

1 **Improvements in estimating bioaccumulation metrics in**
2 **the light of toxicokinetics models and Bayesian inference**

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5

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22 **abstract**

23 The surveillance of chemical substances in the scope of Environmental Risk
24 Assessment (ERA) is classically performed through bio-assays from which data
25 are collected and then modelled. Some statistical analysis base on the fitting
26 of toxicokinetic (TK) models to assess the bioaccumulative capacity of chem-
27 ical substances via the estimation of bioaccumulation metrics as required by
28 regulatory documents. Given that bio-assays are particularly expensive and
29 time consuming, it is of crucial importance to deeply benefit from all informa-
30 tion contained in the data. By revisiting the calculation of bioaccumulation
31 metrics under a Bayesian framework, this paper presents improvements in the
32 classification of the bioaccumulative capacity of chemical substances. A meta-
33 analysis of a data-rich TK database was performed, considering the uncer-
34 tainties around the bioaccumulation metrics. The subsequent results appeared
35 sufficiently statistically robust to propose the replacement of the single me-
36 dian estimate to decide of the class to which assign a chemical substance. The
37 main recommendation is to use the 75th percentile of the uncertainty interval
38 of the bioaccumulation metrics, which revealed a better criterion to classify a
39 chemical substance, and in the same way as the conventional method in 90%
40 of cases.

41 **Keywords:** Environmental Risk Assessment, classification of chemical sub-
42 stances, REACH regulation, accumulation-depuration test.

43 **1 Introduction**

44 Chemical substances, present in the environment as a result of human ac-
45 tivities, are of extreme concern due to their persistence, to their capacity in
46 accumulating within non-target living organisms and to their potential toxic-
47 ity on the different levels of biological organization all along trophic chains
48 (Cousins et al., 2019; Popek, 2018). In particular, this reveals crucial to bring
49 reliable and precise information on the bioaccumulative capacity of the dif-
50 ferent chemical substances, on what depend the concentrations internalized
51 by organisms almost exclusively. This stage then conditions the way in which
52 relevant links can be established between the exposure concentrations and the
53 likely damages on life-history traits (Arnot and Gobas, 2006; Chojnacka and
54 Mikulewicz, 2014; Armitage et al., 2021).

55 In Europe, chemical substances are governed by the REACH regulation,
56 adopted by the European Union to improve the protection of human health
57 and the environment from the risks that can be raised by chemical substances
58 (European Commission, 2006). In principle, REACH applies to all chemical
59 substances, also stipulating the need to reduce animal testing. To comply with
60 the regulation, manufacturers must identify and manage the risks linked to
61 the chemical substances in demonstrating that they can be safely used. For
62 example, in order to evaluate the bioaccumulative capacity of chemical sub-
63 stances, produced or imported above 100 tonnes per year, the REACH regula-
64 tion requires the calculation of bioaccumulation metrics. According to Ratier
65 et al. (2021b), we use the generic expression “bioaccumulation metrics” to de-
66 note either bio-concentration factors (BCF) used when exposure is via water,
67 biota-sediment accumulation factors (BSAF) when exposure is via sediment
68 or biomagnification factors (BMF) when exposure is via food. For substances
69 produced or imported between 10 and 100 tonnes per year, bioaccumulation

70 metrics are not mandatory but still required to classify chemical substances
71 as persistent, bioaccumulative and toxic (abbreviated PBT) or very persistent
72 and very bioaccumulative (abbreviated vPvB). Most of European countries
73 classify substances as bioaccumulative (abbreviated “B”) if the bioaccumula-
74 tion metric is in $[2000; 5000[$, or very bioaccumulative (abbreviated “vB”) if it
75 is > 5000 (European Commission, 2006). Other regulations around the world
76 (Saito et al., 2011; Government of Canada, 1999; Agency, 1979) classify as “B”
77 a chemical substance with a bioaccumulation metric ranging in $[1000; 5000[$.
78 Chemical substances with a bioaccumulation metric in $]1000; 2000[$ are always
79 classified as low bioaccumulative (abbreviated as “ ℓ B”). These classifications
80 are summarized in Wassenaar et al. (2020) and by Hartmann et al. (2014).

