

1 **Eurasian lynx populations in Western Europe: What prospects for the next 50 years?**

2

3 Bauduin Sarah, Germain Estelle, Zimmermann Fridolin, Idelberger Sylvia, Herdtfelder Micha,

4 Heurich Marco, Kramer-Schadt Stephanie, Duchamp Christophe, Drouet-Hoguet Nolwenn,

5 Morand Alain, Blanc Laetitia, Charbonnel Anaïs, and Gimenez Olivier

6

7 Bauduin Sarah, Office français de la biodiversité, 147 avenue de Lodève, 34990 Juvignac,

8 sarah.bauduin@ofb.gouv.fr

9 Germain Estelle, Centre de Recherche et d'Observation sur les Carnivores, Lucy,

10 estelle.germain@croc-crea.org

11 Zimmermann Fridolin, Carnivore Ecology and Wildlife Management. KORA, Thunstrasse 31,

12 CH-3074 Muri bei Bern, Switzerland f.zimmermann@kora.ch; orcid.org/0000-0003-3861-246X

13 Idelberger Sylvia, Stiftung Natur und Umwelt Rheinland-Pfalz, Diether-von-Isenburg-Str. 7,

14 55116 Mainz, Sylvia.Idelberger@snu.rlp.de

15 Herdtfelder Micha, Forest Research Institute of Baden-Wuerttemberg, Wonnhalde 4, 79100

16 Freiburg, micha.herdtfelder@forst.bwl.de

17 Heurich Marco, Chair of Wildlife Ecology and Management, Faculty of Environment and

18 Natural Resources, University of Freiburg, Tennenbacher Straße 4, 79106 Freiburg, Germany

19 Department of Visitor Management and National Park Monitoring, Bavarian Forest National

20 Park, Freyunger Straße 2, 94481 Grafenau, Germany

21 Faculty of Forestry and Wildlife Management, Campus Evenstad, Inland Norway University of

22 Applied Sciences, 2480 Koppang, Norway

23 Marco.Heurich@npv-bw.bayern.de

24 Kramer-Schadt Stephanie, Department of Ecological Dynamics, Leibniz Institute for Zoo and
25 Wildlife Research, D-10315 Berlin, Germany; Department of Ecology, Technische Universität
26 Berlin, Germany, kramer@izw-berlin.de; ORCID: 0000-0002-9269-4446
27 Duchamp Christophe, Office français de la biodiversité, Micropolis - La Bérardie F-05000 Gap,
28 christophe.duchamp@ofb.gouv.fr
29 Drouet-Hoguet Nolwenn, Office français de la biodiversité, 5 Allée de Bethléem, F-38610
30 Gières, nolwenn.drouet-hoguet@ofb.gouv.fr
31 Morand Alain, Centre d'études et d'expertise sur les Risques, l'Environnement, la Mobilité et
32 l'Aménagement, Direction territoriale Est, Ile du Saulcy, 57045 Metz, alain.morand@cerema.fr
33 Blanc Laetitia, CEFÉ, Univ Montpellier, CNRS, EPHE, IRD, Montpellier, France
34 blanc.laetitia.esr34@gmail.com
35 Charbonnel Anaïs, Centre de Recherche et d'Observation sur les Carnivores, Lucy,
36 anais.charbonnel@laposte.net
37 Gimenez Olivier, CEFÉ, CNRS, Univ Montpellier, EPHE, IRD, Montpellier, France
38 olivier.gimenez@cefe.cnrs.fr

39

40 **Abstract**

41 Persistence of viable populations may be uncertain for large carnivore species, especially for
42 those established in human-dominated landscapes. Here, we studied the Eurasian lynx in Western
43 Europe established in the Upper Rhine metapopulation (i.e., Jura, Vosges-Palatinian and Black
44 Forest populations) and in the Alpine population. These populations are currently considered as
45 endangered or critically endangered due to high anthropogenic mortality and isolation. We
46 assessed lynx persistence over a 50-year time horizon by implementing a spatially-explicit
47 individual-based model while accounting for road mortality and habitat selection. For the Alpine

48 and Jura populations, forecasts showed a steady growth rapidly reaching a stabilization phase
49 with high resident female occupancy, and a positive growth for the newly re-established Vosges-
50 Palatinian population. Moreover, this population showed a difference in female occupancy
51 between the northern part, where a recent reintroduction program was conducted in the Palatinate
52 Forest, and the southern part. Only the group of individuals in the Black Forest had an irregular
53 growth probably due to the small number of only male lynx at start of the simulation and poor
54 connectivity to surrounding populations. Exchanges of individuals between populations were
55 limited, and the Jura population played the role of a crossroad. Persistence of lynx in Western
56 Europe seems likely on a large scale over the next 50 years. However, lynx persistence in the
57 southern part of the Vosges-Palatinian population and in the Black Forest appears challenging
58 without long-term conservation management.

59

60 **Keywords:** *Lynx lynx*, population persistence, spatially-explicit individual-based model

61

62 **Introduction**

63

64 The Eurasian lynx (*Lynx lynx*) was eradicated in most of Europe between the 17th and 20th
65 centuries. The main reasons for its disappearance were habitat degradation, human persecution
66 and a decrease in prey availability (Breitenmoser et al., 2000). The species has recently
67 recolonized parts of its historical range in Central and Western Europe thanks to different
68 reintroduction programs which started in the 1970s. Nowadays there are ten or eleven (depending
69 on the author) identified lynx populations in Europe (Chapron et al., 2014; von Arx, 2020), and
70 the species benefits from a conservation status across its whole range area. Although the species
71 is considered as “least concerned” at the European level of the IUCN Red list, its status greatly

72 differs from one population to another, even though they share similar threats, mostly habitat
73 fragmentation, illegal killings and collisions with vehicles (von Arx, 2020). Long-term
74 persistence remains notably uncertain for the Alpine population (France and Switzerland) and for
75 the Upper Rhine meta-population, which encompasses the Jura population (France and
76 Switzerland), the Vosges-Palatinian population (France and Germany) and few individuals
77 located in the Black Forest and its surroundings (e.g., the Swabian Alb, Germany) (Drouet-
78 Hoguet et al., 2021; Germain and Schwoerer, 2021; Herdtfelder et al., 2021; Idelberger et al.,
79 2021; Molinari-Jobin et al., 2021). These populations are currently defined as endangered (Jura
80 and Alpine) or critically endangered (Vosges-Palatinian) (von Arx, 2020). Indeed, the
81 populations forming the Upper Rhine meta-population remain currently small and isolated. Only
82 a few exchanges between them and with the Alpine population are documented, most likely
83 because of habitat fragmentation and little functional connectivity (Morand, 2016; Zimmermann
84 and Breitenmoser, 2007). Low female dispersal rates and movements most likely slow population
85 expansion due to their risk-shy nature compared to male lynx (Port et al., 2021). This situation,
86 added to the common origin of the reintroduced individuals from the Carpathian population
87 (Breitenmoser et al., 1998; Vandel et al., 2006), may lead to high inbreeding within populations
88 (Breitenmoser-Würsten and Obexer-Ruff, 2003; Bull et al., 2016; Premier et al., 2020).

89 In this context, wildlife conservationists, scientists and policy-makers face significant
90 challenges for lynx conservation when individuals inhabit human-dominated landscapes. To
91 inform decision-making processes for population conservation, several studies have used
92 individual-based models (IBMs). These bottom-up models flexibly integrate species demography
93 (dispersal, territory establishment, reproduction, mortality) and how animals interact with their
94 inhomogeneous environment (e.g., habitat selection, collisions, illegal killings) and other
95 individuals, while accounting for individual characteristics (e.g., sex, age, dispersal status).

96 Population-level results emerge from the individual-level simulations (Railsback and Grimm,
97 2012). These models are used for the management and conservation of large and small carnivores
98 (e.g., Hradsky et al., 2019; Kramer-Schadt et al., 2011, 2005, 2004; Marucco and McIntire,
99 2010). Recent applications of IBMs for the lynx species allowed estimating illegal killing
100 (Heurich et al., 2018), linking movements and genetic diversity (Premier et al., 2020) and
101 evaluating reintroduction scenarios (Ovenden et al., 2019; Philips, 2020).

102 Lynx are large carnivores with large spatial requirements (Breitenmoser-Würsten et al.,
103 2007). They are territorial and live in large connected forested areas (Vandel and Stahl, 2005;
104 Zimmermann and Breitenmoser, 2007), which sustain large prey populations (Basille et al.,
105 2009). Natural barriers (e.g., large waterbodies) as well as anthropogenic barriers (e.g., urban
106 areas, highways) limit lynx movements (Schadt et al., 2002). Roads are complex infrastructures
107 acting as semi-permeable movement barriers (Klar et al., 2006; Zimmermann and Breitenmoser,
108 2007) as well as an important source of mortality (Bencin et al., 2019; Ceia-Hasse et al., 2017).
109 Therefore, a major modelling challenge when simulating lynx populations is to account for
110 movements over large areas and consequently, for the impact of landscape characteristics on
111 different demographic processes such as dispersal, territory establishment and survival rates.
112 Because simulation rules in IBMs are defined at the individual level, these models appear as a
113 great tool to integrate the complex interactions between landscape characteristics and both lynx
114 demographic and spatial processes to simulate population changes on a large scale.

115 Our study aimed to assess the long-term persistence of the Eurasian lynx in the large
116 range area delimited by the Upper Rhine meta-population and the Alpine population using the
117 available population status in 2018. We build a spatially-explicit IBM that integrates up-to-date
118 information on lynx ecology while accounting for habitat preferences and the risk of collisions
119 with vehicles. We implement our model in the R programming language and share it publicly

120 with open-source code allowing full reproducibility and its adoption by ecologists and
121 conservationists. We use this model to forecast the fate of lynx populations over a 50-year time
122 horizon. Finally, we provide several population metrics that help in diagnosing the long-term
123 species persistence in its most western range area.

124

125

126 **Material and Methods**

127

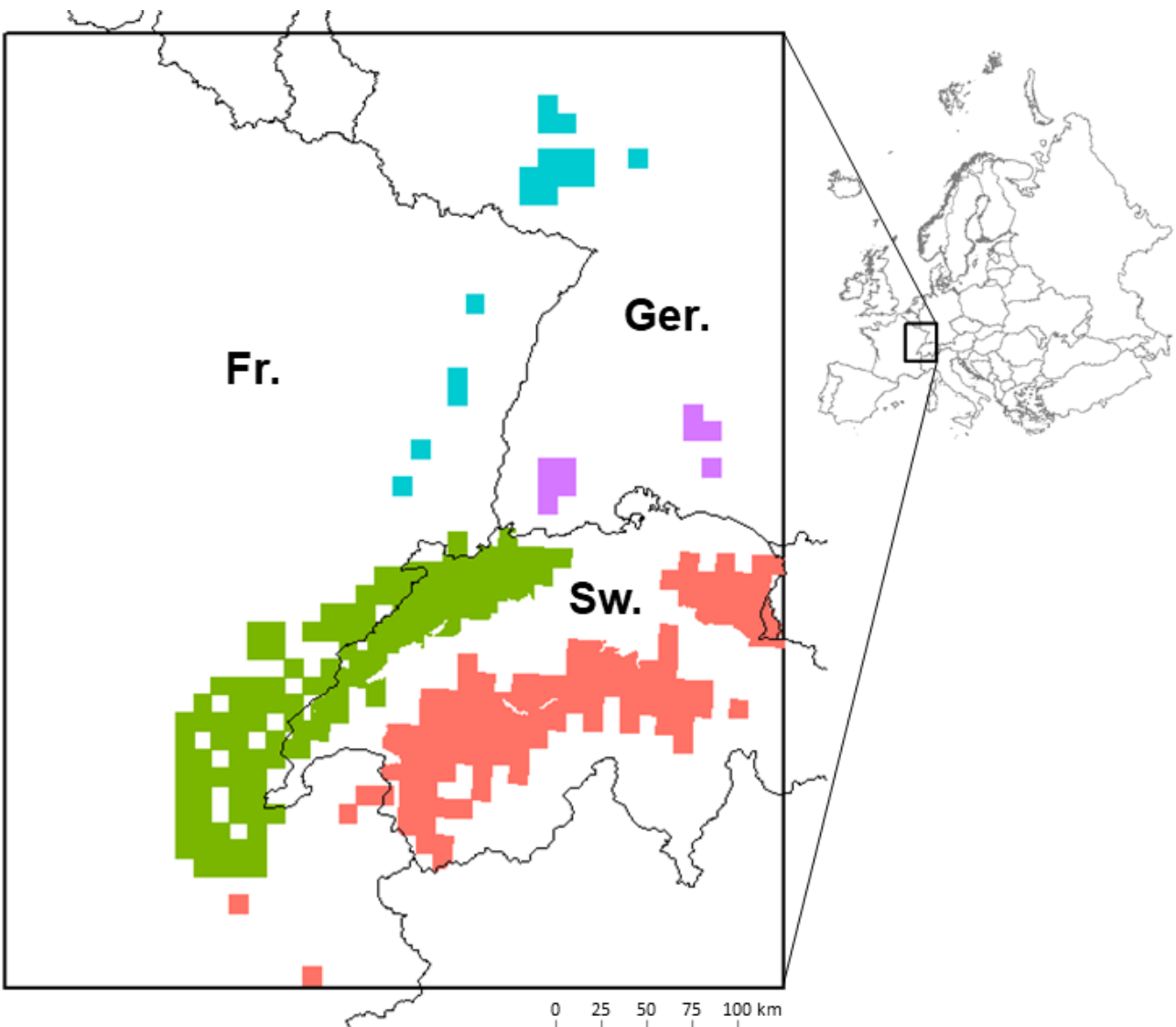
128 *Study area and populations*

129 We conducted the study on Eurasian lynx (*Lynx lynx*) populations located in France, Germany
130 and Switzerland (Fig. 1). The three main populations inhabiting the vast mountainous and
131 forested areas of these countries are the Vosges-Palatinian (France-Germany), the Jura (France-
132 Switzerland) and the Alpine (France-Switzerland) populations (von Arx, 2020). Some individuals
133 were also observed in the Black Forest and its surroundings (e.g., Swabian Alb) in Germany (Fig.
134 1), but were not considered as a lynx population *per se* (von Arx, 2020). Therefore, we should
135 refer to “mountain ranges” when speaking of the Vosges-Palatinian, Jura, Alps and Black Forest
136 with its surroundings, but for simplicity and clarity, we will use “populations” throughout our
137 paper.

138 The Jura and Alpine populations originated from individuals reintroduced in Switzerland
139 in the 1970s (Breitenmoser et al., 1998; Vandel and Stahl, 2005), followed by a natural
140 recolonization of the territories westward. After the complete decline of the Vosges-Palatinian
141 population in the 18th century, a reintroduction of 21 lynx occurred in the southern part of the
142 Vosges Mountains (France) between 1983 and 1993 (Vandel et al., 2006). Only 10 individuals
143 contributed to the lynx local establishment (Vandel et al., 2006) without conclusive stabilization

144 (Charbonnel and Germain, 2020). Since 2016, the lynx has been back in the Palatinate Forest
145 (Germany) thanks to a new reintroduction program finalized in 2020 (Scheid et al., 2021;
146 Schwoerer, 2021) with a few individuals that already moved to the Vosges Mountains (Scheid et
147 al., 2021; Schwoerer, 2021). Finally, only a few male individuals have been observed in the
148 Black Forest area since 2013, most of them coming from the Swiss Jura Mountains
149 (Mitarbeitende der Stiftung KORA, 2017; MLR, 2019; Wölfel et al., 2021). Lynx regular
150 occurrences are yearly documented by the three countries.

151



152

153 **Figure 1:** Eurasian lynx presence as available in 2017-2018 in France (Fr.), Germany (Ger.) and
154 Switzerland (Sw.) in the study area (black rectangle). Data for France cover the period from
155 01/04/2013 to 31/03/2017 (OFB Réseau Loup-Lynx), for Germany from 01/05/2017 to
156 30/04/2018 (Bundesamt für Naturschutz) and for Switzerland from 01/01/2015 to 31/12/2017
157 (KORA). We used a standardized 10x10 km grid from the European Environment Agency for
158 France and Germany, and a grid derived from the 1:25,000 map for Switzerland). The four colors
159 are for the four different populations: the Vosges-Palatinian population (blue; 1,800 km²) with
160 cells in France (Vosges Mountains) and Germany (Palatinate Forest), the Black Forest population
161 (purple; 900 km²) in Germany, and the Jura (green; 12,057 km²) and the Alpine (red; 11,190 km²)
162 populations with cells in France and Switzerland. Top right corner inset: Europe with the location
163 of the study area.

164

165 *Lynx population persistence*

166 Based on previous works by Kramer-Schadt et al. (2004, 2005, 2011), we built a spatially-
167 explicit individual-based model (SE-IBM) to simulate lynx populations dynamics and dispersal,
168 while accounting for the risk of lynx-vehicle collisions and lynx habitat preferences. A SE-IBM
169 is an IBM where individual responses to behavioral rules are constrained by environmental
170 characteristics. Our lynx SE-IBM is made of four components (Appendix A). The first
171 component represents the impact of road network on lynx survival via predicted collision
172 probabilities. The second component represents the impact of land cover on lynx space use with
173 the definition of different lynx habitat types. These first two components are spatial layers
174 influencing the behavioral rules followed by simulated lynx individuals. The third component
175 represents the initial lynx populations made of lynx individual's location and characteristics used
176 to launch the SE-IBM. The fourth component details all SE-IBM rules including lynx

177 demography, dispersal movement and territory establishment. A complete description of the four
178 model components is available in Appendix A, and the SE-IBM structure following the
179 Overview, Design concepts, and Details (ODD) protocol (Grimm et al., 2010, 2006) is provided
180 in Appendix B.

181 We ran 200 replicates of the lynx SE-IBM forecasting the populations over 50 years. We
182 used a different initial population (Appendix A) at each replicate to avoid bias due to initial
183 locations of simulated individuals. Initial populations were created based on real monitoring data
184 (Appendix A). We defined a burn-in phase of 10 years after the start of the simulation (i.e., this
185 phase is not included in the conclusions of the analyses) to let the population settle down and start
186 with a stable system. We removed all mortalities during the first year of simulation to allow all
187 individuals to find a territory, if they can, without dying while doing so. To evaluate lynx
188 persistence, we analyzed a) population growth rates, b) establishments outside of the lynx native
189 population, and c) female territory occupancy.

190 a) We extracted the number of individuals in each population, for every year and for each
191 replicate. Then, we calculated the growth rate for each replicate and each population over the
192 simulated years as the number of individuals at time t divided by the number of individuals at
193 time $t-1$. Furthermore, we calculated the mean growth rate per year and per population, and the
194 95% confidence interval of the mean over the 200 replicates.

195 b) We extracted the number of times individuals established their territory in a population
196 area not corresponding to the one they were born in (“Population layer”, Appendix A) for every
197 year and for each replicate. We did not account for individuals which moved to another
198 population area during their dispersal but finally came back to their native population area to
199 establish, nor those which died in another population area while dispersing. We calculated the
200 cumulative sum of these movements over the simulated years for each replicate and for each

201 movement type (e.g., from the Alpine to the Jura population, from the Jura to the Vosges-
202 Palatinian population). Then, we calculated the mean and 95% confidence interval of the mean
203 across all replicates.

