- 1 The war of corals: patterns, drivers, and implications of changing coral competitive
- 2 performances across reef environments
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11 Abstract

12 Amidst global environmental changes, predicting species responses to future environments is a 13 critical challenge for preserving biodiversity and associated human benefits. We explored the 14 original idea that coral competitive performances, the ability of corals to preempt ecological 15 space on the reef through territorial warfare, serve as indicators of species' ecological niches and 16 environmental windows, and therefore, responses to future environments. Our surveys indicated 17 that coral performances varied with taxonomic-identity, size, and position along environmental 18 gradients, highlighting complex interplays between life-history, warfare-strategy, and niche 19 segregation. Our results forewarn that growing alterations of coastal environments may trigger 20 shifts in coral dominance, with decline of major reef-building taxa like acroporids, and 21 underscore the importance of restraining human impacts for coastal resilience. Our empirical 22 approach untangles the complexity of species' battle-like interactions and can help identify 23 winners and losers in various communities caught in the interplay between ecological niches, 24 environmental windows, and global changes.

25

26 Keywords

27 Coral reef, Competition, Territorial war, Niche segregation, Species coexistence, Global change.

28 Introduction

29 Predicting how global environmental changes will affect species performances in the 30 future is crucial to anticipating biodiversity declines and defining sustainable management. 31 However, large uncertainties blur current predictions of ecosystem trajectory in future 32 environments [1,2]. Finding effective metrics of species responses to changing environments is 33 key, particularly for vulnerable ecosystems in need of rapid intervention, and in developing and 34 island nations where high reliance on natural resources exacerbates socio-ecosystem 35 vulnerability [3–5]. This is particularly true with coral reefs, which support prolific marine life 36 and coastal livelihoods yet stand at the frontline of declining ecosystems due to rapidly altering 37 coastal environments [1,6–9]. Reef degradation from growing coastal development, pollution, 38 fishing, and climate change predominantly involves a gradual decline in coral abundance, 39 composition, and size, with the progressive loss of vulnerable species, particularly at sensitive 40 life-stages, altering key ecosystem functions [7,10–13]. Recent demographic modeling 41 approaches allow characterizing these dynamics on reefs with fine-scale monitoring data [2,14– 42 16]. However, only a few eminent sites, representing an infinitesimal proportion of reefs, benefit 43 from the necessary level of scientific knowledge, leaving out most coral reef ecosystems from 44 such quantitative diagnosis. 45 As an alternative to using demographic modeling, we hypothesize that coral competitive

45 As an alternative to using demographic modernig, we hypothesize that competitive 46 performances, the ability of corals to preempt space on the reef substrate through territorial 47 warfare, could be used as a proxy of species ecological success in different environments. 48 Competition for space and other limiting vital resources is a key process shaping ecological 49 communities in coral reefs, notorious for biodiversity and biotic interactions, and where 50 competition warfare can drive community shifts and ecosystem collapse in altered environments

51 [17–19]. We stipulate that differences in competitive performances across environments can
52 provide insights on species ecological niches, optimal environmental windows and, therefore,
53 potential response to future conditions.

54 On reefs, direct competition for space is ubiquitous where neighboring organisms grow 55 into physical contact, species inevitably engaging in warfare for survival and ecological 56 dominance. On the front line (a.k.a. the battle-zone), corals have the capacity to invade opponent 57 territories by killing enemy living tissues in their vicinity. This battle predominantly takes two 58 forms with either smothering by growing over (a.k.a. overgrowth) or disintegrating using 59 specialized nematocyst-rich attack tentacles (a.k.a. overreach), or a combination of the two 60 (figure 1). In theory, the magnitude of killing varies with fixed characteristics of species life 61 history such as attack mechanism, strength, and reach [14,20–24], but also with additional 62 processes that vary in time, space, and across life-stages such as growth rates and metabolic 63 states of corals [25–29]. As such, while a clear hierarchy of competitive dominance among coral 64 species can be established in a given environment, competitive outcomes in fact appear spatio-65 temporally dynamic [30–34], and are therefore expected to differ in the future with changing reef 66 environments. Because corals are slow growing, habitat-forming species at the foundation of reef 67 ecosystems, even small differences in species abilities to preempt reef space can have strong 68 implications for reef ecosystem structure and functions, and associated services to society. 69 We used observations of coral competitive interactions across the south-western reef 70 system of New Caledonia as indicators of the capacity of coral species to prevail in different 71 environments. The island nation is surrounded by large extents of biodiverse coral reefs,

73 associated with natural environmental variability amplified by human impacts [35,36] (figure

characterized by diverse habitats distributed along pronounced coast-to-ocean gradients

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74 S1). The near shore is most exposed to terrigenous inputs of freshwater, nutrient, and sediment, 75 as well as anthropogenic impacts such as pollution and fishing, and undergoes higher 76 environmental variation with seasons and weather conditions [37–40]. Moving towards the 77 ocean, marked increases in water quality and diminishing human pressure is observed. Portions 78 of the reefs are classified as UNESCO World Heritage site due to their outstanding character for 79 global coral reef conservation. Zonation in the relative abundance of coral species along the 80 coast-to-ocean gradient indicates some degree of niche segregation among dominant taxa [36]. 81 However, the ecological mechanisms underlying such spatial patterns largely remain to be 82 comprehended. Because coral species are expected to occupy delimited ecological niches 83 distinguished by environmental preferences [7,29,41], we tested whether the spatial distribution 84 of ecological windows would be reflected in coral competitive outcomes. In general, a deeper 85 understanding of coral competition can help a better characterization of coral life-strategies, 86 which at this stage predominantly relies on qualitative assumptions of species competitive 87 abilities based on taxonomic traits and demographic performances [28,42]. A better 88 understanding of coral competition can also inform on the ecological processes underlying 89 species coexistence and the exceptional biodiversity of coral reefs. We illuminate our findings 90 with analogies to warfare theory and lessons from human history to untangle the complexity of 91 coral competitive interactions, and discuss implications for coral performances in changing reef 92 environments.