81 Bioaccumulation metrics are calculated from toxicokinetic (TK) parame-
82 ter estimates by fitting a TK model (usually a one-compartment model) to
83 experimental data collected during bioaccumulation tests (*e.g.*, OECD (2008,
84 2012)). Bioaccumulation tests provide internal concentration measurements
85 from two-phase experiments: a first phase (the “accumulation” phase) during
86 which organisms are exposed via one or several uptake route(s) (water, pore
87 water, sediment and/or food) to a given chemical substance, kept constant
88 over time; a second phase (the “depuration” phase) during which organisms
89 are transferred into a clean medium where elimination processes take place.
90 The TK model is expected to account for all uptake routes and elimination
91 processes (including excretion, dilution by growth and/or metabolization) to
92 better describe the overall kinetics in terms of internal concentrations. From
93 TK parameter estimates (namely, uptake and elimination rates), two types of
94 bioaccumulation metrics can be calculated: the “kinetic” bioaccumulation met-
95 ric, as the ratio between uptake and elimination rates; and the “steady-state”
96 bioaccumulation metric as the ratio of the internal concentration at steady
97 state and the constant exposure concentration; the steady state is defined as

98 the plateau concentration that is expected at the end of the accumulation
99 phase. In theory, the “steady-state” bioaccumulation metric should be calcu-
100 lated only if the internal concentrations measured at the last three time points
101 of the accumulation phase are not significantly different.

102 In addition to the experimental procedure of bioaccumulation tests, the
103 OECD guidelines also explain how to obtain bioaccumulation metrics depend-
104 ing on the exposure routes that have been considered within the experiments.
105 If a one-compartment TK model is often sufficient. So, in case the goodness
106 of fit is poor this may be an indication that first order kinetics does not ap-
107 ply suggesting that a more complex model should be employed. For example,
108 one of the most common complexities to account for is fish growth during
109 the bioaccumulation test (OECD, 2012). Surprisingly, in such a case, guide-
110 lines only recommend to seek advice from bio-statistician and/or pharmaco-
111 kineticist experts. In addition, very few tools exist to easily perform TK anal-
112 yses. To name but a few, there are the Excel macro by Gobas et al. (2020),
113 the “bcmfR” package (OECD, 2012) and the free open-source MOSAIC_{bioacc}
114 web service (<http://umr5558-shiny.univ-lyon1.fr/mosaic-bioacc/>) that
115 has been recently updated (Ratier et al., 2021b; Charles et al., 2021a). Only
116 MOSAIC_{bioacc} is entirely generic whatever the species-compound combination
117 of interest, allowing to account for different exposure routes and several elim-
118 ination processes simultaneously, automatically adapting the fitted TK model
119 according to the input experimental data. It provides all possible bioaccumu-
120 lation metrics accordingly, namely BCF, BSAF and/or BMF.

121 Recently, the European Food Safety Authority (EFSA) strongly advocated
122 the need to associate uncertainties with model parameter estimates (EFSA Sci-
123 entific Committee, 2018) in general, emphasizing this need for toxicity indica-
124 tors in particular (Ockleford et al., 2018). This requirement is complementary
125 to the previous regulation with regard to the bioaccumulation metrics. Indeed,

126 for the authorisation dossiers of active chemical substances, the current regu-
127 latory document only ask for a single mean or a median value as bioaccumu-
128 lation metrics (European Commission, 2013). Such a practice is still common
129 today, probably because of a lack of efficient and easy tools to handle com-
130 puter resources, specifically designed to automatically provide uncertainties
131 on any model output. However, recent recommendations have clearly been es-
132 tablished when using toxicokinetic-toxicodynamic (TKTD) models (Baudrot
133 and Charles, 2019), while Charles et al. (2021b) highlighted how critical it
134 is to take uncertainty into account when assessing the toxicity of a chemical
135 substance to a range of non-target terrestrial plants thus revisiting the species
136 sensitivity distribution (SSD) approach.

137 Today, when bioaccumulation metrics need to be estimated, *MOSAIC_{bioacc}*
138 is one of the only tools in support of a facilitated in-depth quantification of
139 the uncertainties; on the contrary, Gobas et al. (2020) tool only gives a stan-
140 dard deviation. While available on-line since May 2020, the number of users
141 of *MOSAIC_{bioacc}* is continuously growing all over the world, whether they
142 are from academia, regulatory bodies or industry (380 recordings these last
143 6 months). A publicly available database accompanies the *MOSAIC_{bioacc}* ser-
144 vices ([http://umr5558-shiny.univ-lyon1.fr/mosaic-bioacc/data/database/
145 index_readme.html](http://umr5558-shiny.univ-lyon1.fr/mosaic-bioacc/data/database/index_readme.html)), with more than 200 accumulation-depuration data sets
146 collected within published scientific papers (Ratier and Charles, 2021). All
147 data sets are automatically analyzed with *MOSAIC_{bioacc}*, full analysis reports
148 being made available via the database directly. This database is dynamically
149 supplemented as new data sets are retrieved from the recent scientific litera-
150 ture, or directly deposited by researchers upon request.