204 c) We extracted female territory locations at the last year of simulation for each replicate.
205 We focused on female territories as male territories are based on those of the females. We
206 assigned a “1” to each cell of the gridded study area (1 km²) that was occupied by a female
207 territory and a “0” otherwise. Finally, we estimated territory occupancy by calculating the mean
208 value per cell from these rescaled maps over the 200 replicates. We also did a visual validation of
209 our model predictions using GPS and VHF tracks from several collared female residents that we
210 overlaid on the resulting territory occupancy map (details in Appendix C).

211

212

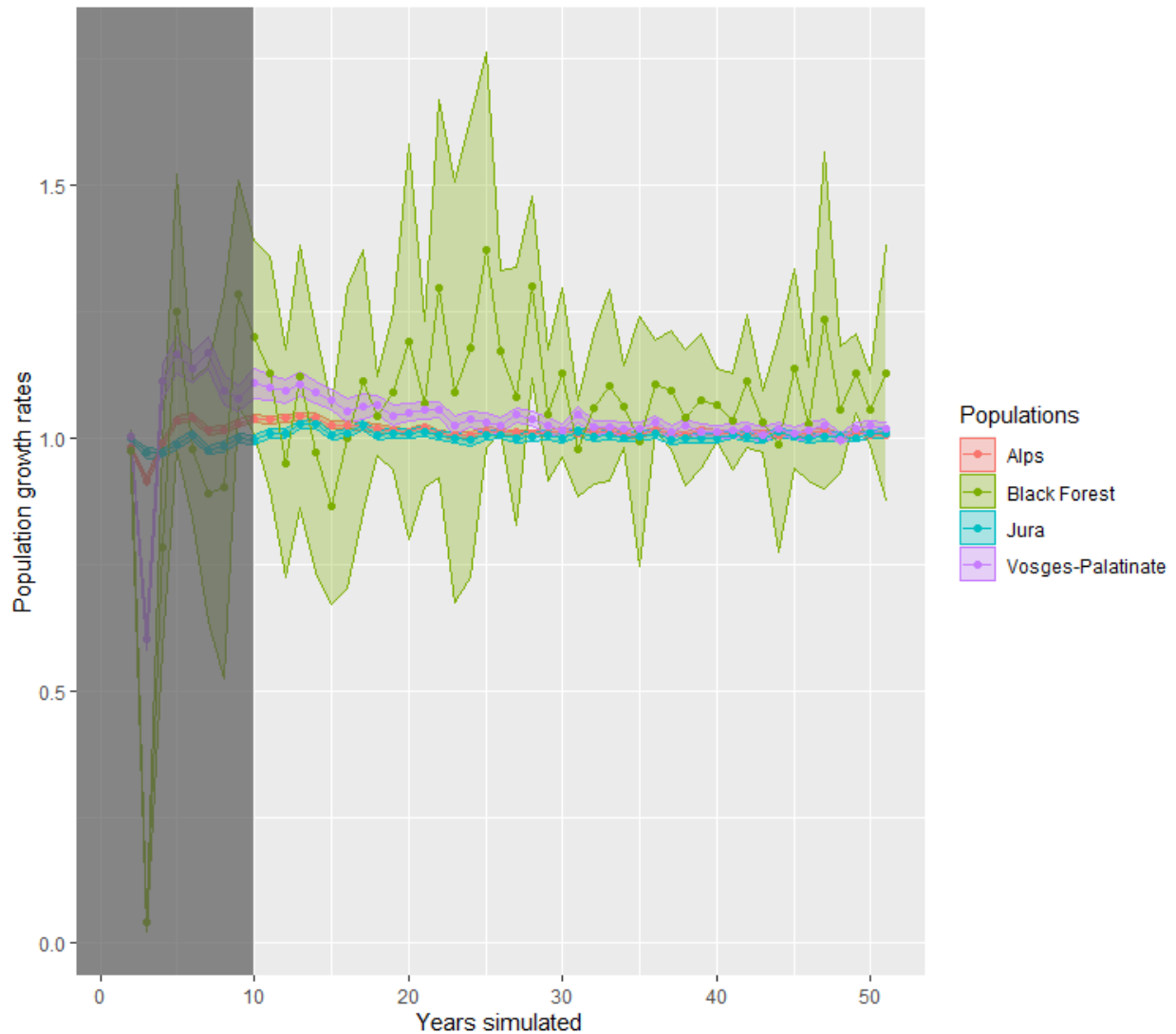
213 **Results**

214

215 *Population growth rates*

216 Simulations predicted similar growth rate patterns for the Alpine, the Jura and the Vosges-
217 Palatinian populations, with a growth rate above 1 (i.e., growing phase) slowly decreasing
218 towards reaching 1 (i.e., stabilization phase). The Alpine and Jura populations had very similar
219 patterns with their maximum growth rates equal to 1.04 (sd = 0.06) and 1.03 (sd = 0.07) at the
220 13th and 14th year of simulation, respectively. They both quickly reached a stabilization phase and
221 fluctuated a little over the final time span of the simulation. The Vosges-Palatinian population
222 had the highest growth rate, equal to 1.11 (sd = 0.21), in the 10th year of the simulation. This
223 population stabilized later than the Alpine and Jura populations. The Black Forest population had
224 a more fluctuating growth rate over the simulation. It reached a minimum of 0.87 (sd = 0.49) at

225 the 15th year of simulation and a maximum of 1.37 (sd = 0.62) at the 25th year of simulation;
226 confidence intervals around mean growth rates were large and much larger than in other
227 populations.

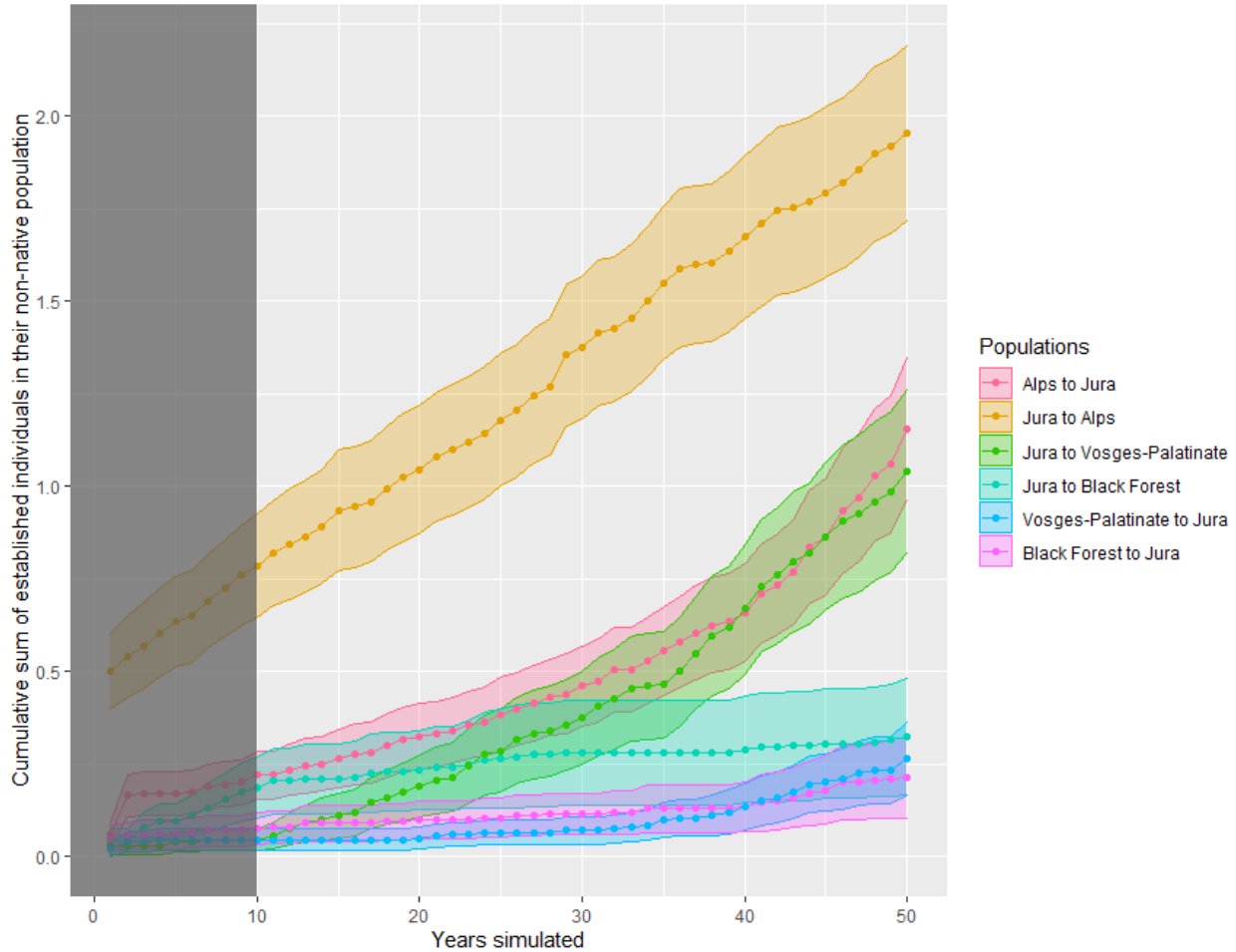


228
229 **Figure 2:** Annual rates of increase over the simulated years for each population. The grey area
230 represents the 10-year burn-in phase. Points are mean over 200 replicates and envelopes are 95%
231 confidence intervals from 200 replicates.

232

233 *Lynx establishments outside their native populations*

234 The cumulative number of lynx establishments in a population area different from their native
235 one showed that the Jura population is at the center of lynx movements between populations. The
236 most important movement was for individuals born in the Jura population going to establish
237 themselves in the Alpine population. After 50 years, the Jura population had, on average across
238 all replicates, 2.0 individuals (sd = 1.7) leaving to settle in the Alpine population over the whole
239 simulated period, 1.0 individuals (sd = 1.6) leaving to settle in the Vosges-Palatinian population,
240 and 0.3 individuals (sd = 1.1) leaving to settle in the Black Forest. The Alpine population showed
241 a movement of 1.2 individuals on average (sd = 1.4) leaving to settle in the Jura population over
242 the simulated period of 50 years. There were only a few movements from the Vosges-Palatinian
243 (0.3 individuals on average, sd = 0.7) or from the Black Forest (0.2 individuals on average, sd =
244 0.8) populations towards the Jura to establish. There was no exchange of individuals simulated
245 between the Alpine, the Vosges-Palatinian and the Black Forest populations.



246

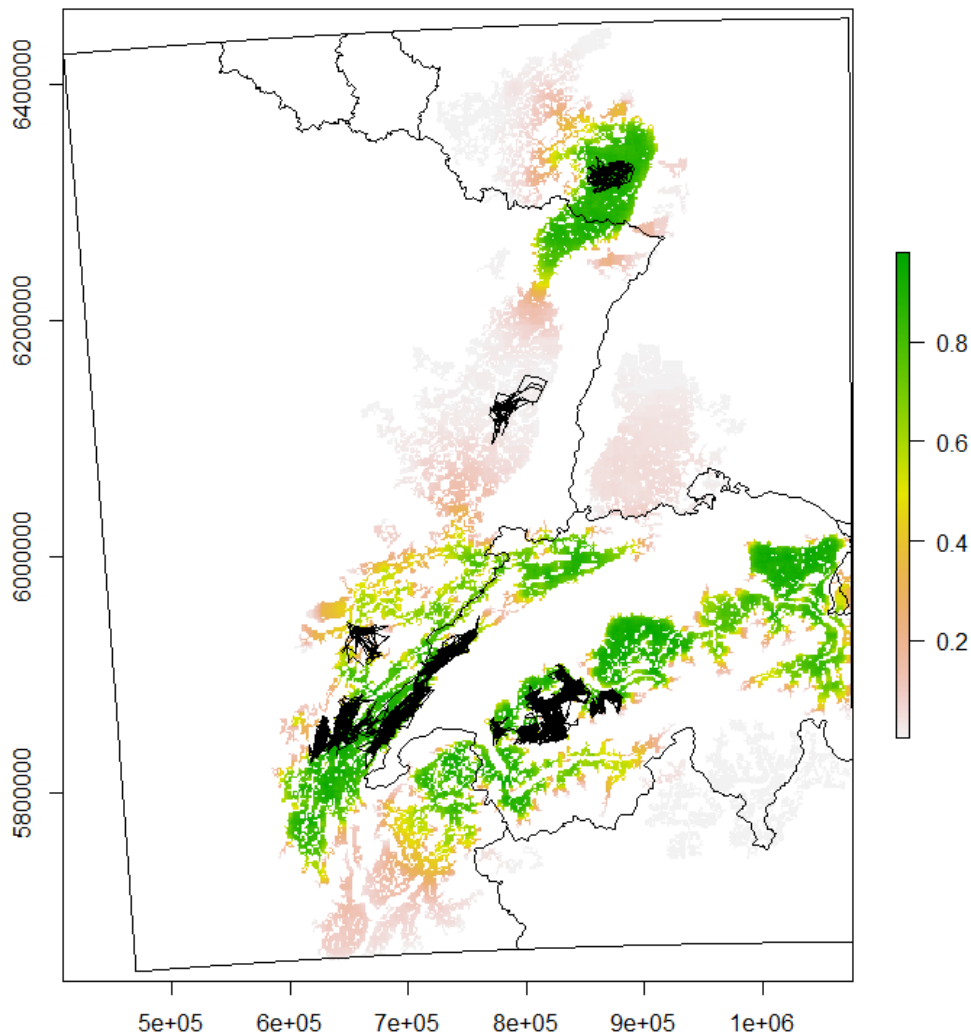
247 **Figure 3:** Cumulative number of lynx establishments outside of their native populations over 50
248 simulated years. Colors represent different directions from the native populations to those where
249 individuals established. We did not represent pairs of populations for which there were no
250 exchange between them. The grey area represents the 10-year burn-in phase. Points are mean
251 over 200 replicates and envelopes are 95% confidence intervals around the mean from 200
252 replicates of cumulative sums over years for each pair of populations.

253

254 *Territory occupancy*

255 At the end of the simulation period, female territories were predicted to cover most of the Jura
256 (mean = 0.45, sd = 0.35) and the Alps (mean = 0.60, sd = 0.27) with a very high frequency across

257 replicates. Most of the area that sustained territories had a mean territory occupancy of over 0.8
258 (maximum value for the Alpine population = 0.98, maximum value for the Jura population =
259 0.97). There were very contrasting values in predicted female territories for the Vosges-Palatinian
260 population (mean = 0.23, sd = 0.33). Female territories were established with a very high
261 frequency of occurrence across the replicates in the Palatinate Forest in Germany (maximum
262 value = 0.94), whereas territories were established less frequently in the southern part of the
263 Vosges Mountains in France. Finally, mean territory occupancy was the lowest for the Black
264 Forest population (mean = 0.03, sd = 0.02, maximum value = 0.09).



265

266 **Figure 4:** Occupancy by female territories over the study area at the last year of simulation.
267 Values between 0 and 1 are mean occupancy probability per cell of 1 km² over 200 replicates
268 (e.g., cell with value equal to 1 were always occupied in all simulation replicates). GPS and VHF
269 recorded paths for female residents are overlaid as thin black lines. A zoom of the different areas
270 with telemetry data is presented in Appendix C.

271

272

273 **Discussion**

274

275 *The well-established Alpine and Jura populations*

276 The model predicted a steady growth of Alpine and Jura populations, quickly reaching a
277 stabilization phase indicating that they may soon be at carrying capacity. However, carrying
278 capacity could be higher than what we found in our simulations, depending on the strength of
279 density-dependence processes (Zimmermann et al., 2007). Although we assumed that the number
280 of individuals an area can support is mainly defined by female territory size, our model does not
281 include a relationship between lynx behavior and density. Yet, studies found that lynx density
282 also influences home range sizes (Pesenti and Zimmermann, 2013), and differently for females
283 and males (Aronsson et al., 2016). For instance, in the North-Western Swiss Alps, lynx territories
284 were much larger in the 80s when lynx density was lower compared to the 90s (Breitenmoser-
285 Würsten et al., 2001). The Alpine and Jura areas could therefore support more individuals if
286 density-dependent mechanisms occur. On the other hand, threats that we did not include in the
287 simulations, such as low genetic diversity (concerning the Alpine population in particular;
288 Stiftung KORA, 2021) or illegal killing (Arlettaz et al., 2021), may slow the predicted growth.

289

290 *The Jura population as the crossroad of Western Europe lynx movements*

291 The Jura population was found to be the only population connected by lynx exchanges with all
292 the other populations. However, individual exchanges were estimated very low, with only a few
293 individuals moving from their native population to establish their territory in another one over the
294 50 years. In Switzerland, only a few lynx movements between the Alpine and Jura populations
295 have been observed until now despite a large camera trapping effort over the years. These two
296 populations also differ genetically, suggesting very few exchanges between them (Breitenmoser-
297 Würsten and Obexer-Ruff, 2003). When inspecting model output maps (Fig. 4), individuals are
298 contained in the Swiss Alps, restricted to an area of good habitat almost totally surrounded by
299 less favorable habitats (Appendix A, Fig. A.3). However, individuals have recently started to
300 settle permanently on the Swiss Plateau (between Alps and the Jura Mountains) predicted with
301 less favorable habitats. Some have even reproduced successfully (F. Zimmerman, pers. comm.),
302 indicating potential connections between the two populations. On the other hand, our model
303 suggests connectivity between the Jura and Alpine populations on the French side with a
304 continuity of a few forested corridors until the Chartreuse Mountains in the Alps (Zimmermann
305 and Breitenmoser, 2007). Many observations were made via camera traps of lynx moving
306 between the southern part of the Jura population and the Chartreuse Mountains (Bailly, 2021).
307 However, movement barriers (e.g., highways) may prevent connections with the rest of the
308 Alpine populations in Switzerland.

309 Lynx movements from the Jura to establish within the Vosges-Palatinian population may
310 be possible but rare (one case on average simulated over 50 years). This is coherent with field
311 monitoring that highlighted the presence of only a male in the south of the Vosges Mountains
312 coming from the Jura Mountains during the winter 2014-2015 (Chenesseau and Briaudet, 2016;
313 Hurstel and Laurent, 2016). Connectivity between these two populations remains far from

314 optimal because of multiple barriers (e.g., highways, railroads and rivers) that impede lynx
315 dispersal between the two mountain ranges and increase collision risk (Morand, 2016;
316 Zimmermann and Breitenmoser, 2007).

317 Regarding the Black Forest (including the adjoining Swabian Alb area), six males are
318 known to have immigrated from the Swiss side of the Jura population since 2013 (unpublished
319 data, Forest Research Institute Baden-Wuerttemberg; Drouet-Hoguet et al., 2021; Mitarbeitende
320 der Stiftung KORA, 2017) as well as two males from the Northeastern Alps (Herdtfelder et al.,
321 2021). Simulated lynx movements towards the Black Forest seem therefore largely under-
322 estimated in our model, probably because we only accounted for individuals who successfully
323 established. Dispersal of individuals from the Jura or the Alps towards the Black Forest area may
324 have been simulated but if individuals did not establish successfully, they were not recorded.
325 Moreover, male establishment is driven by female presence in our model, and because the Black
326 Forest population did not have any female, new males arriving could not establish and were
327 doomed to die while searching for females. A more realistic rule for male establishment would be
328 to let them search for females but still allow them to settle after a defined period of time, even
329 without females. It happens quite regularly that males show territorial behavior even without
330 females (M. Herdtfelder, pers. comm.).

