

1 **The compound topology of a continent-wide interaction network** 2 **explained by an integrative hypothesis of specialization.**

3 Gabriel Moreira Felix^{1,2*} (gabrielfelixmf@gmail.com), Rafael Barros Pereira Pinheiro¹ ([rafael-](mailto:rafael-bpp@hotmail.com)
4 **bpp@hotmail.com**), Robert Poulin³(robert.poulin@otago.ac.nz), Boris R. Krasnov⁴
5 (krasnov@bgu.ac.il), Marco Aurelio Ribeiro de Mello¹ (marmello@gmail.com).

6 ¹Graduate School in Ecology, Conservation, and Wildlife Management, Federal University of Minas
7 Gerais, 31270-901 Belo Horizonte, MG, Brazil. ²Departament of Animal Biology, Institute of Biology,
8 University of Campinas (Unicamp), Campinas, SP, 13083-970, Brazil. ³Department of Zoology,
9 University of Otago, PO Box 56, Dunedin, New Zealand. ⁴Mitrani Department of Desert Ecology, Swiss
10 Institute for Dryland Environmental and Energy Research, Jacob Blaustein Institutes for Desert Research,
11 Ben-Gurion University of the Negev, Sede-Boqer Campus, 8499000 Midreshet Ben-Gurion, Israel.

12 ***Corresponding Author:** Gabriel Moreira Felix, Departament of Animal Biology, Institute of Biology,
13 University of Campinas (Unicamp), Campinas, SP, 13083-970, Brazil. gabrielfelixmf@gmail.com.

14 **Statement of authorship:** All authors conceived and designed the study. BRK and RP provided the
15 database. GMF and RBPP performed the analysis. All authors discussed the results. GMF, RBPP, and
16 MARM prepared the final version of the paper with inputs from all authors.

17 **Keywords:** Trade-offs, Resource breadth, Nestedness, Modularity, Ecological networks, Ecological
18 Specialization

19

20

21

22

23

24

25

26 **ABSTRACT**

27 Is there a prevalent pattern among interaction networks: nestedness or modularity? Must
28 consumers always trade-off generalism for average performance in resource
29 exploitation? These two questions have been addressed in various systems, with
30 contradictory results. A recent integrative hypothesis combines both questions within a
31 common theoretical framework, proposing that ecological specialization is structured by
32 different prevailing processes in smaller and larger network units. This should produce
33 both a compound interaction network, formed by internally nested modules, and a scale-
34 dependence on the relationship between consumer performance and generalism. Here,
35 we confirm both predictions in a large dataset on host-parasite interactions. We show
36 that modules indeed constrain nestedness at the whole network level, and that the
37 relationship between parasite generalism and performance on their hosts changed from
38 negative at large to positive at small scales. Our results shed light on both debates, and
39 provide some clues to their integration and solution.

40

41

42

43

44

45

46

47

48

49 INTRODUCTION

50 Darwin's "tangled bank" of species interactions is one of the most complex phenomena
51 in nature. In the past decades, ecologists have imported analytical tools from network
52 science (Barabási 2016) to disentangle this complexity (Bascompte & Jordano 2013).
53 Despite the progress made in describing pervasive patterns and underlying mechanisms
54 (Vazquez *et al.* 2009), some aspects of the architecture of ecological networks remain
55 controversial, begging for further investigation (Dormann *et al.* 2017). One of those
56 unanswered questions concerns what should be the predominant topology among
57 ecological networks: nestedness or modularity (Thebault & Fontaine 2010).

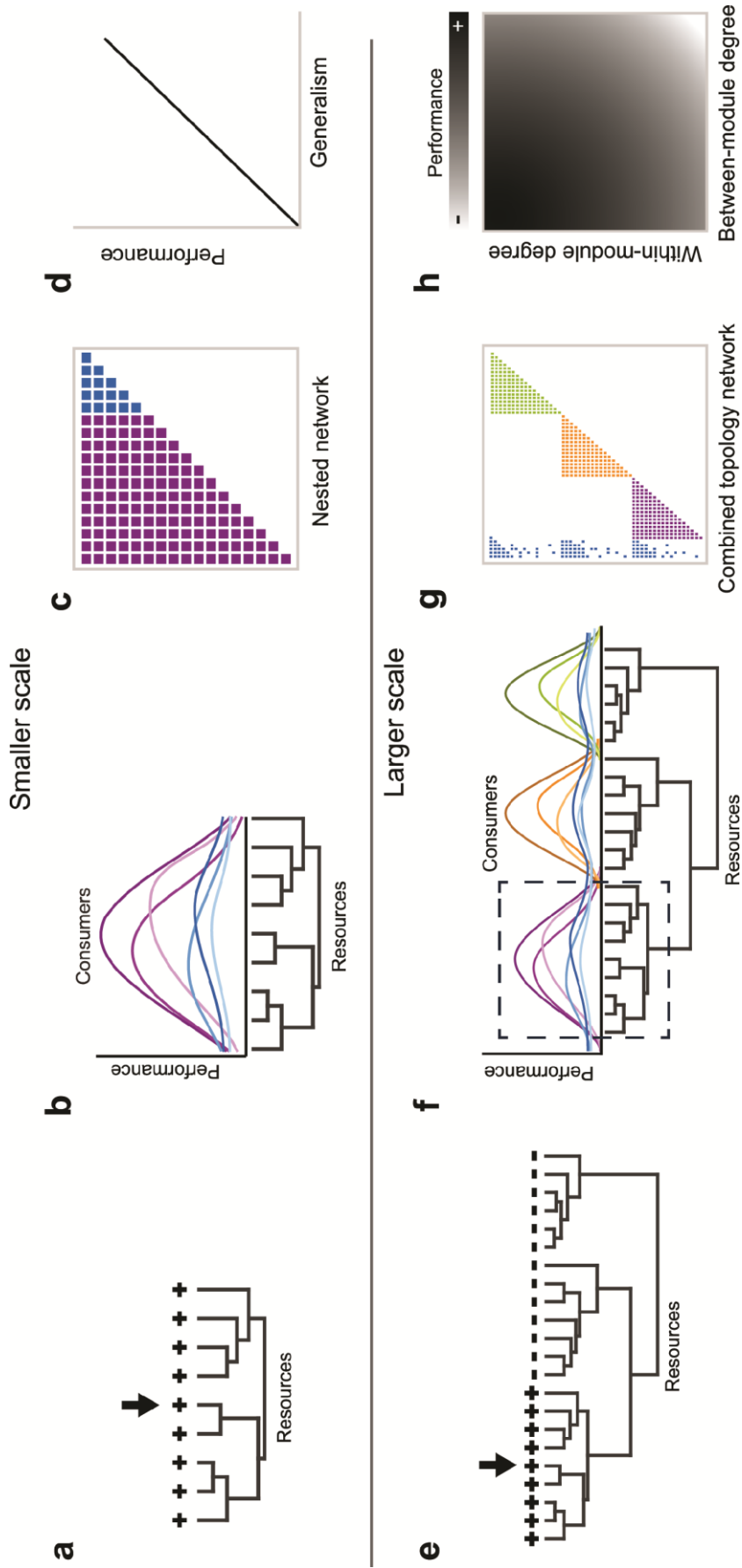
58 By adapting the biogeographic concept of nestedness (Atmar & Patterson 1993) to
59 interaction matrices, Bascompte *et al.* (2003) showed that several plant-animal networks
60 have a nested topology, with the interactions of specialists tending to be subsets of the
61 interactions of generalists. Later studies also found nestedness in many other mutualistic
62 (Ollerton *et al.* 2003, 2007; Guimaraes *et al.* 2006) and antagonistic networks (Vázquez
63 *et al.* 2007; Graham *et al.* 2009). Meanwhile, another topology, modularity, has also
64 been widely reported (Olesen *et al.* 2007; Dupont & Olesen 2009; Mello *et al.* 2011;
65 Krasnov *et al.* 2012). Since a modular network is one composed of modules of species
66 that interact more frequently with one another than with other species of the same
67 network, forbidden interactions between modules should constrain nestedness.
68 Therefore, the two topologies seem to be mutually exclusive. Nevertheless, some
69 ecological networks present high scores of both nestedness and modularity, and a
70 positive relationship between these two topologies has even been reported (Fortuna *et*
71 *al.* 2010).