151 The aim of the present paper is to propose improvements in the estimation
152 of bioaccumulation metrics in association with the quantification of their un-
153 certainty with the perspective to reinforce the statistical foundations leading

154 to the classification of chemical substances as (low, medium or very) bioac-
155 cumulative or not. To this end, we first present the last updates brought to
156 MOSAIC_{bioacc}, especially an innovative prediction device that can be specif-
157 ically used in designing new experiments in full respect of the 3R principles
158 (Replacement, Reduction and Refinement) ensuring animal welfare and quality
159 of science (Prescott and Lidster, 2017). Then, the added value of accounting for
160 uncertainty of bioaccumulation metrics is underlined through a meta-analysis
161 of the TK database associated with MOSAIC_{bioacc}. Finally, we demonstrate
162 how influential may be the consideration of uncertainty when classifying chem-
163 ical substances according to the current regulatory intervals into which the
164 bioaccumulation metric estimates fall. We conclude with a revisited workflow
165 to improve environmental risk assessment if it would be adopted by regulatory
166 bodies.

167 **2 Calculations and predictions of bioaccumulative capacity**

168 This section gives a brief overview of the different features of MOSAIC_{bioacc}.
169 A focus is first made on the calculation of bioaccumulation metrics. Then the
170 new prediction tool is introduced, to close with illustrative case studies.

171 **2.1 Calculations of bioaccumulation metrics**

172 Recent updates have been brought to the MOSAIC_{bioacc} web application to
173 increase the speed of calculations and improve its user-friendliness. Above
174 all, a new R-package is today available on the official CRAN web site (<https://CRAN.R-project.org/package=rbioacc>) that allows to similarly perform
175 all MOSAIC_{bioacc} calculations and graphs directly in the R software with
176 ready-to-use dedicated functions (Ratier et al., 2021a). The new version of
177 MOSAIC_{bioacc} has entirely been rewritten to reduce the length of the source
178

179 code and take advantage of this new package. Above all, *MOSAIC_{bioacc}* is now
180 based on a tabbed presentation that clarifies and facilitates browsing from one
181 step to the next. A special tab gives all bioaccumulation metrics, appropriately
182 calculated according to the input observed data that the user has uploaded.
183 By default, the kinetics BCF, BSAF or BMF values are delivered, displayed
184 via their entire posterior probability distribution then summarized with their
185 median (that is the 50% quantile) and their 95% uncertainty interval (bounded
186 by the 2.5 and 97.5% quantiles). In addition, users can ask for the steady-state
187 corresponding bioaccumulation metrics if they consider it relevant according
188 to the duration of the accumulation phase, and the fact of having actually
189 reached the plateau.

190 2.2 Prediction of bioaccumulation matrices to optimize experiments

191 A new prediction tab have been added to *MOSAIC_{bioacc}* that allows interac-
192 tive simulations of a TK model under a constant or a time-variable exposure
193 profile. The main aim of this prediction tool is to assist experimenters in opti-
194 mizing the design of new experiments, based on previous TK analyses. Indeed,
195 when studying a new chemical substance and/or a new species, when study-
196 ing a new chemical substance and/or a new species, some information needed
197 to be known in advance in full respect of scientific ethic in terms of experi-
198 ment on living organisms and chemical use and recycling. For example, the
199 exposure concentration, the duration of the accumulation and the depuration
200 phases as well as the number of time points at which internal concentrations
201 should be primarily measured need to be defined in advance. Then, based on
202 previous TK analyses for given close species/compound combinations, TK pa-
203 rameter estimates can serve to simulate what could be expected when planing
204 additional time points and/or extending the accumulation phase for example.

205 Above all, benefiting of the Bayesian framework, the uncertainty around pa-
206 rameter estimates can be propagated towards predictions and any function
207 of the parameters (Baudrot and Charles, 2019), in a way that is particularly
208 useful when environmentally realistic exposure scenarios have to be run in
209 numbers.

210 2.3 Illustrations with case studies

211 We provide a collection of case studies as supplementary information (SI, see
212 the .pdf file) to illustrate various situations where the prediction tool can be
213 helpful:

214 **Case study 1** Plan an experiment for an already studied species exposed to
215 a different but chemically similar compound (*i.e.*, with a mode of action
216 expected to be close), without accounting for the parameter uncertainty;

217 **Case study 2** Compare several species exposed to a same chemical substance
218 accounting for the uncertainty around parameter estimates coming from
219 a previous TK analysis conducted on a species phylogenetically (or tax-
220 onomically) close to the new set of interest. In such a case, the user will
221 need to enter the required input information but also a tabular file with
222 the joint posterior distribution of the parameters, either coming from a
223 previous MOSAIC_{bioacc} TK analysis or a home-made TK implementation.

224 **Case study 3** A prediction for a same species/compound combination but
225 for different exposure scenarios for which the user may have observed data
226 to which simulations can be compared as a validation step of the exploited
227 TK model.

