

The behavioral effects of gestational and lactational benzo[a]pyrene exposure vary by sex and genotype in mice with differences at the *Ahr* and *Cyp1a2* loci

Amanda Honaker^{1*}, Angela Kyntchev^{1*}, Emma Foster¹, Katelyn Clough¹, Emmanuella Asiedu¹, Mackenzie Feltner¹, Victoria Ferguson¹, Philip Tyler Forrest¹, Jayasree Mullaguru¹, Mame Diarra Niang¹, Connor Perry¹, Yvonne Sene¹ and Christine Perdan Curran¹

1. Department of Biological Sciences, Northern Kentucky University, 100 Nunn Drive, Highland Heights, Kentucky, USA 41099

* These two authors contributed equally to this work.

Correspondence:

Christine Perdan Curran
Director, Neuroscience Program
SC344 Nunn Drive
Department of Biological Sciences
Northern Kentucky University
Highland Heights, KY 41099
E-mail: curranc1@nku.edu

Highlights

- Gestational and lactational benzo[a]pyrene (BaP) exposure has sex and genotype-specific neurobehavioral effects in mice.
- Female mice were more susceptible to motor deficits following developmental BaP exposure. Males were more susceptible to deficits in reversal learning and memory.
- *Ahr^bCyp1a2(-/-)* knockout mice were more susceptible to spatial learning and memory deficits following developmental BaP exposure.
- Poor-affinity *Ahr^dCyp1a2(-/-)* mice had deficits in spatial learning and memory regardless of treatment.

1 **Abstract**

2 Benzo[a]pyrene (BaP) is a polycyclic aromatic hydrocarbon (PAH) and known
3 carcinogen in the Top 10 on the United States' list of priority pollutants. Humans are
4 exposed through a variety of sources including tobacco smoke, grilled foods and fossil
5 fuel combustion. Recent studies of children exposed to higher levels of PAHs during
6 pregnancy and early life have identified numerous adverse effects on the brain and
7 behavior that persist into school age and adolescence. Our studies were designed to look
8 for genotype and sex differences in susceptibility to gestational and lactational exposure
9 to BaP using a mouse model with allelic differences in the aryl hydrocarbon receptor and
10 the xenobiotic metabolizing enzyme CYP1A2. Pregnant dams were exposed to 10
11 mg/kg/day of BaP in corn oil-soaked cereal or the corn oil vehicle alone from gestational
12 day 10 until weaning at postnatal day 25. Neurobehavioral testing began at P60 using one
13 male and one female per litter. We found main effects of sex, genotype and treatment as
14 well as significant gene x treatment and sex x treatment interactions. BaP-treated female
15 mice had shorter latencies to fall in the Rotarod test. High-affinity *Ahr^bCyp1a2(-/-)* mice
16 had greater impairments in Morris water maze. Interestingly, poor-affinity *Ahr^dCyp1a2(-/-)*
17 mice also had deficits in spatial learning and memory regardless of treatment. We
18 believe our findings provide future directions in identifying human populations at highest
19 risk of early life BaP exposure, because our model mimics known human variation in our
20 genes of interest. Our studies also highlight the value of testing both males and females
21 in all neurobehavioral studies.

22
23
24

251. Introduction

26 Benzo[a]pyrene (BaP) is a Group 1 carcinogen (IARC 2018) and ranked 8th on the U.S.
27 government's Priority Pollutants List while the entire class of compounds (polycyclic aromatic
28 hydrocarbons) ranks 9th (ASTDR 2019). Human exposures are widespread from cigarette smoke,
29 wildfires, fossil fuel combustion, and grilled foods (IARC 2018; ASTDR 1996). In addition to cancer
30 risk, there is accumulating evidence from human and animal studies that BaP is neurotoxic and that the
31 developing brain is particularly susceptible to BaP exposure. Chepelev et al. (2015) even suggested that
32 neurotoxicity might be a more sensitive endpoint for human risk assessments than carcinogenicity.

33 Coke oven workers with high BaP exposure had higher anxiety, impaired reaction time, and more
34 errors on standard tests of neurobehavioral function (Qui et al. 2013). Increased exposure to traffic-related
35 air pollution was associated with a significant decrease in sustained attention in teenagers (Kicinski et al.
36 2015). Air pollution is also considered a risk factor for autism (Lam et al. 2016) and neurodegenerative
37 disorders including Parkinson's disease (Lee et al. 2016) and Alzheimer's disease (Calderón-Garcidueñas
38 et al. 2018).

39 The most detailed studies including personal backpack monitors, cord blood measurements and
40 long-term assessments of children in New York City found consistent evidence linking developmental
41 PAH exposure to IQ deficits and behavioral problems (Perera et al. 2018; 2014; 2012). Peterson et al.
42 (2015) reported dose-dependent reductions in white matter following both prenatal and postnatal exposure
43 to PAHs. The changes were associated with reduced processing speed, ADHD and other behavioral
44 problems in children tracked up to age 9. There was an inverse relationship between cord blood levels of
45 brain-derived neurotrophic factor (BDNF) and prenatal PAH exposure (Perera et al. 2015). Prenatal PAH
46 exposures have now been linked with academic deficits in adolescents based on tests of language, spelling
47 and math (Margolis et al. 2021). Since many regions with high levels of PAH air pollution also suffer
48 from psycho-social stressors, the adverse outcomes can be magnified for many at-risk populations. For
49 example, Pagliaccio et al. (2020) found a significant interaction between prenatal PAH exposures and
50 early life stress that persisted throughout childhood.

51 Not all human studies have measured PAH exposure directly, and many rely on other components
52 of air pollution exposure such as particulate matter or nitrogen oxides (Rivas et al. 2019; Min & Min
53 2017; Sentis et al. 2017; Fuertes et al. 2016). This can make it difficult to pinpoint the specific effects
54 from each type of pollutant, although Jedrychowski et al. (2017) was able to demonstrate that PAH
55 exposure was more important than PM_{2.5} in reducing birth weights. The number of potential confounding
56 factors highlights the value of animal studies in identifying specific neurotoxic compounds, their
57 mechanism of action and modifying factors such as genetics and nutrition. Indeed, numerous studies have
58 reported neurotoxic effects from BaP exposure in multiple model systems from zebrafish to rats (Lin et al.
59 2020; Lyu et al. 2020; McCallister et al. 2016)

60 Our studies were designed to look at genetic differences in the aryl hydrocarbon receptor (AHR)
61 pathway known to alter BaP metabolism (Uno et al. 2006; 2004). We previously found that *Cyp1a2*(-/-)
62 knockout mice were more susceptible to developmental PCB exposure (Colter et al. 2018; Curran et al.
63 2012; 2011), and maternal CYP1A2 can protect offspring from gestational and lactational exposure to
64 AHR ligands (Dragin et al. 2006; Diliberto et al. 1997). Therefore, we hypothesized that wild-type, high-
65 affinity *Ahr^bCyp1a2*(+/+) dams might be better able to sequester and detoxify BaP compared with
66 *Cyp1a2*(-/-) knockouts having either the high-affinity *Ahr^b* allele or the poor-affinity *Ahr^d* allele. Using
67 both knockout lines allowed us to further determine if any observed neurotoxicity was AHR-mediated.
68 We used oral dosing to ensure that offspring were exposed only through the placenta and breast milk,
69 because BaP exposure in food can be a significant route of exposure to pregnant women (Wang et al.
70 2021; Duarte-Salles et al. 2013; Falco et al. 2003) and prior studies indicate oral dosing is the preferred
71 route of exposure for assessing gene-environment interactions related to BaP toxicity (Nebert et al. 2013).

72 **2. Methods**

73 **2.1 Animals**

74 All experiments and care were conducted under protocols approved by the Northern Kentucky
75 University Institutional Animal Care and Use Committee (IACUC) and in accordance with the ARRIVE

76 Guidelines (Percie du Sert et al. 2020; Kilkenny et al., 2010). Animal care and housing was in accordance
77 with the Guide for the Care and Use of Laboratory Animals (8th ed.).

78 **2.1.1. *Animal source and husbandry***

79 Male and female C57BL/6J mice were purchased from The Jackson Laboratory (Bar Harbor,
80 ME). Animals were group housed with a maximum of 4 per cage by sex and treatment group in standard
81 polysulfone shoebox cages with static microfilter lids, corncob bedding and a 12 h:12 h light-dark cycle.
82 Water and Lab Diet 5015 chow (18.9% protein; 11% fat, 32.6% starch) were provided *ad libitum*.
83 Animals were checked daily for health concerns, and cages were changed weekly with one cotton nestlet
84 provided for enrichment. Animals were acclimated to the vivarium for >1 wk prior to beginning
85 experimental treatments. *Cyp1a2(-/-)* knockout mice were originally back-crossed at least eight
86 generations onto a B6 background. Colonies have been maintained in the Northern Kentucky University
87 vivarium, and knockout lines were routinely back-crossed to the B6 breeders from The Jackson
88 Laboratory to avoid confounding by genetic drift. Nulliparous females were mated with males of the same
89 genotype on a four-day breeding cycle. The presence of a vaginal plug was considered evidence of
90 mating, and dams were removed from the breeding cage at that time. Litters were culled or cross-fostered
91 to balance litter size at six pups per dam.

