and Graphics of Microbiome Census Data (M Watson, Ed.). *PLoS ONE* **8**: e61217.
- 515 **McMurdie PJ, Holmes S, Hoffmann C, Bittinger K, Chen Y. 2014.** Waste Not, Want Not:  
516 Why Rarefying Microbiome Data Is Inadmissible (AC McHardy, Ed.). *PLoS Computational*  
517 *Biology* **10**: e1003531.
- 518 **de Mendiburu F. 2016.** agricolae: Statistical Procedures for Agricultural Research.
- 519 **Mitchell C, Brennan RM, Graham J, Karley AJ. 2016.** Plant Defense against Herbivorous

- 520 Pests: Exploiting Resistance and Tolerance Traits for Sustainable Crop Protection. *Frontiers in*  
521 *plant science* **7**: 1132.
- 522 **Naithani KJ, Ewers BE, Adelman JD, Siemens DH. 2014.** Abiotic and biotic controls on local  
523 spatial distribution and performance of *Boechera stricta*. *Frontiers in Plant Science* **5**: 348.
- 524 **Newton RJ, McLellan SL, Dila DK, Vineis JH, Morrison HG, Eren AM, Sogin ML. 2015.**  
525 Sewage reflects the microbiomes of human populations. *mBio* **6**: e02574.
- 526 **Obata T, Fernie AR. 2012.** The use of metabolomics to dissect plant responses to abiotic  
527 stresses. *Cellular and molecular life sciences* □: *CMLS* **69**: 3225–43.
- 528 **Oksanen J. 2015.** Multivariate Analysis of Ecological Communities in R: vegan tutorial.
- 529 **Pineda A, Kaplan I, Bezemer TM. 2017.** Steering Soil Microbiomes to Suppress Aboveground  
530 Insect Pests. *Trends in Plant Science* **17**: 1613–1621.
- 531 **Pineda A, Zheng S-J, van Loon JJA, Pieterse CMJ, Dicke M. 2010.** Helping plants to deal  
532 with insects: the role of beneficial soil-borne microbes. *Trends in Plant Science* **15**: 507–514.
- 533 **Preston GM. 2004.** Plant perceptions of plant growth-promoting *Pseudomonas*. *Philosophical*  
534 *transactions of the Royal Society of London. Series B, Biological sciences* **359**: 907–18.
- 535 **Quast C, Pruesse E, Yilmaz P, Gerken J, Schweer T, Yarza P, Peplies J, Glöckner FO.**  
536 **2013.** The SILVA ribosomal RNA gene database project: improved data processing and web-  
537 based tools. *Nucleic acids research* **41**: D590-6.
- 538 **R Core Team. 2013.** R: A language and environment for statistical computing.
- 539 **Richardson AE, Barea J-M, McNeill AM, Prigent-Combaret C. 2009.** Acquisition of  
540 phosphorus and nitrogen in the rhizosphere and plant growth promotion by microorganisms.  
541 *Plant and Soil* **321**: 305–339.
- 542 **Richardson AE, Simpson RJ. 2011.** Soil microorganisms mediating phosphorus availability  
543 update on microbial phosphorus. *Plant physiology* **156**: 989–96.
- 544 **Schneider CA, Rasband WS, Eliceiri KW. 2012.** NIH Image to ImageJ: 25 years of image