228 **3 Matter of uncertainty in estimating bioaccumulation metrics**

229 Regarding accumulation-depuration data, one output of great interest is the
230 bioaccumulative capacity of a chemical substance which is assessed through
231 the appropriate bioaccumulation metrics according to the exposure route(s)
232 (either the BCF, the BSAF and/or the BMF). So, benefiting of the probability
233 distributions of these metrics (coming from the propagation of the parameter
234 uncertainty) is crucial to catch their precision. This latter may indeed influence
235 the classification of the substance as bioaccumulative or not, and if bioaccu-
236 mulative, influence the choice between “B” or “vB” categories. In order to
237 illustrate this critical issue for ERA, we present below a meta-analysis of the
238 TK database associated to MOSAIC_{bioacc} (Ratier and Charles, 2021).

239 3.1 The accumulation-depuration TK database

240 The TK database currently contains 211 accumulation-depuration data sets
241 collected from a literature review and corresponding to a total of 56 studies.
242 The 211 data sets encompass 52 genus, 124 chemical substances, three different
243 exposure routes (water being the main one, sediment and food), 34 data sets
244 with also biotransformation data (that is metabolization data). Figure 1 shows
245 several tree maps performed with the `treemap` R-package (Tennekes, 2017)
246 allowing to visualise the proportion of each chemical category (Figure 1-a) and
247 each genus (Figure 1-b) among the 211 data sets. Pesticides and hydrocarbons
248 are the most represented chemical substances within the TK database, what
249 can be explained by the predominance of data sets on freshwater invertebrates
250 and fish. *Gammarus* and *Daphnia* are the most represented genus, probably
251 due to less ethic exigence with them, while fish studies are rarer. *Gammarus*
252 and *Daphnia* are also the genus for which the most biotransformation data are
253 available.

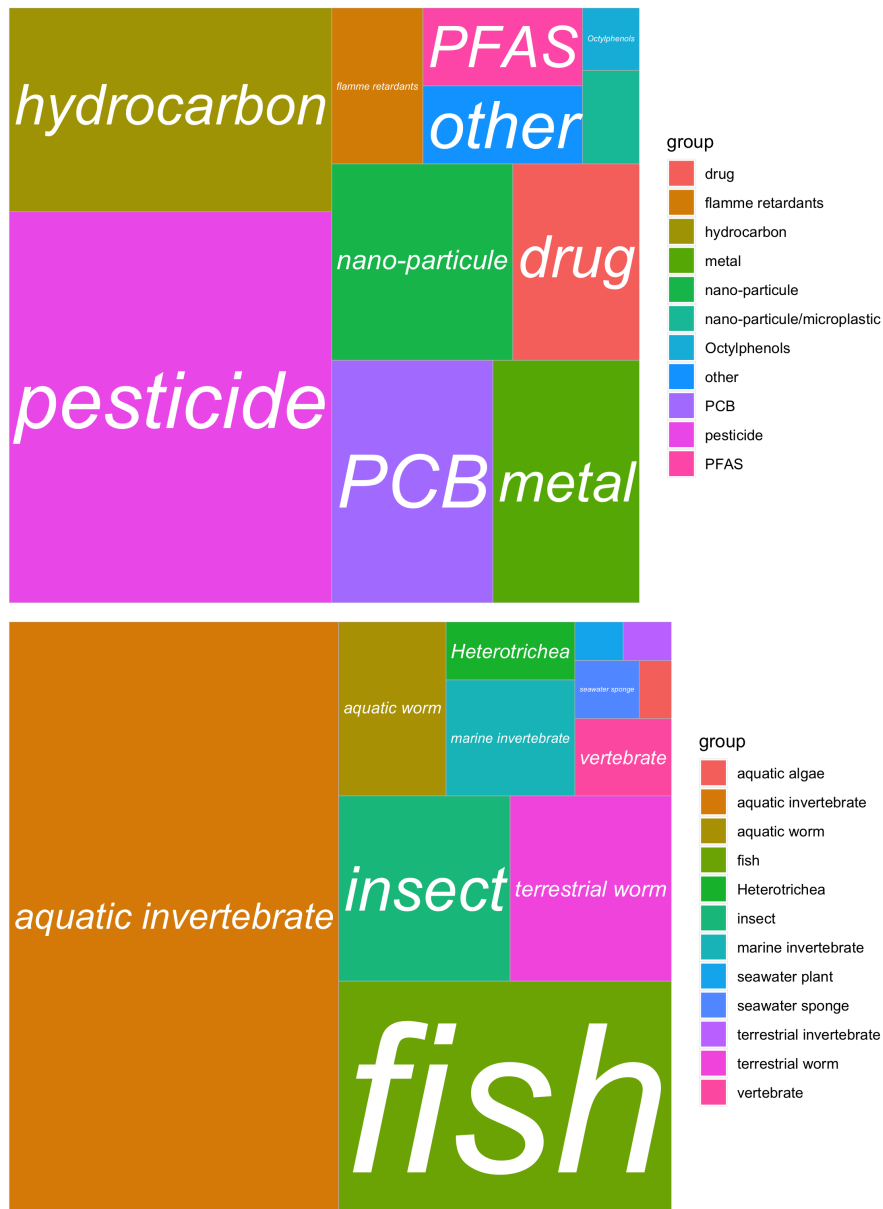


Fig. 1 Tree maps of chemical categories (upper panel) and genus categories (lower panel) available in the TK database available at http://lbbe-shiny.univ-lyon1.fr/mosaic-bioacc/data/database/TK_database.html.