72 Thus, a new question arose: how can a network be both modular and nested at the same
73 time? A possible answer states that a dual nested-modular structure would arise if each

74 topology predominates at different structural scales of the network (Lewinsohn *et al.*
75 2006). Specifically, the authors suggested that some plant-animal networks are modular
76 at the scale of the entire network, but their modules are internally nested. This kind of
77 multi-scale architecture was named a compound (or combined) topology. Later, some
78 empirical studies found evidence of compound topologies in pollination (Bezerra *et al.*
79 2009), seed dispersal (Sarmiento *et al.* 2014), and phage-bacteria networks (Flores *et al.*
80 2013). In addition, theoretical studies confirmed this topology in simulated host-parasite
81 networks (Beckett & Williams 2013; Leung & Weitz 2016).

82 Recently, an “integrative hypothesis of specialization” (IHS) was advanced, which
83 proposed a mechanism by which a compound topology might emerge in ecological
84 networks (Pinheiro *et al.* 2016). If its logic is correct, this hypothesis would also help to
85 solve another important ecological controversy: what is the expected relationship
86 between the resource range (generalism) of a given consumer and its average
87 performance in exploring these resources (Futuyma & Moreno 1988)? The IHS states
88 that this relationship should change with scale, from negative across the network as a
89 whole to positive within each module, which should lead to the emergence of an
90 interaction network formed by internally nested modules. The rationale of the IHS is
91 briefly described in Box 1 and Fig. 1.

92 Despite their potential to help solve two important ecological debates, and to improve
93 our understanding of the structure of ecological communities, neither the compound
94 topology nor the IHS has been widely tested. Here, we address these two issues in an
95 extensive host-parasite data set composed of flea-mammal interactions in 15 Palearctic
96 regions. Since this data set was collected at large phylogenetic and geographic scales, it
97 is a good model to test the relationship between generalism and performance, and also
98 the existence of a compound topology.



99

100 **Figure 1:** The integrative hypothesis of specialization. Explanations are given in Box 1.

101 First, we adapted the method described by Flores *et al.* (2013) and developed a general
102 framework to test nestedness at different network scales. Then, we tested whether the
103 flea-mammal networks have a compound topology. This question was addressed in both
104 the global network (formed by pooling together interactions reported in different
105 regions) and the local networks (formed by interactions reported for each region).
106 Second, we tested whether the relationship between host range (generalism) and
107 performance in fleas is scale-dependent, changing from positive within clusters of
108 similar resources (within each module) to negative between clusters (between modules).
109 This second question was addressed using only the local networks, since constraints in
110 the global network are mainly geographic (two species need to co-occur in the same site
111 to interact with one another) and would not reflect trade-offs in specialization. Our
112 results shed light on both debates, and provide some clues to their integration and
113 solution.

114 **METHODS**

115 **Data set**

116 We used an extensive host-parasite data set that has been analyzed in several studies on
117 ecological interactions (*e.g.*, Krasnov *et al.* 2004, 2008; Vázquez *et al.* 2007; Fortuna *et*
118 *al.* 2010). It is composed of dozens of flea-mammal interaction matrices sampled all
119 around the world, from which we selected 15 Palearctic regions (see Table S1 in
120 Appendix S2) to maximize two parameters: the size of the matrix (at least 10 parasite
121 species and 10 host species per region), and the number of hosts sampled (more than
122 1,000 individual mammals per region).

123 The global matrix with all 15 regions pooled has a size of 263 species (nodes: 161 fleas
124 and 102 mammals), and contains 1,200 interaction records (links). The local networks
125 have an average size of 45.06 ± 12.64 species (mean \pm standard deviation), with $26.26 \pm$

126 9.42 fleas and 18.8 ± 4.79 mammals, and contain on average 129.6 ± 57.22 interaction
127 records (Appendix S2: Table S1).

128 Furthermore, the global matrix and some local matrices produced networks with more
129 than one component, *i.e.*, a cluster of species totally separated from the other nodes of
130 the network. In most of these networks, there is a giant component comprising most of
131 the network nodes and one or few minor components, each one including a small
132 number of nodes. The analyses below, at global and local scales, were carried out by
133 using only the respective binary version of the largest component of each matrix.

134 **Network topology**

135 *Modularity*

136 The first step to test for a compound topology is to unfold the modular structure of the
137 network. We did this computing the Barber modularity (Q) (Barber 2007) optimized by
138 the DIRTLPAwb+ algorithm (Beckett 2016), through the *computeModules* function of
139 the *bipartite* package (Dormann *et al.* 2008) for R (R Development Core Team 2017).
140 Modularity (Q) varies from 0 to 1, and the algorithm reveals also the number and
141 composition of the modules found in the network.

142 *Nestedness*

143 A nested matrix has its interactions arranged in a particular way: the interactions of the
144 least connected species are proper subsets of the interactions of the more connected
145 species (Ulrich *et al.* 2009). NODF is a metric that aims to synthesize this pattern in a
146 single number (Almeida-Neto *et al.* 2008). In its default procedure, a NODF score is
147 computed for each pair of species (independently for consumers and resources, *i.e.*,
148 rows and columns of the interaction matrix) and, then, averaged to calculate the NODF
149 score of the whole matrix. This procedure implicitly assumes that nestedness is evenly

150 distributed in the matrix. However, as pointed out by the authors of NODF themselves,
151 it is important to “*explore whether nestedness is a general pattern of the community or*
152 *derives from some particular species subsets*” (Almeida-Neto *et al.* 2008). If different
153 species subsets of the matrix have different degrees of nestedness, an overall NODF is
154 not an appropriate summary of the matrix structure (Gotelli & Ulrich 2012). It turns out
155 that this is exactly the case if the network has a compound topology, where nestedness
156 between pairs of species of the same module should be much higher than nestedness
157 between pairs of species of different modules.

158 In order to solve this problem, we adapted the method described in Flores *et al.* (2013),
159 and averaged nestedness independently between pairs of species of the same module
160 ($NODF_{SM}$), and between pairs of species of different modules ($NODF_{DM}$), and
161 compared those values with those expected by species degrees under two scenarios (null
162 models): in the absence of a modular structure (free null model) and in the presence of a
163 modular structure (restricted null model). The rationale behind using these two null
164 models is explained below, while the detailed instructions for performing both null
165 models are presented in Appendix S1.

166 *Predictions*

167 If nestedness and modularity coexist at large network scales as two sides of the same
168 coin, we expected nestedness between species of different modules ($NODF_{DM}$) to be
169 equal to or higher than expected by their degrees (*i.e.*, the free null model). Otherwise, if
170 modularity constrains nestedness between pairs of species of different modules, we
171 expect $NODF_{DM}$ to be smaller than expected by their degrees.

172 Notice, however, that NODF between pairs of species of the same module ($NODF_{SM}$)
173 will be higher than expected by species degrees whether or not modules are internally
174 nested. This would happen since, by definition, species of the same module share more

175 interactions with one another than expected by their degrees, regardless of those
176 interactions being nested or not. Hence, a $NODF_{SM}$ value higher than expected by the
177 free null model is a necessary, but not sufficient condition, for a network to have a
178 compound topology. This is the reason why we need the restricted null model – which
179 also conserves the modular structure of the original matrix when generating the null
180 matrices (Additional Information 1) –, to test whether $NODF_{SM}$ is higher than expected
181 *given* the modular structure. On the one hand, if the network is formed by modules that
182 are not internally nested, $NODF_{SM}$ should be higher than expected by the free null
183 model, but lower than expected by the restricted null model. On the other hand, if the
184 network is formed by internally nested modules (*i.e.*, a compound topology), $NODF_{SM}$
185 should be higher than expected by both the free and restricted null models.