- 545 analysis. *Nature methods* **9**: 671–5.
- 546 **Schranz ME, Manzaneda AJ, Windsor AJ, Clauss MJ, Mitchell-Olds T. 2009.** Ecological  
547 genomics of *Boechera stricta*: identification of a QTL controlling the allocation of methionine-  
548 vs branched-chain amino acid-derived glucosinolates and levels of insect herbivory. *Heredity*  
549 **102**: 465–74.
- 550 **Silveira RL, Stoyanov SR, Gusarov S, Skaf MS, Kovalenko A. 2013.** Plant Biomass  
551 Recalcitrance: Effect of Hemicellulose Composition on Nanoscale Forces that Control Cell Wall  
552 Strength. *Journal of the American Chemical Society* **135**: 19048–19051.
- 553 **Sze MA, Schloss PD. 2016.** Looking for a Signal in the Noise: Revisiting Obesity and the  
554 Microbiome. *mBio* **7**: e01018-16.
- 555 **Tétard Jones C, Kertesz MA, Gallois P, Preziosi RF. 2007.** Genotype by Genotype  
556 Interactions Modified by a Third Species in a Plant–Insect System. *The American Naturalist*  
557 **170**: 492–499.
- 558 **Urbanczyk-Wochniak E, Fernie AR. 2005.** Metabolic profiling reveals altered nitrogen  
559 nutrient regimes have diverse effects on the metabolism of hydroponically-grown tomato  
560 (*Solanum lycopersicum*) plants. *Journal of Experimental Botany* **56**: 309–321.
- 561 **van de Voorde TFJ, van der Putten WH, Bezemer TM. 2012.** Soil inoculation method  
562 determines the strength of plant–soil interactions. *Soil Biology and Biochemistry* **55**: 1–6.
- 563 **de Vos M, Kim JH, Jander G. 2007.** Biochemistry and molecular biology of Arabidopsis–  
564 aphid interactions. *BioEssays* **29**: 871–883.
- 565 **Wagner M, Mitchell-Olds T. 2017.** Plasticity Of Plant Defense And Its Evolutionary  
566 Implications In Wild Populations Of *Boechera stricta*. *bioRxiv*.
- 567 **War AR, Paulraj MG, Ahmad T, Buhroo AA, Hussain B, Ignacimuthu S, Sharma HC.**  
568 **2012.** Mechanisms of plant defense against insect herbivores. *Plant signaling & behavior* **7**:  
569 1306–20.
- 570 **Wetzel WC, Kharouba HM, Robinson M, Holyoak M, Karban R. 2016.** Variability in plant

571 nutrients reduces insect herbivore performance The performance and population dynamics of  
572 insect herbivores depend on the nutritive and defensive traits of their host plants.

573 **Wickham H. 2009.** *ggplot2: Elegant Graphics for Data Analysis*. New York: Springer-Verlag  
574 New York.

575 **Windsor AJ, Reichelt M, Figuth A, Svatoš A, Kroymann J, Kliebenstein DJ, Gershenzon J,**  
576 **Mitchell-Olds T. 2005.** Geographic and evolutionary diversification of glucosinolates among  
577 near relatives of *Arabidopsis thaliana* (Brassicaceae). *Phytochemistry* **66**: 1321–1333.

578 **Wittstock U, Burow M. 2010.** Glucosinolate breakdown in *Arabidopsis*: mechanism, regulation  
579 and biological significance. *The Arabidopsis book* **8**: e0134.

580 **Xia J, Wishart DS, Xia J, Wishart DS. 2016.** Using MetaboAnalyst 3.0 for Comprehensive  
581 Metabolomics Data Analysis. *Current Protocols in Bioinformatics*. Hoboken, NJ, USA: John  
582 Wiley & Sons, Inc., 14.10.1-14.10.91.

583 **Zebelo SA, Maffei ME. 2015.** Role of early signaling events in plant-insect interactions.  
584 *Journal of Experimental Botany* **66**: 435–448.

585 **Zhou S, Lou Y-R, Tzin V, Jander G. 2015.** Alteration of Plant Primary Metabolism in  
586 Response to Insect Herbivory. *Plant physiology* **169**: 1488–98.

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595 **Tables and Figures**

596 **Table 1:** Differences in glucosinolate concentration between populations and microbiome  
 597 treatments. Letters denote significant differences between population and microbiome  
 598 treatments.