92 **2.1.2. *Benzo[a]pyrene treatments***

93 Benzo[a]pyrene (> 96% purity) was purchased from Millipore-Sigma and dissolved in corn oil.
94 At gestational day 10, pregnant dams were randomly assigned to receive either 10/mg/kg/day BaP in corn
95 oil-soaked cereal (Cap'n Crunch peanut butter) or the corn oil vehicle until pups were weaned at postnatal
96 day 25 (P25). The dose was based on our preliminary studies to avoid overt toxicity and similar rodent
97 studies designed to mimic known human exposures (Chen et al. 2012; Zhang et al. 2016) while
98 accounting for the higher metabolic rate of laboratory mice.

99 **2.1.3. *Verification of genotype***

100 Tail snips (~ 1 cm) were taken, and PCR was used to verify the genotypes of all dams, sires and
101 offspring used in behavioral tests. Data from animals not matching the expected genotype was excluded
102 from analysis (n=4).

103 **2.2 Behavioral testing**

104 To verify that offspring received equivalent care, we used two tests of maternal behavior during
105 the first week following birth. One male and one female pup from each litter were randomly selected at
106 P25 for behavioral testing that began at P60. Experiments are presented in the order in which they were
107 conducted. Mice underwent no more than one behavioral test per day during the light cycle, and testing
108 was restricted to a four-hour time block to avoid confounding by circadian rhythms. Personnel handling
109 animals for behavior experiments were blinded to treatment groups. Lighting was set at 50% of full
110 brightness in all rooms used for behavioral testing. Male and female mice were tested separately on each
111 apparatus.

112 **2.2.1 Dam behavior**

113 We conducted snapshot ethograms of dam behavior within 1 h of lights on and 1 h of lights off to
114 assess general care of pups and maternal behavior. The percent time spent on nest during each 1 h
115 observation period was compared. We also used a scale of 0-3 to rate nest quality with 0 being the
116 absence of a nest and 3 being a fully developed nest including roof.

117 **2.2.2 Open field locomotor activity**

118 Mice were placed in a square plexiglass chamber (41 cm x 41 cm) for 60 min while a Photobeam
119 Activity System (San Diego Instruments, San Diego, CA) tracked animal activity during 12 five-min
120 intervals. The number of beam breaks was used as a measure of ambulatory activity. A second row of
121 photobeams detected rearing movements. Chambers were cleaned with 70% ethanol between each test.

122 **2.2.3 Acoustic startle with pre-pulse inhibition**

123 The SR-Lab apparatus (San Diego Instruments, San Diego, CA) was used to test baseline startle
124 response and sensorimotor gating with a 120 db startle stimulus and pre-pulses of 74 and 76 db. A Latin
125 Squares design was used with four trial types randomly presented over a total of 48 trials. Peak response

126 amplitudes (V_{max}) were recorded, and the percent attenuation following the pre-pulse tones was
127 calculated (Curran et al. 2012).

128 **2.2.4 Pole climb**

129 To test motor coordination, mice were placed at the top of a 50 cm pole facing upwards. The time
130 required to turn downward and total time to descend to the home cage were recorded (Colter et al. 2018).

131 **2.2.5 Rotarod**

132 After two days of acclimation, mice were tested for five days on a Rotarod (IITC Life Science
133 Inc., Woodland Hills, CA.) to assess motor coordination and motor learning. The Rotarod was
134 programmed to accelerate from 0-20 rpm over 180 s with a 300 s maximum trial time. Mice were given
135 three trials/day and the latency to fall or slip was recorded. The apparatus was cleaned between each trial
136 to avoid odor interference (Colter et al. 2018).

137 **2.2.6 Novel object recognition**

138 Mice were acclimated to a circular arena (91 cm) with no objects present for two days followed
139 by two days of acclimation with objects present similar in size to test objects, but with distinctly different
140 shapes and colors. On the fifth day, mice accumulated 30 s of exploration of two identical objects in the
141 familiarization phase. One hour later, visual recognition memory was tested by presenting a copy of the
142 familiar object and a novel test object. Again, each mouse accumulated 30 s total exploration time. The
143 percent time exploring the novel object was compared. Familiar and test objects were counter-balanced to
144 avoid confounding by object preference (Brown et al. 2020).

145 **2.2.7 Morris water maze**

146 The Morris water maze was used to assess hippocampal dependent spatial learning and memory
147 using a four-phase paradigm. (Brown et al. 2020, Curran et al., 2011, 2012; Vorhees and Williams, 2006).
148 In each phase, mice were placed in random start locations around a 112 cm tank with the goal of finding
149 an escape platform. Water temperature was maintained at $22^{\circ}\text{C} \pm 2^{\circ}\text{C}$, and the water was colored with
150 non-toxic white tempera paint to obscure the platform location. In the Cued phase, curtains were drawn
151 and an orange ping-pong ball on a pole in the center of a 10 cm circular platform served as a proximal cue

152 to the escape platform's location. The Cued phase demonstrated that mice could swim and could use
153 visual cues to find the escape platform. All mice successfully completed this phase and moved into the
154 Hidden Platform phases. Mice received 4 trials/day with a 15 min inter-trial interval for 6 days followed
155 by a 30 s Probe trial with the platform removed on day 7. Curtains were opened to reveal distal cues
156 around the room. Platform size decreased each week from 10 cm to 7 cm to 5 cm, and each week the
157 platform was moved to a new location. Latency to escape, distance traveled, swim speed, and average
158 distance to the escape platform were recorded by AnyMaze™ software (Stoelting, Inc.). Mice that did not
159 reach the platform within 60 s were guided to the platform and kept on the platform for 15 s. For probe
160 trials, average distance to the target and target zone crossings were recorded.

161 **2.2.8 Data analysis**

162 Data were analyzed by the KY-INBRE Applied Statistics Laboratory at the University of Kentucky
163 using SAS Proc Mixed and slice effects when significant differences were found. Corrections were made
164 for multiple post-hoc analyses. Swim speed was used as a covariate with latency to escape in Morris
165 water maze. A repeated measures design was used for Rotarod and Morris water maze with day as the
166 repeated measure and in Open Field Locomotor with interval as the repeated measure. Significance was
167 set as $P < 0.05$, and trends are noted if $P < 0.1$. Data are presented as LS means, and error bars represent
168 the standard error of the mean. Based on prior experience with mouse behavior, we used 15-20 litters per
169 group to generate sufficient power for analyses.

170 **3. Results**

171 **3.1 Dam behavior**

172 There were no significant differences in maternal care based on genotype or treatment when
173 comparing time spent on nest and nest quality ($P > 0.05$; data not shown).

174 **3.2 Open field locomotor activity**

175 There was a significant gene x treatment interaction when comparing time spent in the periphery v.
176 the central region of the arena. BaP-treated *Ahr^dCyp1a2(-/-)* mice spent less time in the central region,
177 which is generally considered an indication of increased anxiety-like behavior ($P < 0.05$; Fig. 1). There

178 was also a main effect of genotype with *Ahr^bCyp1a2(-/-)* mice spending significantly more time in the
179 central region compared with the other two genotypes. When comparing total beam breaks in each 5-
180 minute interval, BaP-treated high-affinity *Ahr^bCyp1a2(-/-)* mice were more active than their corn oil
181 controls while BaP-treated poor-affinity *Ahr^dCyp1a2(-/-)* mice were less active than their corn oil
182 controls. However, the differences were only significant in one of the 12 intervals. All groups showed
183 normal habituation over the 1 h period with lower activity in the later intervals (data not shown).
184