186 But why did we not individualize each module and, then, test its nestedness
187 independently? Although this would be a valid procedure to test $NODF_{SM}$, it would not
188 allow to test whether interactions between pairs of species of different modules
189 ($NODF_{DM}$) are more nested than expected *given* the constraints imposed by the
190 modules. This can be done by comparing the observed $NODF_{DM}$ with that expected by
191 the restricted null model.

192 *Z-Score*

193 For each main component of the 16 networks (the global network and 15 local
194 networks), we generated 1,000 random matrices using the free null model and another
195 1,000 matrices using the restricted null model. Next, for each random matrix, we
196 computed its overall NODF and decomposed it into $NODF_{SM}$ and $NODF_{DM}$ using the
197 observed partitions of their corresponding real network.

198 Finally, for all combination of matrices (16 in total: 1 global and 15 local), null models
199 (2: free and restricted), and NODF metrics (3: NODF, $NODF_{SM}$ and $NODF_{DM}$), a Z-

200 score was calculated as $Z = [\text{Value}_{\text{obs}} - \text{mean}(\text{Value}_{\text{sim}})] / \sigma(\text{Value}_{\text{sim}})$, where $\text{Value}_{\text{obs}}$ is
201 the observed value of the metric and $\text{Value}_{\text{sim}}$ represents the values of the metric in the
202 randomized matrices. Observed and expected modularity values were also compared
203 using Z-scores, but only for the free null model, as it does not make sense to compare
204 observed and expected modularities with a null model that fixes the modules.

205 Nestedness and modularity standardized by null models will be called relative
206 nestedness and relative modularity, respectively. For simplicity, they will be represented
207 here as Z_F or Z_R , depending on the null model, followed by the metric name (*e.g.*, Z_FQ
208 and $Z_F\text{NODF}_{\text{SM}}$ represent, respectively, relative modularity and relative nestedness
209 between pairs of species of the same module, when standardized by the free null model).

210 Our goal was to see how modularity and nestedness interact with each other in a
211 continuous way. Therefore, in all analyses we used the original Z-scores, without
212 classifying them as significant and non-significant.

213 *Matrix plotting*

214 The interaction matrices were reorganized to maximize between- and within-module
215 nestedness as done in previous studies (Flores *et al.* 2013, 2016). Briefly, we first
216 reordered the matrix rows and columns by degree without disrupting its modular
217 structure and, then, permuted the modules in order to find the arrangement of modules
218 which maximizes the overall NODF of the matrix. This procedure facilitates the
219 visualization of a compound topology, if one exists.

220 **Specialization versus performance at different scales**

221 *Performance index*

222 In the host-parasite literature, the performance of a parasite in a host is usually
223 quantified indirectly through some metric assumed to reflect it: *e.g.*, prevalence,

224 intensity, or abundance (Poulin 2007). We chose abundance: the average number of
225 individual fleas per individual mammal (calculated including infected and uninfected
226 hosts). This choice is justified as abundance is considered a good measure of
227 performance in host-parasite systems (Krasnov *et al.* 2006), since it integrates intensity
228 of infestation and prevalence in a single metric (abundance = intensity of infestation
229 *times* prevalence), measuring different aspects of parasite performance. While intensity
230 of infestation is the average number of individual parasites per infected individual host,
231 prevalence is the proportion of infected individuals in the host population.

232 *Generalism within modules and between modules*

233 For each flea species, specialization within and between modules was measured through
234 its cartographic position in the network: its within-module degree (Z , which should not
235 be confused with the Z -score of the null models) and participation coefficient (P)
236 (Guimerà & Amaral 2005).

237 These two metrics define the functional role of a species in a network, and they are
238 respectively related to the number of interactions a species makes with other species of
239 its own module and with species of other modules. Z and P values were calculated
240 independently for each local network.

241 *Mixed models*

242 We used mixed models (Bolker *et al.* 2009) to test whether flea performances are
243 positively correlated with their within-module degrees (Z) and negatively correlated
244 with their participation coefficients (P). Linear mixed models (LMMs) were built by the
245 *lmer* function of the *lme4* package (Bates *et al.* 2015), and fitted by restricted maximum
246 likelihood (REML).

247 We used the log-transformed abundance of each parasite species in each host species in

248 each region as the response variable, Z and P values for each flea species in each region
249 as the explanatory variables, and host species, parasite species and region as crossed
250 random factors. We decided to use parasite abundances per host species, rather than
251 average it between all hosts exploited by a flea, to control for host characteristics known
252 to affect abundance (e.g., carrying capacity, susceptibility, and richness of parasite
253 fauna) (Krasnov *et al.* 2005). Averaging would also decrease the power of the analysis
254 (Hopkins 1982; Schank & Koehnle 2009).

255 In addition, as pointed out in Box1 (see also Fig. 4 in Pinheiro *et al.* (2016)), in local
256 networks composed of very similar resources, we should not expect to find either a
257 negative relationship between abundance and P or modules. In those networks, the
258 modules recovered by the DIRTLPawb+ algorithm will be spurious, not imposing
259 constraints to interactions, and the measured $NODF_{DM}$ should be higher than expected
260 by species degrees (the free null model). To test this prediction, we included an
261 interaction between $Z_F NODF_{DM}$ and the fixed factors of the model (P and Z). We expect
262 $Z_F NODF_{DM}$ not to have an influence on the effect of Z on abundance, but to influence
263 the effect of P. Specifically, we expected that the effect of P on abundance should be
264 negative only in local networks in which the modular structure constrains nestedness
265 between pairs of species of different modules, that is, in local networks with negative
266 values of $Z_F NODF_{DM}$.

267 We used backward stepwise regression to select fixed and random effects, following the
268 procedure suggested by Bolker *et al.* (2009). We used the *anova* function of the *stats*
269 package to perform a likelihood ratio (LR) test on the random effects (to which the
270 models are refitted with maximum likelihood) and, then, used the *Anova* function of the
271 *car* package to perform Wald X^2 tests on the fixed effects. To tell apart the variance
272 explained by either the fixed or random factors in the minimal selected model, we used

273 the *r.squaredGLMM* function of the *MuMIn* package for R to compute both marginal
274 and conditional R squared (Nakagawa & Schielzeth 2013). The confidence intervals of
275 the parameters were obtained by bootstrapping using the *confint.merMod* function of
276 *lme4* package. The confidence interval of the conditional effect of P given $Z_F \text{NODF}_{DM}$
277 was also computed by simulation using the *interplot* function of *interplot* package.