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94 Methods

We evaluated coral competitive performances by inspecting natural occurrences of direct
physical interactions among corals, as well as with other sessile benthic species (figure S2).

97 Surveys were performed on 20 sites distributed along pronounced cross-shelf environmental 98 gradients (figure S1), and over a five-month period to capture seasonal variability with a shift 99 from the warm (January) to the cold (June) season (decreasing water temperature, 27-23°C). For 100 each interaction haphazardly encountered, the taxonomic identities, morphotypes, and three-101 dimensional size (length, width, and height) of each organism (the focal coral plus all its direct 102 competitors) were recorded, along with the contact perimeter characterizing the battle zone, and 103 overreach and overgrowth distances as short- and long-term metrics of competitive outcomes 104 (figure 1). Because coral demographic performances in survival, growth, and reproduction vary 105 with size [28,43], changes in coral abilities to defend their territories were related to species 106 fitness (i.e. chances of ecological success). We focused specifically on direct competition for 107 space (a.k.a. territorial war), leaving aside indirect competition for light, food, and other 108 resources.

109 As competitive interactions regularly consisted of bilateral attacks, net overreach and 110 overgrowth performances were calculated as the difference between the observed maximum 111 conquered and ceded distances along the frontline. These measurements of net performances 112 differ from most studies on competition, which typically characterize species interactions as a 113 simplistic binomial (win or loss) or trinomial (win or loss or standoff) outcome (e.g. [23,24,31– 114 33]). Further characteristics of the interaction and environment that could also potentially 115 influence competitive outcomes, such as distance to the coastal city Noumea, water depth, 116 competitor abundance, total competing perimeter, and differences in height among competitors, 117 were also recorded (figure 1I, table S1).

Because coral performances were expected to be influenced by various ecological factors
acting at different scales and that in concert shape species responses (figure 1I), we used

120 generalized additive models to characterize changes in competitive outcome as a function of 121 different candidate ecological descriptors (table S1) in a non-linear, multi-dimensional account 122 [44]. Model parametrization was designed to capture a set of common ecological processes 123 regulating coral performances, such as size- and density-dependence (e.g. covariates *Coral-size* 124 and *Competitor-abundance*), as well as taxonomic deviations in such processes (e.g. interaction 125 $Coral-size \times Coral-taxon$) to account for evolutionary differences among species [28]. The 126 degree of non-linearity of model terms was optimized based on semi-parametric spline-127 penalization (see [28,45] for details), and non-significant model terms were sequentially 128 excluded during the model selection process [44] (figures S3 and S4). Among the multitude of 129 possible models resulting from combinations of the explanatory variables, the models best 130 describing competitive outcomes were identified using Akaike information criteria, a measure of 131 trade-off between model performance and complexity [46]. The final models (tables S2 and S3) 132 explained variability in coral competitive performances at 66.6% in terms of net overreach and 133 79.0% in terms of net overgrowth (figures S3 and S4). This is relatively high compared to 134 previous attempts (e.g. [23,24,32]) and considering the many additional biological and 135 environmental factors that may influence competitive outcome between two living organisms 136 (genotype, age, health, metabolic state, disturbance history, etc.), suggesting that our models 137 accounted for key ecological gradients influencing coral competitive performances in our study 138 system. Restricting data to the most abundant taxa resulted in similar patterns (figures S5 and 139 S6), confirming the prevalence of the identified mechanisms among dominant species. 140 A total of 1073 competitive interactions were recorded encompassing 41 taxa and 8 141 morphotypes (table S1). All surveys were performed by the same observer using SCUBA,

142 occasionally assisted by another diver. Analyses and graphing were coded in R statistical

143 software complemented by the mgcv package [44]. Over-dispersed variables were log-

144 transformed, and model residuals were systematically checked for normality and

145 homoscedasticity.

146

147 **Results and Discussion**

148 Identifying drivers of coral competitive performance

149 Of the 1073 coral competitive interactions inspected, 84.3% (905) involved traces of 150 tentacle deployment along the frontline, out of which 7.4% (67, or 6.2% of all interactions) were 151 bilateral. In 8.9% (6) of these bilateral overreach attacks, net space intrusions were tied between 152 competitors (i.e. net overreach = 0). Similarly, 69.8% (749) of all interactions involved 153 overgrowth, out of which 2.0% (15, or 1.4% of all interactions) were bilateral. In 13.3% (2) of 154 these bilateral overgrowth attacks, net space invasions were tied between competitors (i.e. net 155 overgrowth = 0). Only 0.5% (5) of all interactions were characterized as standoffs for both 156 overgrowth and overreach, demonstrating the complementary nature of these two metrics to 157 assess competitive wars among corals.

158 Coral competitive performances varied with attributes relative to individuals, 159 interactions, and environments, highlighting the interactive importance of biological and 160 environmental factors in defining competitive outcomes. Overgrowth and overreach 161 performances were both contingent on taxonomic identity, morphology, size, competitor 162 abundance, and shelf-position, whereas perimeter of contact and height differences only 163 influenced overgrowth, and day of year only overreach (tables S2 and S3). As a metric of short-164 term competitive interactions, overreach reflects recently deployed battle strategies such as 165 spontaneous attack-tentacle developments into opponent territories, brief skirmishes along the

166 frontline that were seasonally variable in several taxa (figure 2C) and may reveal transitory in 167 long-running competitive battles [22,25,30,32,34]. In contrast, as a more integrated measure of 168 competitive interactions over time, overgrowth accounts for additional ecological mechanisms 169 that prevail across successive battles in war strategy. This includes the capacity to sustain siege 170 and lead large battlefields, sometimes for long times and simultaneously on multiple fronts 171 (figure 1), performances that rely heavily on resource provisions and differed across taxa.