Glucosinolate	Population ( $\mu\text{mol g dry weight}^{-1}$ )		Microbiome Treatment ( $\mu\text{mol g dry weight}^{-1}$ )	
	Crow Creek	Road 234	Intact	Disrupted
methylethyl	200.1 $\pm$ 14.8 <b>A</b>	205.9 $\pm$ 9.8 <b>A</b>	210.2 $\pm$ 8.2 <b>A</b>	191.7 $\pm$ 17.9 <b>A</b>
GLS_6.6	0.8 $\pm$ 0.1 <b>A</b>	1.3 $\pm$ 0.1 <b>B</b>	1.1 $\pm$ 0.1 <b>A</b>	1.1 $\pm$ 0.1 <b>A</b>
GLS_7.4	3.1 $\pm$ 0.3 <b>A</b>	4.2 $\pm$ 0.2 <b>B</b>	3.8 $\pm$ 0.2 <b>A</b>	3.8 $\pm$ 0.4 <b>A</b>
methylpropyl	25.6 $\pm$ 5.3 <b>A</b>	43.8 $\pm$ 6.1 <b>B</b>	33.5 $\pm$ 5.8 <b>A</b>	46.2 $\pm$ 7.0 <b>A</b>
6- methylsulfinylhexyl	35.3 $\pm$ 4.3 <b>A</b>	45.8 $\pm$ 4.0 <b>A</b>	45.3 $\pm$ 3.3 <b>A</b>	36.4 $\pm$ 6.4 <b>A</b>
GLS_9.6	1.1 $\pm$ 0.3 <b>A</b>	1.7 $\pm$ 0.2 <b>A</b>	1.2 $\pm$ 0.1 <b>A</b>	2.1 $\pm$ 0.3 <b>B</b>
7- methylsulfinylheptyl	0.5 $\pm$ 0.01 <b>A</b>	0.4 $\pm$ 0.02 <b>B</b>	0.4 $\pm$ 0.03 <b>A</b>	0.4 $\pm$ 0.04 <b>A</b>
GLS_12.6	0.5 $\pm$ 0.1 <b>A</b>	0.9 $\pm$ 0.1 <b>B</b>	0.8 $\pm$ 0.1 <b>A</b>	0.8 $\pm$ 0.1 <b>A</b>
GLS_13.9	0.9 $\pm$ 0.1 <b>A</b>	3.7 $\pm$ 0.9 <b>B</b>	3.3 $\pm$ 0.1 <b>A</b>	2.3 $\pm$ 1.0 <b>A</b>

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603 **Table 2:** Population and microbiome treatment influence several primary metabolic pathways.  
604 Parentheses indicate the population or treatment in which metabolites were in the highest  
605 concentration.

<b>Population</b>		<b>Microbiome Treatment</b>	
Pathway	P – Value	Pathway	P – Value
Ascorbate and aldarate metabolism ( <b>Crow Creek</b> )	0.002	Citrate cycle ( <b>Disrupted</b> )	0.002
Phenylpropanoid biosynthesis ( <b>Crow Creek</b> )	0.009	Pentose and glucuronate interconversions ( <b>Intact</b> )	0.012
Glyoxylate and dicarboxylate metabolism ( <b>Crow Creek</b> )	0.041	Glyoxylate and dicarboxylate metabolism ( <b>Disrupted</b> )	0.025
		Pentose phosphate pathway ( <b>Disrupted</b> )	0.027

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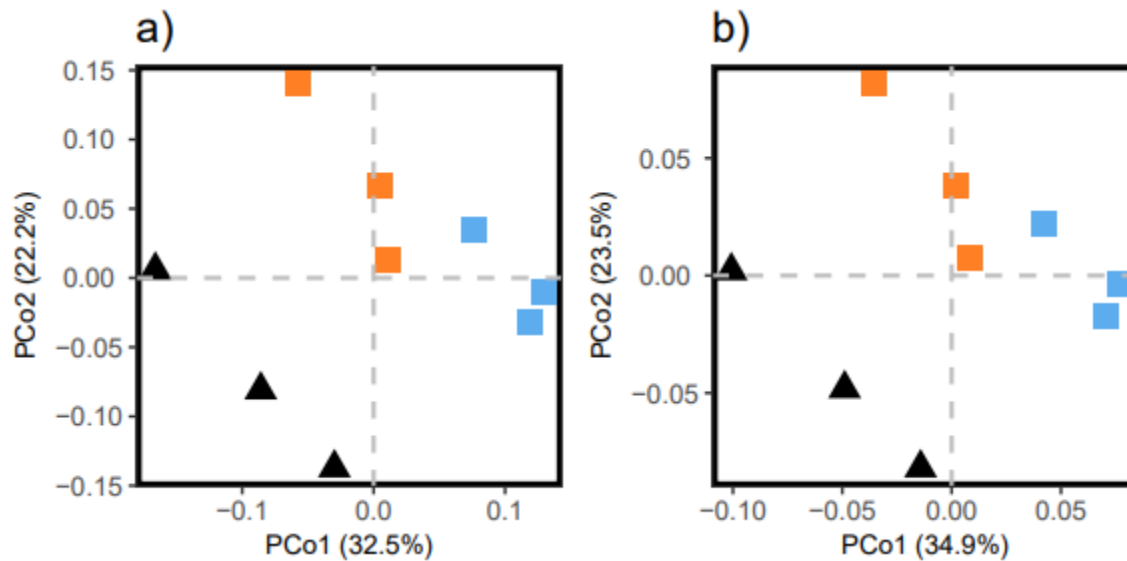