278 **RESULTS**

279 **Topology**

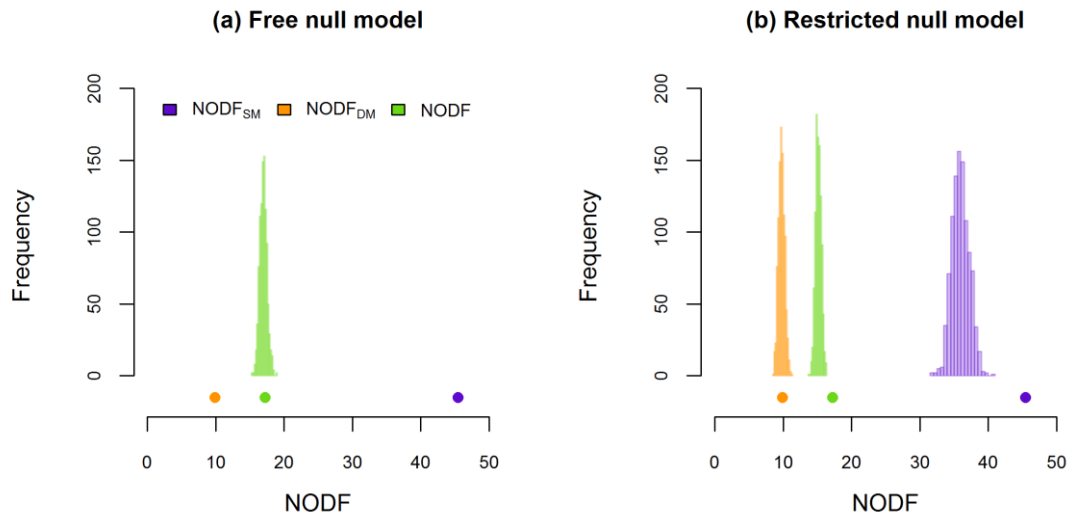
280 *Global network*

281 The global network presented higher modularity ($Z_F Q = 51.13$) and overall nestedness
282 equal ($Z_F \text{NODF} = 0.39$) to that expected by the free null model. However, the observed
283 scores of nestedness were much higher between pairs of species of the same module
284 than between pairs of species of different modules ($\text{NODF}_{SM} = 45.47$, $\text{NODF}_{DM} = 9.85$).
285 In addition, as expected if the modules constrain nestedness between species of different
286 modules, NODF_{DM} was smaller than expected by the free null model ($Z_F \text{NODF}_{DM} = -$
287 13.46) (Fig. 2a), but equal to expected by the restricted null model ($Z_R \text{NODF}_{DM} = 0.22$)
288 (Fig. 2b). Finally, nestedness between pairs of species at the same module was higher
289 than expected by both null models ($Z_F \text{NODF}_{SM} = 53.89$, $Z_R \text{NODF}_{SM} = 22.08$) (Fig. 2a-
290 b).

291 Those results strongly support the hypothesis that the global flea-mammal network has
292 a compound topology, which can be easily seen when we plot the interaction matrix
293 maximizing nestedness without disrupting its modular structure (Fig 3).

294 *Local networks*

295 Nestedness and modularity varied widely among local networks, which, in general,
296 were more nested ($Z_F \text{NODF} = 2.03 \pm 1.94$) than modular ($Z_F Q = 0.36 \pm 1.31$) (Fig. 4a).



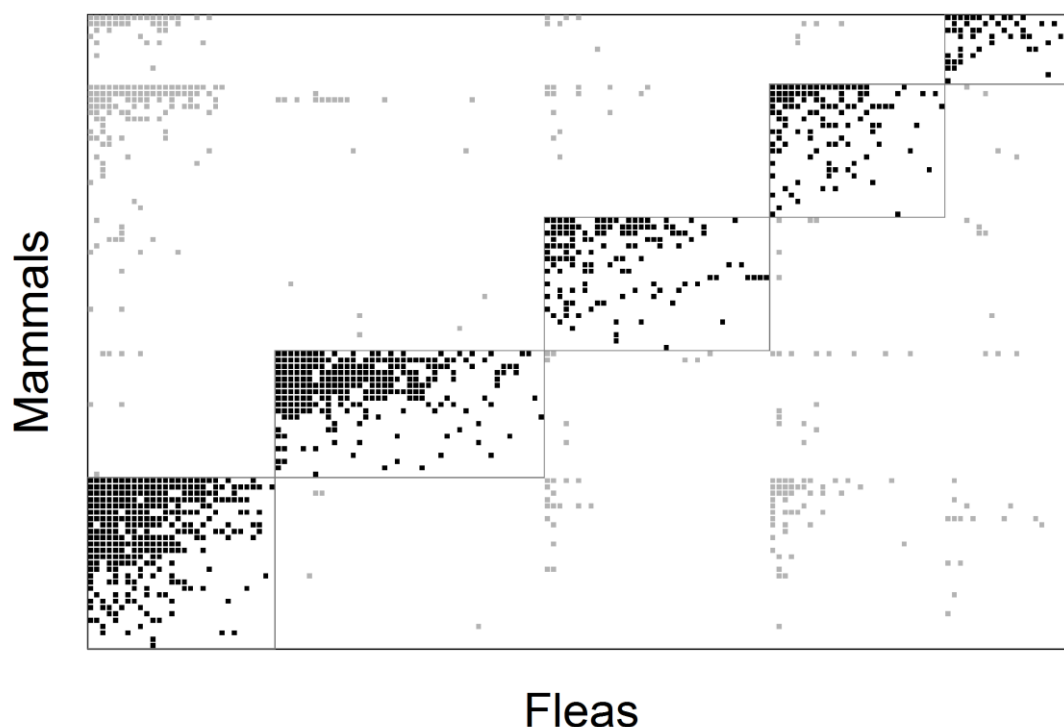
297

298 **Figure 2:** Observed values (dots) of NODF, NODF_{SM}, and NODF_{DM} in the global network contrasted
299 with the values expected by species degrees (distributions) in the absence (a: free null model) or presence
300 (b: restricted null model) of a modular structure. NODF: overall nestedness. NODF_{SM}: nestedness
301 between pairs of species of the same module. NODF_{DM}: nestedness between pairs of species of different
302 modules. As expected if the global network has a compound topology, NODF_{SM} is higher than expected
303 by both null models, and NODF_{DM} is smaller than expected by the free null model and equal to that
304 expected by the restricted null model.

305 However, some relationships between these two topologies were evident. First, both
306 observed and relative overall nestedness (NODF) (Fig. 4a, c and e; and Appendix S2:
307 Table S2) decreased as modularity increased. Second, the same pattern was true for
308 nestedness between pairs of species of different modules (NODF_{DM}) (Fig. 4a,c and e;
309 and Appendix S2: Table S2). Third, although nestedness between pairs of species of the
310 same module (NODF_{SM}) was not significantly related to modularity in any case
311 (Appendix S2: Table S2), the Z_F NODF_{SM} showed a trend to increase with Z_F Q (Fig. 4c).

312 In addition, as for the global networks, observed and relative values of NODF_{SM} were
313 higher than those of NODF_{DM} (Fig. 4b,d and f), and the difference between them
314 increased with modularity. Finally, NODF_{SM} values were higher than expected by both
315 null models, while NODF_{DM} values were smaller than expected by the free null model

316 but equal to that expected by the restricted null model (Fig. 4b,d and f).