172

173 Individual-level attributes

174 Coral overgrowth and overreach performances differed among taxa as expected for 175 species exhibiting contrasting life history traits, with differences in growth form and rate, in 176 tentacle size and reach, etc. [23,34,42]. Yet, contrasting responses to ecological gradients 177 provided deeper insights into distinct life-strategies as reflected by different patterns of 178 competitive performance across life-stages, as well as contrasting susceptibilities to 179 environmental variation as reflected by segregated environmental optima. Larger corals 180 generally exhibited higher net overgrowth, though many taxa deviated from a common size-181 dependent pattern in overgrowth and overreach (figures 2A and 3A). While evidence of size-182 dependent variability in coral competitive performance is not new [25,26], our comparative study 183 indicates maximum competitive capacities occur at different stages among species, suggesting 184 differences in size-specific investments in competition. Some taxa showed higher overgrowth 185 and overreach at small sizes, potentially in a strategy to secure enough space early on, until 186 reaching a size-refuge that guarantees survival and investment in other demographic processes 187 such as reproduction [28,33,43]. This was the case for *Porites*, in which competitive 188 performances declined with colony size (-0.3 cm in net overreach and -1.5 cm in net overgrowth

189	across the size-range), with an inflection point at a size of ~15 cm diameter (figures 2A and 3A).
190	Other taxa performed better at intermediate or larger sizes, which corresponds with higher ability
191	in allocating large energetic resources to competitive battles. A marked positive effect of colony
192	size was detected in mean overreach of Merulina (+14 cm across the size-range), Montastrea
193	(+12 cm), and the soft-coral Sarcophyton (+7 cm), as well as in mean overgrowth of Goniastrea
194	(+11 cm), <i>Hydnophora</i> (+9 cm), <i>Merulina</i> (+10 cm), and the soft-coral <i>Nephthea</i> (+70 cm). The
195	latter taxon reveals being a particularly fierce competitor of reef-building corals with an
196	unmatched ability to overgrow them (figure 3A) [21].
197	
198	Interaction-level attributes
199	Coral competitive outcomes were influenced by several characteristics of species
200	interactions, namely competitor abundance (a.k.a. number of enemies), contact perimeter (a.k.a.

201 battlefield stretch), and height difference among competitors (a.k.a. unequal battlegrounds).

202 Competitor abundance was associated with changes in overreach performances of four coral taxa 203 (figure 2B), and generally influenced size-specific overgrowth, with larger corals showing higher 204 capacities in leading multi-front wars (figure 3B). Larger battlefields were associated with higher 205 overgrowth in some taxa, such as *Millepora* and *Montipora* (+2 cm in net overgrowth), showing 206 high capacities in waging large-scale competitive endeavors to the detriment of others such as 207 *Porites*, in which a threshold in the capacity to hold space against competitors was observed (-10 208 cm in net overgrowth, with again a size-threshold at ~15 cm diameter, figure 3B). Overall, 209 extended battlefield perimeters were associated with higher overgrowth performance in 210 encrusting and columnar species and decreasing in branching and massive taxa, reflecting 211 evolutionary differences in competitive abilities among morphological groups [24,42].

212

213 Environmental attributes

214	Battlefield environments also influenced coral competitive performances as reflected by
215	the effects of shelf position, depth, and time, revealing differing environmental preferences
216	among taxa (tables S2 and S3). Cross-shelf variability in coral performances was detected in five
217	taxa, among which peak overgrowth and/or overreach performances were spatially segregated.
218	Pocillopora showed higher net overreach (+1 cm) near the coast, whereas Acropora (+0.6 cm)
219	and Favia (+4 cm) peaked in mid-lagoon, Porites exhibited highest overgrowth (+2.0 cm)
220	towards the barrier-reef, and Echinopora showed contrasting spatial patterns between
221	overgrowth and overreach metrics (figures 2C and 3C). Similarly, overreach performances
222	varied seasonally in six taxa with a marked temporal segregation (figure 2C). Acropora (+0.4
223	cm) and Montipora (+0.2 cm) showed higher performance in warm season, Pocillopora (+0.7
224	cm) in cool season, Isopora (+1.5 cm) and Porites (+0.3 cm) during the inter-season, whereas
225	Galaxea (+3.5 cm) peaked in each season. While the spatial differences in competitive
226	performances among taxa confirm contrasting environmental optima along the coast-to-ocean
227	gradient, the temporal patterns identified may reflect differences in environmental preferences
228	per se (e.g. differing temperature optima) or different timings of investments in other
229	demographic processes such as growth and reproduction (i.e. differing temporal windows).
230	Indeed, territorial wars as well as growth and reproduction are energetically costly processes, and
231	some corals may show temporal tradeoffs in their investment in these endeavors [22,26,28,32].
232	For example, in New Caledonia as in the neighboring Great Barrier Reef, acroporids including
233	Acropora and Montipora reproduce at the onset of the warm season following a 6-month period
234	of gametogenesis, whereas other species such as <i>Pocillopora</i> reproduce throughout the year [47–

50]. In addition, portions of the temporal trends identified may be attributable to response to
external stimuli, which may explain the higher overreach observed in *Porites* during the interseason between the performance peaks of acroporids and *Pocillopora*, and conversely high
overreach in *Galaxea* during both seasons potentially as retaliation to attacks from these
dominant taxa [32,36].

240

241 Species baseline attributes

242 When the effects of ecological gradients were isolated, species with encrusting and 243 massive morphologies were associated with higher overreach (figure 2A), whereas encrusting, 244 columnar, and foliaceous species exhibited higher overgrowth (figure 3A). These differences 245 may reflect differing evolutionary pathways among species. In contrast to other growth forms 246 that enable refuging via vertical ascension, encrusting species are fully exposed to competition 247 and their survival relies fundamentally on their capacity to preempt space on a two-dimensional 248 substrate [14,22,33]. Similarly, several coral species with massive growth forms exhibit large 249 polyps able to rebuff competitors on longer distances by developing long-range tentacles (figure 250 1) [20,23,24,34]. A maximum overreach distance of 5.7 cm performed by a massive *Galaxea* on 251 a branching *Pocillopora* was recorded in this study, and the three massive taxa *Euphyllia*, 252 Lobophyllia, and Montastrea exhibited high baseline overreach (figure 2A). Noticeably, marked 253 negative effects of competitors on coral overreach and overgrowth were only detected from hard-254 and soft- corals, whereas interactions with ascidians, sponges, and algae were characterized 255 solely by positive deviations from the average patterns (figures 2A and 3A). This suggests a less 256 substantial effect of chemical wars alone as employed by these later taxa, compared to additional

uses of physical wars involving tentacle-attacks as deployed by cnidarians, in direct competitiveinteractions [27,51,52].