331

332 *The increasing Vosges-Palatinian population*

333 Projections for the Vosges-Palatinian population showed a growth over the next 50 years.
334 However, there was a difference between the northern part of the population range (Northern
335 Vosges in France and Palatinate Forest in Germany) predicted with high resident female
336 occupancy compared to the low occupancy predicted from Central to Southern Vosges. The lynx
337 reintroduction program in the Palatinate Forest conducted since 2016 and the arrival of new

338 individuals colonizing the Northern Vosges and regions more to the south (Schwoerer 2021 ;
339 Scheid et al. 2021) are currently one of the main inputs. Lynx abundance was initially a bit higher
340 in this smaller part of the Vosges-Palatinian population (i.e., 11 individuals in the Palatinate
341 Forest compared to 8.4 individuals on average in the Vosges Mountains in France). The habitat
342 quality in the Vosges Mountains and the Palatinate Forest appears favorable for lynx
343 establishment (Appendix A, Fig. A.3). However, the lack of functional connectivity between the
344 Jura and the Vosges-Palatinian populations on one hand, and within the Vosges Mountains on the
345 other hand, may explain the difficulty for lynx to colonize the Vosges Mountains elsewhere than
346 in the Northern Vosges from the Palatinate Forest. Indeed, lynx movements between Jura and
347 Vosges-Palatinian populations are possible, but so far very scarce. In the same way, functional
348 connectivity between the Northern Vosges and the rest of the area may also be altered by
349 anthropogenic barriers, for example at the *Col de Saverne* where this forest bottleneck is
350 fragmented by both a highway and high-speed railway (Klar et al., 2006; Morand, 2016; Scheid
351 et al., 2021). However, movement is possible as lynx from the Palatinate Forest are known to
352 have crossed this pass, some even going back and forth (Idelberger et al., 2021; Scheid et al.,
353 2021). Lynx reintroduced in the Palatinate Forest could therefore reinforce the Vosges Mountains
354 part of the population if functional connectivity would allow it.

355 Because our model does not include human attitudes towards the species, we warn against
356 its blind use to assess the effect of reinforcement on lynx long-term persistence without
357 accounting for the human dimension component. Illegal killings, as they occurred after the 90's
358 reintroduction program in the southern part of Vosges Mountain (Vandel et al., 2006) as well as
359 more recent ones (Germain, 2020) may be an additional mortality that the model does not
360 account for. The extent of acceptance towards species reinforcement from some local
361 stakeholders is an important element (Charbonnel and Germain, 2020). In this case, we

362 recommend that our model be extended to include the dynamics of the whole socio-ecosystem
363 (Behr et al., 2017; Guerrero et al., 2018).

364

365 *The isolated Black Forest*

366 Growth rates were very heterogeneous along the simulated time period and confidence intervals
367 around the mean were large for this population. This is due to the demographic stochasticity
368 which impacted the Black Forest population a lot because of its small size when starting the
369 simulation compared to other populations. The model simulated only a few lynx movements from
370 the Jura population to the Black Forest, which did not contribute much to population growth, and
371 female established with a very low rate after 50 years. These results tend to show that the Black
372 Forest population is unlikely to be soon recolonized by lynx, especially considering that observed
373 individuals are only males (Drouet-Hoguet et al., 2021; Port et al., 2021). For instance, no female
374 lynx has been observed until now dispersing from Jura to Black forest crossing the Rhine valley
375 that separates the Black Forest from the Jura population, probably due to their risk-shy nature
376 compared to males (Port et al., 2021). Monitoring data from 2021 indicate that one female from
377 the very eastern Alpine population crossed this barrier for the first time (M. Herdtfelder, pers.
378 comm.). Using IMBs, Herdtfelder (2012) showed that population reinforcement with lynx
379 females might be one solution considering that habitat is of good quality in this area. Again, if
380 our model was to be used to assess this strategy, it would need to incorporate human dimensions.

381

382 *Model limitations*

383 Although model predictions passed the validation step (Appendix C), several aspects of our IBM
384 could still be improved. The habitat layer is defined in categories with strict “barriers” that the
385 simulated lynx cannot cross whereas the species may be tolerant of human activities (Basille et

386 al., 2009; Bouyer et al., 2015). Lynx may live near urban areas and cross small lakes (F.
387 Zimmermann, pers. comm.). In that context, the habitat layer could be improved by being defined
388 as a continuous variable representing a degree of preference or permeability of the landscape to
389 avoid prohibiting the movement through certain landscape elements. The habitat layer could also
390 be improved by accounting for roads and their associated structures as movement barriers (e.g.,
391 highways difficult to cross, over-pass facilitating the movement) instead of only mortality sources
392 (Klar et al., 2006; Marchand et al., 2017). In our model, lynx movement is not impeded by roads.
393 The inclusion of permeability of these linear barriers to the lynx movements may help to refine
394 the behavior rules (Marchand et al., 2017). We could also redefine the movement behavior rules
395 to be sex or age-dependent. Indeed, females seem more conservative (i.e., disperse close to their
396 natal range) compared to males and some males are able to disperse over long distances (Port et
397 al., 2021).

398

399 *Perspectives for assessing lynx conservation strategies*

400 Our model used to evaluate lynx population persistence could be applied to better understand
401 aspects of lynx conservation that we did not include yet, such as illegal killings (Heurich et al.,
402 2018). Moreover, thanks to its individual-based structure, a genetic component could also be
403 included to track relatedness between individuals and allow studying inbreeding risk and Allee
404 effects (Premier et al., 2020). Our model could also be used to test the effect of different
405 scenarios, either by modifying the lynx populations (e.g., illegal killings, reintroduction
406 programs) or the landscape (e.g., green bridges, roads construction, habitat destruction or
407 restoration) and assess the potential benefits or negative effects on lynx population persistence.
408 Modifications of road networks to improve connectivity, such as the removal of road segments or
409 the addition of overpasses (i.e., green bridges), could be tested. Then, their effect could be

410 included through a new layer of collision probabilities and the population persistence calculated
411 accordingly. Similarly, other modifications of land cover (e.g., restoring forest areas) could also
412 be tested. This model could be very useful for stakeholders working on corridors and the
413 reduction of lynx-vehicle collisions as well as reintroduction programs and species acceptance.

414

415 **Conclusion**

416 In this paper, we built and analyzed a spatially-explicit individual-based model to forecast the
417 fate of lynx populations over the next 50 years. Our results suggest that exchanges of individuals
418 between populations to establish new territories were limited, and emphasize that the Jura
419 population plays the role of a Western Europe crossroad. Overall, lynx persistence in the Upper
420 Rhine meta-population and the Alpine population over the next 50 years seems likely on a large
421 scale. However, lynx persistence in the southern part of the Vosges-Palatinian population and in
422 the Black Forest appears more challenging without long-term conservation planning like
423 improving connectivity or reintroduction programs.

424

425 **References**

- 426 Arlettaz, R., Chapron, G., Kéry, M., Klaus, E., Mettaz, S., Roder, S., Vignali, S., Zimmermann,
427 F., Braunisch, V., 2021. Poaching threatens the establishment of a lynx population,
428 highlighting the need for a centralized judiciary approach. *Front. Conserv. Sci.* 2, 665000.
429 <https://doi.org/10.3389/fcosc.2021.665000>
- 430 Aronsson, M., Low, M., López-Bao, J. V., Persson, J., Odden, J., Linnell, J.D.C., Andrén, H.,
431 2016. Intensity of space use reveals conditional sex-specific effects of prey and conspecific
432 density on home range size. *Ecol. Evol.* 6, 2957–2967.
- 433 Bailly, J., 2021. Le Lynx en Chartreuse - Épine - Mont du Chat : bilan du suivi photographique

434 2017-2020 réalisé d'après les données collectées dans le cadre du réseau Loup-Lynx de
435 l'OFB.

436 Basille, M., Herfindal, I., Santin-Janin, H., Linnell, J.D.C., Odden, J., Andersen, R., Arild Høgda,
437 K., Gaillard, J.M., 2009. What shapes Eurasian lynx distribution in human dominated
438 landscapes: selecting prey or avoiding people? *Ecography (Cop.)*. 32, 683–691.
439 <https://doi.org/10.1111/j.1600-0587.2009.05712.x>

440 Behr, D.M., Ozgul, A., Cozzi, G., 2017. Combining human acceptance and habitat suitability in a
441 unified socio-ecological suitability model: a case study of the wolf in Switzerland. *J. Appl.*
442 *Ecol.* 54, 1919–1929. <https://doi.org/10.1111/1365-2664.12880>

443 Bencin, H.L., Prange, S., Rose, C., Popescu, V.D., 2019. Roadkill and space use data predict
444 vehicle-strike hotspots and mortality rates in a recovering bobcat (*Lynx rufus*) population.
445 *Sci. Rep.* 9, 15391. <https://doi.org/10.1038/s41598-019-50931-5>

446 Bouyer, Y., Gervasi, V., Poncin, P., Beudels-Jamar, R.C., Odden, J., Linnell, J.D.C., 2015.
447 Tolerance to anthropogenic disturbance by a large carnivore: the case of Eurasian lynx in
448 south-eastern Norway. *Anim. Conserv.* 18, 271–278. <https://doi.org/10.1111/acv.12168>

449 Breitenmoser-Würsten, C., Obexer-Ruff, G., 2003. Population and conservation genetics of two
450 re-introduced lynx (*Lynx lynx*) populations in Switzerland – a molecular evaluation 30 years
451 after translocation, in: *Proceedings of the 2nd Conference on the Status and Conservation of*
452 *the Alpine Lynx Population (SCALP)*. Amden, Switzerland, pp. 28–31.

453 Breitenmoser-Würsten, C., Vandel, J.-M., Zimmermann, F., Breitenmoser, U., 2007.
454 Demography of lynx *Lynx lynx* in the Jura Mountains. *Wildlife Biol.* 13, 381–392.
455 [https://doi.org/10.2981/0909-6396\(2007\)13](https://doi.org/10.2981/0909-6396(2007)13)

456 Breitenmoser-Würsten, C., Zimmermann, F., Ryser, A., Capt, S., Laass, J., Siegenthaler, A.,
457 Breitenmoser, U., 2001. Untersuchungen zur Luchspopulation in den Nordwestalpen der

- 458 Schweiz 1997–2000, KORA Bericht.
- 459 Breitenmoser, U., Breitenmoser-Würsten, C., Capt, S., 1998. Re-introduction and present status
460 of the lynx (*Lynx lynx*) in Switzerland. *Hystrix* 10, 17–30. [https://doi.org/10.4404/hystrix-](https://doi.org/10.4404/hystrix-10.1-4118)
461 10.1-4118
- 462 Breitenmoser, U., Breitenmoser-würsten, C., Okarma, H., Kaphegyi, T., Kaphygyi, U., Müller,
463 U.M., 2000. Action Plan for the conservation of the Eurasian Lynx (*Lynx lynx*) in Europe,
464 Nature and environment.
- 465 Bull, J.K., Heurich, M., Saveljev, A.P., Schmidt, K., Fickel, J., Förster, D.W., 2016. The effect of
466 reintroductions on the genetic variability in Eurasian lynx populations: the cases of
467 Bohemian–Bavarian and Vosges–Palatinian populations. *Conserv. Genet.* 17, 1229–1234.
468 <https://doi.org/10.1007/s10592-016-0839-0>
- 469 Ceia-Hasse, A., Borda-de-Água, L., Grilo, C., Pereira, H.M., 2017. Global exposure of carnivores
470 to roads. *Glob. Ecol. Biogeogr.* 26, 592–600. <https://doi.org/10.1111/geb.12564>
- 471 Chapron, G., Kaczensky, P., Linnell, J.D.C., Von Arx, M., Huber, D., Andrén, H., López-Bao,
472 J.V., Adamec, M., Álvares, F., Anders, O., Balečiauskas, L., Balys, V., Bedő, P., Bego, F.,
473 Blanco, J.C., Breitenmoser, U., Brøseth, H., Bufka, L., Bunikyte, R., Ciucci, P., Dutsov, A.,
474 Engleder, T., Fuxjäger, C., Groff, C., Holmala, K., Hoxha, B., Iliopoulos, Y., Ionescu, O.,
475 Jeremić, J., Jerina, K., Kluth, G., Knauer, F., Kojola, I., Kos, I., Krofel, M., Kubala, J.,
476 Kunovac, S., Kusak, J., Kutal, M., Liberg, O., Majjić, A., Männil, P., Manz, R., Marboutin,
477 E., Marucco, F., Melovski, D., Mersini, K., Mertzanis, Y., Mysłajek, R.W., Nowak, S.,
478 Odden, J., Ozolins, J., Palomero, G., Paunović, M., Persson, J., Potočnik, H., Quenette,
479 P.Y., Rauer, G., Reinhardt, I., Rigg, R., Ryser, A., Salvatori, V., Skrbinšek, T., Stojanov, A.,
480 Swenson, J.E., Szemethy, L., Trajçe, A., Tsingarska-Sedefcheva, E., Váňa, M., Veeroja, R.,
481 Wabakken, P., Wölfel, M., Wölfel, S., Zimmermann, F., Zlatanova, D., Boitani, L., 2014.

- 482 Recovery of large carnivores in Europe's modern human-dominated landscapes. *Science*
483 (80-.). 346, 1517–1519. <https://doi.org/10.1126/science.1257553>
- 484 Charbonnel, A., Germain, E., 2020. Plan Régional d'Actions en faveur du Lynx boréal (Lynx
485 lynx) dans le Massif des Vosges □: rétablir le Lynx dans un état de conservation favorable
486 dans le cadre d'une démarche participative, concertée et partagée avec les acteurs du
487 territoire. *Lucy* (57), France.
- 488 Chenesseau, D., Briaudet, P.-E., 2016. Destin de lynx, trombinoscope insolite de lynx identifiés
489 par piégeage photographique □: Bingo □! Du massif jurassien aux Vosges, il n'y a qu'un
490 pas... de lynx. *Bull. Lynx du Réseau* 20, 9–10.
- 491 Drouet-Hoguet, N., Chenesseau, D., Kunz, F., Zimmermann, F., 2021. Situation of the lynx in the
492 Jura Mountains. *Cat News Spec. Issue* 14, 29–34.
- 493 Germain, E., 2020. Lettre d'information n°1 du PRA en faveur du Lynx dans le Massif des
494 Vosges. Janvier-Juin 2020. *Lucy* (57), France.
- 495 Germain, E., Schwoerer, M.-L., 2021. Situation of the Eurasian lynx in the Vosges Mountains.
496 *Cat News Spec. Issue* 14, 34–37.
- 497 Grimm, V., Berger, U., Bastiansen, F., Eliassen, S., Ginot, V., Giske, J., Goss-Custard, J., Grand,
498 T., Heinz, S.K., Huse, G., Huth, A., Jepsen, J.U., Jørgensen, C., Mooij, W.M., Müller, B.,
499 Pe'er, G., Piou, C., Railsback, S.F., Robbins, A.M., Robbins, M.M., Rossmanith, E., Rüger,
500 N., Strand, E., Souissi, S., Stillman, R.A., Vabø, R., Visser, U., DeAngelis, D.L., 2006. A
501 standard protocol for describing individual-based and agent-based models. *Ecol. Modell.*
502 198, 115–126. <https://doi.org/10.1016/j.ecolmodel.2006.04.023>
- 503 Grimm, V., Berger, U., DeAngelis, D.L., Polhill, J.G., Giske, J., Railsback, S.F., 2010. The ODD
504 protocol: A review and first update. *Ecol. Modell.* 221, 2760–2768.
505 <https://doi.org/10.1016/j.ecolmodel.2010.08.019>

- 506 Guerrero, A.M., Bennett, N.J., Wilson, K.A., Carter, N., Gill, D., Mills, M., Ives, C.D., Selinske,
507 M.J., Larrosa, C., Bekessy, S., Januchowski-Hartley, F.A., Travers, H., Wyborn, C.A.,
508 Nuno, A., 2018. Achieving the promise of integration in social-ecological research: a review
509 and prospectus. *Ecol. Soc.* 23, 38. <https://doi.org/10.5751/ES-10232-230338>
- 510 Herdtfelder, M., Schraml, U., Suchant, R., 2021. Steps towards a lynx population in the Black
511 Forest? *Cat News Spec. Issue* 14, 45–46.
- 512 Heurich, M., Schultze-naumburg, J., Piacenza, N., Magg, N., Čeverny, J., Engleder, T.,
513 Herdtfelder, M., Sladova, M., Kramer-Schadt, S., 2018. Illegal hunting as a major driver of
514 the source-sink dynamics of a reintroduced lynx population in Central Europe. *Biol.*
515 *Conserv.* 224, 355–365. <https://doi.org/S0006320717314003>
- 516 Hradsky, B.A., Kelly, L.T., Robley, A., Wintle, B.A., 2019. FoxNet: An individual-based model
517 framework to support management of an invasive predator, the red fox. *J. Appl. Ecol.* 56,
518 1460–1470. <https://doi.org/10.1111/1365-2664.13374>
- 519 Hurstel, A., Laurent, A., 2016. Première preuve de dispersion du lynx d’Eurasie (*Lynx lynx*) du
520 Jura vers les Vosges. *Ciconia* 40, 1–6.
- 521 Idelberger, S., Krebühl, J., Back, M., Ohm, J., Prüssing, A., Sandrini, J., Huckschlag, D., 2021.
522 Reintroduction of Lynx in the Palatinate Forest Biosphere Reserve, Germany. *Cat News*
523 *Spec. Issue* 14, 38–42.
- 524 Klar, N., Hermann, M., Kramer-Schadt, S., 2006. Effects of roads on a founder population of
525 lynx in the biosphere reserve “Pfalzerwald-Vosges du Nord.” *Naturschutz und*
526 *Landschaftsplan.* 38, 330–337.
- 527 Kramer-Schadt, S., Kaiser, T.S., Frank, K., Wiegand, T., 2011. Analyzing the effect of stepping
528 stones on target patch colonisation in structured landscapes for Eurasian lynx. *Landsc. Ecol.*
529 26, 501–513. <https://doi.org/10.1007/s10980-011-9576-4>

- 530 Kramer-Schadt, S., Revilla, E., Wiegand, T., 2005. Lynx reintroductions in fragmented
531 landscapes of Germany: Projects with a future or misunderstood wildlife conservation? *Biol.*
532 *Conserv.* 125, 169–182. <https://doi.org/10.1016/j.biocon.2005.02.015>
- 533 Kramer-Schadt, S., Revilla, E., Wiegand, T., Breitenmoser, U., 2004. Fragmented landscapes,
534 road mortality and patch connectivity: modelling influences on the dispersal of Eurasian
535 lynx. *J. Appl. Ecol.* 41, 711–723.
- 536 Marchand, P., Garel, M., Bourgoïn, G., Duparc, A., Dubray, D., Maillard, D., Loison, A., 2017.
537 Combining familiarity and landscape features helps break down the barriers between
538 movements and home ranges in a non-territorial large herbivore. *J. Anim. Ecol.* 86, 371–
539 383. <https://doi.org/10.1111/1365-2656.12616>
- 540 Marucco, F., McIntire, E.J.B., 2010. Predicting spatio-temporal recolonization of large carnivore
541 populations and livestock depredation risk: wolves in the Italian Alps. *J. Appl. Ecol.* 47,
542 789–798. <https://doi.org/10.1111/j.1365-2664.2010.01831.x>
- 543 Mitarbeitende der Stiftung KORA, 2017. KORA Jahresbericht 2017. Muri.
- 544 MLR, 2019. Wildtierbericht für Baden-Württemberg 2018. Stuttgart.
- 545 Molinari-Jobin, A., Breitenmoser, U., Breitenmoser-Würsten, C., Cerne, R., Drouet-Hoguet, N.,
546 Fuxjäger, C., Kos, I., Krofel, M., Marucco, F., Molinari, P., Nägele, O., Rauer, G., Sindjic,
547 M., Trbojevic, I., Wölfl, M., Wölfl, S., Zimmermann, F., 2021. SCALP: Monitoring the
548 Eurasian lynx in the Alps and beyond. *Cat News Spec. Issue* 14, 50–52.
- 549 Morand, A., 2016. Le Lynx □: risques routiers et mesures correctrices – état des lieux et
550 recommandations.
- 551 Ovenden, T.S., Palmer, S.C.F., Travis, J.M.J., Healey, J.R., 2019. Improving reintroduction
552 success in large carnivores through individual-based modelling: How to reintroduce
553 Eurasian lynx (*Lynx lynx*) to Scotland. *Biol. Conserv.* 234, 140–153.