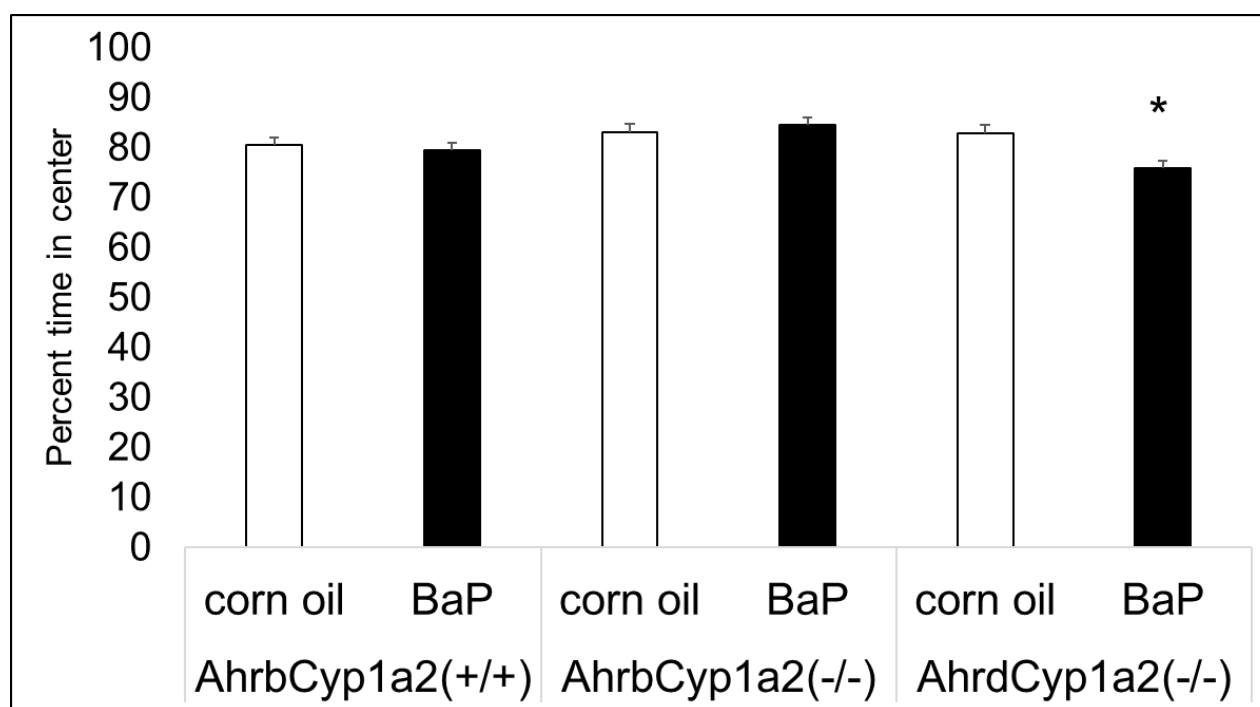
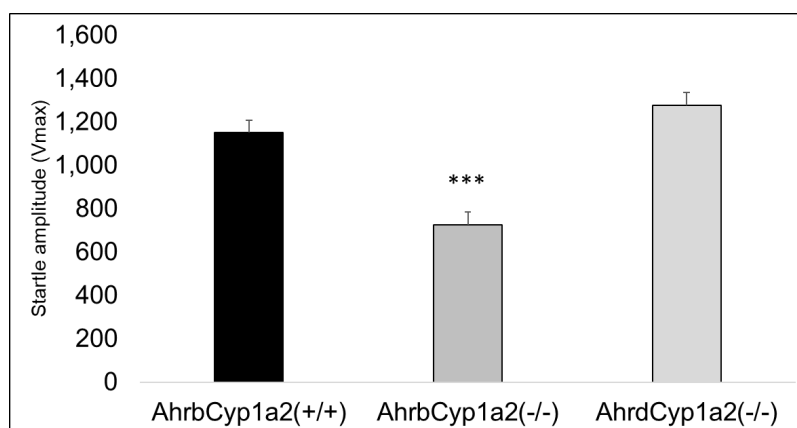


Fig. 1 Open field locomotor activity. BaP-exposed *Ahr^dCyp1a2(-/-)* mice spent significantly less time in the central area of the apparatus compared with all other groups. * P < 0.05

185 3.3 Acoustic startle with pre-pulse inhibition

186 We found the expected sex difference in the baseline startle response with heavier males
187 having a higher amplitude compared with females (P < 0.001) and a significant main effect of
188 genotype with high-affinity *Ahr^bCyp1a2(-/-)* mice having a much lower baseline startle response
189 compared with the other two genotypes (P < 0.001; Fig.2). There was no effect of treatment or

190 genotype on sensorimotor gating with all groups showing attenuation under the two pre-pulse
191 conditions.

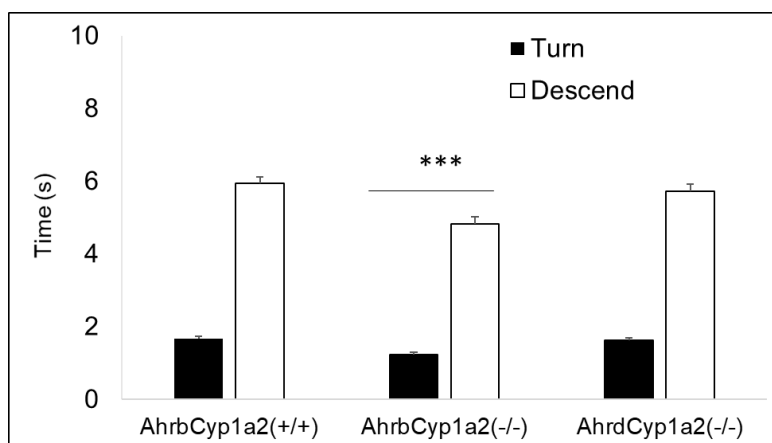


192

Fig. 2 Baseline startle response. *Ahr^bCyp1a2(-/-)* mice had a significantly lower response to the 120 db startle stimulus compared with the other two genotypes. *** $P < 0.001$

193 3.4 Pole climb

194 There was a main effect of genotype in the pole climb experiment. High-affinity
195 *Ahr^bCyp1a2(-/-)* mice had significantly shorter latencies to turn and to descend the pole ($P <$
196 0.001 ; Fig. 3). There was also a main effect of sex with females having shorter latencies to turn
197 ($P < 0.05$) and descend ($P < 0.01$), but there was no effect of BaP treatment ($P > 0.05$; data not
198 shown).



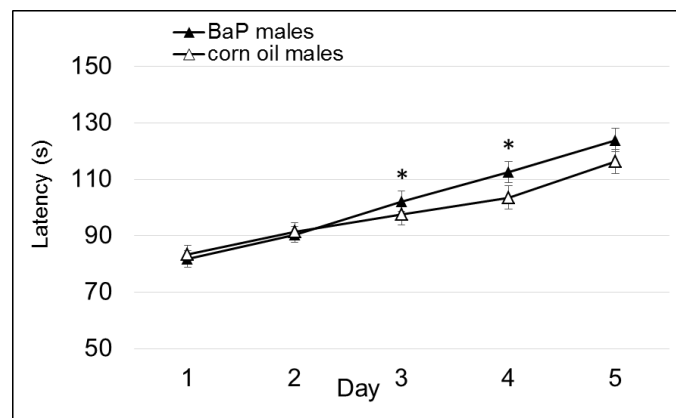
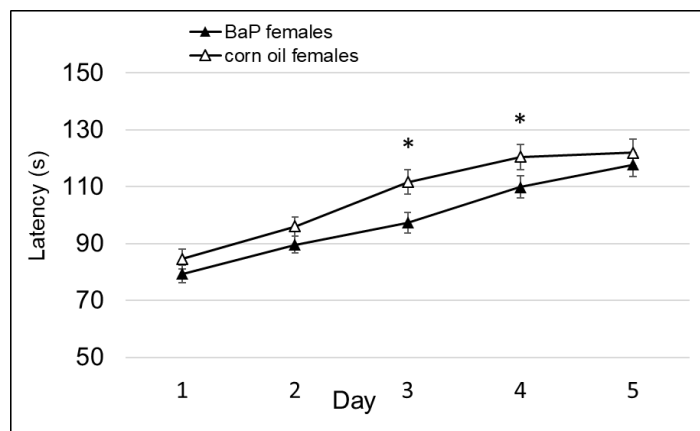
199

Fig. 3 Pole climb results. *Ahr^bCyp1a2(-/-)* mice had a significantly faster times to turn and descend a 50 cm pole. *** $P < 0.001$

200 3.5 Rotarod

201 There was a significant sex x treatment interaction in the Rotarod test. BaP-treated female
202 mice had shorter latencies to fall off the rotarod on all five days of testing compared with control
203 females; however, the differences were only significant on two of five days of testing ($P < 0.05$;
204 Fig. 4A). Interestingly, BaP-treated males had longer latencies than corn oil-treated males on
205 days 3-5 of testing and significantly longer latencies on two days of testing indicating greater
206 motor learning over the 5-day test (Fig. 4B).

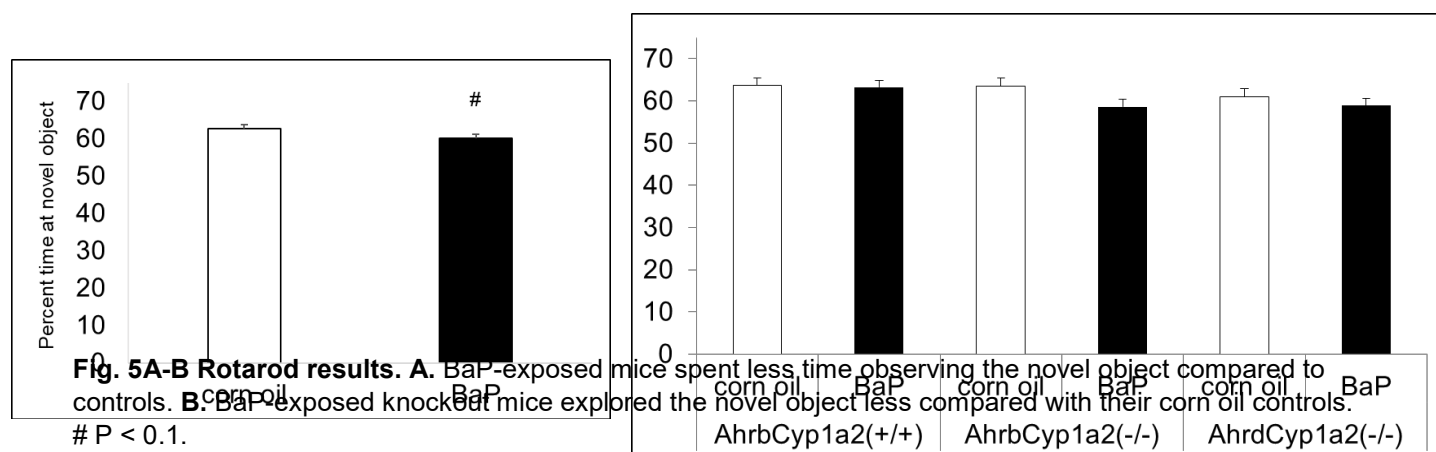
207



208 3.6 Novel object recognition

Fig. 4A-B Rotarod results. A. BaP-exposed female mice had shorter latencies to fall off the rotarod on days 3 and 4 of testing. **B.** BaP-exposed males had longer latencies on days 3 and 4. * $P < 0.05$

209 There was a trend for significance in Novel Object Recognition with BaP-treated mice
210 spending less time exploring the novel object ($P = 0.095$; Fig. 5A). Interestingly, the greatest
211 differences were seen in the two lines of *Cyp1a2*(-/-) knockout mice although there was not a
212 significant gene x treatment interaction ($P > 0.05$; Fig. 5B).