254 The 211 data sets of the TK database were fully analysed with MOSAIC_{bioacc}.
255 Corresponding reports are available from the database itself, as well as refer-
256 ences from which data have been extracted. From these analyses, we conducted
257 a meta-analysis of the bioaccumulation metrics to specifically illustrate how
258 much accounting for uncertainty matters when characterizing the bioaccumu-
259 lation capacities of chemical substances.

260 3.2 Meta-analysis of the TK database

261 Even if not required within regulatory documents, associate the uncertainty
262 to a bioaccumulation metric is of crucial importance (Wassenaar et al., 2020).
263 Among the 211 data sets, a total of 137 corresponds to an exposure via wa-
264 ter for which the MOSAIC_{bioacc} analysis then provides a BCF probability
265 distribution. Based on the median and the 95% uncertainty interval of these
266 BCF estimates (Figure 2), aquatic invertebrates have the highest values, when
267 predominantly exposed to pesticides or metals, among which the genus *Gam-*
268 *marus* has the most BCF estimates greater than 5000, classifying the corre-
269 sponding chemical substances as “vB”. The SI (see the .html file) provides an
270 additional figure with all the 211 estimated bioaccumulation metrics, without
271 distinguishing BCF from BSAF and BMF estimates. Note that from this Fig-
272 ure, we would have concluded to similar trends. On the other hand, based on
273 the median of the BCF estimates as required by the regulation, only 22.7%
274 of the BCF values classify chemical substances as “B”, that is with BCF me-
275 dians higher than 1000. Nevertheless, as a matter of fact, this classification
276 does not account for the precision of the BCF estimates. Hence, this raises the
277 question of using another criterion to decide whether a chemical substance is
278 bioaccumulative or not, and at which bioaccumulative capacity.