- 554 <https://doi.org/10.1016/j.biocon.2019.03.035>
- 555 Pesenti, E., Zimmermann, F., 2013. Density estimations of the Eurasian lynx (*Lynx lynx*) in the
556 Swiss Alps. *J. Mammal.* 94, 73–81. <https://doi.org/10.1644/11-MAMM-A-322.1>
- 557 Philips, I., 2020. An agent based model to estimate lynx dispersal if re-introduced to Scotland.
558 *Appl. Spat. Anal. Policy* 13, 161–185. <https://doi.org/10.1007/s12061-019-09297-4>
- 559 Port, M., Henkelmann, A., Schröder, F., Waltert, M., Middelhoff, L., Anders, O., Jokisch, S.,
560 2021. Rise and fall of a Eurasian lynx (*Lynx lynx*) stepping-stone population in central
561 Germany. *Mammal Res.* 66, 45–55. <https://doi.org/10.1007/s13364-020-00527-6>
- 562 Premier, J., Fickel, J., Heurich, M., Kramer-Schadt, S., 2020. The boon and bane of boldness:
563 movement syndrome as saviour and sink for population genetic diversity. *Mov. Ecol.* 8, 1–
564 17. <https://doi.org/10.1186/s40462-020-00204-y>
- 565 Railsback, S.F., Grimm, V., 2012. Agent-based and individual-based modeling: a practical
566 introduction. Princeton University Press, Princeton, NJ.
- 567 Schadt, S., Revilla, E., Wiegand, T., Knauer, F., Kaczensky, P., Breitenmoser, U., Bufka, L.,
568 Cerveny, J., Koubek, P., Huber, T., Stanisa, C., Trepl, L., 2002. Assessing the suitability of
569 central European landscapes for the reintroduction of Eurasian lynx. *J. Appl. Ecol.* 39, 189–
570 203.
- 571 Scheid, C., Germain, E., Schwoerer, M.-L., 2021. Les Lynx (*Lynx lynx*) du Pfälzerwald
572 s’installent progressivement dans le Massif des Vosges. *Ann. Sci. Rés. Bios. Trans. Vosges*
573 *du Nord.* 20, 104–125.
- 574 Schwoerer, M.-L., 2021. Axe 3□: Suivi et conservation du Lynx, Action 10 / Bilan 2019-2020 du
575 suivi conduit par le Réseau Loup-Lynx. PRA Lynx Massif des Vosges, Lett. d’information
576 n2 7–14.
- 577 Stiftung KORA, 2021. 50 Jahre Luchs in der Schweiz.

- 578 Vandel, J.-M., Stahl, P., 2005. Distribution trend of the Eurasian lynx *Lynx lynx* populations in
579 France. *Mammalia* 69, 145–158. <https://doi.org/10.1515/mamm.2005.013>
- 580 Vandel, J.-M., Stahl, P., Herrenschmidt, V., Marboutin, E., 2006. Reintroduction of the lynx into
581 the Vosges mountain massif: From animal survival and movements to population
582 development. *Biol. Conserv.* 131, 370–385. <https://doi.org/10.1016/j.biocon.2006.02.012>
- 583 von Arx, M., 2020. *Lynx lynx* (amended version of 2018 assessment).
- 584 Wölfel, S., Anders, O., Middelhoff, T.L., Hohmann, U., Back, M., Idelberger, S., Krebühl, J.,
585 Ohm, J., Prüssing, A., Herdtfelder, M., Böcker, F., Erretkamps, J., Kopaniak, L., Wölfel, M.,
586 Jokisch, S., Hucht-Ciorga, I., Teubner, J., Trost, M., Zschille, J., Jeß, E., Steinberg, C., 2021.
587 Status des Luchses in Deutschland. *Natur und Landschaft* 96.
- 588 Zimmermann, F., Breitenmoser-Würsten, C., Breitenmoser, U., 2007. Importance of dispersal for
589 the expansion of a Eurasian lynx *Lynx lynx* population in a fragmented landscape. *Oryx* 41,
590 358–368. <https://doi.org/10.1017/S0030605307000712>
- 591 Zimmermann, F., Breitenmoser, U., 2007. Potential distribution and population size of the
592 Eurasian lynx *Lynx lynx* in the jura Mountains and possible corridors to adjacent ranges.
593 *Wildlife Biol.* 13, 406–416. [https://doi.org/10.2981/0909-](https://doi.org/10.2981/0909-6396(2007)13[406:PDAPSO]2.0.CO;2)
594 [6396\(2007\)13\[406:PDAPSO\]2.0.CO;2](https://doi.org/10.2981/0909-6396(2007)13[406:PDAPSO]2.0.CO;2)

595

596

597 **Acknowledgements**

598 We thank all the volunteers from the “Réseau Loup-Lynx” who collected data on the field. SB
599 and OG were funded by French National Research Agency (ANR-16-CE02-0007). SB was
600 funded as well by OFB and OG was funded by CNRS and the “Mission pour l’Interdisciplinarité”
601 through the “Osez l’Interdisciplinarité” initiative. CEFÉ, Cerema and CROC were funded by

602 CIL&B, MTES (ITTECOP) and FRB through the research program ERC-Lynx. Cerema received
603 support from the “ Direction générale des infrastructures de transports et de la mer (Ministère de
604 la transition écologique) ”. CROC was funded in 2019/2020 by the European Union within the
605 framework of the Operational Program FEDER-FSE “Lorraine et Massif des Vosges 2014–
606 2020”, the “Commissariat à l’Aménagement du Massif des Vosges” for the FNADT (“Fonds
607 National d’Aménagement et de Développement du Territoire”), the DREAL Grand Est
608 (“Direction Régionale pour l’Environnement, l’Aménagement et le Logement”), the “Région
609 Grand Est” and the “Fondation d’entreprise UEM”.

610

611

612 **Appendix A – Detail of the lynx spatially-explicit individual-based model**

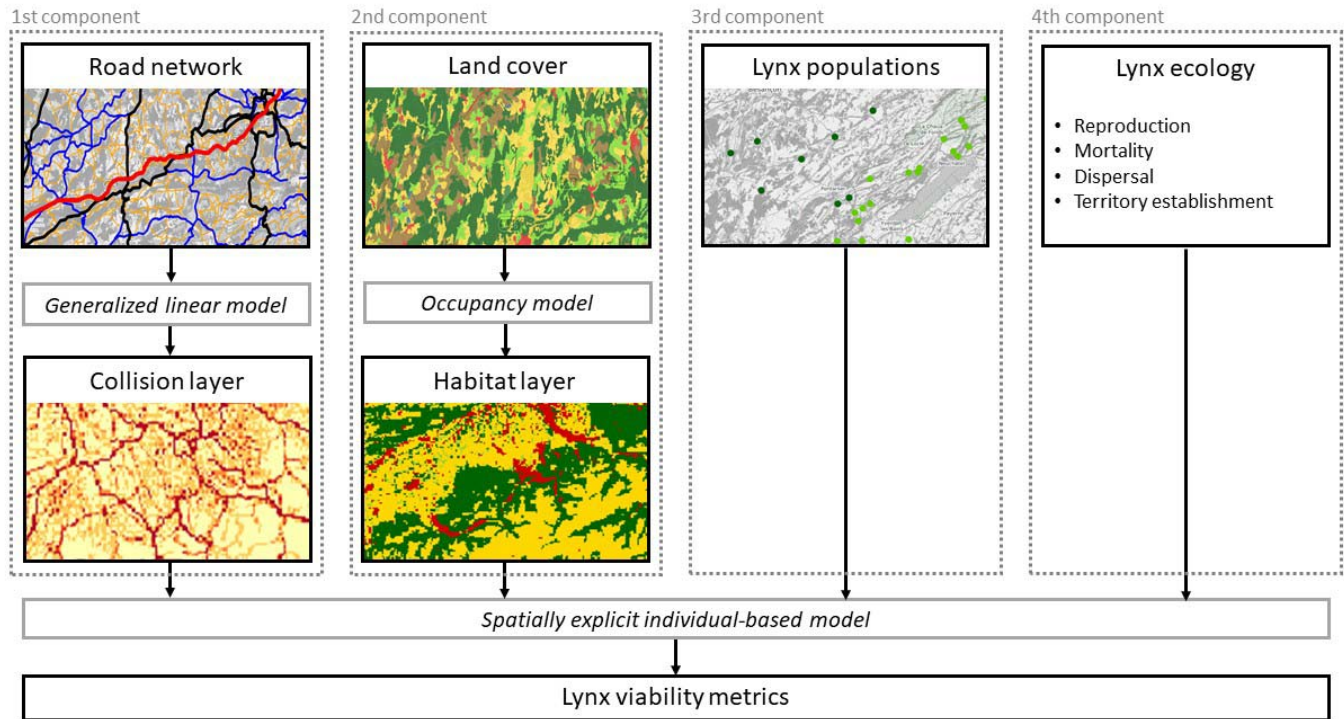
613

614 A complete description of the model following the Overview, Design concepts, and Details
615 (ODD) protocol (Grimm et al., 2010, 2006) is provided in Appendix B.

616

617 Our lynx spatially-explicit individual-based model is made of four components (Fig. A.1). The
618 first component represents the impact of road network on lynx survival via predicted collision
619 probabilities. The second component represents the impact of land cover on lynx space use with
620 the definition of different lynx habitat types. These first two components are spatial layers
621 influencing the behavioral rules followed by simulated lynx individuals. The third component
622 represents the initial lynx populations made of lynx individual's location and characteristics used
623 to launch the SE-IBM. The fourth component details all SE-IBM rules including lynx
624 demography, dispersal movement and establishment.

625



626

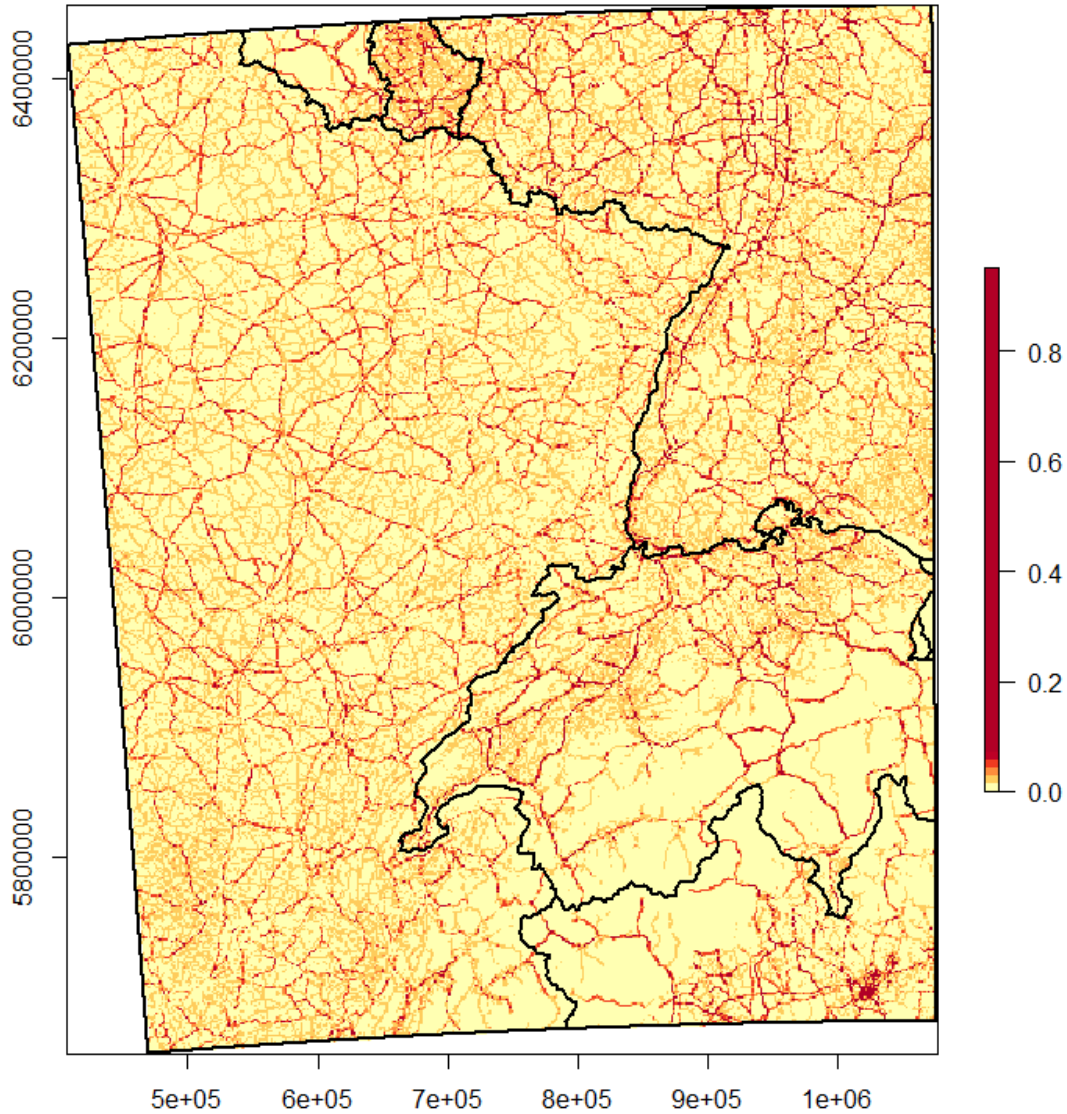
627 **Figure A.1:** The four components of the lynx SE-IBM. The first component represents the
628 impact of road network on lynx survival through vehicle collision. A generalized linear model
629 predicted collision probabilities (“Collision layer”, Fig. A.2). The second component represents
630 the impact of land cover on the lynx populations. A site-occupancy model predicted lynx habitats
631 influencing lynx movement (“Habitat layer”, Fig. A.3). The third component represents lynx
632 populations with individuals’ locations and characteristics (Fig. A.4). The fourth component
633 encompasses all ecological rules and SE-IBM parameters. All four components are included in
634 the SE-IBM to simulate lynx populations and assess their persistence with different metrics.

635

636 *Impact of road network (SE-IBM 1st component)*

637 We first built a risk model to predict collision probabilities between lynx and vehicles within the
638 1 km² resolution gridded study area. We used lynx mortality events recorded by the wolf-lynx
639 monitoring framework implemented in France (Duchamp et al., 2012), Réseau Loup-lynx

640 <https://www.loupfrance.fr/suivi-du-loup/reseau-loup-lynx/>) to train a logistic regression
641 explaining lynx collisions using lynx presence and both road and environmental characteristics
642 (Visintin et al., 2018, 2017). We used the IGN route500[©] for France (IGN ROUTE 500, 2018)
643 and OpenStreetMap[©] for the other countries (Geofabrik OpenStreetMap, 2014) as data sources to
644 extract total road length per cell and the type of road of the longest road segment in each cell. We
645 classified the road segments as “highways” (i.e., “Type autoroutier” in route500[©] data,
646 “motorway”, “motorway_link”, “trunk” and “trunk_link” in OpenStreetMap data), “main road”
647 (i.e., “Liaison principale” in route500[©] data, “primary” and “primary_link” in OpenStreetMap
648 data), “secondary road” (i.e., “Liaison régionale” in route500[©] data, and “secondary” and
649 “secondary_link” in OpenStreetMap data) and “local road” (i.e., “Liaison locale” in route500[©]
650 data, “tertiary”, “tertiary_link” and “unclassified” in OpenStreetMap data). For environmental
651 characteristics, we used Corine Land Cover[©] at 100 m of resolution for all Europe (Copernicus,
652 2012) to calculate the proportion of urban area in each cell (i.e., human presence). The best
653 model identified included total road length, type of road of the longest road segment and the
654 proportion of urban area. Distance to highways and human density were also tested to explain
655 lynx-vehicles collisions but were not significant. We used this model to predict a collision
656 probability between 0 and 1 in each cell of our study area to create the “Collision layer” (Fig.
657 A.2). Lynx presence was also included as explanatory variable in the model. To predict collision
658 probabilities, we defined the lynx as present everywhere on the gridded study area because
659 simulated individuals in the SE-IBM suffer from collision probability on the cells they are
660 located. Study area grid cells without road intersecting them have a zero collision probability.
661

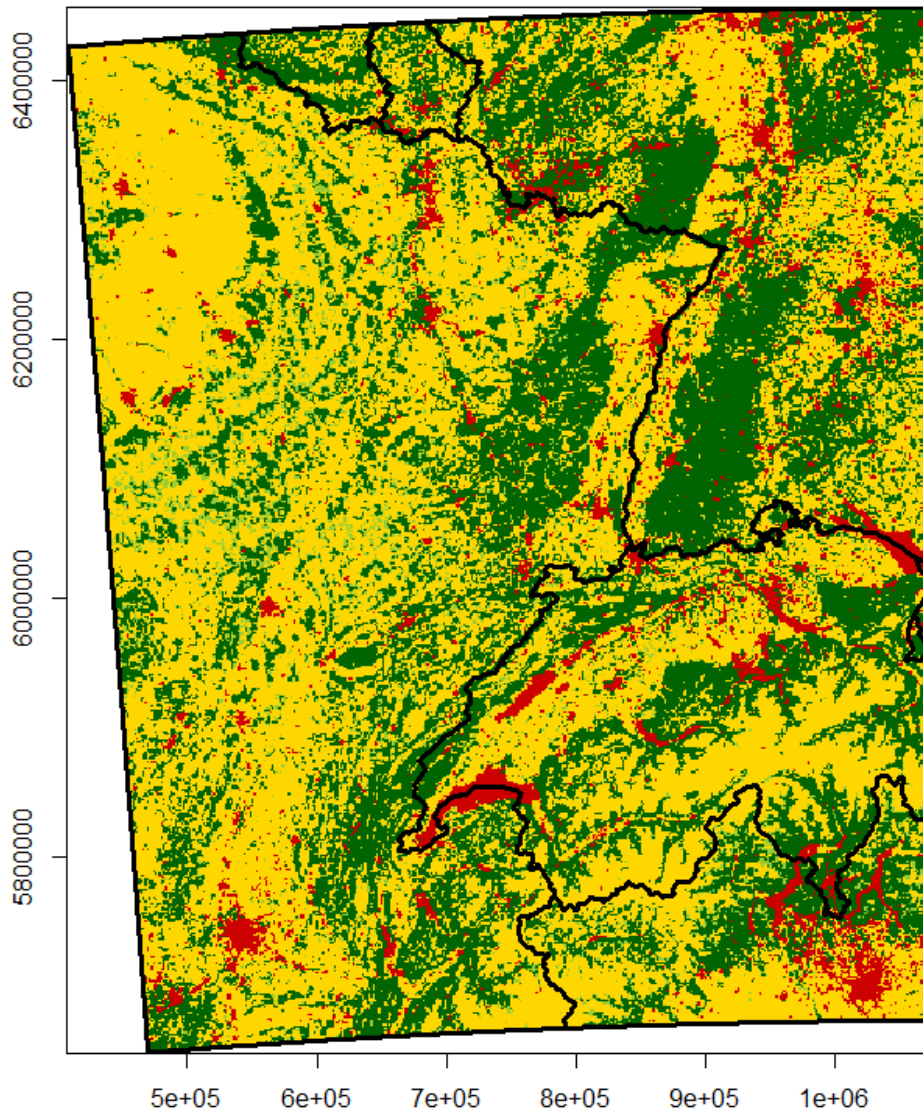


662
663 **Figure A.2:** “Collision layer” with collision probabilities between lynx and vehicles estimated
664 between 0 and 1 (yellow to red color scale). Collision probabilities are assigned to each cell of
665 the gridded landscape over the whole study area (black rectangle) used in the lynx SE-IBM using
666 the collision model. Limits of the countries (main text, Fig. 1) are overlaid in black over the map.