213

214 3.7 Morris water maze

215 In the Cued phase with a visible cue on the platform, all mice were able to swim and
 216 locate the escape platform. However, poor-affinity *Ahr^dCyp1a2(-/-)* mice had significantly longer
 217 escape latencies on four of six days of testing (Fig. 6). There was a main effect of treatment (P <
 218 0.05) on day2 and 6 with BaP-treated mice having longer latencies compared with corn oil-
 219 treated controls.

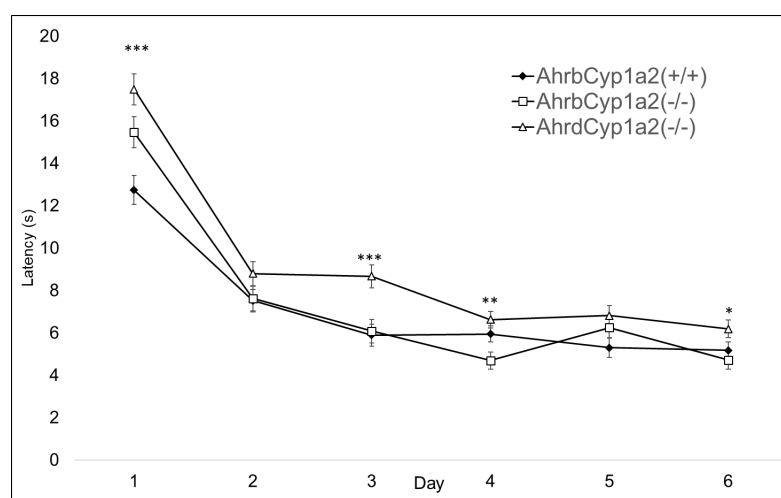
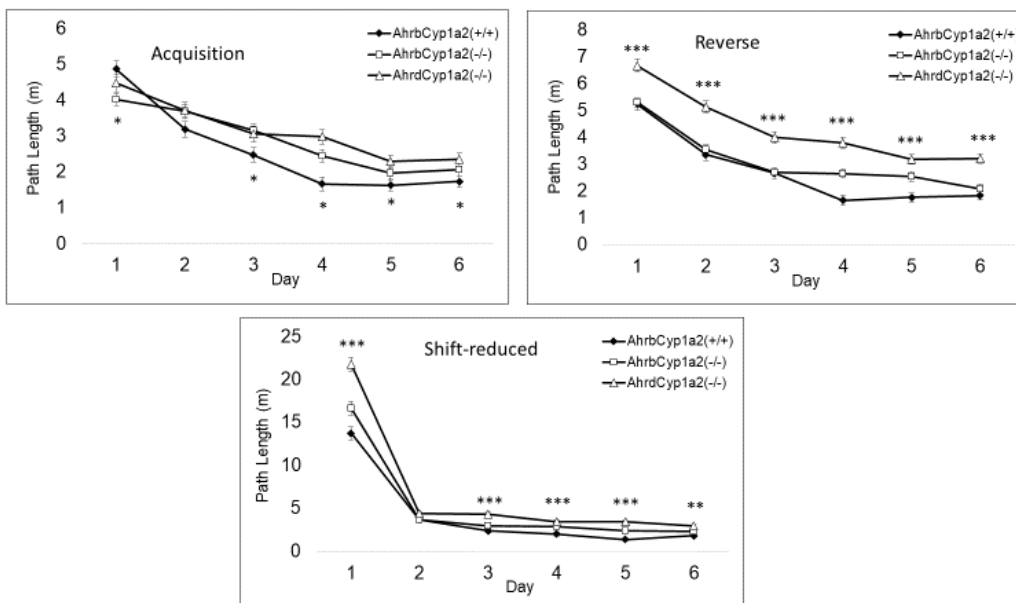


Fig.6 Morris water maze Cued Phase. Poor affinity *Ahr^dCyp1a2(-/-)* knockout mice had longer latencies to escape on days 1, 3, 4 and 6. * P < 0.05, ** P < 0.01, *** P < 0.001

220

221 In the hidden platform phases, there was a significant main effect of genotype in all three
222 phases with poor-affinity *Ahr^dCyp1a2(-/-)* mice having longer path lengths to the escape
223 platform on all but two days of testing (Figs 7A-C). In the Acquisition phase of the hidden
224 platform trials, there was a main effect of treatment on three of the first four days of testing,
225 indicating it took BaP-treated mice longer to learn the location of the escape platform. There was
226 a significant gene x treatment interaction on two days of testing and a trend toward significance
227 on two other days with BaP-treated *Ahr^bCyp1a2(-/-)* knockouts showing the greatest
228 impairments (Figs. 8A-C). In the Reverse phase, there was a significant main effect of treatment
229 on four days of testing with BaP-treated mice having longer average distances to the escape
230 platform (Fig. 9) and a significant sex x treatment interaction ($P < 0.05$) on two days of testing
231 with BaP-treated males having longer path lengths compared to corn oil-treated control males.
232 The opposite trend was seen in females. There was a trend for significance of treatment on Day 5
233 of the Shift-reduced phase ($P = 0.051$) and a significant difference on Day 6 for average distance
234 to the platform ($P < 0.01$).