610 **Table 3:** ANOVA table for measures of insect herbivore presence and damage.

	df	F value	P
<b>Aphid</b>			
<b>Experiment 1</b>			
<i>Population</i>	1, 86	2.29	0.025
<i>Microbiome Status</i>	2, 86	26.96	> 0.001
<i>Interaction</i>	2, 86	0.94	0.39
<b>Experiment 2</b>			
<i>Microbiome</i>	1, 34	6.56	0.015
<b>Flea Beetles</b>			
<b>Leaf Damage</b>			
<i>Population</i>	4, 133	2.20	0.072
<i>Microbiome Status</i>	3, 133	28.09	> 0.001
<i>Interaction</i>	12, 133	0.96	0.488
<b>Whole Plant Damage</b>			
<i>Population</i>	4, 140	0.28	0.89
<i>Microbiome Status</i>	3, 140	11.76	> 0.001
<i>Interaction</i>	12, 140	0.60	0.84

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614 **Figure 1:** Rhizosphere communities differ between intact and disrupted microbiome treatments.

615 Squares represent plants grown in the intact microbiome treatments and triangles represent plants

616 grown in the disrupted microbiome treatment. Plants were grown in sterilized potting mix

617 inoculated with Crow Creek (orange) or Road 234 (blue) microbiomes, or the disrupted

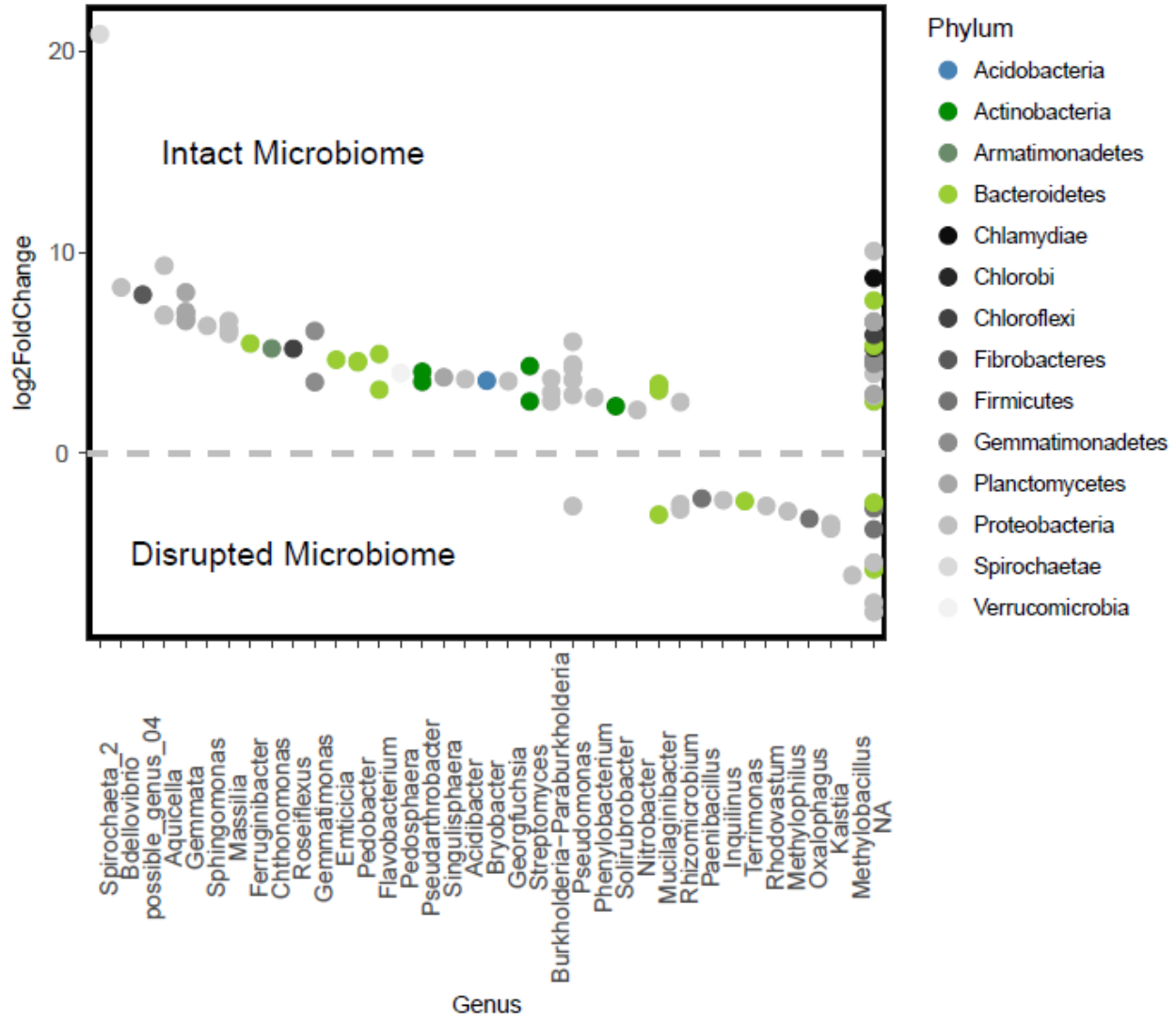
618 microbiome (black). **a)** Principal coordinate analysis of Jaccard dissimilarities ( $n = 9$ ). Intact and