259

260 Comprehending coral competitive interactions

261 Our study shows that coral competitive performances are governed by a complex 262 interplay between who is involved and how, where, and when the interactions occur, with 263 outcomes in terms of net overreach and overgrowth that are largely predictable (figures 2 and 3). Species baseline performances associated with inherited taxonomic traits (e.g. tentacle reach) are 264 265 modulated by a set of ecological gradients related to intrinsic characteristics of organisms (e.g. 266 size, evolutionary life-strategy) as well as extrinsic features of their interactions and 267 environments that vary in time and space (e.g. competitor abundance, contact perimeter, 268 seasonality). The mechanisms governing coral competitive performances are therefore 269 fundamentally analogous to those prevailing in human warfare where concerted military power 270 (weapon abundance, deadliness, and reach), war strategy (attack, defense, skirmish tactics), 271 battle characteristics (stretch of battlefronts, duration of conflicts, number of enemies and allies), 272 and battlefield features (battleground evenness, weather conditions) influence outcomes [53,54]. 273 Notorious instances when battle characteristics and environmental condition influenced war 274 outcome and sealed the fate of human history include the deleterious effects of multi-front wars 275 for Napoleon's endeavor to expand the French empire across Europe between 1805 and 1815 276 [55], and the contribution of wintery weather conditions to Hitler's army's defeat at the Soviet 277 frontline in 1941 [56]. Another historical example relates to the Battle of Agincourt in 1415, a 278 turn in the Hundred Years' War for the dominion of France and England, where muddy terrain

following rainfall severely handicapped the heavily armored knights of the numerically superior
French army to the advantage of Henri V [57].

281

282 Niche segregations and implications in changing environments

283 The contrasted responses of coral taxa as identified across multiple ecological gradients 284 provide new insights into the variety of mechanisms underlying niche segregation in biodiverse 285 species assemblages. Indeed, the diversity of ecological windows occupied by species is 286 reflected in the contrasting environmental preferences revealed by differing performances in time 287 and space, as well as the divergent evolutionary pathways as indicated by different responses to 288 individual and interaction level attributes (figure 4). These differences may explain how the 289 species coexist as a result of distinct demographic life-strategies, environmental heterogeneity, 290 and competitive interactions, resulting in the exceptional biodiversity observed on reefs 291 [17,23,28,42]. Nevertheless, several key coral taxa were sensitive to environmental variability as 292 reflected by distance to the coast and seasonality, indicating that their competitive success, and 293 perhaps overall fitness, may be affected by alterations of coastal environments. Our findings 294 suggest that warmer oceanic conditions, similar to those presently observed in summer, may 295 advantage higher competitive performances of acroporids to the detriment of pocilloporids, 296 although further anthropization of coastal habitats, currently restricted to the coastline, benefits 297 pocilloporids over acroporids (figure 4). In New Caledonia and globally, acroporids contribute exceptionally to coral reef structural complexity, biodiversity, and calcification [13,36,58-60]. 298 299 Despite high capabilities to dominate reefs in peri-optimal environments, acroporids are 300 particularly sensitive to environmental stressors such as warming and declining water quality, 301 with community shifts from acroporid dominance to pocilloporids or poritids often observed in

302 sub-optimal conditions [7,10,11,29,41]. Widespread acroporid declines have been associated 303 with reef environment degradation in various regions including the Caribbean, Persian Gulf, 304 Great Barrier Reef, and French Polynesia [12,15,16,61]. In contrast to other regions where 305 acroporids appear to face their upper temperature limits [11,29], our study suggests acroporid 306 performances may actually increase in a warmer climate in the sub-tropical reef system of New 307 Caledonia where few large-scale bleaching events have been recorded. Nevertheless, restraining 308 anthropization of coastal environments appears key to preserving near-shore acroporid 309 populations and their unique contributions to reef accretion and resilience. 310 311 **Conclusions** 312 As global changes increasingly alter coastal marine environments, some ecosystems are 313 inexorably expected to collapse while others may transform to new community compositions,

structures, and functions, changes that remain hard to predict [1,3,4,11,]. Nevertheless, present
trajectories of coastal degradation indicate future reef environments may increasingly resemble

those found near dense human concentrations today [6–8]. Our study suggests that such
anthropization results in lower abilities of some major coral taxa in preempting reef space via
direct competition. Because corals are slow growing habitat-forming species at the basis of reef

319 ecosystems, such differences in competitive performances may result in extirpations of

320 vulnerable populations, with implications for reef ecosystem biodiversity and services to society.

Overall, competitive performance appears as an effective, widespread, and accessible indicator of species performances across ecological gradients. It can help identify the biological and environmental constraints underlying ecological niches and environmental windows that define species distributions, coexistence, and therefore biodiversity patterns. Given the many

- 325 ecological pathways that link species performances to their environment, we encourage similar
- 326 quantitative investigations to further understanding of determinants of species interactions at the
- 327 interplay between evolutionary traits, life-strategies, and global changes, and implications for the
- 328 dynamics of ecosystems in a changing environment.
- 329