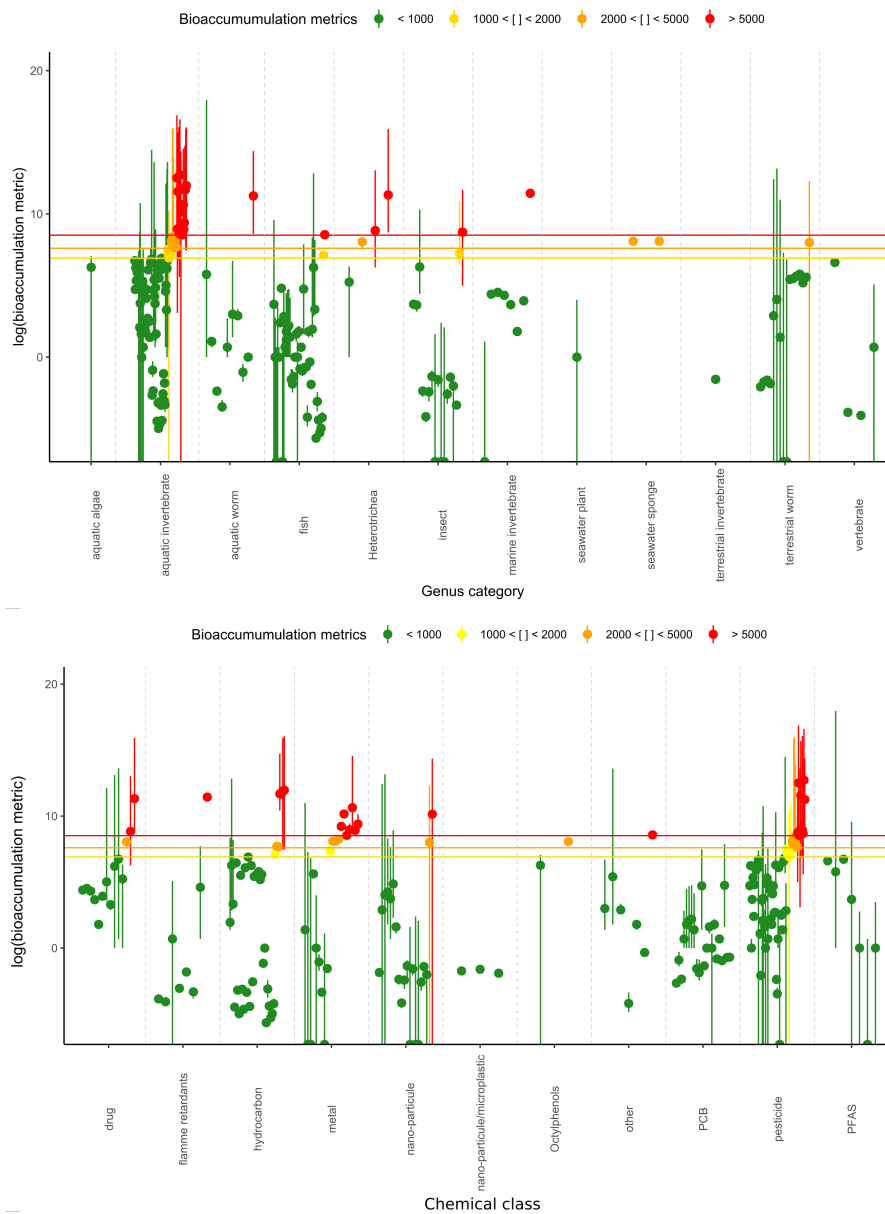


Fig. 2 Bioaccumulation metrics (in \log_{10} scale) according to genus (upper panel) and chemical (lower panel) categories. Dots represent medians of the bioaccumulation metrics, while vertical segments represent the associated 95% uncertainty intervals. Horizontal lines delineate regulatory threshold values used in the regulation (1000, 2000 and 5000, respectively) to classify chemical substances according to their bioaccumulative capacity. Bioaccumulation metrics are colored accordingly: in green when the metric is < 1000 ; in yellow if the metric is in $[1000; 2000]$; in orange if the metric is in $[2000; 5000]$; in red if the metric is > 5000 .

279 3.3 Towards improvements in ERA

280 Accounting for the precision of the bioaccumulation estimates relies on the use
281 of their probability distribution, that can be brought into play to define a new
282 decision criterion for the classification of chemical substances.

283 Instead of the median only, we first envisaged to consider the upper bound
284 of the uncertainty interval around the bioaccumulation metric estimates, that
285 is the 97.5% quantiles of their posterior probability distributions (referred as
286 Q97.5 in Table 1). This raised the question of an overestimation of the bioac-
287 cumulative capacity of the chemical substances. Indeed, the number of BCF
288 values doubled from 22 to 45 in the “vB” category (Table 1). Such a criterion
289 would classify 36.5% of the chemical substances as bioaccumulative ($n = 77$),
290 against only 22.7% ($n = 48$) with the usual criterion based on the median
291 estimate (referred as Q50 in Table 1). In terms of safety prediction, this could
292 be acceptable but still based on a biased interpretation of the bioaccumulative
293 capacity.

Table 1 Number of chemical substances in each bioaccumulative capacity class according to the three decision criteria built on the 50%, the 75% or the 97.5% quantiles of the posterior probability distributions of the bioaccumulation metric estimates: Q50, Q75 and Q97.5, respectively. Abbreviations ℓ B stands for low bioaccumulative, B for bioaccumulative and vB for very bioaccumulative, respectively.

Criterion	ℓ B ($BCF \in [1000; 2000]$)	B ($BCF \in [2000; 5000]$)	vB ($BCF > 5000$)
Q50	10	16	22
Q75	11	20	30
Q97.5	10	22	45

294 Again in the perspective of exploiting the uncertainty on bioaccumulation
295 metric estimates to build a new classification criterion of chemical substances,

296 we compared the use of the 75% quantile (referred as Q75 in Table 1) to the
297 use of the usual Q50 or the previous Q97.5, as an alternative compromise.
298 As illustrated on Figure 3, concerning the species *Daphnia magna* exposed to
299 phenanthrene (Wang et al., 2021), the Q50 classifies as “nB” ($BCF < 1000$),
300 while both the Q75 and the Q97.5 criteria classify it as “B” ($BCF > 1000$).
301 This example illustrates the need to consider the uncertainty to avoid “false
302 negatives”, that is chemical substances as non bioaccumulative while they
303 have 75% of chance to be actually. Over the 211 chemical substances in the
304 TK database, the Q75 criterion classifies a total of 28.9% of the chemical
305 substances as “B” ($n = 61$).