619 disrupted microbiome treatments differ significantly in the presence-absence of bacterial taxa ( $P$

620 = 0.014). **b)** Principal coordinate analysis of Bray-Curtis dissimilarities ( $n = 9$ ). Intact and

621 disrupted microbiome treatments differ significantly in the abundance of bacterial taxa ( $P =$

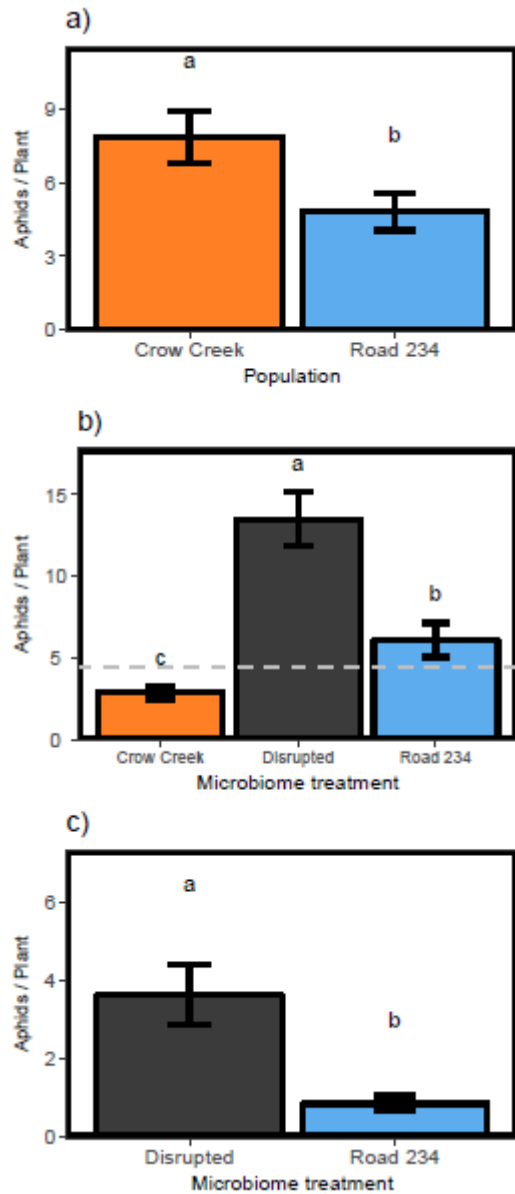
622 0.011).