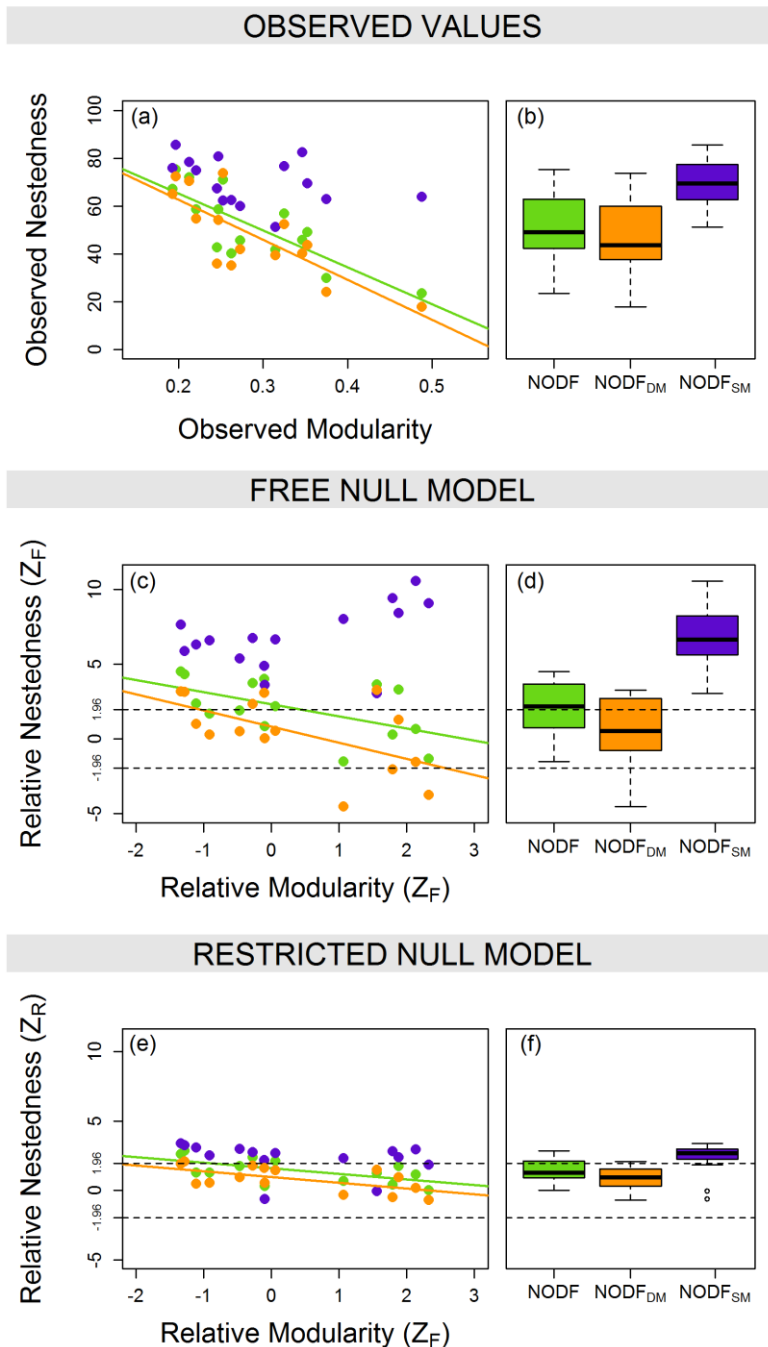
317

318 **Fig. 3:** Interaction matrix reorganized to maximize between- and within-module nestedness
319 without disrupting the modular structure of the network. Interactions within modules (delimited
320 by boxes) are showed in black, while those outside modules are showed in gray. Flea species
321 are represented in rows and mammals in columns. The compound topology of the global
322 network is evident.

323 **Specialization versus performance**

324 Random factors explained a significant portion of the variance in flea abundance
325 (Appendix S2: Table S3). In addition, as expected, only the within module degree (Z),
326 participation coefficient (P), and the interaction between P and $Z_{FNODF_{DM}}$ significantly
327 affects flea abundances. Neither $Z_{FNODF_{DM}}$ nor its interaction with Z was retained in
328 the minimum selected model (Appendix S2: Table S4 and Table S5).

329



330

331 **Figure 4:** Relationship between observed and relative (Z_F and Z_R) scores of nestedness
332 components (NODF, NODF_{SM} and NODF_{DM}) and modularity in local networks. Z_F and Z_R
333 represent the relative score of a metric (nestedness or modularity) standardized by the score
334 expected in the absence (the free null model) or in the presence (restricted null model) of the
335 modular structure, respectively. NODF (green): overall nestedness. NODF_{SM} (purple):
336 nestedness between pairs of species of the same module. NODF_{DM} (orange): nestedness between

337 pairs of species of different modules. Left panels (a, c and e): relationship between observed and
338 relative nestedness and modularity scores. Right panels (b, d and f): box plots of observed and
339 relative nestedness scores. Notice that relative modularity is always standardized by modularity
340 expected by the free null model, both in c and e. Lines show significant relationships ($p < 0.05$).

341 On the one hand, flea abundances were always positively correlated with within-module
342 degree (Z) (Fig. 5). On the other hand, as expected, the relationship between flea
343 abundance and participation coefficient (P) changed from positive to negative as the
344 $NODF_{DM}$ becomes smaller than expected by the free null model, crossing zero at
345 $Z_{FNODF_{DM}} \approx -2$ (Fig. 5). In addition, the predicted positive effect of participation
346 coefficient (P) on flea performance was higher than that of within-module degree (Z)
347 when $NODF_{DM}$ becomes equal to or higher than expected by the free null model.

348 Finally, although the complete models explained a large amount of data variance, the
349 fixed factors were responsible for only a very small fraction of the explanation ($R^2_{(m)} =$
350 0.086 , $R^2_{(c)} = 0.53$).

351 **DISCUSSION**

352 In the present study, we provide strong support for the integrative hypothesis of
353 specialization (Pineiro *et al.* 2016) using a continent-wide host-parasite network. We
354 confirmed both (i) the emergence of a compound topology in the local and global
355 networks (Fig. 2, 3 and 4); and (ii) the scale-dependence of the relationship between
356 specialization and performance (Fig. 5). Our results unite two long-standing debates in
357 the ecological literature within the same theoretical framework and provide clues to
358 their solution.

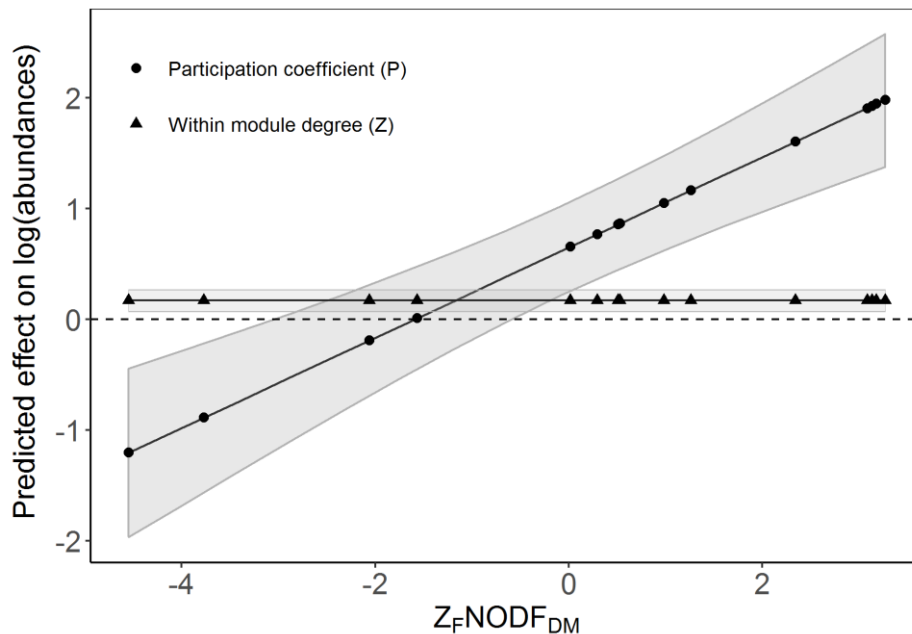
359 *Nestedness versus modularity*

360 What is the predominant topology in ecological networks: nestedness or modularity? In

361 the past decades, conflicting results have been reported, leaving us with three possible
362 scenarios. First, different systems, taxa, and interaction types lead to networks with
363 different topologies. For instance, it has already been suggested that antagonistic
364 networks tend to be modular, while mutualistic networks tend to be nested (Thebault &
365 Fontaine 2010). Second, both topologies coexist as two sides of the same coin (Fortuna
366 *et al.* 2010), with the nested structure superimposed over the modular structure. Third,
367 alternative topologies should become manifest at different network scales, resulting in a
368 compound topology (Lewinsohn *et al.* 2006). Our results provide strong support for the
369 third scenario.

370 At first glance – considering only that the global and some local flea-mammal networks
371 presented scores of nestedness and modularity higher than or equal to those expected by
372 their species degrees (free null model) (Fig. 2a and Fig. 4c-d) – one could conclude that
373 those two topologies coexist in the flea-mammal networks as two emergent properties
374 of the same underlying phenomenon. However, the observed and relative values of
375 modularity and overall nestedness were negatively correlated in the local networks (Fig.
376 4a,c and e). And while nestedness between pairs of species of the same module was
377 higher than expected by the species degrees ($NODF_{SM}$), the opposite was true for
378 nestedness between pairs of species of different modules ($NODF_{DM}$) (Fig. 4d). In
379 addition, the difference between these two sets of nestedness increased with modularity
380 (Fig. 4a,c and e). Together, those results strongly suggest that modularity constrains
381 nestedness at large topological scales of this host-parasite system.

382



383

384 **Figure 5:** The predicted effects of within-module degree (Z) and participation coefficient (P) on
385 flea performances (abundance) for different values of $Z_{FNODF_{DM}}$ (relative nestedness between
386 pairs of species at different modules in a given local network when compared to that expected
387 by the restricted null model). Lines show the predicted effects of P and Z on abundance (and its
388 95% confidence intervals), while dots indicate the fifteen local networks. As expected by the
389 IHS, the effect of P changes from negative to positive (becoming higher than the positive effect
390 of Z) as nestedness between species in different modules increases. The effect of Z on
391 abundance in a given local network is independent of $Z_{FNODF_{DM}}$.