667
668 *Impact of land cover (SE-IBM 2nd component)*
669 We built a habitat model to define a habitat type for each cell within the 1 km² resolution gridded
670 study area. We used a multi-year occupancy model (Isaac et al., 2014; Outhwaite et al., 2018) to

671 explain regular lynx presence using land cover types important to lynx (agricultural fields, forest
672 and open land), the distance to highways (Basille et al., 2013), and human density. The model
673 was calibrated using French data for the lynx presence from 1994 to 2017 (data: Réseau Loup-
674 Lynx <https://www.loupfrance.fr/suivi-du-loup/reseau-loup-lynx/>) and then predicted over our
675 entire study area. We used used Corine Land Cover[®] at 100 m of resolution for all Europe
676 (Copernicus, 2012) to calculate the proportion of agricultural and cultivated fields, forest, pasture
677 and open land in each cell. We used the IGN route500[®] for France (IGN ROUTE 500, 2018) and
678 OpenStreetMap[®] for the other countries (Geofabrik OpenStreetMap, 2014) as data sources to
679 calculate the distance from each cell center to the nearest highway. We also used IGN[®] data for
680 France (IGN ADMIN-EXPRESS-COG, 2018) and NASA[®] data for the other countries (CIESIN
681 GPWv4, 2015) to calculate the mean human density per cell. The best model identified to explain
682 lynx presence included presence of agricultural fields, forest, and open land, distance to highways
683 and human density. Shrub cover and road length were also tested to explain lynx presence but
684 were not significant. We used this model to predict lynx occupancy probability in our entire
685 gridded study area for the year 2017. We then categorized the cells to obtain a habitat layer with
686 the four habitat types defined by the habitat categorization of Kramer-Schadt et al. (2004):
687 “breeding”, “dispersal”, “matrix” and “barrier” (Fig. A.3). All cells where lynx occupancy was
688 predicted as non-null (above or equal a threshold of 0.01) were considered “breeding” habitat. All
689 forested areas not already “breeding” habitat were considered as “dispersal” habitat. Cells with
690 more than half of their surface covered by water or urban area defined by Corine Land Cover[®]
691 (Copernicus, 2012) were considered “barrier” for lynx movements. The rest was considered
692 “matrix” area (i.e., habitat not favorable for lynx but that can be traversed by dispersers). The
693 map of categorized cells for our study area defined the “Habitat layer” used in the SE-IBM (Fig.
694 A.3).

695



696

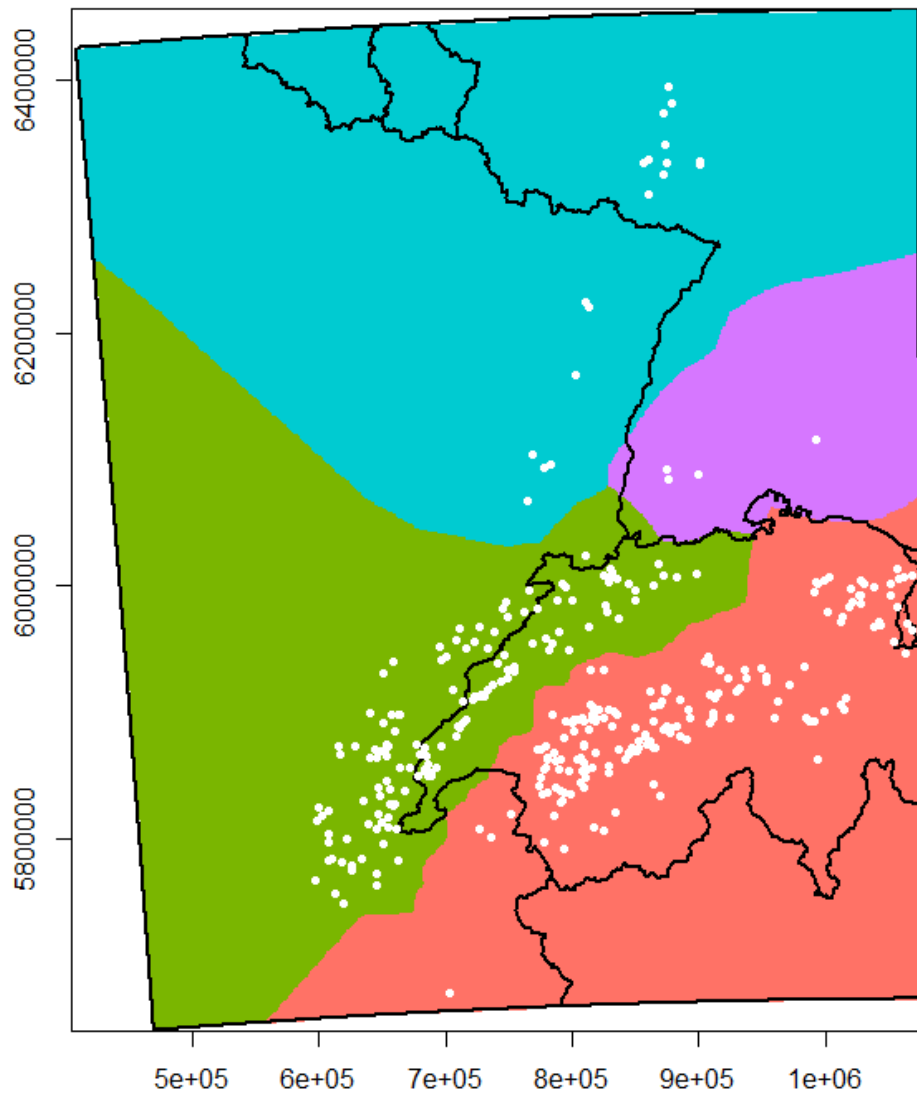
697 **Figure A.3:** “Habitat layer” with lynx habitat types as breeding habitat (dark green), dispersal
698 habitat (light green), matrix (yellow) and barrier (red). Habitat types are assigned to each cell of
699 the gridded landscape over the whole study area (black rectangle) used in the lynx SE-IBM using
700 the habitat model. Limits of the countries (main text, Fig. 1) are overlaid in black over the map.

701

702 *Lynx initial populations (SE-IBM 3rd component)*

703 The initial populations to launch the SE-IBM were built with generated locations and
704 characteristics for the lynx populations using the best data available at the time of the study (Fig.
705 A.4). We used cells of regular lynx presence for France (for the time period of 01/04/2013 to
706 31/03/2017 from the Réseau Loup-Lynx, main text, Fig. 1), Switzerland (for the time period of
707 01/01/2015 to 31/12/2017 from the KORA, main text, Fig. 1) and Germany (for the time period
708 of 01/05/2017 to 30/04/2018 from the Bundesamt für Naturschutz, main text, Fig. 1). In these
709 cells, we extracted “breeding” and “dispersal” areas and sampled random lynx locations in these
710 areas to locate individuals within the different populations. In France, we used the most reliable
711 lynx density estimate (1.14 lynx per 100 km²; Gatti et al., 2014) regardless of differences in local
712 densities (Gimenez et al., 2019) to convert the area of regular presence into a number of
713 individuals to create. Ninety-two individuals were generated and dispatched over “breeding” and
714 “dispersing” areas in France: in the Vosges Mountains (500 km² of presence), in the French Jura
715 (7,700 km² of presence) and in the French Alps (500 km² of presence). A density of 1.14 lynx per
716 100 km² may be an over- or an under-estimate of the lynx density in certain areas but in the
717 absence of local density for each French population, we used this mean value. On average, this
718 method led to 4.5 individuals in the French Alps (sd = 2.1), 77.0 individuals in the French Jura
719 (sd = 3.4), and 8.4 individuals in the Vosges Mountains (sd = 2.7). In Switzerland, we generated
720 230 individuals distributed in the different areas of presence according to estimated local
721 population sizes in 2017 (data: F. Zimmermann, pers. comm.). In Germany, 11 lynx (i.e.,
722 reintroduced lynx still alive and their offspring) were identified in the Palatinate at the end of
723 April 2018 (Scheid et al., 2021) and four male lynx were identified in the Black Forest (Wölfel et
724 al., 2021). Therefore, we generated 11 lynx in Palatinate Forest and 4 lynx in the Black Forest
725 area.

726 Except for reintroduced individuals for which we knew their characteristics, we randomly
727 assigned individuals' sex (male or female) according to ratio 1:1 usually observed at birth
728 (Breitenmoser et al., 1993; Jedrzejewski et al., 1996). We also randomly assigned an age between
729 2 and 15 as lynx usually live until 15-17 years old (e.g., Breitenmoser-Würsten et al., 2007a).
730 Age is defined as 0 for the first year of life, 1 during the second year of life, etc. For the
731 Palatinate part of the Vosges-Palatinian population, we assigned known age and sex (5 females
732 and 6 males). In the Vosges Mountains, only males were detected in 2017 thanks to the camera
733 trap survey design implemented (Charbonnel and Germain, 2020) and no case of reproduction
734 was reported during the years preceding our analyzes. We then only defined males for the
735 individuals located in the Vosges part of the Vosges-Palatinian population, with ages randomly
736 generated. We also only defined males for the Black Forest population. All generated individuals
737 were defined as “disperser” to avoid the bias of defining territories by ourselves. They find their
738 territories on their own as defined by the SE-IBM rules.
739



740
741 **Figure A.4:** “Population layer” with the example of an initial population over the whole study
742 area (black rectangle) used in the lynx SE-IBM using the habitat model. Limits of the countries
743 (main text, Fig. 1) are overlaid in black over the map. Areas for the Vosges-Palatinian (blue),
744 Black Forest (purple), Jura (green) and Alpine (red) populations were defined using the cells of
745 lynx regular presence as in 2017-2018. Each cell of the gridded study area was assigned the
746 population from its closest cell of lynx regular presence. White dots represent one initial
747 population (i.e., simulated lynx released at the beginning of one simulation replicate) generated
748 using the cells of lynx regular presence, the density or number of lynx in each population and the

749 habitat layer to place these theoretical individuals in breeding or dispersal habitats. A different
750 initial population was used at each simulation replicate; we used the same cells of lynx regular
751 presence, the same density or number of lynx in each population and the same habitat layer, only
752 the generated locations, and chosen sex and age for unknown individuals, were different.

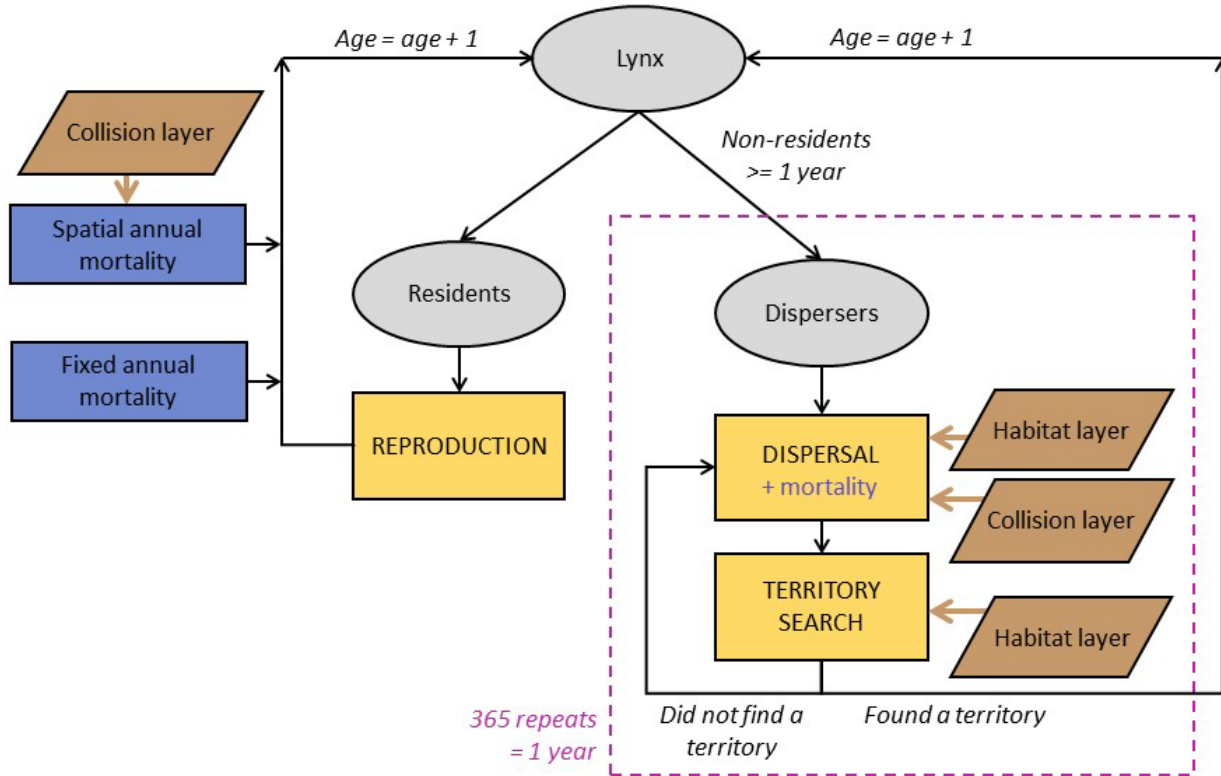
753

754 *SE-IBM rules according to lynx ecology (SE-IBM 4th component)*

755 Lynx individuals are simulated over a landscape represented as a grid of 1 km² resolution
756 covering the whole study area and encompassing the four lynx populations (main text, Fig. 1).
757 The gridded study area resolution corresponds to lynx's perceptual range (Haller and
758 Breitenmoser, 1986) as well as to the resolution of previous lynx IBMs (Kramer-Schadt et al.,
759 2011, 2005, 2004). Two variables characterize the gridded study area: a probability of lynx-
760 vehicle collision between 0 and 1 ("Collision layer", Fig. A.2) and a habitat type among
761 "breeding", "dispersal", "matrix", and "barrier" ("Habitat layer", Fig. A.3). Simulated individuals
762 are characterized by their "disperser" (i.e., not established on a territory and in search of one) or
763 "resident" status (i.e., established on a defined territory), their age and their sex.

764 Simulated resident individuals follow rules on a yearly time step (Fig. A.5). They do not
765 move (i.e., their movement inside their territory is not simulated), they hold a territory and they
766 may reproduce once a year. They suffer two types of annual mortality: a fixed baseline mortality
767 and a spatial one derived from the collision layer. Simulated dispersing lynx follow rules on a
768 daily time step (Fig. A.5). They do not have a territory yet and move every day along the gridded
769 study area, searching for a place to establish themselves. Their dispersal movement (Fig. A.6)
770 and search for a territory (Fig. A.7) are driven by the habitat layer. At each step individuals can
771 die from the spatial mortality derived from the collision layer, and daily from a fixed baseline
772 mortality.

773



774

775 **Figure A.5:** Diagram of the main structure of the SE-IBM with the main events affecting both
 776 resident and dispersing simulated lynx. Grey circles represent individuals. Yellow boxes
 777 represent SE-IBM main events detailed in model description. Blue boxes and blue writing
 778 represent different mortality causes individuals suffer. Brown boxes represent environmental
 779 layers used in the SE-IBM and brown arrows point where behavioral rules are constrained by
 780 these layers. Black arrows indicate the flow of the model. Dispersers can go through dispersal
 781 and territory search every day during a year, as long as they do not find a territory (Figs. A.6 and
 782 A.7).