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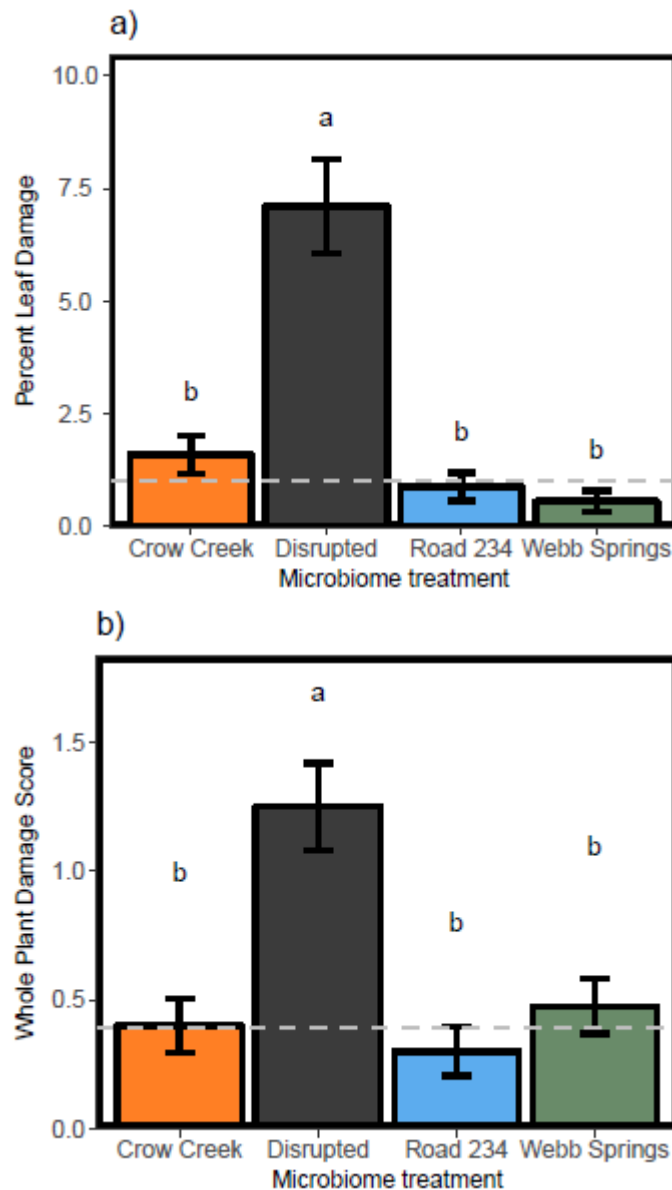
624 **Figure 2:** Bacterial taxa differentially abundant between intact (above the line) and disrupted

625 (below the line) microbiome treatments.



626

627 **Figure 3:** Population and microbiome treatment influences aphid prevalence. **a)** In the *fall 2015*  
628 experiment, plants from the Road 234 population ( $n = 44$ ) had significantly fewer aphids than  
629 plants from the Crow Creek population ( $P = 0.027$ ;  $n = 48$ ). **b)** In the *fall 2015* experiment, plants  
630 grown in the disrupted microbiome treatment ( $n = 20$ ) had significantly more aphids per plant  
631 than plants grown in the intact Crow Creek ( $n = 37$ ) and Road 234 ( $n = 35$ ) microbiome  
632 treatments ( $P < 0.001$ ). The gray dashed line is the mean number of aphids per plant in intact  
633 microbiome treatments. **c)** In the *fall 2016* experiment, plants grown in the disrupted microbiome  
634 treatment ( $n = 35$ ) had significantly more aphids than plants grown in the intact Road 234 ( $n =$   
635 30) microbiome treatment ( $P = 0.003$ ).



636

637 **Figure 4:** Microbiome treatment influences plant susceptibility to flea beetle damage. Gray  
638 dashed lines represent the mean damage for plants grown in the intact microbiome treatment. **a)**  
639 Plants grown in the disrupted microbiome treatment ( $n = 37$ ) experienced significantly more leaf  
640 damage than plants grown in the Crow Creek ( $n = 38$ ), Road 234 ( $n = 39$ ), and Webb Springs ( $n$   
641  $= 39$ ) microbiome treatments ( $P < 0.001$ ). **b)** Plants grown in the disrupted microbiome treatment  
642 ( $n = 40$ ) experienced significantly more whole plant damage than plants grown in the intact  
643 Crow Creek ( $n = 40$ ), Road 234 ( $n = 40$ ), and Webb Springs ( $n = 40$ ) microbiome treatments ( $P <$   
644  $0.001$ ).

645 **Supplemental Tables and Figures**

646 **Supplemental Table 1:** Sites where seeds and soils for inoculate were collected.

647 **Supplemental Table 2:** Number of reads at each step of sequence data processing.

648 **Supplemental Table 3:** Primary metabolites significantly affected by microbiome status,  
649 population and/or their interaction.

650 **Supplemental Figure 1:** Populations experienced similar amounts of leaf and whole plant flea  
651 beetle damage.