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335 **References**

- 1. Jouffray JB, Blasiak R, Norström AV, Österblom H, Nyström M. 2020 The Blue Acceleration:
- the trajectory of human expansion into the ocean. One Earth 2(1), 43-54.
- doi:10.1016/j.oneear.2019.12.016
- 2. Carturan BS, Pither J, Maréchal JP, Bradshaw CJ, Parrott L. 2020 Combining agent-based,
- trait-based and demographic approaches to model coral-community dynamics. *Elife* 9,
- 341 e55993. doi:10.7554/eLife.55993
- 342 3. WWF. 2020 Living Planet Report 2020 Bending the curve of biodiversity loss. Almond
- 343 REA, Grooten M, Petersen T, eds. WWF, Gland. https://livingplanet.panda.org
- 344 4. Díaz S et al. 2020. Summary for policymakers of the global assessment report on biodiversity
- 345 and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and
- 346 *Ecosystem Services*. doi:10.5281/zenodo.3553579
- 5. Kayal M, Lewis H, Ballard J, Kayal E. 2019 Humanity and the 21st century's resource
- 348 gauntlet: a commentary on Ripple et al.'s article "World scientists' warning to humanity: a
- 349 second notice". *Rethinking Ecology* 4, 21–30. doi:10.3897/rethinkingecology.4.32116
- 350 6. Cinner JE et al. 2020. Meeting fisheries, ecosystem function, and biodiversity goals in a
- human-dominated world. *Science* 368(6488), 307–311. doi:10.1126/science.aax9412
- 352 7. Darling ES et al. 2019 Social-environmental drivers inform strategic management of coral
- reefs in the Anthropocene. *Nature Ecology & Evolution* 3(9), 1341–1350.
- doi:10.1038/s41559-019-0953-8
- 8. Duprey NN *et al.* 2020. Megacity development and the demise of coastal coral communities:
- Evidence from coral skeleton δ 15N records in the Pearl River estuary. *Global Change Biology*
- 357 26(3), 1338–1353. doi:10.1111/gcb.14923

- 9. Souter D, Planes S, Wicquart J, Logan M, Obura D, Staub F. 2021. Status of Coral Reefs of
- 359 *the World: 2020.* ICRI, GCRMN, Australia Institute of Marine Science, UNEP.
- 360 www.unep.org/resources/status-coral-reefs-world-2020
- 361 10. Adjeroud M, Kayal M, Iborra-Cantonnet C, Vercelloni J, Bosserelle P, Liao V, Chancerelle
- 362 Y, Claudet J, Penin L. 2018 Recovery of coral assemblages despite acute and recurrent
- 363 disturbances on a South Central Pacific reef. *Scientific Reports* 8(1), 1–8.
- doi:10.1038/s41598-018-27891-3
- 365 11. Hughes TP et al. 2018. Global warming transforms coral reef assemblages. Nature
- 366 556(7702):492–496. doi:10.1038/s41586-018-0041-2
- 367 12. Cramer KL, Jackson JB, Donovan MK, Greenstein BJ, Korpanty CA, Cook GM, Pandolfi
- 368 JM. 2020 Widespread loss of Caribbean acroporid corals was underway before coral
- 369 bleaching and disease outbreaks. *Science Advances* 6(17), eaax9395.
- doi:10.1126/sciadv.aax9395
- 371 13. Carlot J et al. 2021 Juvenile corals underpin coral reef carbonate production after
- 372 disturbance. *Global Change Biology* 27(11), 2623–2632. doi:10.1111/gcb.15610
- 373 14. Wakeford M, Done TJ, Johnson CR. 2008 Decadal trends in a coral community and evidence
- of changed disturbance regime. *Coral Reefs* 27(1), 1–13. doi:10.1007/s00338-007-0284-0
- 375 15. Kayal M, Lenihan HS, Brooks AJ, Holbrook SJ, Schmitt RJ, Kendall BE. 2018 Predicting
- 376 coral community recovery using multi-species population dynamics models. *Ecology Letters*
- 377 21(12), 1790–1799. doi:10.1111/ele.13153
- 16. Riegl B, Johnston M, Purkis S, Howells E, Burt J, Steiner SC, Sheppard CR, Bauman A.
- 379 2018 Population collapse dynamics in *Acropora downingi*, an Arabian/Persian Gulf

- 380 ecosystem-engineering coral, linked to rising temperature. *Global Change Biology* 24(6),
- 381 2447–2462. doi:10.1111/gcb.14114
- 382 17. Connell JH. 1978 Diversity in tropical rain forests and coral reefs. *Science* 199(4335), 1302–
- 383 1310. doi:10.1126/science.199.4335.1302
- 18. Menge BA, Sutherland JP. 1987 Community regulation: variation in disturbance,
- 385 competition, and predation in relation to environmental stress and recruitment. *The American*
- 386 *Naturalist* 130(5), 730–757. www.jstor.org/stable/2461716
- 387 19. Mumby PJ, Hastings A, Edwards HJ. 2007 Thresholds and the resilience of Caribbean coral
- 388 reefs. *Nature* 450(7166), 98–101. doi:10.1038/nature06252
- 389 20. Sheppard CRC. 1979 Interspecific aggression between reef corals with reference to their
- distribution. *Marine Ecology Progress Series* 1, 237–247. www.jstor.org/stable/24812948
- 391 21. Sammarco PW, Coll JC, La Barre S. 1985 Competitive strategies of soft corals (Coelenterata:
- 392 Octocorallia). II. Variable defensive responses and susceptibility to scleractinian corals.
- *Journal of Experimental Marine Biology and Ecology* 91, 199–215. doi:10.1016/0022-
- 394 0981(85)90176-5
- 395 22. Romano SL. 1990 Long-term effects of interspecific aggression on growth of the reef-
- 396 building corals *Cyphastrea ocellina* (Dana) and *Pocillopora damicomis* (Linnaeus). *Journal*
- 397 of Experimental Marine Biology and Ecology 140(1-2), 135–146. doi:10.1016/0022-
- 398 0981(90)90087-S
- 399 23. Precoda K, Allen AP, Grant L, Madin JS. 2017 Using traits to assess nontransitivity of
- 400 interactions among coral species. *The American Naturalist* 190(3), 420–429.
- 401 doi:10.1086/692758