392 However, as pointed out in the Methods section, to say that the modules constrain
393 nestedness between species of different modules and increase nestedness between
394 species of the same module is a truism, a logical consequence of the definition of
395 modules itself. In ecological terms, it is like saying that species preferences
396 (specialization) constrain resource breadth processes (Brown 1984), which is trivial. A
397 much more interesting issue would be to evaluate if interactions are more nested than

398 expected *given* the species preferences, which are reflected in the network modular
399 structure. This raises two questions.

400 First, is nestedness between pairs of species of the same module higher than expected
401 *given* that they have similar dietary preferences, *i.e.*, that they belong to the same
402 module? By comparing the observed $NODF_{SM}$ to that expected by the species degrees in
403 the presence of a modular structure (the restricted null model), we showed that this is
404 true for the flea-mammal network, which is formed by internally nested modules (Fig.
405 3).

406 Second, is nestedness between pairs of species of different modules higher than
407 expected given that they have dissimilar dietary preferences, *i.e.*, that they belong to
408 different modules? That is, when a consumer c_i of module A pervades the modular
409 structure and consumes resources from module B , will c_i consume the most consumed
410 resources of module B ? By comparing $NODF_{DM}$ to that expected by the restricted null
411 model, we show that this is also the case in the studied host-parasite system, since both
412 in the global (Fig. 2b) and the local networks (Fig. 4f) the interactions between species
413 of different modules were equally nested as expected by the restricted null model.

414 This scenario ($NODF_{SM}$ and $NODF_{DM}$ equal to or higher than expected given the
415 modular structure) suggests that, once the constraints imposed by modules are
416 overcome, the same processes that structure interactions within modules also structure
417 the few interactions outside them. This does not necessarily need to be true, since
418 competitive exclusion would predominate over resource availabilities outside the
419 modules. For example, if parasites have poor performances in hosts that do not belong
420 to their modules, diffused competition would prohibit parasites of module B to
421 successfully establish themselves in the most exploited hosts of module A . If this
422 happened, parasites would become supertramps (Diamond 1975) when exploiting hosts

423 of other modules, and an antinested pattern of interactions would emerge between pairs
424 of species of different modules (Poulin & Guégan 2000). Although it is not true in the
425 flea-mammal network, it is an interesting question to be addressed in other systems.

426 Understanding the processes governing the spillover of species interactions between
427 modules has important practical implications. It may help, for example, to predict
428 which species are the most likely to invade new habitats or which parasite species are
429 most likely to emerge in new hosts (Pimm 1991).

430 *The hypotheses of trade-offs and resource breadth*

431 What is the expected relationship between the resource range of a species and its
432 average performance at exploiting these resources (Futuyma & Moreno 1988)? As in
433 the case of network topology, conflicting results have been reported, and two scenarios
434 are possible. First, the relationship between resource breadth and average performance
435 varies among systems, taxa, places, and interaction types. Second, this relationship
436 should change at different community scales, from negative at larger scales to positive
437 at smaller scales, as predicted by the IHS (Pinheiro *et al.* 2016).

438 Our results provide strong support for the second scenario (Fig 5). In local networks
439 where modules represent a significant restriction to interactions, the relationship
440 between flea abundance and generalism changed from positive within to negative
441 among modules, as predicted by the IHS. In addition, the effect of P on abundance
442 became more negative as the modular structure imposed more constraints on the
443 interactions. Therefore, if the community (or, at least, the part we sampled) is composed
444 of more than one module, a multi-scale relationship between specialization and
445 performance should emerge. Otherwise, if the community is composed of very similar
446 resources (*i.e.*, just one internally nested module), we expect a simple positive
447 relationship between generalism and performance. A further step would be to test if the

448 contradictory results reported by previous studies which addressed the relationship
449 between performance and generalism would also be explained by differences in the
450 scale of each community studied. While some of them focused on different populations
451 of the same resource species (e.g, Szollósi *et al.* 2011) others sampled entire resource
452 communities (Poulin 1998; Hellgren *et al.* 2009).

453 Fort *et al.* (2016) showed that higher abundance implies greater generalism in
454 ecological networks, and not the contrary (but see Dorman *et al.* (2017)). That is,
455 abundant species (the most available ones) are generalist because they have higher
456 probability of finding potential interaction partners. This makes sense in the context of
457 the IHS: the species with higher performances in exploiting a given set of similar
458 resources reach higher abundances and, then, interact freely with a large number of that
459 resources in the absence of trade-offs. Alternatively, although the nestedness between
460 pairs of species of different modules was equal to the expected one *given* the modular
461 structure in most of the local networks, the negative relationship between P and
462 abundance suggests that generalism between modules has a negative effect on
463 abundance, due to trade-offs. Therefore, although a species that is a hub in its own
464 module would eventually spill over on the most connected species of other modules, if
465 this species evolves adaptations to interact with a broad spectrum of dissimilar species,
466 which makes it a hub in the entire network, this would have a negative impact on its
467 performance.

468 *Concluding remarks*

469 Some authors have pointed out that discontinuities are much more common in nature
470 than previously thought. In addition, recent theoretical models (Holt 2006; Scheffer &
471 van Nes 2006) suggest that, contrary to the principle of limiting similarity (MacArthur &
472 Levins 1967), a balance between neutral and niche processes might generate self-

473 organized clusters of similar species, which have been called emergent groups (Héroult
474 2007). The IHS and the compound topology are in complete agreement with this view.
475 In fact, the idea that trade-offs generate modules, while a coupling between the
476 availabilities of resources and consumers determines the dynamics inside each module,
477 generating nestedness, suggests that there is a balance between niche processes at large
478 scales and neutral processes at small scales.

479 Finally, we are not proposing that every ecological network should, necessarily, have a
480 compound topology. Although it is likely that all nested matrices are modular at a larger
481 scale, because any interactions should be constrained at some point, not all modules
482 need to be internally nested. Several other patterns might be observed within a module,
483 each resulting from different underlying mechanisms (Presley *et al.* 2010). The methods
484 and the conceptual framework developed in our study provide tools to investigate these
485 patterns.

486 **BOX 1 - The integrative hypothesis of specialization**

487 Earlier coined the “integrative hypothesis of parasite specialization” (IHPS) (Pinheiro *et*
488 *al.* 2016), the hypothesis was proposed in the context of host-parasite interactions as a
489 solution to an old controversy in the parasitological literature: what is the expected
490 relationship between generalism and performance of parasites (Poulin 2007)?

491 However, since its assumptions are broad enough to apply to other interactions systems,
492 we describe it here as valid for any consumer-resource system and call it “the
493 integrative hypothesis of specialization” (IHS). The IHS states that consumer-resource
494 interactions are structured by a balance between resource breadth processes (Brown
495 1984) at small and trade-offs at large community scales.

496 On the one hand, at small community scales, resources should be very similar to one
497 another and an adaptation (the arrow in Fig. 1a) that increases the intrinsic performance
498 of a consumer in exploiting a specific resource type should also be an adaptation to all
499 other resource types (the “+” in Fig 1a). Therefore, we do not expect to find preferences
500 (*e.g.*, phylogenetic, phenetic, or geographic signals) in consumer-resource interactions at
501 small scales. Instead, a coupling between resource availabilities and intrinsic consumer
502 performances should produce a nested pattern of realized performances: the resources
503 with highest availability should be more strongly exploited by all consumers, in
504 proportion to the intrinsic performances of the consumers on those resources (Fig. 1b).

505 Consequently, by sampling small scales of a community, one should find both a nested
506 pattern of interactions (Fig. 1c), and a positive relationship between generalism and
507 performance (Fig. 1d). This last pattern (Fig. 1d) is similar to the positive occupancy-
508 abundance relationships widely reported for biogeographic data (Gaston & Blackburn
509 2000), normally explained by Brown’s resource breadth hypothesis (Brown 1984).
510 Since the mechanisms described above are very similar to that proposed by Brown, but
511 adapted to a context of interactions, we summarize them as resource breadth processes.