236

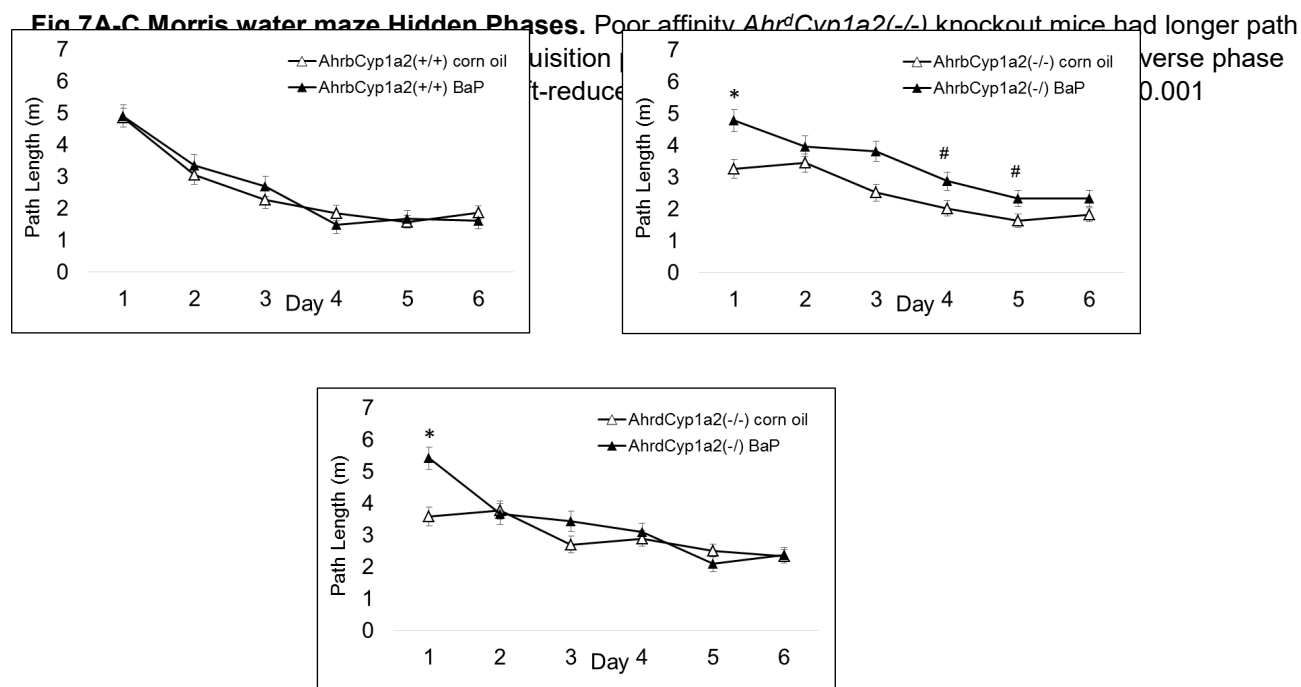
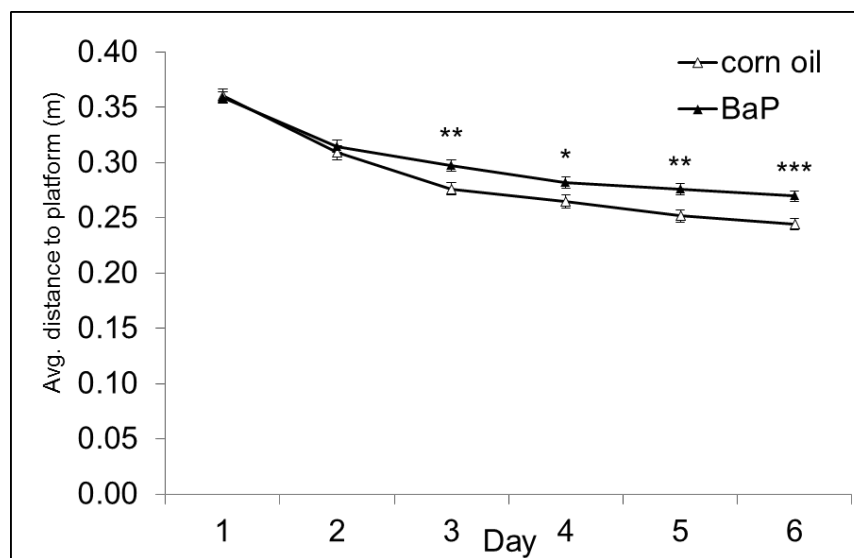
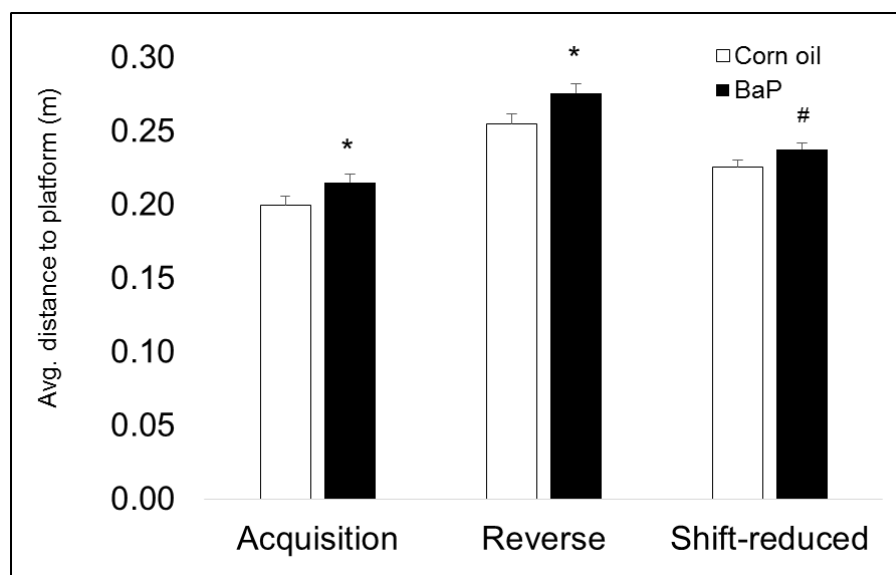


Fig.8A-C Morris Acquisition gene x treatment interaction. BaP-exposed high-affinity *Ahr^bCyp1a2(-/-)* knockout mice had significantly longer path lengths on days 1 and a trend for significance on days 4 and 5. # $P < 0.1$. * $P < 0.05$



237

238 In the Probe trials, BaP- treated mice had significantly longer average distances to the
239 platform in the Acquisition and Reverse phases ($P < 0.05$) and a trend for significance in the
240 Shift-reduced phase ($P = 0.68$, Fig.10). There was a significant gene x treatment interaction for
241 target zone crossings in the Shift-reduced phase with BaP-treated *Ahr^bCyp1a2(+/+)* wild type
242 mice having more crossings than their corn oil-treated controls. In contrast, BaP-treated



243 *Ahr^bCyp1a2(-/-)* knockout mice had fewer crossings than their corn oil-treated controls ($P <$
244 0.05).

245

246 **4. Discussion**

247 Using a mouse model with allelic differences at the *Ahr* and *Cyp1a2* loci, we examined adult offspring
248 exposed to benzo[a]pyrene (BaP) during gestation and lactation, a time of critical brain development. We
249 found a variety of treatment, sex and genotype effects and multiple interactions using a battery of behavioral
250 tests to assess locomotor activity, sensorimotor gating, motor function, and learning and memory. Our
251 findings of BaP neurotoxicity are consistent with previous rodent studies (Patel et al. 2016; McCallister et

Fig. 10 Morris Probe trials. BaP exposed mice had significantly longer average distances to the platform location in the Acquisition and Reverse probe trials and a trend for significance in the Shift-reduced phase. * $P < 0.05$, # $P < 0.1$

252 al. 2016; Maciel et al. 2014). Importantly, there were no differences in maternal care based on two measures
253 of dam behavior. Therefore, it is highly likely that the adverse effects reported here were due to BaP
254 exposure and the reported interactions of genotype x treatment and sex x treatment have helped to identify
255 the groups most susceptible to early life BaP exposure.

256 BaP-treated poor-affinity *Ahr^dCyp1a2(-/-)* knockout mice spent significantly less time in the central
257 area of the open field locomotor apparatus compared with all other groups; however, the percentage of time
258 in the center was still well above 70%. Therefore, it would be difficult to conclude that this indicates
259 increased anxiety-like behavior. There was no effect of BaP treatment in the Pole Climb test, but there was
260 a sex x treatment interaction in Rotarod. BaP-treated females were impaired relative to corn oil-treated
261 females whereas BaP-treated males out-performed corn oil-treated males overall. Liu et al. (2002) and Patel
262 et al. (2016) both reported impairments in Rotarod following BaP treatment in male Wistar rats, but to our
263 knowledge, this is the first report of sex-specific effects in Rotarod following developmental BaP exposure.
264 In human studies, Dix-Cooper et al. (2012) reported impaired fine motor skills in school-aged children
265 exposed to wood smoke in cooking fires and carbon monoxide. Our findings indicate that PAH exposure
266 should also be considered as a possible causative agent.

267 It will also be important to look for sex differences in BaP tissue levels. Singh et al. (1997)
268 demonstrated that higher expression of three glutathione S-transferases in male A/J mice protected them
269 from tumors following BaP exposure compared with females. Although all mice in this study are on a
270 C57BL/6J background, it's plausible that these females were also less able to detoxify BaP, resulting in a
271 higher body burden and greater neurotoxicity compared with males. On the other hand, there was a sex x
272 treatment interaction in the Reverse phase of Morris water maze with BaP-exposed males having greater
273 impairments than females. Therefore, future studies should include neurotransmitter analyses of all relevant
274 brain regions to determine if there are sex-dependent effects that could explain the observed differences in
275 susceptibility.

276 BaP-exposed high-affinity *Ahr^bCyp1a2(-/-)* knockout mice had longer path lengths compared to
277 corn oil control mice in the Reverse Phase of Morris water maze, and both lines of BaP-exposed knockout

278 mice spent less time exploring the novel object compared with wild type mice. This indicates that both
279 *Cyp1a2* and *Ahr* genotype are important when considering susceptibility to developmental BaP exposure
280 on measures of hippocampal dependent learning and memory. Both genes are highly variable in the human
281 population (Zhou et al. 2010; Nebert 2017; Nebert et al. 2013), and our mouse model was designed to
282 mimic that genetic variation. Further work is needed to determine how genotype affects the transfer of BaP
283 from dam to fetus and offspring during gestation and lactation, but previous work with AHR agonists clearly
284 indicate that maternal genotype is critical to sequestering and metabolizing persistent organic pollutants
285 such as coplanar PCBs and dioxins (Curran et al. 2011; Dragin et al. 2006; Diliberto et al. 1997). CYP1A1
286 induction will be higher in *Ahr^b* mice compared with *Ahr^d* mice, and both CYP1A1 and CYP1A2 are
287 expressed in liver and the intestines (Uno et al. 2008). Therefore, it is likely we will find different levels of
288 the parent BaP compound and metabolites in tissues from each genotype tested and that differential
289 toxicokinetics could account for the observed differences in neurotoxicity. This is important, because Patel
290 et al. (2016) reported that acute neonatal exposure to BaP increased oxidative stress in the hippocampus, a
291 region essential for normal learning and memory. We would expect to see higher levels of oxidative stress
292 in animals with a higher body burden during early brain development, which would suggest a potential
293 mechanism for differential neurotoxicity.

294 Beyond the BaP effects, we found poor-affinity *Ahr^dCyp1a2(-/-)* knockout mice had significant
295 impairments in the Cued and Reverse phases of Morris water maze. This is consistent with our previous
296 findings of impaired spatial learning and memory in this line of mice (Curran et al. 2012). High-affinity
297 *Ahr^bCyp1a2(-/-)* knockout mice had shorter latencies to turn and descend in the Pole Climb test. Although
298 this cannot be considered an impairment, it does suggest a difference in motivation or potentially anxiety-
299 like behavior that can be further explored with tests such as Zero Maze and Marble Burying. These results
300 emphasize the importance of thoroughly characterizing the neurobehavioral phenotype of knockout mice
301 prior to drawing conclusions relative to environmental exposures. Given recent findings regarding the role
302 of AHR activation and the microbiome (Korecka et al. 2016; Nebert 2017) and the effects of BaP on the

303 gut microbiome (Ribi re et al. 2016), these mice are likely to be extremely useful in understanding normal
304 and perturbed signaling along the gut-brain axis.