- 402 24. Alvarez-Noriega M, Baird AH, Dornelas M, Madin JS, Connolly SR. 2018 Negligible effect
- 403 of competition on coral colony growth. *Ecology* 99(6), 1347–1356. doi:10.1002/ecy.2222
- 404 25.Russ GR. 1982 Overgrowth in a marine epifaumal community: competitive hierarchies and
- 405 competitive networks. *Oecologia* 53(1), 12–19. doi:10.1007/BF00377130
- 406 26. Rinkevich B., Loya Y. 1983 Intraspecific competitive networks in the Red Sea coral
- 407 Stylophora pistillata. Coral Reefs 1(3), 161–172. doi:10.1007/BF00571193
- 408 27. Porter JW, Targett NM. 1988 Allelochemical interactions between sponges and corals. The

409 *Biological Bulletin* 175(2), 230–239. doi:10.2307/1541563

- 410 28. Kayal M, Vercelloni J, Wand MP, Adjeroud M. 2015 Searching for the best bet in life-
- 411 strategy: A quantitative approach to individual performance and population dynamics in reef-
- 412 building corals. *Ecological Complexity* 23, 73–84. doi:10.1016/j.ecocom.2015.07.003
- 413 29. Vajed Samiei J, Saleh A, Mehdinia A, Shirvani A, Kayal M. 2015 Photosynthetic response of
- 414 Persian Gulf acroporid corals to summer versus winter temperature deviations. *PeerJ* 3,
- 415 e1062. doi:10.7717/peerj.1062
- 416 30. Logan A. 1984 Interspecific aggression in hermatypic corals from Bermuda. Coral Reefs
- 417 3(3), 131–138. doi:10.1007/BF00301957
- 418 31. Sheppard CRC. 1985 Unoccupied substrate in the central Great Barrier Reef: role of coral
- 419 interactions. *Marine Ecology Progress Series* 25, 259–268.
- 420 32. Chornesky EA. 1989 Repeated reversals during spatial competition between corals. *Ecology*
- 421 70(4), 843–855. doi:10.2307/1941353
- 422 33. Barott KL, Williams GJ, Vermeij MJ, Harris J, Smith JE, Rohwer FL, Sandin SA. 2012
- 423 Natural history of coral-algae competition across a gradient of human activity in the Line
- 424 Islands. *Marine Ecology Progress Series* 460, 1–12. doi:10.3354/meps09874

425	34. Ferriz-Domínguez N, Horta-Puga G. 2001 Short-term aggressive behavior in scleractinian
426	corals from La Blanquilla reef, Veracruz Reef System. Revista de Biología Tropical 49(1),
427	67–75. pmid:11795171
428	35. Andréfouët S, Torres-Pulliza D. 2004 Atlas des récifs coralliens de Nouvelle-Calédonie.
429	IFRECOR Nouvelle-Calédonie, IRD, Nouméa. www.documentation.ird.fr/hor/fdi:010045824
430	36. Adjeroud M, Poisson E, Peignon C, Penin L, Kayal M. 2019 Spatial patterns and short-term
431	changes of coral assemblages along a cross-shelf gradient in the Southwestern Lagoon of New
432	Caledonia. Diversity 11(2), 21. doi:10.3390/d11020021
433	37. Jacquet S, Delesalle B, Torréton JP, Blanchot J. 2006 Response of phytoplankton
434	communities to increased anthropogenic influences (southwestern lagoon, New Caledonia).
435	Marine Ecology Progress Series 320, 65–78. www.int-res.com/abstracts/meps/v320/p65-78/
436	38. Fichez R, Chifflet S, Douillet P, Gérard P, Gutierrez F, Jouon A, Ouillon S, Grenz C. 2010
437	Biogeochemical typology and temporal variability of lagoon waters in a coral reef ecosystem
438	subject to terrigeneous and anthropogenic inputs (New Caledonia). Marine Pollution Bulletin
439	61(7-12), 309–322. doi:10.1016/j.marpolbul.2010.06.021
440	39. Ouillon S et al. 2010. Circulation and suspended sediment transport in a coral reef lagoon:
441	The south-west lagoon of New Caledonia. Marine Pollution Bulletin 61, 269–296.
442	doi:10.1016/j.marpolbul.2010.06.023
443	40. Gonson C, Pelletier D, Alban F, Giraud-Carrier C, Ferraris J. 2017 Influence of settings
444	management and protection status on recreational uses and pressures in marine protected
445	areas. Journal of Environmental Management 200, 170–185.
446	doi:10.1016/j.jenvman.2017.05.051