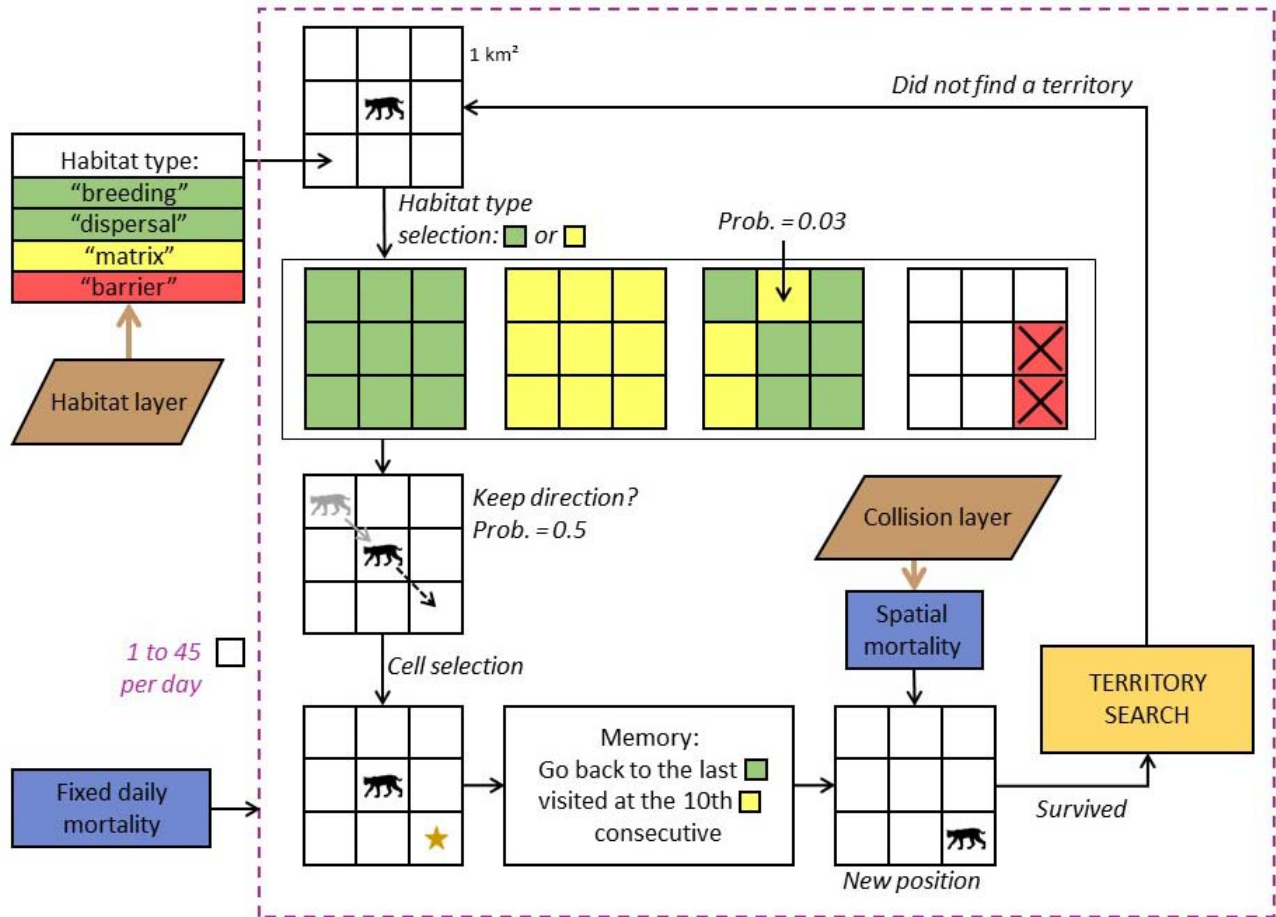
783

784 Lynx are solitary carnivores (i.e., each resident holds its own territory), except during
 785 reproduction (i.e., mating between male and female) and when females are raising their kittens

786 (Breitenmoser et al., 1993; Stahl and Vandel, 1998). Female residents reproduce with a
787 probability of 0.81 if a resident male occurs on their territory (Breitenmoser-Würsten et al.,
788 2007a; López-Bao et al., 2019). Females have 0.5 chance to be sexually mature at 1 year old, and
789 they are all sexually mature at 2 years old. Males have 0.5 chance to be sexually mature at 2 years
790 old, all being sexually mature at 3 years old (Breitenmoser-Würsten et al., 2007a; Kramer-Schadt
791 et al., 2005; Kvam, 1991). In the wild, lynx litter size averages two kittens and can be up to four
792 (Breitenmoser-Würsten et al. 2001; Lopez-Bao et al. 2019). However, around 50% of lynx kittens
793 die before reaching the age of dispersal (Breitenmoser-Würsten et al., 2007a, 2001). We
794 simulated that resident females up to 11 years old (“young” female in the SE-IBM model)
795 produce 1 or 2 kittens, with a probability of 0.5 for each litter size, that will survive until
796 becoming dispersers the following year (Henriksen et al., 2005; Kramer-Schadt et al., 2005).
797 Senescence reduces litter size, therefore “old” females in the model (12 years old and older)
798 produce 0 or 1 kitten with an equal probability of 0.5 (Henriksen et al., 2005; Kramer-Schadt et
799 al., 2005). We defined that residents die from a fixed annual mortality (i.e., baseline mortality)
800 with a probability equal to 0.1 (Breitenmoser-Würsten et al., 2007a; Heurich et al., 2018;
801 Kramer-Schadt et al., 2005), which does not include the mortality due to vehicle collisions (i.e.,
802 spatial mortality added separately with the collision model) and illegal killing (i.e., unavailable
803 estimates). We did not define an increase of mortality due to senescence but we set an age
804 maximum of 20 years (Breitenmoser-Würsten et al., 2007a; Stahl and Vandel, 1998; von Arx et
805 al., 2017). Residents can also die from a spatial annual mortality due to vehicle collisions inside
806 their territory. This spatial mortality is specific to each resident and corresponds to mean collision
807 probability inside their territory. If residents survive, they age by one year and the annual loop
808 starts again and continues as long as the simulation lasts or until they die.

809 Dispersers are individuals of 1 year old and older that do not hold a territory yet. Both
810 dispersing males and dispersing females move on the gridded study area, one cell at a time, from
811 one to 45 times per day (Fig. A.6), following the same rules. The number of steps individuals
812 move per day is sampled, each day for each disperser, from a non-linear distribution defined by
813 (Kramer-Schadt et al., 2004). Dispersers follow a correlated habitat-dependent walk in a two-step
814 process: first, they favor their habitat preferences and then, maintain their previous direction with
815 a certain probability (i.e., correlation factor; Kramer-Schadt et al., 2004). This movement process
816 has been rigorously tested with inverse fitting, and “pattern-oriented modelling” (Kramer-Schadt
817 et al., 2007) using telemetry data of 6 dispersing lynx (5 females and one male) followed in the
818 Swiss Jura Mountains between 1988 and 1991. First, dispersers choose in which habitat type they
819 will move next. Dispersers favor “breeding” and “dispersal” habitats without distinction between
820 the two when moving and tend to avoid “matrix” habitats. By contrast, they never use “barrier”
821 habitats. The choice to move into the habitat “breeding/dispersal” or “matrix” depends on the
822 types of their 9 available cells for their next step (i.e., their 8 surrounding cells plus the one they
823 are currently on as they also can choose not to move). If the 9 cells are all of one type, excluding
824 the “barrier”, so they are either only “breeding/dispersal” (with or without “barrier”) or only
825 “matrix” (with or without “barrier”), the only available habitat type is selected. If the available
826 cells for an individual are a mix between “breeding/dispersal” and “matrix” habitats (with or
827 without “barrier”), there is a probability of 0.03 times the number of “matrix” cells among the 9
828 ones, to choose the “matrix” habitat for the next step (Kramer-Schadt et al., 2004). For example,
829 if an individual has 3 “matrix” cells available, there is a probability of 0.09 that it will choose a
830 “matrix” cell for its following location. Second, once the habitat type is selected, the choice of the
831 particular cell to move on, among the ones of the selected habitat type, is given by the correlation
832 part of the movement. Individuals follow a correlated movement with a probability equal to 0.5,

833 except for the first step of the day where there is no correlated movement (Kramer-Schadt et al.,
834 2004). If the movement is not correlated, the choice of the next cell among the ones of the
835 selected habitat type is random. If the movement is correlated, the chosen cell is the one
836 maintaining the most of the individual's current direction among the ones of the selected habitat
837 type. The chosen cell is then where the simulated lynx is moving to. Dispersers try to minimize
838 their time spent in "matrix" habitat and they can use their memory to return to a previously
839 visited "breeding/dispersal" habitat when needed. Lynx do not move more than 9 cells inside
840 "matrix" habitats, so if an individual already stepped 9 consecutive times in "matrix" cells and
841 the chosen cell for its next step is again of "matrix" type, it will use its memory and return to the
842 last "breeding/dispersal" cell visited (Kramer-Schadt et al., 2004). Finally, dispersers rotate
843 towards their chosen cell and move on their center. Once dispersers move to their next cell, they
844 may die from the spatial mortality due to vehicle collision. This spatial mortality is the collision
845 probability from the collision layer for the cell of their new location. If dispersers survive, they
846 search for a territory. If found, they stop moving, establish a territory, and become resident.
847 Dispersers that do not find a territory to establish on their new location keep moving, as many
848 steps during the day as simulated for them at the beginning of the day. At the end of the day, all
849 individuals that dispersed during the day may die from a fixed daily mortality probability of
850 0.0007 (i.e., baseline mortality estimated by inverse fitting; Kramer-Schadt et al., 2004).
851 Dispersers that have not established during the day can move and search for a territory every day
852 during the year. At the end of the year, all individuals (i.e., the ones that are still dispersers and
853 the former dispersers that found a territory and became residents) age by one year. Individuals
854 still dispersing will do again this same loop and the new residents will do the annual loop, as long
855 as the simulation lasts or until they die.
856



857

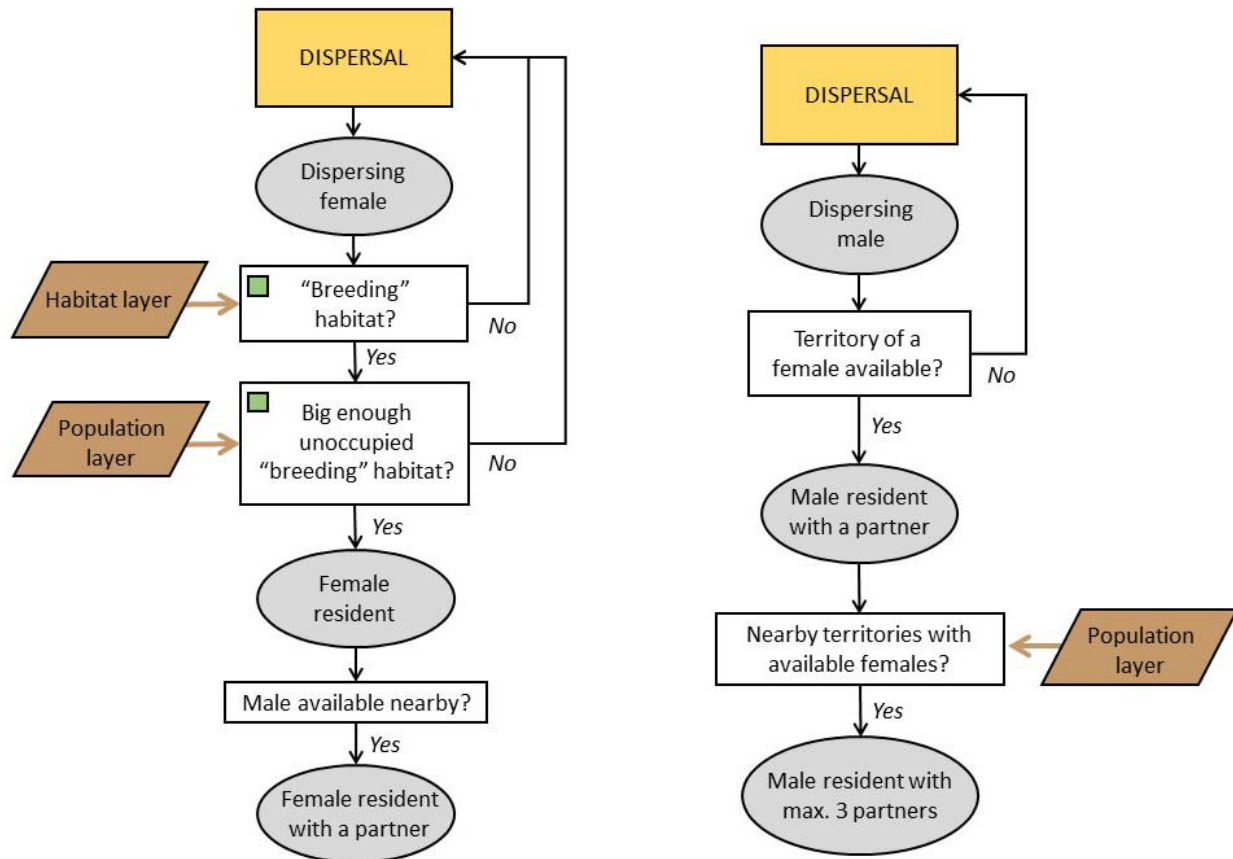
858 **Figure A.6:** Diagram of simulated lynx dispersal movement in the lynx SE-IBM. The loop
 859 represents one step. Individuals move one cell at the time per step, up to 45 a day or until they
 860 find a territory to establish (yellow box). Individual movement is constrained by habitat types in
 861 their surroundings defined by habitat layer (brown box), probabilities to step into different habitat
 862 types and to keep a direction (i.e., correlated movement), and their memory. Blue boxes represent
 863 mortalities an individual suffers, a spatial one each time reaching a new location, derived from
 864 the collision layer (brown box) and a daily one.

865

866 Dispersing individuals arriving at a new location have a different strategy to search for a
 867 new territory to establish regarding their sex (Fig. A.7). Females mainly look for good habitat

868 (which represent high prey availability) while males search for breeding opportunities and seek
869 female presence (Breitenmoser-Würsten et al., 2007b). Female dispersers need to be on a cell of
870 habitat type “breeding”, defined by the habitat layer, and with enough unoccupied “breeding”
871 habitats around their position to establish their territory. The size of the territory females need to
872 define represents the territory core area (Kramer-Schadt et al., 2005) and depends on the
873 population area where they are located in. We calculated these territory sizes to reach as:
874 $\text{mean}(95\% \text{ kernel}) - \text{sd}(95\% \text{ kernel})$, where “95% kernel” represents the territory size of female
875 residents calculated using a 95% kernel density (Heurich et al., 2018). When a female builds a
876 territory, the cells used as the territory core area are no longer available for other females. Only
877 the core areas are exclusive whereas an overlap between female territories on their edges is
878 possible, as observed with telemetry data (Breitenmoser-Würsten et al., 2007b, 2001;
879 Breitenmoser et al., 1993). Territory size to reach for the Alpine population is equal to 43.5 km^2
880 (mean home range computed with a 95% kernel = 76 km^2 , standard deviation = 32 km^2 ;
881 Breitenmoser-Würsten et al., 2001) and to 73 km^2 for the Jura population (mean home range
882 computed with a 95% kernel = 119 km^2 , standard deviation = 46 km^2 ; Breitenmoser-Würsten et
883 al., 2007b). As we did not have a reliable estimate for the Vosges-Palatinian and the Black Forest
884 populations, we assigned them the value of the Jura population due to the similarity of the habitat
885 structure (Fig. A.3). The “population” layer in which population area individuals are located in
886 (Fig. A.4). For that, each cell of the gridded study area was assigned the population of its closest
887 cell of regular lynx presence. Once a female has established a territory, she becomes resident. If a
888 resident male already established his territory nearby, the male includes the new resident female’s
889 territory into his own, if he has less than three females already associated. Then, the two
890 individuals may potentially reproduce the following year. That means, the male’s dispersal
891 strategy to establish is female-dependent. At their new location, dispersing males check if they

892 arrived inside the territory of a resident female available (i.e., with no male associated). If a male
893 finds a female, he pairs with her and his territory becomes the same as the one from the female je
894 joined. He also looks at nearby territories for up to three other available females to pair with.
895 Males can pair with females whose territories are within the size of their maximum territories.
896 The maximum distance between the male location and a female territory (i.e., a cell of her
897 territory) is the radius of the mean territory size (estimated by 95% kernel density). It is equal to
898 6.6 km for males in the Alpine population (mean home range computed with a 95% kernel = 137
899 km²; Breitenmoser-Würsten et al., 2001) and to 8.5 km for the other populations (mean home
900 range computed with a 95% kernel = 226 km²; Breitenmoser-Würsten et al., 2007b). If there are
901 several females available for a particular male, the closest one(s) are chosen. The male can
902 potentially reproduce the following year with every three females. The male territory is defined
903 by the union of all the female territories he has paired with. If a disperser cannot find a territory
904 to create (for female) or to join (for male) at their location, it continues dispersing. If all the
905 females a male is paired with die, the male start to disperse again in search of new available
906 female.
907



908

909 **Figure A.7:** Diagram of the territory establishment in the lynx SE-IBM separated by sex. The
 910 successive grey circles represent the successive status of the individual. Females are constrained
 911 by their surrounding habitat defined by the habitat layer (brown box). Males are constrained by
 912 female presence. The territory sizes females need to establish in “breeding” habitat and the
 913 maximum distance to which males can look for available female residents are constrained by the
 914 population in which the individuals are located (population layer; brown box).

915

916 The model is coded using R 4.1.0 (R Core Team, 2014). We used the package NetLogoR
 917 (Bauduin et al., 2019) to facilitate the IBM structure implementation in R language and the
 918 package SpaDES (Chubaty and McIntire, 2018) to schedule SE-IBM rules with different time

919 units. We also used the packages `data.table` (Dowle and Srinivasan, 2019), `randomcoloR`
920 (Ammar, 2019), `raster` (Hijmans and Van Etten, 2018), and `testthat` (Wickham et al., 2019).
921
922

923 **Appendix B - Complete description of the lynx spatially-explicit individual-based model**
924 **following the ODD protocol (Overview, Design concepts, and Details) developed by Grimm**
925 **et al. (2006, 2010)**

926

927 **Overview**

928 *Purpose*

929 The model simulates lynx population dynamics and dispersal, accounting for the impact of the
930 road network via the risk of vehicle collisions and of the land cover to represent the lynx habitat
931 preferences (Kramer-Schadt et al., 2005). Layers (i.e., maps) of collision probabilities, lynx
932 habitats and lynx populations (Appendix A) are combined with SE-IBM rules simulating lynx
933 dispersal (Kramer-Schadt et al., 2007, 2004) and demography (Kramer-Schadt et al., 2011, 2005).

934

935 *Entities, state variables, and scales*

936 The mobile entities of the model represent lynx individuals. Each simulated individual holds
937 several characteristics:

938 - *id*: each lynx is unique and has a unique numerical identity (*id*);

939 - *heading*: direction of the lynx in degrees from 0 to 360 (0 is heading North) to one of its 8

940 neighboring cells (i.e., 0°, 45°, 90°, 135°, 180°, 225°, 270°, 315°);

941 - *location*: coordinates of the lynx current position;

942 - *previous location*: lynx hold in memory the coordinates of their previous position;

943 - *population*: the population area where the individuals are born among Vosges-Palatinat, Black
944 Forest, Jura or Alps;

945 - *sex*: male or female;

946 - *age*: numerical value to represent the lynx age. This number represents the age as for humans
947 with 0 for the first year of life, 1 for the second year of life, etc.;

948 - *status*: resident (i.e., established with a territory) or disperser (i.e., without a territory);

949 - *steps*: number of steps dispersing lynx have to do during the current day;

950 - *last dispersing location*: dispersing lynx hold in memory the coordinates of the last cell of
951 habitat type “breeding” or “dispersal” they visited;

952 - *number matrix*: how many consecutive steps dispersing lynx move in “matrix” habitat;

953 - *male id*: for resident female only, the *id* of their male associated if they have one;

954 - *number females*: for resident male only, the number of resident females associated they have. It
955 cannot be more than 3;

956 - *road mortality territory*: mean collision probability inside its territory;

957

958 Simulated lynx progress on a gridded study area of 1 km² resolution encompassing the four
959 populations of interest: Vosges-Palatinian, Black Forest, Jura and Alpine, over Germany, France
960 and Switzerland (main text, Fig. 1). The gridded study area resolution corresponds to the lynx
961 perceptual range (Haller and Breitenmoser, 1986) and original resolution for the lynx individual-
962 based models (Kramer-Schadt et al., 2011, 2005, 2004). The gridded study area holds four
963 different variables:

964 - *collision probability* (from the collision layer): cells crossed by roads have an estimated
965 probability of fatal collision with lynx, a value between 0 and 1. The collision probability is equal
966 to 0 for cells with no road in it;

967 - *habitat type* (from the habitat layer): each cell is of one of the following habitat types:
968 “breeding”, “dispersal”, “matrix” or “barrier”;

969 - *individual territory*: this variable holds the *id* of the female residents for the cells included in
970 their territories.