305 Given the widespread of exposure to BaP through fossil fuel combustion, wildfires, and grilled
306 food (ASTDR 2019; Jedrychowski et al. 2012), the impact of PAH exposure on human health will remain
307 an important public health concern. Recent studies have strongly associated prenatal and early life PAH
308 exposure with adverse behavioral and neurological outcomes in school-aged children and adolescents
309 (Margolis et al. 2021; Perera et al. 2018; Peterson et al. 2015). Teasing out the causative agent and threshold
310 for adverse effects becomes more difficult when studying human populations because there are typically
311 multiple co-exposures such as ozone, nitrogen oxides and particulate matter. Individuals living in urban
312 areas with high levels of traffic-related air pollution are also more likely to face additional psycho-social
313 stressors and the lingering contamination from lead paint and lead water lines.

314 **5. Conclusions**

315 We found expected BaP-mediated impairments in learning and memory with high-affinity
316 *Ahr^bCyp1a2(-/-)* showing the greatest susceptibility to developmental BaP exposure. We identified sex as
317 a key variable for motor function deficits with BaP-exposed females having impairments whereas males
318 did not. Intriguingly, we found numerous differences based on genotype alone. These findings replicate and
319 extend our previous work with *Cyp1a2(-/-)* knockout mice and suggest that our genes of interest have a
320 normal function in brain development or function. This emphasizes the importance of using intact animal
321 models to gain a deeper understanding of neurodevelopment and neurobehavior and the potential value of
322 these lines of research.

323 **6. Author contributions**

324 All authors made substantial contributions to the experimental work and to the
325 development of the manuscript. The two first authors prepared the draft manuscript and assisted
326 with treatments, animal care, data collection and preparation of data for statistical analysis. The
327 other co-authors all assisted with treatments, animal care, neurobehavioral testing and review of

328 the manuscript. The senior author Curran served as the principal investigator responsible for
329 training, experimental design and oversight of the project.

330 ***7. Acknowledgements***

331 We acknowledge support from multiple funding sources including NIH grants
332 R15ES030541, R15ES020053, P20GM103436 and NSF RSF-034-07, Society of Toxicology
333 internship grants, and the following Northern Kentucky University sources: College of Arts and
334 Sciences Collaborative Faculty-Student Project Awards, Faculty Development Project Grants,
335 Center for Integrative Natural Sciences and Mathematics (CINSAM) Research Grants, CINSAM
336 UR-STEM Fellowships, Greaves and Herrmann Fellowships, and Student Undergraduate
337 Research and Creative Activities awards. We thank Dr. Daniel W. Nebert (University of Cincinnati
338 Medical Center) for the generous donation of *Cyp1a(-/-)* knockout mice. We gratefully
339 acknowledge statistical assistance from Melinda MacDougall. Finally, the encouragement,
340 guidance and wisdom of Dr. Philip J. Bushnell must be acknowledged as well as his legacy of
341 fostering the highest level of rigor and reproducibility in behavioral toxicology.

342 ***8. Conflicts of interest***

343 The authors have no conflicts to declare.

344 ***9. Data statement***

345 All data are available upon request as well as detailed protocols for all experiments
346 described herein. Portions of these data were presented previously at the annual meetings of the
347 Society of Toxicology, the Developmental Neurotoxicology Society and the Society for Birth
348 Defects Research and Prevention.

349

350

351 **Literature Cited**

- 352 ATSDR (Agency for Toxic Substances and Disease Registry). 1995. Toxicological profile for
353 polycyclic aromatic hydrocarbons. Atlanta, GA: U.S. Department of Health and Human
354 Services, Public Health Service.
- 355 ATSDR (Agency for Toxic Substances and Disease Registry). 2018. Substance Priority List.
356 <https://www.atsdr.cdc.gov/spl/index.html#2019spl>
- 357 Brown, J., Villalona, Y., Weimer, J., Ludwig, C.P., Hays, B.T., Massie, L., Marczynski, C.A.,
358 Curran, C.P., 2020. Supplemental taurine during adolescence and early adulthood has sex-
359 specific effects on cognition, behavior and neurotransmitter levels in C57BL/6J mice
360 dependent on exposure window. *Neurotoxicol Teratol* 79, 106883.
361 <https://doi.org/10.1016/j.ntt.2020.106883>
- 362 Calderón-Garcidueñas, L., González-Maciél, A., Reynoso-Robles, R., Delgado-Chávez, R.,
363 Mukherjee, P.S., Kulesza, R.J., Torres-Jardón, R., Ávila-Ramírez, J., Villarreal-Ríos, R.,
364 2018. Hallmarks of Alzheimer disease are evolving relentlessly in Metropolitan Mexico
365 City infants, children and young adults. APOE4 carriers have higher suicide risk and higher
366 odds of reaching NFT stage V at ≤ 40 years of age. *Environ Res* 164, 475–487.
367 <https://doi.org/10.1016/j.envres.2018.03.023>
- 368 Chen, C., Tang, Y., Jiang, X., Qi, Y., Cheng, S., Qiu, C., Peng, B., Tu, B., 2012. Early postnatal
369 benzo(a)pyrene exposure in Sprague-Dawley rats causes persistent neurobehavioral
370 impairments that emerge postnatally and continue into adolescence and adulthood. *Toxicol*
371 *Sci* 125, 248–261. <https://doi.org/10.1093/toxsci/kfr265>
- 372 Chepelev, N.L., Moffat, I.D., Bowers, W.J., Yauk, C.L., 2015. Neurotoxicity may be an
373 overlooked consequence of benzo[a]pyrene exposure that is relevant to human health risk
374 assessment. *Mutat Res Rev Mutat Res* 764, 64–89.
375 <https://doi.org/10.1016/j.mrrev.2015.03.001>
- 376 Colter, B.T., Garber, H.F., Fleming, S.M., Fowler, J.P., Harding, G.D., Hooven, M.K., Howes,
377 A.A., Infante, S.K., Lang, A.L., MacDougall, M.C., Stegman, M., Taylor, K.R., Curran,
378 C.P., 2018. Ahr and Cyp1a2 genotypes both affect susceptibility to motor deficits
379 following gestational and lactational exposure to polychlorinated biphenyls.
380 *Neurotoxicology* 65, 125–134. <https://doi.org/10.1016/j.neuro.2018.01.008>
- 381 Curran, C.P., Altenhofen, E., Ashworth, A., Brown, A., Kamau-Cheggeh, C., Curran, M., Evans,
382 A., Floyd, R., Fowler, J., Garber, H., Hays, B., Kraemer, S., Lang, A., Mynhier, A.,
383 Samuels, A., Strohmaier, C., 2012. Ahrd Cyp1a2(-/-) mice show increased susceptibility
384 to PCB-induced developmental neurotoxicity. *Neurotoxicology* 33, 1436–1442.
385 <https://doi.org/10.1016/j.neuro.2012.08.005>
- 386 Curran, C.P., Vorhees, C.V., Williams, M.T., Genter, M.B., Miller, M.L., Nebert, D.W., 2011. In
387 utero and lactational exposure to a complex mixture of polychlorinated biphenyls: toxicity