512 On the other hand, at larger community scales, the IHS states that there are trade-offs in
513 the capacity to exploit resources of different clusters. Specifically, adaptations to a
514 specific resource type (the arrow in Fig. 1e) should also be adaptations to other similar
515 resource types of the same cluster (the “+” in Fig. 1e), but maladaptations to dissimilar
516 resources of other clusters (the “-” in Fig. 1e). Thus, the pattern of realized
517 performances described above (Fig. 1 b) should be restricted to within each resource
518 cluster (Fig. 1f). In this context, the different clusters of resources are the real units of
519 specialization and the true generalist consumers are those that can exploit resources at
520 several clusters (blue species in figure 1). Consequently, by sampling large scales of a

521 community, we do not expect to find a completely nested network, but rather a modular
522 network with internally nested modules (i.e., a *compound topology*) (Lewinsohn *et al.*
523 2006) (Fig. 1g). In addition, although we expect a positive relationship between
524 performance and generalism within each cluster of similar resources, we also expect
525 trade-offs to result in a negative relationship between performance and capacity to
526 exploit resources of different clusters (modules in the network) (Fig. 1h).

527 Krasnov *et al.* (2004) have already suggested that the relationship between performance
528 and generalism should be negative in communities composed of dissimilar resources,
529 but positive in communities composed of similar resources. Indeed, Krasnov *et al.*'s
530 suggestion and the IHS have the same underlying rationale: the probability of an
531 adaptation to a given resource of being also an adaptation to other resources (*i.e.*, the
532 resource breadth hypothesis) is higher the more similar the resources are. The IHS is
533 just a more inclusive hypothesis that also predicts the relationship between
534 specialization and performance in communities in which resource dissimilarity is not
535 gradually structured, that is, in communities with clusters of similar resources separated
536 from one another by gaps of dissimilarity (Allen 2006).

537 **SUPPLEMENTARY FILES**

538 **Appendix S1**

539 Details of both free and restricted null models computations.

540 **Appendix S2**

541 Supplementary Tables S1-S5

542 **ACKNOWLEDGEMENTS**

543 We thank our institutions and many colleagues, who helped us in different ways during
544 this project. Elisabeth Kalko, Judith Bronstein, Pedro Jordano, Nico Blüthgen, Carsten

545 Dormann, Thomas Lewinsohn, Leonardo Rê-Jorge, Cang Hui, Adriano Paglia,
546 Fernando Silveira and Erika Braga helped us with exciting discussions about ecological
547 interactions, complex networks, and community assembly rules. We thank specially
548 Prof. Thomas Lewinsohn for so many uplifting conversations about network topologies,
549 which contributed substantially to conceptual and methodological issues of this work.
550 The Graduate School in Ecology of the Federal University of Minas Gerais (ECMVS)
551 provided us with scholarships granted to G.M.F. and R.B.P.P., and with infrastructure.
552 M.A.R.M. was funded by the Alexander von Humboldt Foundation (AvH: 3.4-
553 8151/15037), Minas Gerais Research Foundation (FAPEMIG: PPM-00324-15),
554 Research Dean of the Federal University of Minas Gerais (UFMG-PRPq: 02/2014),
555 Brazilian Council for Scientific and Technological Development (CNPq: 472372/2013-
556 0), Brazilian Coordination for the Improvement of Higher Education Personnel
557 (CAPES, grad student scholarships), and Research Program of the Biodiversity of the
558 14 Atlantic Forest (PPBio-MA/CNPq: 457458/2012-7).

559 REFERENCES

- 560 Allen, C.R. (2006). Discontinuities in ecological data. *Proc. Natl. Acad. Sci.*, 103,
561 6083–6084.
- 562 Almeida-Neto, M., Guimarães, P., Guimarães, P.R., Loyola, R.D. & Ulrich, W. (2008).
563 A consistent metric for nestedness analysis in ecological systems: reconciling
564 concept and measurement. *Oikos*, 117, 1227–1239.
- 565 Atmar, W. & Patterson, B.D. (1993). The measure of order and disorder in the
566 distribution of species in fragmented habitat. *Oecologia*, 96, 373–382.
- 567 Barabási, A.L. (2016). *Network science*. Cambridge University Press.
- 568 Bascompte, J. & Jordano, P. (2013). *Mutualistic networks*. Princeton University Press.

- 569 Bascompte, J., Jordano, P., Melian, C.J. & Olesen, J.M. (2003). The nested assembly of
570 plant-animal mutualistic networks. *Proc. Natl. Acad. Sci.*, 100, 9383–9387.
- 571 Bates, D., Mächler, M., Bolker, B. & Walker, S. (2015). Fitting Linear Mixed-Effects
572 Models Using lme4. *J. Stat. Softw.*, 67, 48.
- 573 Beckett, S.J. & Williams, H.T.P. (2013). Coevolutionary diversification creates nested-
574 modular structure in phage-bacteria interaction networks. *Interface Focus*, 3,
575 20130033–20130033.
- 576 Bezerra, E.L.S., Machado, I.C. & Mello, M.A.R. (2009). Pollination networks of oil-
577 flowers: a tiny world within the smallest of all worlds. *J. Anim. Ecol.*, 78, 1096–
578 1101.
- 579 Bolker, B.M., Brooks, M.E., Clark, C.J., Geange, S.W., Poulsen, J.R., Stevens, M.H.H.,
580 *et al.* (2009). Generalized linear mixed models: a practical guide for ecology and
581 evolution. *Trends Ecol. Evol.*, 24, 127–135.
- 582 Brown, J.H. (1984). On the relationship between abundance and distribution of species.
583 *Amer. Natur.*, 124, 255–79.
- 584 Diamond, J.M. (1975). Assembly of species communities. In: *Ecology and Evolution of*
585 *Communities*. pp. 342–444.
- 586 Dormann, C.F., Fründ, J. & Schaefer, H.M. (2017). Identifying Causes of Patterns in
587 Ecological Networks: Opportunities and Limitations. *Annu. Rev. Ecol. Evol. Syst.*,
588 48, 559–584.
- 589 Dormann, C.F., Gruber, B. & Fründ, J. (2008). Introducing the bipartite Package:
590 Analysing Ecological Networks. *R News*, 8, 8–11.
- 591 Dupont, Y.L. & Olesen, J.M. (2009). Ecological modules and roles of species in

- 592 heathland plant-insect flower visitor networks. *J. Anim. Ecol.*, 78, 346–353.
- 593 Flores, C.O., Poisot, T., Valverde, S. & Weitz, J.S. (2016). BiMat : a MATLAB package
594 to facilitate the analysis of bipartite networks. *Methods Ecol. Evol.*, 7, 127–132.
- 595 Flores, C.O., Valverde, S. & Weitz, J.S. (2013). Multi-scale structure and geographic
596 drivers of cross-infection within marine bacteria and phages. *ISME J.*, 7, 520–32.
- 597 Fort, H., Vázquez, D.P. & Lan, B.L. (2016). Abundance and generalisation in
598 mutualistic networks: solving the chicken-and-egg dilemma. *Ecol. Lett.*, 19, 4–11.
- 599 Fortuna, M.A., Stouffer, D.B., Olesen, J.M., Jordano, P., Mouillot, D., Krasnov, B.R., *et*
600 *al.* (2010). Nestedness versus modularity in ecological networks: Two sides of the
601 same coin? *J. Anim. Ecol.*, 79, 811–817.
- 602 Futuyma, D.J. & Moreno, G. (1988). The evolution of ecological specialization. *Annu.*
603 *Rev. Ecol. Syst.*, 19, 207–233.
- 604 Gaston, K.J. & Blackburn, T.M. (2000). *Pattern and process in macroecology. Pattern*
605 *Process Macroecology*. Blackwell Science Ltd, Malden, MA, USA.
- 606 Gotelli, N.J. & Ulrich, W. (2012). Statistical challenges in null model analysis. *Oikos*,
607 121, 171–180.
- 608 Graham, S.P., Hassan, H.K., Burkett-Cadena, N.D., Guyer, C. & Unnasch, T.R. (2009).
609 Nestedness of Ectoparasite-Vertebrate Host Networks. *PLoS One*, 4, e7873.
- 610 Guimaraes, P.R., Rico-Gray, V., Furtado dos Reis, S. & Thompson, J.N. (2006).
611 Asymmetries in specialization in ant-plant mutualistic networks. *Proc. R. Soc. B*
612 *Biol. Sci.*, 273, 2041–2047.
- 613 Guimerà, R. & Amaral, L.A.N. (2005). Cartography of complex networks: modules and
614 universal roles. *J. Stat. Mech. Theory Exp.*, 2005, P02001.