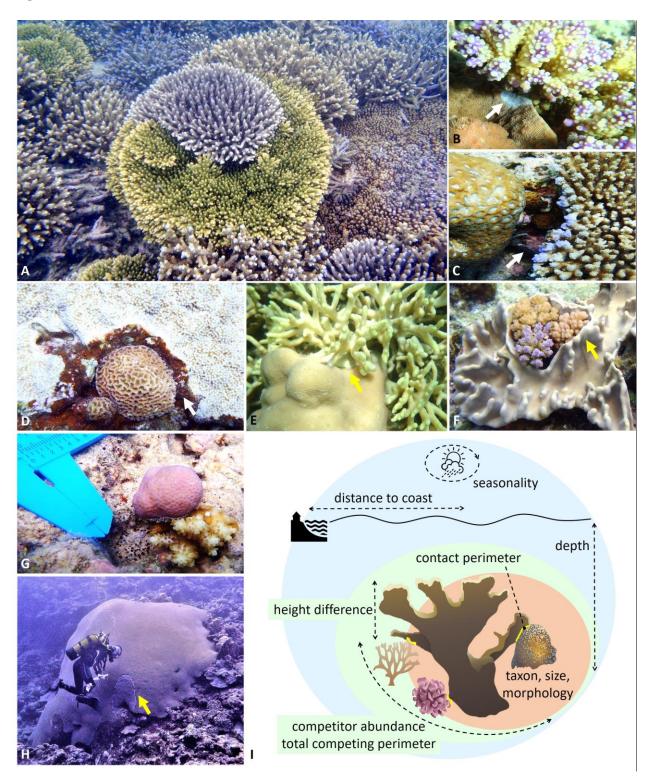
- 447 41. Moustaka M, Mohring MB, Holmes T, Evans RD, Thomson D, Nutt C, Stoddart J, Wilson
- 448 SK. 2019 Cross-shelf heterogeneity of coral assemblages in Northwest Australia. *Diversity*
- 449 11(2), 15. doi:10.3390/d11020015
- 450 42. Darling ES, Alvarez-Filip L, Oliver TA, McClanahan TR, Côté IM. 2012 Evaluating life-
- 451 history strategies of reef corals from species traits. *Ecology Letters* 15(12), 1378–1386.
- 452 doi:10.1111/j.1461-0248.2012.01861.x
- 453 43. Hall VR, Hughes TP. 1996 Reproductive strategies of modular organisms: comparative
- 454 studies of reef-building corals. *Ecology* 77(3), 950–963. doi:10.2307/2265514
- 455 44. Wood, S. 2017. Generalized additive models: an introduction with R. CRC press.
- 456 doi:10.1201/9781315370279
- 457 45. Ruppert D, Wand MP, Carroll RJ. 2003 Semiparametric regression. Cambridge University
- 458 Press, Cambridge. isbn:0-521-78050-0
- 459 46. Akaike H. 1978 A Bayesian analysis of the minimum AIC procedure. Annals of the Institute
- 460 *of Statistical Mathematics* 30, 9–14. doi:10.1007/BF02480194
- 461 47. Heyward AJ, Collins JD. 1985 Growth and sexual reproduction in the scleractinian coral
- 462 *Montipora digitata* (Dana). *Marine and Freshwater Research* 36(3), 441–446.
- 463 doi:10.1071/MF9850441
- 464 48. Baird AH, Kospartov MC, Purcell S. 2010 Reproductive synchrony in *Acropora* assemblages
- 465 on reefs of New Caledonia. *Pacific Science* 64(3), 405–412. doi:10.2984/64.3.405
- 466 49. Keith SA *et al.* 2016 Coral mass spawning predicted by rapid seasonal rise in ocean
- 467 temperature. *Proceedings of the Royal Society B: Biological Sciences*, 283(1830), 20160011.
- 468 doi:10.1098/rspb.2016.0011

- 469 50. Smith HA, Moya A, Cantin NE, van Oppen MJ, Torda G. 2019 Observations of
- 470 simultaneous sperm release and larval planulation suggest reproductive assurance in the coral
- 471 Pocillopora acuta. Frontiers in Marine Science 6, 362. doi:10.3389/fmars.2019.00362
- 472 51. Suchanek THD, Green J. 1981 Interspecific competition between *Palythoa caribaeorum* and
- 473 other sessile invertebrates on St. Croix reefs, US Virgin Islands. *Proceedings of the 4th*
- 474 International Coral Reef Symposium, 679–684.
- 475 52. Smith JE et al. 2006 Indirect effects of algae on coral: algae-mediated, microbe-induced
- 476 coral mortality. *Ecology Letters* 9(7), 835–845. doi:10.1111/j.1461-0248.2006.00937.x
- 477 53. Tzu S, Tzu S, Sun W, Vu SC. 1971 *The art of war*. Oxford University Press, Oxford.
- 478 54. Diamond J. 2005 *Collapse: How societies choose to fail or succeed*. Penguin, London.
- 479 55. Roberts A. 2014 Napoleon: A life. Penguin, London.
- 480 56. Burleigh M. 2000 *The Third Reich: a new history*. Pan.
- 481 57. Michôd D. 2019 *The King*. Film. 140 min.
- 482 58. Fenner D, Muir P. 2008 Reef corals of the northwestern lagoon of Grande-Terre, New
- 483 Caledonia. A Rapid Marine Biodiversity Assessment of the Coral Reefs of the Northwest
- 484 Lagoon, Between Koumac and Yandé, Province Nord, New Caledonia. Conservation
- 485 International. doi:10.1896/054.053.0102
- 486 59. Blanchon P, Granados-Corea M, Abbey E, Braga JC, Braithwaite C, Kennedy DM, Spencer
- 487 T, Webster JM, Woodroffe CD. 2014 Postglacial fringing-reef to barrier-reef conversion on
- 488 Tahiti links Darwin's reef types. *Scientific Reports* 4(1), 1–9. doi:10.1038/srep04997
- 489 60. Hongo C, Wirrmann D. 2015 Preliminary identification of key coral species from New
- 490 Caledonia (Southwest Pacific Ocean), their significance to reef formation, and responses to
- 491 environmental change. *Island Arc* 24(1), 31–46. doi:10.1111/iar.12090

- 492 61. Clark TR, Roff G, Zhao JX, Feng YX, Done TJ, McCook LJ, Pandolfi JM. 2017 U-Th dating
- 493 reveals regional-scale decline of branching *Acropora* corals on the Great Barrier Reef over the
- 494 past century. *Proceedings of the National Academy of Sciences* 114(39), 10350–10355.
- 495 doi:10.1073/pnas.1705351114

496 Figures

497



499 Figure 1. Photographs illustrating encountered coral competitive interactions (A-H), and
500 schematic (I) indicating how they were characterized by taking into account a set of factors

- 501 relating to individual organisms (beige), interactions (green), and environments (blue). By
- 502 quantifying the dead-zones left by sweeper attack-tentacles on opponent coral skeletons in the
- 503 aftermath of competitive battles, overreach distances (white arrows) reflect short-term
- 504 competitive outcomes resulting from recently deployed assaults (hours to months preceding
- 505 observations). In contrast, given the slow growth of corals, overgrowth distances (yellow arrows)
- 506 often integrate competitive interactions over several years. See appendix 1 for further
- 507 information, and figure S2 for visualizations of the raw data.