- 388 in pups dependent on the Cyp1a2 and Ahr genotypes. *Toxicol Sci* 119, 189–208.
389 <https://doi.org/10.1093/toxsci/kfq314>
- 390 Diliberto, J.J., Burgin, D., Birnbaum, L.S., 1997. Role of CYP1A2 in hepatic sequestration of
391 dioxin: studies using CYP1A2 knock-out mice. *Biochem Biophys Res Commun* 236, 431–
392 433. <https://doi.org/10.1006/bbrc.1997.6973>
- 393 Dix-Cooper, L., Eskenazi, B., Romero, C., Balmes, J., Smith, K.R., 2012. Neurodevelopmental
394 performance among school age children in rural Guatemala is associated with prenatal and
395 postnatal exposure to carbon monoxide, a marker for exposure to woodsmoke.
396 *Neurotoxicology* 33, 246–254. <https://doi.org/10.1016/j.neuro.2011.09.004>
- 397 Dragin, N., Dalton, T.P., Miller, M.L., Shertzer, H.G., Nebert, D.W., 2006. For dioxin-induced
398 birth defects, mouse or human CYP1A2 in maternal liver protects whereas mouse CYP1A1
399 and CYP1B1 are inconsequential. *J Biol Chem* 281, 18591–18600.
400 <https://doi.org/10.1074/jbc.M601159200>
- 401 Duarte-Salles, T., Mendez, M.A., Meltzer, H.M., Alexander, J., Haugen, M., 2013. Dietary
402 benzo(a)pyrene intake during pregnancy and birth weight: associations modified by
403 vitamin C intakes in the Norwegian Mother and Child Cohort Study (MoBa). *Environ Int*
404 60, 217–223. <https://doi.org/10.1016/j.envint.2013.08.016>
- 405 Falcó, G., Domingo, J.L., Llobet, J.M., Teixidó, A., Casas, C., Müller, L., 2003. Polycyclic
406 aromatic hydrocarbons in foods: human exposure through the diet in Catalonia, Spain. *J*
407 *Food Prot* 66, 2325–2331. <https://doi.org/10.4315/0362-028x-66.12.2325>
- 408 Fuertes, E., Standl, M., Forns, J., Berdel, D., Garcia-Aymerich, J., Markevych, I., Schulte-Koerne,
409 G., Sugiri, D., Schikowski, T., Tiesler, C.M.T., Heinrich, J., 2016. Traffic-related air
410 pollution and hyperactivity/inattention, dyslexia and dyscalculia in adolescents of the
411 German GINIplus and LISApplus birth cohorts. *Environ Int* 97, 85–92.
412 <https://doi.org/10.1016/j.envint.2016.10.017>
- 413 Genkinger, J.M., Stigter, L., Jedrychowski, W., Huang, T.-J., Wang, S., Roen, E.L., Majewska,
414 R., Kieltyka, A., Mroz, E., Perera, F.P., 2015. Prenatal polycyclic aromatic hydrocarbon
415 (PAH) exposure, antioxidant levels and behavioral development of children ages 6-9.
416 *Environ Res* 140, 136–144. <https://doi.org/10.1016/j.envres.2015.03.017>
- 417 IARC (International Agency for Research on Cancer. 2018. Monograph on Benzo[a]Pyrene.
418 <https://monographs.iarc.who.int/wp-content/uploads/2018/06/mono100F-14.pdf>
- 419 ILAR (Institute for Laboratory Animal Research) 2011. Guide for the Care and Use of Laboratory
420 Animals, 8th edition. National Research Council (US) Committee for the Update of the
421 Guide for the Care and Use of Laboratory Animals. Washington (DC): National Academies
422 Press (US); 2011.
- 423 Iyer, S., Perera, F., Zhang, B., Chanock, S., Wang, S., Tang, D., 2014. Significant interactions
424 between maternal PAH exposure and haplotypes in candidate genes on B[a]P-DNA
425 adducts in a NYC cohort of non-smoking African-American and Dominican mothers and
426 newborns. *Carcinogenesis* 35, 69–75. <https://doi.org/10.1093/carcin/bgt339>

- 427 Iyer, S., Wang, Y., Xiong, W., Tang, D., Jedrychowski, W., Chanock, S., Wang, S., Stigter, L.,
428 Mróz, E., Perera, F., 2016. Significant interactions between maternal PAH exposure and
429 single nucleotide polymorphisms in candidate genes on B[a]P-DNA adducts in a cohort
430 of non-smoking Polish mothers and newborns. *Carcinogenesis* 37, 1110–1115.
431 <https://doi.org/10.1093/carcin/bgw090>
- 432 Jedrychowski, W.A., Majewska, R., Spengler, J.D., Camann, D., Roen, E.L., Perera, F.P., 2017.
433 Prenatal exposure to fine particles and polycyclic aromatic hydrocarbons and birth
434 outcomes: a two-pollutant approach. *Int Arch Occup Environ Health* 90, 255–264.
435 <https://doi.org/10.1007/s00420-016-1192-9>
- 436 Jedrychowski, W.A., Perera, F.P., Camann, D., Spengler, J., Butscher, M., Mroz, E., Majewska,
437 R., Flak, E., Jacek, R., Sowa, A., 2015. Prenatal exposure to polycyclic aromatic
438 hydrocarbons and cognitive dysfunction in children. *Environ Sci Pollut Res Int* 22, 3631–
439 3639. <https://doi.org/10.1007/s11356-014-3627-8>
- 440 Korecka, A., Dona, A., Lahiri, S., Tett, A.J., Al-Asmakh, M., Braniste, V., D'Arienzo, R.,
441 Abbaspour, A., Reichardt, N., Fujii-Kuriyama, Y., Rafter, J., Narbad, A., Holmes, E.,
442 Nicholson, J., Arulampalam, V., Pettersson, S., 2016. Bidirectional communication
443 between the Aryl hydrocarbon Receptor (AhR) and the microbiome tunes host metabolism.
444 *NPJ Biofilms Microbiomes* 2, 16014. <https://doi.org/10.1038/npjbiofilms.2016.14>
- 445 Lam, J., Sutton, P., Kalkbrenner, A., Windham, G., Halladay, A., Koustas, E., Lawler, C.,
446 Davidson, L., Daniels, N., Newschaffer, C., Woodruff, T., 2016. A Systematic Review and
447 Meta-Analysis of Multiple Airborne Pollutants and Autism Spectrum Disorder. *PLoS One*
448 11, e0161851. <https://doi.org/10.1371/journal.pone.0161851>
- 449 Lee, J., Kalia, V., Perera, F., Herbstman, J., Li, T., Nie, J., Qu, L.R., Yu, J., Tang, D., 2017. Prenatal
450 airborne polycyclic aromatic hydrocarbon exposure, LINE1 methylation and child
451 development in a Chinese cohort. *Environ Int* 99, 315–320.
452 <https://doi.org/10.1016/j.envint.2016.12.009>
- 453 Li, C., Wang, J., Su, Q., Yang, K., Chen, C., Jiang, X., Han, T., Cheng, S., Mo, T., Zhang, R.,
454 Peng, B., Guo, Y., Baker, P.N., Tu, B., Xia, Y., 2018. Postnatal Subacute Benzo(a)Pyrene
455 Exposure Caused Neurobehavioral Impairment and Metabolomic Changes of Cerebellum
456 in the Early Adulthood Period of Sprague-Dawley Rats. *Neurotox Res* 33, 812–823.
457 <https://doi.org/10.1007/s12640-017-9832-8>
- 458 Lin, Y.-C., Wu, C.-Y., Hu, C.-H., Pai, T.-W., Chen, Y.-R., Wang, W.-D., 2020. Integrated Hypoxia
459 Signaling and Oxidative Stress in Developmental Neurotoxicity of Benzo[a]Pyrene in
460 Zebrafish Embryos. *Antioxidants (Basel)* 9, E731. <https://doi.org/10.3390/antiox9080731>
- 461 Liu, S.-H., Wang, J.-H., Chuu, J.-J., Lin-Shiau, S.-Y., 2002. Alterations of motor nerve functions
462 in animals exposed to motorcycle exhaust. *J Toxicol Environ Health A* 65, 803–812.
463 <https://doi.org/10.1080/00984100290071144>
- 464 Lyu, Y., Ren, X.-K., Zhang, H.-F., Tian, F.-J., Mu, J.-B., Zheng, J.-P., 2020. Sub-chronic
465 administration of benzo[a]pyrene disrupts hippocampal long-term potentiation via

- 466 inhibiting CaMK II/PKC/PKA-ERK-CREB signaling in rats. *Environ Toxicol* 35, 961–
467 970. <https://doi.org/10.1002/tox.22932>
- 468 Margolis, A.E., Herbstman, J.B., Davis, K.S., Thomas, V.K., Tang, D., Wang, Y., Wang, S.,
469 Perera, F.P., Peterson, B.S., Rauh, V.A., 2016. Longitudinal effects of prenatal exposure
470 to air pollutants on self-regulatory capacities and social competence. *J Child Psychol*
471 *Psychiatry* 57, 851–860. <https://doi.org/10.1111/jcpp.12548>
- 472 Margolis, A.E., Ramphal, B., Pagliaccio, D., Banker, S., Selmanovic, E., Thomas, L.V., Factor-
473 Litvak, P., Perera, F., Peterson, B.S., Rundle, A., Herbstman, J.B., Goldsmith, J., Rauh, V.,
474 2021. Prenatal exposure to air pollution is associated with childhood inhibitory control and
475 adolescent academic achievement. *Environ Res* 202, 111570.
476 <https://doi.org/10.1016/j.envres.2021.111570>
- 477 McCallister, M.M., Li, Z., Zhang, T., Ramesh, A., Clark, R.S., Maguire, M., Hutsell, B., Newland,
478 M.C., Hood, D.B., 2016. Revealing Behavioral Learning Deficit Phenotypes Subsequent
479 to In Utero Exposure to Benzo(a)pyrene. *Toxicol Sci* 149, 42–54.
480 <https://doi.org/10.1093/toxsci/kfv212>
- 481 Min, J.-Y., Min, K.-B., 2017. Exposure to ambient PM10 and NO2 and the incidence of attention-
482 deficit hyperactivity disorder in childhood. *Environ Int* 99, 221–227.
483 <https://doi.org/10.1016/j.envint.2016.11.022>
- 484 Nebert, D.W., 2017. Aryl hydrocarbon receptor (AHR): “pioneer member” of the basic-
485 helix/loop/helix per-Arnt-sim (bHLH/PAS) family of “sensors” of foreign and endogenous
486 signals. *Prog Lipid Res* 67, 38–57. <https://doi.org/10.1016/j.plipres.2017.06.001>
- 487 Nebert, D.W., Shi, Z., Gálvez-Peralta, M., Uno, S., Dragin, N., 2013. Oral benzo[a]pyrene:
488 understanding pharmacokinetics, detoxication, and consequences--Cyp1 knockout mouse
489 lines as a paradigm. *Mol Pharmacol* 84, 304–313. <https://doi.org/10.1124/mol.113.086637>
- 490 Pagliaccio, D., Herbstman, J.B., Perera, F., Tang, D., Goldsmith, J., Peterson, B.S., Rauh, V.,
491 Margolis, A.E., 2020. Prenatal exposure to polycyclic aromatic hydrocarbons modifies the
492 effects of early life stress on attention and Thought Problems in late childhood. *J Child*
493 *Psychol Psychiatry* 61, 1253–1265. <https://doi.org/10.1111/jcpp.13189>
- 494 Patel, B., Das, S.K., Das, S., Das, L., Patri, M., 2016. Neonatal exposure to benzo[a]pyrene induces
495 oxidative stress causing altered hippocampal cytomorphometry and behavior during early
496 adolescence period of male Wistar rats. *Int J Dev Neurosci* 50, 7–15.
497 <https://doi.org/10.1016/j.ijdevneu.2016.01.006>
- 498 Percie du Sert, N., Hurst, V., Ahluwalia, A., Alam, S., Avey, M.T., Baker, M., Browne, W.J.,
499 Clark, A., Cuthill, I.C., Dirnagl, U., Emerson, M., Garner, P., Holgate, S.T., Howells,
500 D.W., Karp, N.A., Lazic, S.E., Lidster, K., MacCallum, C.J., Macleod, M., Pearl, E.J.,
501 Petersen, O.H., Rawle, F., Reynolds, P., Rooney, K., Sena, E.S., Silberberg, S.D., Steckler,
502 T., Würbel, H., 2020. The ARRIVE guidelines 2.0: updated guidelines for reporting animal
503 research. *BMJ Open Sci* 4, e100115. <https://doi.org/10.1136/bmjos-2020-100115>