- 615 Hellgren, O., Pérez-Tris, J. & Bensch, S. (2009). A jack-of-all-trades and still a master
616 of some: prevalence and host range in avian malaria and related blood parasites.
617 *Ecology*, 90, 2840–2849.
- 618 Hérault, B. (2007). Reconciling niche and neutrality through the Emergent Group
619 approach. *Perspect. Plant Ecol. Evol. Syst.*, 9, 71–78.
- 620 Holt, R. (2006). Emergent neutrality. *Trends Ecol. Evol.*, 21, 531–533.
- 621 Hopkins, K.D. (1982). The unit of analysis: Group means versus individual
622 observations. *Am. Educ. Res. J.*, 19, 5–18.
- 623 Krasnov, B.R., Fortuna, M.A., Mouillot, D., Khokhlova, I.S., Shenbrot, G.I. & Poulin,
624 R. (2012). Phylogenetic Signal in Module Composition and Species Connectivity
625 in Compartmentalized Host-Parasite Networks. *Am. Nat.*, 179, 501–511.
- 626 Krasnov, B.R., Mouillot, D., Khokhlova, I.S., Shenbrot, G.I. & Poulin, R. (2008). Scale-
627 invariance of niche breadth in fleas parasitic on small mammals. *Ecography*
628 (*Cop.*), 31, 630–635.
- 629 Krasnov, B.R., Mouillot, D., Shenbrot, G.I., Khokhlova, I.S. & Poulin, R. (2005).
630 Abundance patterns and coexistence processes in communities of fleas parasitic on
631 small mammals. *Ecography (Cop.)*, 28, 453–464.
- 632 Krasnov, B.R., Poulin, R., Shenbrot, G.I., Mouillot, D. & Khokhlova, I.S. (2004).
633 Ectoparasitic “Jacks-of-All-Trades”: Relationship between Abundance and Host
634 Specificity in Fleas (Siphonaptera) Parasitic on Small Mammals. *Am. Nat.*, 164,
635 506–516.
- 636 Krasnov, B.R., Shenbrot, G.I., Khokhlova, I.S. & Poulin, R. (2006). Is abundance a
637 species attribute? An example with haematophagous ectoparasites. *Oecologia*, 150,
638 132–140.

- 639 Leung, C.Y. (Joey) & Weitz, J.S. (2016). Conflicting attachment and the growth of
640 bipartite networks. *Phys. Rev. E*, 93, 32303.
- 641 Lewinsohn, T.M., Inácio Prado, P., Jordano, P., Bascompte, J. & M. Olesen, J. (2006).
642 Structure in plant-animal interaction assemblages. *Oikos*, 113, 174–184.
- 643 MacArthur, R. & Levins, R. (1967). The Limiting Similarity, Convergence, and
644 Divergence of Coexisting Species. *Am. Nat.*, 101, 377–385.
- 645 Mello, M.A.R., Marquitti, F.M.D., Guimarães, P.R., Kalko, E.K.V., Jordano, P. & de
646 Aguiar, M.A.M. (2011). The modularity of seed dispersal: differences in structure
647 and robustness between bat– and bird–fruit networks. *Oecologia*, 167, 131–140.
- 648 Nakagawa, S. & Schielzeth, H. (2013). A general and simple method for obtaining R^2
649 from generalized linear mixed-effects models. *Methods Ecol. Evol.*, 4, 133–142.
- 650 Olesen, J.M., Bascompte, J., Dupont, Y.L. & Jordano, P. (2007). The modularity of
651 pollination networks. *Proc. Natl. Acad. Sci.*, 104, 19891–19896.
- 652 Ollerton, J., Johnson, S.D., Cranmer, L. & Kellie, S. (2003). The pollination ecology of
653 an assemblage of grassland asclepiads in South Africa. *Ann. Bot.*, 92, 807–834.
- 654 Ollerton, J., McCollin, D., Fautin, D.G. & Allen, G.R. (2007). Finding NEMO:
655 nestedness engendered by mutualistic organization in anemonefish and their hosts.
656 *Proc. R. Soc. B Biol. Sci.*, 274, 591–598.
- 657 Pimm, S.L. (1991). *The balance of nature? Ecological issues in the conservation of*
658 *species and communities*. University of Chicago Press.
- 659 Pinheiro, R.B.P., Félix, G.M.F., Chaves, A. V., Lacorte, G.A., Santos, F.R., Braga, É.M.,
660 *et al.* (2016). Trade-offs and resource breadth processes as drivers of performance
661 and specificity in a host–parasite system: a new integrative hypothesis. *Int. J.*

- 662 *Parasitol.*, 46, 115–121.
- 663 Poulin. (1998). Large-scale patterns of host use by parasites of freshwater fishes. *Ecol.*
664 *Lett.*, 1, 118–128.
- 665 Poulin, R. (2007). *Evolutionary Ecology of Parasites*. Princeton University Press,
666 Princeton, New Jersey.
- 667 Poulin, R. & Guégan, J.-F. (2000). Nestedness, anti-nestedness, and the relationship
668 between prevalence and intensity in ectoparasite assemblages of marine fish: a
669 spatial model of species coexistence. *Int. J. Parasitol.*, 30, 1147–1152.
- 670 Presley, S.J., Higgins, C.L. & Willig, M.R. (2010). A comprehensive framework for the
671 evaluation of metacommunity structure. *Oikos*, 119, 908–917.
- 672 Sarmiento, R., Alves-Costa, C.P., Ayub, A. & Mello, M.A.R. (2014). Partitioning of seed
673 dispersal services between birds and bats in a fragment of the Brazilian Atlantic
674 Forest. *Zool.*, 31, 245–255.
- 675 Schank, J.C. & Koehnle, T.J. (2009). Pseudoreplication is a pseudoproblem. *J. Comp.*
676 *Psychol.*, 123, 421–433.
- 677 Scheffer, M. & van Nes, E.H. (2006). Self-organized similarity, the evolutionary
678 emergence of groups of similar species. *Proc. Natl. Acad. Sci.*, 103, 6230–6235.
- 679 Szollósi, E., Cichón, M., Eens, M., Hasselquist, D., Kempnaers, B., Merino, S., *et al.*
680 (2011). Determinants of distribution and prevalence of avian malaria in blue tit
681 populations across Europe: separating host and parasite effects. *J. Evol. Biol.*, 24,
682 2014–2024.
- 683 Thebault, E. & Fontaine, C. (2010). Stability of Ecological Communities and the
684 Architecture of Mutualistic and Trophic Networks. *Science (80-.)*, 329, 853–856.

685 Ulrich, W., Almeida-Neto, M. & Gotelli, N.J. (2009). A consumer's guide to nestedness
686 analysis. *Oikos*, 118, 3–17.

687 Vazquez, D.P., Bluthgen, N., Cagnolo, L. & Chacoff, N.P. (2009). Uniting pattern and
688 process in plant-animal mutualistic networks: a review. *Ann. Bot.*, 103, 1445–1457.

689 Vázquez, D.P., Melián, C.J., Williams, N.M., Blüthgen, N., Krasnov, B.R. & Poulin, R.
690 (2007). Species abundance and asymmetric interaction strength in ecological
691 networks. *Oikos*, 116, 1120–112

692