971

972 *Process overview and scheduling*

973 At the beginning of each simulated year, lynx individuals are differentiated by their resident or
974 disperser *status* (Appendix A, Fig. A.5). Residents progress on a yearly time step while dispersers
975 progress on a daily time step. Resident lynx may reproduce and kittens are born. Then, residents
976 may die from a fixed annual baseline mortality and from a spatial annual mortality, the latter
977 being defined by the mean *collision probability* inside their territories. All these happen once in
978 the year. Dispersing lynx move on the gridded study area (Appendix A, Fig. A.6) and their
979 simulation time step is daily. A number of *steps* to move per day is generated for each dispersing
980 lynx at the beginning of each day. All dispersers move simultaneously, one step (i.e., cell) at the
981 time, following a correlated habitat-dependent movement influenced by the *habitat type*. After
982 each step, dispersing lynx may suffer from a spatial mortality given by the *collision probability* at
983 their new location. Then, the surviving ones search for a place where establish their territory
984 (Appendix A, Fig. A.7). Dispersing females are constrained by the *habitat type* in their
985 surroundings to establish a new territory while dispersing males are constrained by the presence
986 of available resident females for reproduction (Breitenmoser-Würsten et al., 2007b). Dispersing
987 lynx that do not establish a territory continue to disperse during their set number of *steps* per day,
988 and every day during the year until establish one. At the end of each day, lynx that dispersed
989 suffer from a fixed daily baseline dispersal mortality. At the end of a year, all lynx *age* are
990 incremented by one and their *status* is updated if applicable.

991 Simulated lynx *id*, *population*, and *sex* variables do not change during a simulation. Their
992 *heading*, *location*, *previous location*, *last dispersing location*, *number matrix*, *male id*, *number*

993 *females* and *road mortality territory* may change at every dispersal step. The *steps* variable is
994 updated each day, and *age* and *status* are updated once a year. There is no modeling of the
995 landscape during the simulation, therefore the *collision probability* and *habitat type* are constant
996 during one simulation. However, the *territory id* may change at every step (i.e., each time there is
997 a new territory created).

998

999 **Design concepts**

1000 *Basic principles*

1001 Our lynx SE-IBM is an assemblage of four components: three spatial layers and an existing
1002 individual-based model (Kramer-Schadt et al., 2011, 2007, 2005, 2004) parameterized with field
1003 data and inverse model fitting (Appendix A, Fig. A.1). The *collision probability* layer and the
1004 *habitat type* layer are output maps from a collision model and a habitat model (Appendix A). The
1005 population layer was generated using presence data and estimated number of lynx (Appendix A).

1006

1007 *Emergence*

1008 Simulated males and females follow the same rules, except for the territory establishment where
1009 females are constrained by the habitat and males by the presence of females (Breitenmoser-
1010 Würsten et al., 2007b). In an area with few individuals, females have many possibilities of empty
1011 land to establish. On the other hand, males will have to move further to find one of the few
1012 females available, inducing longer dispersal movement and higher chances of dying. However, in
1013 a very dense area, this can be reverse. Females will have to move further to find free land
1014 whereas males, if they are not too numerous, may move less as they may have many females
1015 around them to pair with. If there are many males, all females may be already taken, inducing
1016 again longer dispersal for the males.

1017 To breed, residents are constrained by the presence of a partner on their territory and both
1018 individuals need to be in age of reproduction. The presence of a suitable partner depends on their
1019 past dispersal movement and territory search, and if they could establish a new territory or not.
1020 Reproduction success is therefore stochastic, as well as the number of kittens. All this will
1021 modify the number of new individuals in the population.

1022 Residents and dispersers are subject to a fixed baseline and spatial mortality. Both
1023 mortalities are stochastic and the spatial one depends on the location of the individuals (i.e.,
1024 *collision probability* at their location). The location of the individuals depends on their current or
1025 past dispersal movement that depends on the *habitat type*. All this will modify the number of
1026 individuals dying each year.

1027

1028 *Adaptation*

1029 There is no adaptation in the SE-IBM (e.g., no density-dependent rules). All individuals follow
1030 the same behavioral rules regarding dependent on their inner and environmental characteristics.

1031

1032 *Objectives*

1033 Simulated lynx do not have adaptive traits and they do not have a final goal to reach. They take
1034 decisions at each time step based only their current characteristics and their environment to fulfill
1035 an immediate goal (e.g., where to move next).

1036

1037 *Learning*

1038 Simulated lynx try to minimize their time spent in poor habitat quality. When they step into
1039 “matrix” habitat, they store the position they just left (i.e., their last position in good habitat so
1040 either in “breeding” or “dispersing” habitat) and start counting how many consecutive steps into

1041 “matrix” they do. At the 10th step, if the chosen cell is still located into “matrix”, simulated lynx
1042 will use their memory to find and move back to their last location of good habitat.

1043

1044 *Prediction*

1045 Simulated lynx cannot predict the future nor have a global view of the study area. Both residents
1046 and dispersers sense the current state of their environment and population in their immediate
1047 surroundings and react to this information only. For example, dispersing lynx move, making
1048 decisions based only on their immediate surroundings. They cannot anticipate their path nor have
1049 a global view of their environment to make advantageous decisions for the long term (e.g., they
1050 cannot find a better territory and/or more quickly).

1051

1052 *Sensing*

1053 Simulated lynx can sense the *habitat types* and the *individual territories* of their surroundings.
1054 Dispersers establishing a territory and trying to find a partner can sense other individuals: it
1055 knows who hold nearby territories and the characteristics of these individuals, such as their *sex*
1056 and the availability to pair with.

1057

1058 *Interaction*

1059 There is a tacit competition between dispersing individuals during the territory search. Dispersers
1060 evaluate their surroundings before establishing one at the time (in a random order each time). The
1061 first female in the list looking for a territory has therefore more chance to find empty space to
1062 establish than the last female. Similarly, the first male in the list looking for an available female
1063 to join has more chance to find a partner than the last male.

1064 Residents also interact with each other for reproduction. If a male and a female resident in
1065 age of reproduction are on the same territory, they may reproduce and produce new individuals.
1066 However, this interaction does not affect any of the individuals (i.e., a reproduction event does
1067 not change the characteristics of the concerned individuals).

1068

1069 *Stochasticity*

1070 The model involves several stochastic processes. Regarding demography, both reproduction (i.e.,
1071 the reproduction event itself and the number of kittens) and mortality (both fixed baseline and
1072 spatial for both residents and dispersers) are defined with probabilities and therefore outcomes
1073 vary between individuals. During dispersal, the choice of the *habitat type* to select when there is a
1074 mix of “breeding/dispersal” and “matrix” habitats in the lynx surroundings depends on the
1075 probability to step into “matrix”. Then, there is also stochasticity in the choice of the next
1076 location within the selected *habitat type*, as the choice to follow a correlated movement or not are
1077 defined by probabilities, as well as the choice of the next location among those available if the
1078 movement is not correlated. Finally, the number of steps dispersing simulated lynx must move
1079 during a day is sampled from a probability distribution and so, the outcomes are not the same for
1080 all individuals all the time.

1081

1082 *Collectives*

1083 When a resident male finds one or several available resident females, he may pair and reproduce
1084 with them. Every year, the pairs persist unless one of the lynx dies and therefore frees the other
1085 one to find another partner.

1086 Individuals belong to populations. The two rules different between populations are the
1087 size of the territory females have to reach and the maximum size of territory males can hold (i.e.,

1088 how far can they look for females). Other than that, no rule is defined at the population level and
1089 the population affiliation does not affect the individual progress. Individual progress according to
1090 their own characteristics.

1091

1092 *Observation*

1093 The state of all alive simulated individuals (i.e., their location and characteristics) and of the
1094 *individual territory* map are available at the end of each yearly time step. All disperser
1095 movements are also recorded on a single map over the whole simulation time. Events changing
1096 the demography are saved each year: which individuals died and how (i.e., from the fixed
1097 baseline mortality or from collision), which individuals reproduce, who are the kittens, and which
1098 individuals became residents.

1099

1100 **Details**

1101 *Initialization*

1102 At the beginning of the simulation, a grid covering the whole study area (main text, Fig. 1) with a
1103 resolution of 1 km² is created. The values from the habitat layer and the collision layer are
1104 transferred to the gridded study area variables *habitat type* and *collision probability*. The variable
1105 *individual territory* is set to missing value (NA) everywhere. *Collision probability* on the borders
1106 of the grid are set to 1 so that individuals reaching the borders of the study area die instead of
1107 bouncing on the border. We assumed that these individuals would have left the study area and
1108 therefore are removed from the population even if it is not unlikely that they, or others, may come
1109 back.

1110 Using the population layer, a lynx population is created with unique *id* for each
1111 individual, *location*, *population*, as well with a *sex* and *age* (when known from field data) for

1112 some individuals. A simulated individual cannot start in “matrix” or “barrier” habitats. Therefore,
 1113 if some individuals’ *location* are located in these habitats, they are relocated to the closest cell of
 1114 “breeding” or “dispersal” habitat. At the start of the simulation, individuals cannot have a
 1115 *previous location* or a *last dispersing location*, so we used their current *location* for these two
 1116 variables. For individuals with unknown *sex* or *age* (no field data), a random *sex* (male or female
 1117 with ratio 1:1) and *age* (between 2 and 15) is randomly given. A random *heading* (i.e., initial
 1118 direction) is given to each individual, between 0° and 360°. At the start, all individuals have a
 1119 *disperser status*, 0 *steps* to do, 0 steps done in “matrix” (*number matrix*) and 0 for *road mortality*
 1120 *territory*. As all individuals are dispersers, no couples are made so *male id* for the females are set
 1121 to NA and the *number females* for males are set to 0 (Table B1).

1122 The parameter values used in the SE-IBM to represent the dynamics and dispersal of the
 1123 Eurasian lynx in our study area (main text, Fig. 1) are provided in Table B.1.

1124

Parameter	Explanation	Value	Reference
endTime	Number of years to simulate the lynx populations.	50	By default
pRepro	Probability of reproduction for a couple (one resident female and one resident male associated on the same territory, both in age of reproduction).	0.81	(Breitenmoser-Würsten et al., 2007a; López-Bao et al., 2019)
nKittyYoungF	Number of kittens that survive until the age of dispersal when the female is considered “young” (maxAgeYoungF	1 or 2 (with equal	(Breitenmoser-Würsten et al., 2007a, 2001;

	parameter).	probability for each value)	Henriksen et al., 2005; Kramer- Schadt et al., 2005; López-Bao et al., 2019)
nKittyOldF	Number of kittens that survive until the age of dispersal when the female is considered “old” (maxAgeYoungF parameter).	0 or 1 (with equal probability for each value)	(Breitenmoser- Würsten et al., 2007a, 2001; Henriksen et al., 2005; Kramer- Schadt et al., 2005)
maxAgeYoungF	Age maximum for a female to be considered “young”.	11	(Henriksen et al., 2005)
minAgeReproF	Age minimum at which all resident females can reproduce (i.e., are sexually mature). One year younger, only half of the females are mature.	2 (3 rd year of life)	(Breitenmoser- Würsten et al., 2007a; Henriksen et al., 2005; Kvam, 1991)
minAgeReproM	Age minimum at which all resident males can reproduce (i.e., are sexually mature). One year younger, only half of the males are mature.	3 (4 th year of life)	(Kvam, 1991)

pMortRes	Fixed annual baseline probability of mortality for resident lynx (without the risk of collisions included).	0.1	(Breitenmoser-Würsten et al., 2007a; Heurich et al., 2018)
ageMax	Age maximum a lynx can reach.	20	(von Arx et al., 2017)
xPs	Exponent value of the power function to define the daily step distribution. This distribution is used to sample a number of steps per day to move for dispersing lynx.	11	(Kramer-Schadt et al., 2004)
sMaxPs	Maximum number of steps dispersing lynx can do in one day.	45	(Kramer-Schadt et al., 2004)
pMat	Probability of stepping into “matrix” cells. This probability is multiplied by the number of “matrix” cells available for a dispersing lynx to define its probability to choose this habitat type for its next location.	0.03	(Kramer-Schadt et al., 2004)
pCorr	Movement correlation probability for dispersal movement.	0.5	(Kramer-Schadt et al., 2004)
pMortDisp	Fixed daily baseline probability of mortality for dispersing lynx (without the risk of collisions included).	0.0007	(Kramer-Schadt et al., 2004)

nMatMax	Maximum number of consecutive steps a dispersing lynx can do into “matrix” habitat. After this number of steps, disperser can use its memory to find and move in the last good habitat visited (“breeding” or “dispersal” habitat).	9	(Kramer-Schadt et al., 2004)
coreTerrSizeFAlps	Territory size (in km ²) females located in the Alpine population need to reach.	43.5	(Breitenmoser-Würsten et al., 2001; Heurich et al., 2018)
coreTerrSizeFJura	Territory size (in km ²) females located in the Jura population need to reach.	73	(Breitenmoser-Würsten et al., 2007b; Heurich et al., 2018)
coreTerrSizeFBF	Territory size (in km ²) females located in the Black Forest population need to reach.	73	(Breitenmoser-Würsten et al., 2007b; Heurich et al., 2018)
coreTerrSizeFVP	Territory size (in km ²) females located in the Vosges-Palatinian population need to reach.	73	(Breitenmoser-Würsten et al., 2007b; Heurich et al., 2018)
terrSizeMAlps	Maximum territory size (in km ²) males	137	(Breitenmoser-

	located in the Alpine population can reach.		Würsten et al., 2001)
terrSizeMJura	Maximum territory size (in km ²) males located in the Jura population can reach.	226	(Breitenmoser-Würsten et al., 2007b)
terrSizeMBF	Maximum territory size (in km ²) males located in the Black Forest population can reach.	226	(Breitenmoser-Würsten et al., 2007b)
terrSizeMVP	Maximum territory size (in km ²) males located in the Vosges-Palatinian population can reach.	226	(Breitenmoser-Würsten et al., 2007b)
nFem	Maximum number of resident females a resident male can pair with and potentially reproduce with.	3	(Kramer-Schadt et al., 2005)

1125 **Table B.1:** Parameter values used in the SE-IBM.

1126

1127 *Input data*

1128 See Appendix A.

1129

1130 *Submodels*

1131 Reproduction: A female resident in *age* of reproduction (1 or 2 years old, see Table B1) with a

1132 resident male associated, also in *age* of reproduction (2 or 3 years old, see Table B1), may

1133 reproduce and have offspring. There is a Bernoulli trial for the reproduction success with a mean

1134 probability of 0.81 (Table B1). If the couple reproduces, the female can produce 1 or 2 kittens
1135 with equal probability (Table B1) if she is “young” (11 years old or younger; Table B1) or 0 or 1
1136 kitty with equal probability if she is “old” (12 years old or older; Table B1). Kittens obtain a
1137 unique *id*. The *sex* of the kittens is randomly chosen between male and female with ratio 1:1
1138 (Table B1). Their *location*, *heading* and *last dispersing location* are the ones from their mother.
1139 Their *previous location* is NA. Their *population* is not the one of their parents but the one of the
1140 territory where they are born. Their *age* is 0, their *status* is resident (i.e., they stay with their
1141 mother and cannot disperse at this age), their *steps* and *number matrix* are 0, their *male id* is NA,
1142 and their *number females* and *road mortality territory* are 0.

1143
1144 Mortality: Resident individuals may die from a fixed baseline annual mortality and a spatial
1145 mortality using a Bernoulli trial. Mortality rate for all residents is equal to 0.1 (Table B1) for the
1146 fixed baseline mortality. We did not simulate an effect of senescence but we set an age maximum
1147 (20 years; Breitenmoser-Würsten et al., 2007a; Stahl and Vandel, 1998; von Arx et al., 2017) and
1148 all individuals reaching this age die. Resident individuals may also die from spatial mortality, the
1149 mortality rate for female residents is the mean *collision probability* of the cells inside of their
1150 territory (*road mortality territory*). For male residents, spatial mortality rate is the mean *road*
1151 *mortality territory* of all his paired females.

1152 When a resident female dies, her territory disappears (i.e., her *id* is removed from the *individual*
1153 *territory study area variable*) and the local male resident loses one female. If this male did not
1154 have other females, he becomes disperser again to look for new available female resident as male
1155 home ranges are adjusted to mate availability (Breitenmoser-Würsten et al., 2007b). If the dead
1156 female had kittens during the year, they all die too. When a male resident dies, its paired females
1157 become available to reproduce with other males.

1158
1159 Demography: All individuals' *age* increment of 1. The *status* of the offspring of the year changes
1160 from resident to disperser.

1161
1162 Dispersal: A number of steps to do during the day is sampled every day from the distribution
1163 defined by Kramer-Schadt et al. (2004), one per dispersing individual. Dispersers move one cell
1164 per step, into one of their 8 neighboring cell or their current cell (i.e., so a total of 9 potential cells
1165 for their next location). First, there is a selection of the habitat type each disperser chooses for its
1166 next location. Among those 9 cells, all cells of "barrier" type are removed from the potential
1167 choices. As individuals cannot be on "barrier" habitat, it cannot happen that all 9 cells are
1168 "barrier". For a dispersing individual, if all his cells are of "breeding" or "dispersal" type, this
1169 type ("breeding/dispersal" without distinction between the two) is selected. If all his cells are of
1170 "matrix" (i.e., habitat not favorable for lynx), this type is selected. If the cells are a mix of
1171 "breeding/dispersal" and "matrix", there is a Bernoulli trial of mean 0.03 (Table B1) times the
1172 number of "matrix" cells among the 9 to select the "matrix" habitat type.

1173 Second, there is the choice of the cell for the next location, among all cells of the selected
1174 habitat type. If there is only one cell of the selected type, the disperser goes on this one (for
1175 example, if the selected cell type is "breeding/dispersal" and there is only one cell of
1176 "breeding/dispersal" habitat, all other cells being "matrix" or "barrier", therefore the only cell of
1177 "breeding/dispersal" is selected). Otherwise, there is a choice either random or governed by a
1178 correlation movement to select the cell. The first step of the day is never correlated, otherwise
1179 there is a Bernoulli trial of mean 0.5 (Table B1) to determine if the movement is correlated for
1180 each individual for each step. If there is no correlation of the movement, the choice for the cell is
1181 done with an equal probability for all the cells of the selected type. If the movement is correlated,

1182 the rotating angle between the individual's heading and each cell of the selected type is
1183 calculated. There is a preference for the cell minimizing this rotating angle. If the cell that the
1184 individual is on is of the selected habitat type (i.e., this cell can be selected for the individual's
1185 next step which will make the lynx stay at its current location), the preference for this cell is
1186 equal to the preference for the cells inducing a rotation of +/- 90 degrees, to favor a forward
1187 movement. The selected cell is the one with the highest preference (i.e., smallest rotation
1188 movement).

1189 Third, if the selected cell is of "matrix" type, the individuals check how many consecutive
1190 steps they have done in "matrix" habitat (*number matrix*). If a disperser has already step 9
1191 consecutive times (Table B1) in "matrix" and this 10th consecutive step is in "matrix" too, it used
1192 its memory and select, for its next location, the last visited cell of good habitat ("breeding" or
1193 "dispersal"), stored in his memory (*last dispersing location*) (Kramer-Schadt et al., 2004).