- 504 Perera, F., Phillips, D.H., Wang, Y., Roen, E., Herbstman, J., Rauh, V., Wang, S., Tang, D., 2015.
505 Prenatal exposure to polycyclic aromatic hydrocarbons/aromatics, BDNF and child
506 development. *Environ Res* 142, 602–608. <https://doi.org/10.1016/j.envres.2015.08.011>
- 507 Perera, F., Weiland, K., Neidell, M., Wang, S., 2014. Prenatal exposure to airborne polycyclic
508 aromatic hydrocarbons and IQ: estimated benefit of pollution reduction. *J Public Health*
509 *Policy* 35, 327–336. <https://doi.org/10.1057/jphp.2014.14>
- 510 Perera, F.P., Chang, H., Tang, D., Roen, E.L., Herbstman, J., Margolis, A., Huang, T.-J., Miller,
511 R.L., Wang, S., Rauh, V., 2014. Early-life exposure to polycyclic aromatic hydrocarbons
512 and ADHD behavior problems. *PLoS One* 9, e111670.
513 <https://doi.org/10.1371/journal.pone.0111670>
- 514 Perera, F.P., Tang, D., Wang, S., Vishnevetsky, J., Zhang, B., Diaz, D., Camann, D., Rauh, V.,
515 2012. Prenatal polycyclic aromatic hydrocarbon (PAH) exposure and child behavior at age
516 6-7 years. *Environ Health Perspect* 120, 921–926. <https://doi.org/10.1289/ehp.1104315>
- 517 Perera, F.P., Wheelock, K., Wang, Y., Tang, D., Margolis, A.E., Badia, G., Cowell, W., Miller,
518 R.L., Rauh, V., Wang, S., Herbstman, J.B., 2018. Combined effects of prenatal exposure
519 to polycyclic aromatic hydrocarbons and material hardship on child ADHD behavior
520 problems. *Environ Res* 160, 506–513. <https://doi.org/10.1016/j.envres.2017.09.002>
- 521 Peterson, B.S., Rauh, V.A., Bansal, R., Hao, X., Toth, Z., Nati, G., Walsh, K., Miller, R.L., Arias,
522 F., Semanek, D., Perera, F., 2015. Effects of prenatal exposure to air pollutants (polycyclic
523 aromatic hydrocarbons) on the development of brain white matter, cognition, and behavior
524 in later childhood. *JAMA Psychiatry* 72, 531–540.
525 <https://doi.org/10.1001/jamapsychiatry.2015.57>
- 526 Qiu, C., Peng, B., Cheng, S., Xia, Y., Tu, B., 2013. The effect of occupational exposure to
527 benzo[a]pyrene on neurobehavioral function in coke oven workers. *Am J Ind Med* 56, 347–
528 355. <https://doi.org/10.1002/ajim.22119>
- 529 Ribière, C., Peyret, P., Parisot, N., Darcha, C., Déchelotte, P.J., Barnich, N., Peyretailade, E.,
530 Boucher, D., 2016. Oral exposure to environmental pollutant benzo[a]pyrene impacts the
531 intestinal epithelium and induces gut microbial shifts in murine model. *Sci Rep* 6, 31027.
532 <https://doi.org/10.1038/srep31027>
- 533 Rivas, I., Basagaña, X., Cirach, M., López-Vicente, M., Suades-González, E., Garcia-Esteban, R.,
534 Álvarez-Pedrerol, M., Dadvand, P., Sunyer, J., 2019. Association between Early Life
535 Exposure to Air Pollution and Working Memory and Attention. *Environ Health Perspect*
536 127, 57002. <https://doi.org/10.1289/EHP3169>
- 537 Sentís, A., Sunyer, J., Dalmau-Bueno, A., Andiarena, A., Ballester, F., Cirach, M., Estarlich, M.,
538 Fernández-Somoano, A., Ibarluzea, J., Íñiguez, C., Lertxundi, A., Tardón, A.,
539 Nieuwenhuijsen, M., Vrijheid, M., Guxens, M., INMA Project, 2017. Prenatal and
540 postnatal exposure to NO₂ and child attentional function at 4-5years of age. *Environ Int*
541 106, 170–177. <https://doi.org/10.1016/j.envint.2017.05.021>

- 542 Singh, S.V., Benson, P.J., Hu, X., Pal, A., Xia, H., Srivastava, S.K., Awasthi, S., Zaren, H.A.,
543 Orchard, J.L., Awasthi, Y.C., 1998. Gender-related differences in susceptibility of A/J
544 mouse to benzo[a]pyrene-induced pulmonary and forestomach tumorigenesis. *Cancer Lett*
545 128, 197–204. [https://doi.org/10.1016/s0304-3835\(98\)00072-x](https://doi.org/10.1016/s0304-3835(98)00072-x)
- 546 Uno, S., Dalton, T.P., Derkenne, S., Curran, C.P., Miller, M.L., Shertzer, H.G., Nebert, D.W.,
547 2004. Oral exposure to benzo[a]pyrene in the mouse: detoxication by inducible cytochrome
548 P450 is more important than metabolic activation. *Mol Pharmacol* 65, 1225–1237.
549 <https://doi.org/10.1124/mol.65.5.1225>
- 550 Uno, S., Dalton, T.P., Dragin, N., Curran, C.P., Derkenne, S., Miller, M.L., Shertzer, H.G.,
551 Gonzalez, F.J., Nebert, D.W., 2006. Oral benzo[a]pyrene in Cyp1 knockout mouse lines:
552 CYP1A1 important in detoxication, CYP1B1 metabolism required for immune damage
553 independent of total-body burden and clearance rate. *Mol Pharmacol* 69, 1103–1114.
554 <https://doi.org/10.1124/mol.105.021501>
- 555 Uno, S., Dragin, N., Miller, M.L., Dalton, T.P., Gonzalez, F.J., Nebert, D.W., 2008. Basal and
556 inducible CYP1 mRNA quantitation and protein localization throughout the mouse
557 gastrointestinal tract. *Free Radic Biol Med* 44, 570–583.
558 <https://doi.org/10.1016/j.freeradbiomed.2007.10.044>
- 559 Vorhees, C.V., Williams, M.T., 2006. Morris water maze: procedures for assessing spatial and
560 related forms of learning and memory. *Nat Protoc* 1, 848–858.
561 <https://doi.org/10.1038/nprot.2006.116>
- 562 Wang, Y., Jiao, Y., Kong, Q., Zheng, F., Shao, L., Zhang, T., Jiang, D., Gao, X., 2021. Occurrence
563 of polycyclic aromatic hydrocarbons in fried and grilled fish from Shandong China and
564 health risk assessment. *Environ Sci Pollut Res Int*. [https://doi.org/10.1007/s11356-021-](https://doi.org/10.1007/s11356-021-13045-y)
565 13045-y
- 566 Zhang, H., Nie, J., Li, X., Niu, Q., 2013. Association of aryl hydrocarbon receptor gene
567 polymorphism with the neurobehavioral function and autonomic nervous system function
568 changes induced by benzo[a]pyrene exposure in coke oven workers. *J Occup Environ Med*
569 55, 265–271. <https://doi.org/10.1097/JOM.0b013e318278272f>
- 570 Zhou, S.-F., Wang, B., Yang, L.-P., Liu, J.-P., 2010. Structure, function, regulation and
571 polymorphism and the clinical significance of human cytochrome P450 1A2. *Drug Metab*
572 *Rev* 42, 268–354. <https://doi.org/10.3109/03602530903286476>