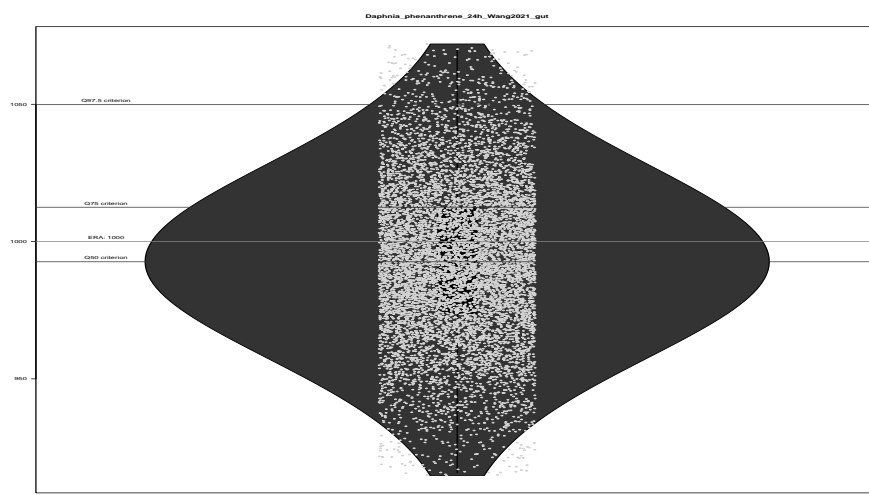


Fig. 3 Violin plot of the bioaccumulation metric posterior probability distribution for *Daphnia magna* exposed to phenanthrene (Wang et al., 2021). The green line symbolizes the threshold value at 1000 according to regulatory ERA, while grey lines stand for the 50th (Q50 criterion in table 1), 75th (Q75 criterion) and the 97.5th (Q97.5 criterion) percentiles of the distribution, respectively.

306 Figure 4 shows another example with the genus *Enchytraeus* exposed to
307 silver nano-particles (Topuz and van Gestel, 2015), illustrating the possible
308 miss-classification of the bioaccumulative capacity of a chemical substance
309 according to the Q97.5 criterion. Indeed, instead of being classified as “B”
310 ($BCF \in [2000; 5000]$), silver nano-particles are considered as non bioaccumu-
311 lative by both the Q50 (the ERA criterion) and the alternative Q75 criteria
312 ($BCF < 1000$). In this case, the difference in the classification comes from a
313 lack of precision of the BCF estimate which is associated with a large uncer-
314 tainty range, that is the coefficient of variation (abbreviated as CV) is far over
315 0.5. This means that the BCF was delivered with a very dispersed probability
316 distribution (Figure 4. The formula for CV writes as follows:

317
$$CV \simeq \frac{BCF_{\text{median}}}{97.5\% \text{quantile} - 2.5\% \text{quantile}}$$

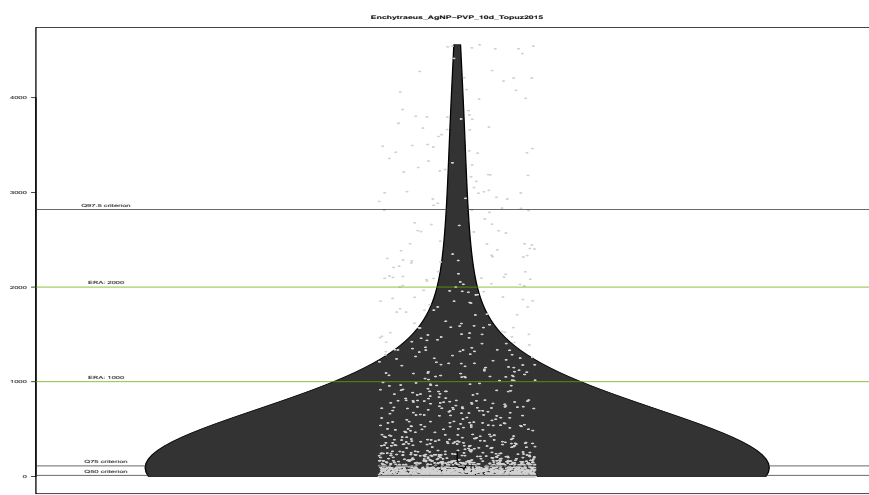


Fig. 4 Violin plot of the bioaccumulation metric posterior probability distribution for *Enchytraeus* exposed to silver nano-particles (Topuz and van Gestel, 2015). The green lines symbolize the threshold value at 1000 and 2000 according to regulatory ERA, while grey lines stand for the 50th (Q50 criterion), 75th (Q75 criterion) and the 97.5th (Q97.5 criterion) percentiles of the distribution, respectively (Table 1).

318 Considering all the 211 data sets available, we performed the classification
319 of all chemical substances based on each of the Q50, Q75 and Q97.5 crite-
320 ria, and compared the results in order to formulate sufficiently well-founded
321 recommendations. As illustrated on Figure 5 concerning the genus *Anax* ex-
322 posed to chlorpyrifos (Rubach et al., 2010), the classification was the same
323 whatever the criteria. In particular, we assigned more than 90% ($n = 190$)
324 of the chemical substances in the same class as done with the classical Q50
325 criterion (Table 1). These results support the use of the 75% quantile of the
326 posterior probability distribution of the bioaccumulation metric estimates as
327 a reasonable compromise between the classical median as currently required
328 by the regulatory ERA, and the upper bound of the uncertainty range, while
329 still accounting for the precision of the bioaccumulation metric estimates via
330 the three-quarter quantile.