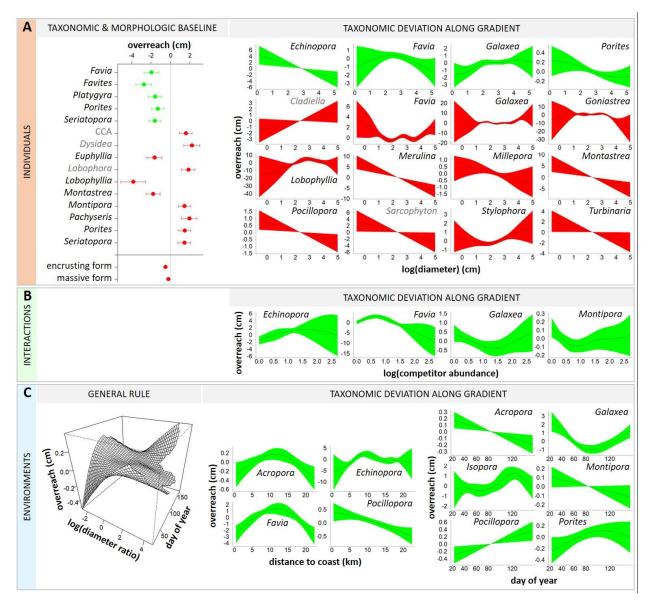


Figure 2. Changes in coral competitive performance as measured by net overreach distance along multiple ecological gradients. Plots illustrate partial contributions of different covariables to variation in net overreach of focal corals (mean ± standard error). Covariables are organized by scale, characterizing which organisms are involved (A, individuals), and how (B, interactions) and where/when (C, environments) the interactions occur. Covariables measured on focal corals are displayed in green (e.g. a positive effect of focal coral diameter on focal coral performance) and those on competing organisms in red (e.g. a negative effect of competitor diameter on focal

- 518 coral performance). Taxonomic and morphologic baselines identify differences in performance
- among species and growth-forms once the effects of other ecological gradients are accounted for.
- 520 Three-dimensional plot illustrates the interactive effects of two ecological gradients on the
- 521 response of all species, while other plots indicate deviations specific to some taxa and growth
- 522 forms. Note differences in axes ranges. Texts in grey distinguish non hard-coral species (CCA
- 523 for crustose coralline algae). Only significant effects are illustrated (table S2).
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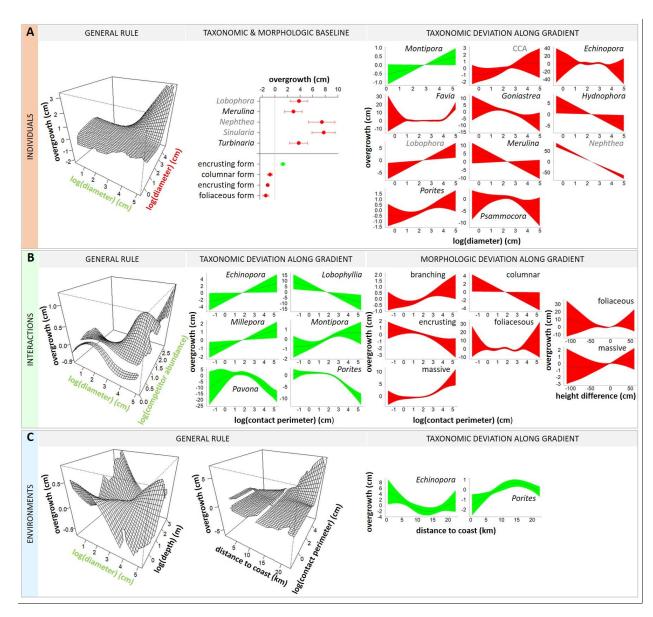
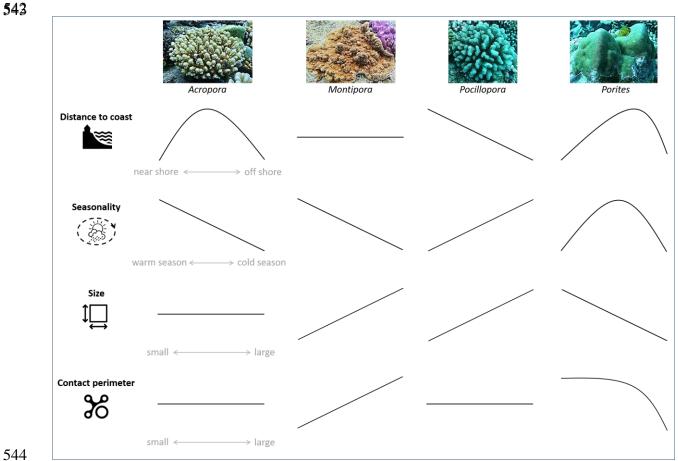


Figure 3. Changes in coral competitive performance as measured by net overgrowth distance along multiple ecological gradients. Plots illustrate partial contributions of different covariables to variation in net overgrowth of focal corals (mean ± standard error). Covariables are organized by scale, characterizing which organisms are involved (A, individuals), and how (B, interactions) and where/when (C, environments) the interactions occur. Covariables measured on focal corals are displayed in green (e.g. a positive effect of focal coral diameter on focal coral performance) and those on competing organisms in red (e.g. a negative effect of competitor diameter on focal

- 536 coral performance). Taxonomic and morphologic baselines identify differences in performance
- among species and growth-forms once the effects of other ecological gradients are accounted for.
- 538 Three-dimensional plots illustrate the interactive effects of two ecological gradients on the
- response of all species, while other plots indicate deviations specific to some taxa and growth
- 540 forms. Note differences in axes ranges. Texts in grey distinguish non hard-coral species (CCA
- 541 for crustose coralline algae). Only significant effects are illustrated (table S3).





545 Figure 4. Multi-dimensional niche segregation among the four major reef-building coral taxa as 546 revealed by variability in their competitive performances. The response patterns (summarized 547 from figures 2 and 3) indicate segregation in time and space (different environmental

548 preferences) as well as in life-strategies (different optimal sizes and warfare capacities).