1194 Finally, all dispersers rotate towards their selected cell and move to the center of it. They
1195 update their *previous location* with the coordinates of the cell they were before moving. If their
1196 new location is of type "breeding" or "dispersal", this location is stored in their memory for the
1197 *last dispersing location*, and their *number matrix* is reset to 0. If their new location is of type
1198 "matrix", their *number matrix* increments of 1.

1199 Now, disperser may be also subject to spatial mortality. There is a Bernoulli trial for each
1200 disperser with mean equal to the *collision probability* value of their new location. Dispersers that
1201 do not die from spatial mortality search for a new home range to establish (*searchTerritory* sub-
1202 model). At the beginning of the simulation, all initial individuals were created as dispersers
1203 because we did not have data on lynx territories. Thus, we decided to let simulated individuals
1204 establish their initial territories. Individuals need some time to do so, therefore we did not apply
1205 any mortality during the first year of simulation to allow dispersers to establish their territories.

1206 The dispersal movement is done until all dispersing individuals have reached their number
1207 of steps to do during the day, or they have established a territory and are now resident, or they
1208 died from spatial mortality. At the end of the day, all individuals that were dispersers at the
1209 beginning of the day (even the ones that established a territory during the day) may die from a
1210 fixed daily baseline mortality. There is a Bernoulli trial of mean 0.0007 (Table B1) for each
1211 individual. If some of the individuals dying established a territory during this day, there is an
1212 update of their characteristics similarly as the death of resident individuals. As the spatial
1213 mortality, there is no fixed daily baseline mortality applied the first year of the simulation.
1214
1215 SearchTerritory: Dispersing individuals arriving at a new location search for a territory to
1216 establish, one at the time, and females first (i.e., male establishment depending on female
1217 territories). One at the time, in a random order each time, females evaluate their surroundings to
1218 establish their territory. Females needs to be on a cell of “breeding” type that is not already
1219 included in any territory to start creating their own. Then, they need to have enough empty
1220 contiguous cells of “breeding” habitat to create a big enough territory (Table B1). When enough
1221 empty contiguous cells of “breeding” habitat are present, females establish their territory and
1222 becomes resident. Otherwise, they stay disperser and keep dispersing. After a female establishes
1223 her territory, all the cells of this territory become occupied and are not available anymore for the
1224 other females. The female *id* is given to the *individual territory* study area variable for all cells in
1225 her territory. The mean *collision probability* of the territory is calculated and given to the female
1226 for her *road mortality territory*. Then, the female checks if there is a resident male around. If
1227 there is a male not further than the radius of the maximum territory size for the male (Table B1)
1228 and he has less than three females associated, she pairs with him. Then, she obtains the male *id*
1229 for her *male id* and the *number females* of the male increments of 1.

1230 After all dispersing females tried to establish, dispersing males try to disperse of their
1231 own, one at the time in a random order each time. Each male checks if the cell they are currently
1232 on is located inside the territory of an available female (i.e., without male associated yet). If there
1233 is one, the male pairs with this female and becomes resident. His territory is the same one as the
1234 female and its *number females* is set to 1. The female obtains the male *id* for her *male id*
1235 variables. Then, the male also looks if there are other nearby available females to pair with, as far
1236 as the radius of its maximum territory size (Table B1). He can pair with up to three available
1237 females. The male territory represents the union of all the female territory with which he paired
1238 with. His *road mortality territory* is the mean value of the *road mortality territory* of all the
1239 females he paired with.

1240

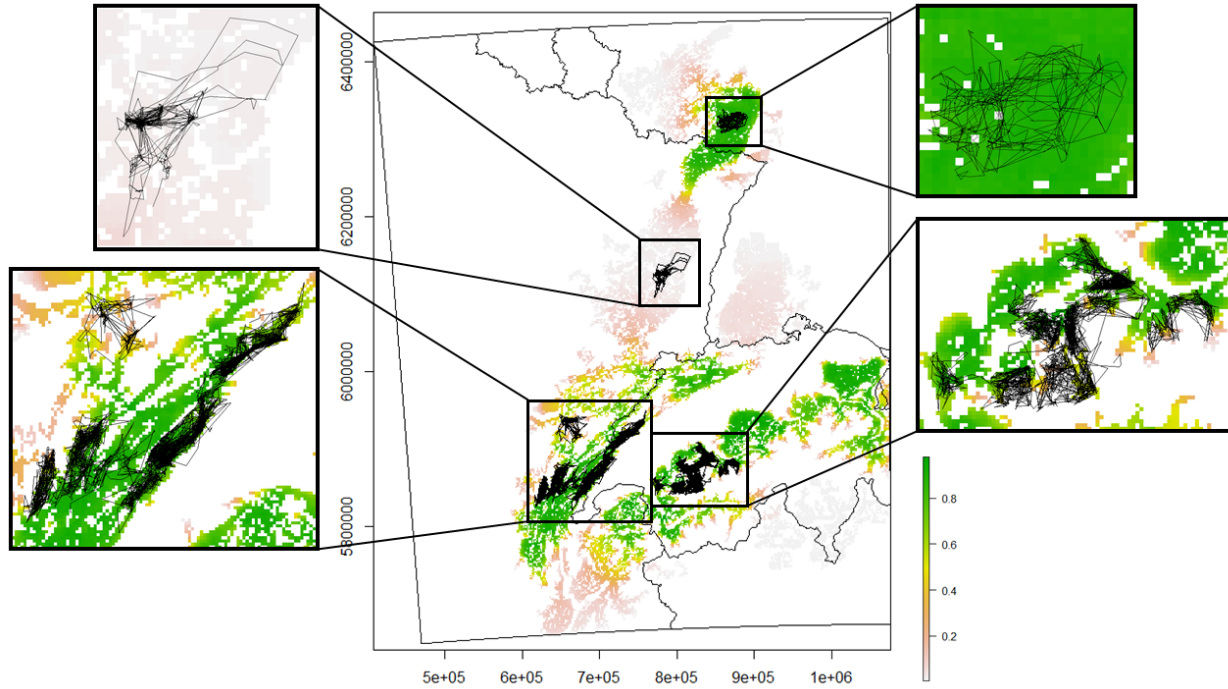
1241

1242 **Appendix C – Model validation using telemetry data**

1243

1244 We performed a qualitative validation of the lynx model predictions using tracks from collared
1245 female residents. The map below (Fig. C.1) shows the simulated territory occupancy with darker
1246 green colors representing areas most often occupied by female territories across simulation
1247 replicates (main text, Fig. 4). We overlaid GPS and VHF paths from collared female residents on
1248 the map. Data for the Vosges-Palatinat consist in 2 females in the Palatinat Forest (Germany)
1249 and 1 in the Vosges Mountains (France) (locations recorded in 2017-2018; Scheid et al., 2021).
1250 Data in the Jura represent 4 females in France and 13 in Switzerland (locations recorded in 1988-
1251 1998; Breitenmoser-Würsten et al., 2007b). Data in the Alps (Switzerland) represent 16 females
1252 (locations recorded in 1997-2000; Breitenmoser-Würsten et al., 2001). We show the full study
1253 area and 4 zoomed areas on the telemetry locations. Overall, the realized movements and territory
1254 emergence inferred from field data were coherent with the frequencies across simulation
1255 replicates of the territories defined by simulated lynx. Most of the recorded tracks were located in
1256 cells with a high frequency of territory occupancy and very few were located in cells where no
1257 territories were simulated.

1258



1259

1260 **Figure C.1:** Occupancy by female territories at the last year of simulation over the whole study
1261 area (black rectangle) used in the lynx SE-IBM using the habitat model. Limits of the countries
1262 (main text, Fig. 1) are overlaid in black over the map. Values between 0 and 1 are mean
1263 occupancy probability per cell of 1 km² over 200 replicates (e.g., cell with value equal to 1 were
1264 always occupied in all simulation replicates). GPS and VHF recorded paths for female residents
1265 are overlaid as thin black lines. A zoom of different areas are presented on the sides of the central
1266 map.

1267

1268

1269 **References**

- 1270 Ammar, R., 2019. randomcoloR: Generate Attractive Random Colors.
- 1271 Basille, M., Van Moorter, B., Herfindal, I., Martin, J., Linnell, J.D.C., Odden, J., Andersen, R.,
1272 Gaillard, J.M., 2013. Selecting habitat to survive: the impact of road density on survival in a
1273 large carnivore. PLoS One 8, e65493. <https://doi.org/10.1371/journal.pone.0065493>
- 1274 Bauduin, S., McIntire, E.J.B., Chubaty, A.M., 2019. NetLogoR: a package to build and run
1275 spatially explicit agent-based models in R. Ecography (Cop.). 42, 1841–1849.
1276 <https://doi.org/10.1111/ecog.04516>
- 1277 Breitenmoser-Würsten, C., Vandel, J.-M., Zimmermann, F., Breitenmoser, U., 2007a.
1278 Demography of lynx *Lynx lynx* in the Jura Mountains. Wildlife Biol. 13, 381–392.
1279 [https://doi.org/10.2981/0909-6396\(2007\)13](https://doi.org/10.2981/0909-6396(2007)13)
- 1280 Breitenmoser-Würsten, C., Zimmermann, F., Ryser, A., Capt, S., Laass, J., Siegenthaler, A.,
1281 Breitenmoser, U., 2001. Untersuchungen zur Luchspopulation in den Nordwestalpen der
1282 Schweiz 1997–2000, KORA Bericht.
- 1283 Breitenmoser-Würsten, C., Zimmermann, F., Stahl, P., Vandel, J.-M., Molinari-Jobin, A.,
1284 Molinari, P., Capt, S., Breitenmoser, U., 2007b. Spatial and social stability of a Eurasian
1285 lynx *Lynx lynx* population: an assessment of 10 years of observation in the Jura Mountains.
1286 Wildlife Biol. 13, 365–380. [https://doi.org/10.2981/0909-6396\(2007\)13\[365:sassoa\]2.0.co;2](https://doi.org/10.2981/0909-6396(2007)13[365:sassoa]2.0.co;2)
- 1287 Breitenmoser, U., Kavczensky, P., Dötterer, M., Breitenmoser-Würsten, C., Capt, S., Bernhart,
1288 F., Liberek, M., 1993. Spatial organization and recruitment of lynx (*Lynx lynx*) in a re-
1289 introduced population in the Swiss Jura Mountains. J. Zool. London 231, 449–464.
- 1290 Charbonnel, A., Germain, E., 2020. Plan Régional d'Actions en faveur du Lynx boréal (*Lynx*
1291 *lynx*) dans le Massif des Vosges: rétablir le Lynx dans un état de conservation favorable
1292 dans le cadre d'une démarche participative, concertée et partagée avec les acteurs du

- 1293 territoire. Lucy (57), France.
- 1294 Chubaty, A.M., McIntire, E.J.B., 2018. Package “SpaDES”: Develop and Run Spatially Explicit
1295 Discrete Event Simulation Models.
- 1296 Dowle, M., Srinivasan, A., 2019. data.table: Extension of “data.frame.”
- 1297 Duchamp, C., Boyer, J., Briaudet, P.E., Leonard, Y., Moris, P., Bataille, A., Dahier, T., Delacour,
1298 G., Millisher, G., Miquel, C., Poillot, C., Marboutin, E., 2012. A dual frame survey to assess
1299 time- and space-related changes of the colonizing wolf population in France. *Hystrix* 23, 1–
1300 12. <https://doi.org/10.4404/hystrix-23.1-4559>
- 1301 Gatti, S., Blanc, L., Gimenez, O., Marboutin, E., 2014. Estimation des densités de lynx dans le
1302 massif du Jura entre 2011 et 2014. *Faune Sauvag.* 304, 4–8.
- 1303 Gimenez, O., Gatti, S., Duchamp, C., Germain, E., Laurent, A., Zimmermann, F., Marboutin, E.,
1304 2019. Spatial density estimates of Eurasian lynx (*Lynx lynx*) in the French Jura and Vosges
1305 Mountains. *Ecol. Evol.* 9, 11707–11715. <https://doi.org/10.1002/ece3.5668>
- 1306 Grimm, V., Berger, U., Bastiansen, F., Eliassen, S., Ginot, V., Giske, J., Goss-Custard, J., Grand,
1307 T., Heinz, S.K., Huse, G., Huth, A., Jepsen, J.U., Jørgensen, C., Mooij, W.M., Müller, B.,
1308 Pe’er, G., Piou, C., Railsback, S.F., Robbins, A.M., Robbins, M.M., Rossmanith, E., Rüger,
1309 N., Strand, E., Souissi, S., Stillman, R.A., Vabø, R., Visser, U., DeAngelis, D.L., 2006. A
1310 standard protocol for describing individual-based and agent-based models. *Ecol. Modell.*
1311 198, 115–126. <https://doi.org/10.1016/j.ecolmodel.2006.04.023>
- 1312 Grimm, V., Berger, U., DeAngelis, D.L., Polhill, J.G., Giske, J., Railsback, S.F., 2010. The ODD
1313 protocol: A review and first update. *Ecol. Modell.* 221, 2760–2768.
1314 <https://doi.org/10.1016/j.ecolmodel.2010.08.019>
- 1315 Haller, H., Breitenmoser, U., 1986. Zur Raumorganisation der in den Schweizer Alpen
1316 wiederangesiedelten Population des Luchses *Lynx lynx*. *Z. Säugetierk* 51, 289–311.

- 1317 Henriksen, H.B., Andersen, R., Hewison, A.J.M., Gaillard, J.-M., Bronndal, M., Jonsson, S.,
1318 Linnell, J.D.C., Odden, J., 2005. Reproductive biology of captive female Eurasian lynx,
1319 *Lynx lynx*. *Eur. J. Wildl. Res.* 51, 151–156. <https://doi.org/10.1007/s10344-005-0104-1>
- 1320 Heurich, M., Schultze-naumburg, J., Piacenza, N., Magg, N., Čeverny, J., Engleder, T.,
1321 Herdtfelder, M., Sladova, M., Kramer-Schadt, S., 2018. Illegal hunting as a major driver of
1322 the source-sink dynamics of a reintroduced lynx population in Central Europe. *Biol.*
1323 *Conserv.* 224, 355–365. <https://doi.org/S0006320717314003>
- 1324 Hijmans, R.J., Van Etten, J., 2018. Package “raster”: Geographic Data Analysis and Modeling.
- 1325 Isaac, N.J.B., van Strien, A.J., August, T.A., de Zeeuw, M.P., Roy, D.B., 2014. Statistics for
1326 citizen science: extracting signals of change from noisy ecological data. *Methods Ecol. Evol.*
1327 5, 1052–1060. <https://doi.org/10.1111/2041-210X.12254>
- 1328 Jedrzejewski, W., Jedrzejewska, B., Okarma, H., Schmidt, K., Bunevich, A.N., Miłkowski, L.,
1329 1996. Population dynamics (1869-1994), demography, and home ranges of the Lynx in
1330 Białowieża Primeval Forest (Poland and Belarus). *Ecography (Cop.)*. 19, 122–138.
- 1331 Kramer-Schadt, S., Kaiser, T.S., Frank, K., Wiegand, T., 2011. Analyzing the effect of stepping
1332 stones on target patch colonisation in structured landscapes for Eurasian lynx. *Landsc. Ecol.*
1333 26, 501–513. <https://doi.org/10.1007/s10980-011-9576-4>
- 1334 Kramer-Schadt, S., Revilla, E., Wiegand, T., 2005. Lynx reintroductions in fragmented
1335 landscapes of Germany: Projects with a future or misunderstood wildlife conservation? *Biol.*
1336 *Conserv.* 125, 169–182. <https://doi.org/10.1016/j.biocon.2005.02.015>
- 1337 Kramer-Schadt, S., Revilla, E., Wiegand, T., Breitenmoser, U., 2004. Fragmented landscapes,
1338 road mortality and patch connectivity: modelling influences on the dispersal of Eurasian
1339 lynx. *J. Appl. Ecol.* 41, 711–723.
- 1340 Kramer-Schadt, S., Revilla, E., Wiegand, T., Grimm, V., 2007. Patterns for parameters in

- 1341 simulation models. *Ecol. Modell.* 204, 553–556.
- 1342 <https://doi.org/10.1016/j.ecolmodel.2007.01.018>
- 1343 Kvam, T., 1991. Reproduction in the European lynx, *Lynx lynx*. *Zeitschrift für Säugetierkd.* 56,
1344 146–158.
- 1345 López-Bao, J.V., Aronsson, M., Linnell, J.D.C., Odden, J., Persson, J., Andrén, H., 2019.
1346 Eurasian lynx fitness shows little variation across Scandinavian human-dominated
1347 landscapes. *Sci. Rep.* 9, 8903. <https://doi.org/10.1038/s41598-019-45569-2>
- 1348 Outhwaite, C.L., Chandler, R.E., Powney, G.D., Collen, B., Gregory, R.D., Isaac, N.J.B., 2018.
1349 Prior specification in Bayesian occupancy modelling improves analysis of species
1350 occurrence data. *Ecol. Indic.* 93, 333–343. <https://doi.org/10.1016/j.ecolind.2018.05.010>
- 1351 R Core Team, 2014. R: A language and environment for statistical computing.
- 1352 Scheid, C., Germain, E., Schwoerer, M.-L., 2021. Les Lynx (*Lynx lynx*) du Pfälzerwald
1353 s’installent progressivement dans le Massif des Vosges. *Ann. Sci. Rés. Bios. Trans. Vosges*
1354 *du Nord.* 20, 104–125.
- 1355 Stahl, P., Vandel, J.-M., 1998. Le Lynx boréal □: *Lynx lynx* (Linné, 1758). *Encycl. des Carniv.*
1356 *Fr.*
- 1357 Visintin, C., Golding, N., van der Ree, R., McCarthy, M.A., 2018. Managing the timing and
1358 speed of vehicles reduces wildlife-transport collision risk. *Transp. Res. Part D* 59, 86–95.
1359 <https://doi.org/10.1016/j.trd.2017.12.003>
- 1360 Visintin, C., van der Ree, R., Mccarthy, M.A., 2017. Consistent patterns of vehicle collision risk
1361 for six mammal species. *J. Environ. Manage.* 201, 397–406.
- 1362 von Arx, M., Breitenmoser-Würsten, C., Zimmermann, F., Kunz, F., Vogt, K., Ryser, A., Struch,
1363 M., Breitenmoser, U., 2017. Der Luchs im Jura - unter besonderer Berücksichtigung des
1364 Solothurner Juras. *Naturforschende Gesellschaft des Kantons Solothurn* 43, 177–234.

- 1365 Wickham, H., RStudio, R Core Team, 2019. Package “testthat.”
- 1366 Wöfl, S., Anders, O., Middelhoff, T.L., Hohmann, U., Back, M., Idelberger, S., Krebühl, J.,
1367 Ohm, J., Prüssing, A., Herdtfelder, M., Böcker, F., Erretkamps, J., Kopaniak, L., Wöfl, M.,
1368 Jokisch, S., Hucht-Ciorga, I., Teubner, J., Trost, M., Zschille, J., Jeß, E., Steinberg, C., 2021.
1369 Status des Luchses in Deutschland. Natur und Landschaft 96.
- 1370
- 1371