331 To summarize the whole database based on our recommended Q75 crite-
332 rion, we finally classified 5.2% ($n = 11$), 9.5% ($n = 20$) and 14.2% ($n = 30$)
333 of the chemical substances as “ ℓ B”, “B” and “vB”, respectively, confirming
334 71% of the chemical substances as non bioaccumulative. Regarding the preci-
335 sion of all bioaccumulation metrics, we got a high variability among species
336 and chemical substances (see SI, see the .html file). Only considering the Q75
337 criterion, 37.4% ($n = 79$) of the chemical substances are associated with a
338 $CV < 0.5$, with the most high CV among the “nB”-classified chemical sub-
339 stances (38.9%, $n = 82/150$). This reveals that when a chemical substances is
340 considered as bioaccumulative, what ever the class, the corresponding bioac-
341 cumulation metric is precisely estimated in most of the cases.

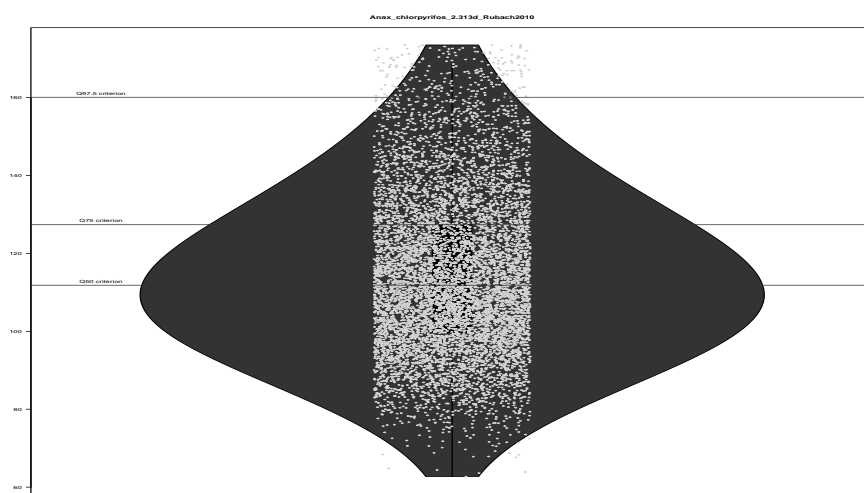


Fig. 5 Violin plot of the bioaccumulation metric posterior probability distribution for *Anax* exposed to chlorpyrifos (Rubach et al., 2010). The green lines symbolize the threshold value at 1000 and 2000 according to regulatory ERA, while grey lines stand for the 50th (Q50 criterion), 75th (Q75 criterion) and the 97.5th (Q97.5 criterion) percentiles of the distribution, respectively (Table 1).

342 **4 Conclusion**

343 Based on a meta-analysis of a TK database comprising 211 data sets, this
344 paper establishes how crucial it is to consider the uncertainty in classifying
345 the chemical substances according to their bioaccumulative capacity. Thanks
346 to the fitting of TK models under a Bayesian framework, delivering bioac-
347 cumulation metrics as probability distributions, this paper gathers together
348 statistically-founded results towards the adoption a new criterion for this clas-
349 sification. Indeed, it can be recommended to use the 75% quantile of the bioac-
350 cumulation metric distributions in order to more precisely assign the chemical
351 substances into the four regulatory ERA categories, that is as non-, low-,
352 medium-, or very- bioaccumulative, and 90% in the same way as the current

353 approach required today. Also associated with precise estimates of bioaccu-
354 mulation metrics, this new criterion would strongly improve ERA if it would
355 be adopted by regulatory bodies. In addition, it could easily be implemented
356 into the MOSAIC_{bioacc} web service in support of the daily work of regulators
357 when they need to quickly classify a set of several chemical substances.

358 **Conflict of interest**

359 The authors declare that they have no conflict of interest.

360 **Consent for Publication**

361 This manuscript has original research that has not been published previously
362 and is not under consideration for publication elsewhere, in whole or in part.

363 **Author Contributions**

364 All authors contributed to the investigation of the TK database. Raw data
365 collection and first analyses were performed by Aude Ratier and Sandrine
366 Charles. The first draft of the manuscript was written by Aude Ratier and all
367 authors commented on previous versions of the manuscript. All authors read
368 and approved the final manuscript.

369 **Supplementary Information and data availability**

370 Supplementary information is available at [https://doi.org/10.5281/zenodo.](https://doi.org/10.5281/zenodo.5865150)
371 [5865150](https://doi.org/10.5281/zenodo.5865150). All data used in this paper is downloadable from the MOSAIC_{bioacc}
372 web tool, directly from the associated TK database freely accessible at [http://](http://lbbe-shiny.univ-lyon1.fr/mosaic-bioacc/data/database/TK_database.html)
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