

The rise of grasslands is linked to atmospheric CO₂ decline in the late Palaeogene

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ABSTRACT

1
2 Grasslands are predicted to experience a major biodiversity change by the year 2100 in part due to recent and projected
3 increases in atmospheric CO₂ concentration. A better understanding of how grasslands have responded to past environmental
4 changes will help predict the outcome of current and future environmental changes. Here, we explore the relationship between
5 past atmospheric CO₂ and temperature fluctuations and the shifts in diversification rate of grasses (Poaceae) and daisies
6 (Asteraceae), two exceptionally species-rich grassland families (~11,000 and ~23,000 species, respectively). To this end, we
7 developed a novel Bayesian approach that simultaneously estimates diversification-rates through time from time-calibrated
8 phylogenies and correlations between environmental variables and diversification rates. Additionally, we developed a new
9 statistical approach that incorporates the information of the distribution of missing species in the phylogeny. We found
10 strong evidence supporting a simultaneous increase in diversification rates for daisies and grasses after the most significant
11 reduction of atmospheric CO₂ in the Cenozoic (~34 Mya). The fluctuations of paleo-temperatures, however, appear not to have
12 had a significant relationship with the diversification of these grassland families. Overall, our results shed new light on our
13 understanding of the origin of grasslands in the context of past environmental changes.

14 Introduction

15 The grassland biome (steppes, savannas and prairies) covers vast areas of the Earth's surface and today accounts for as much as
16 one-third of the net primary production on land^{1,2}. Although grasses (Poaceae) comprise the bulk of the biomass and plant
17 population in grasslands, other plant families—in particular the daisies (Asteraceae)—are usually as much as (or even more)
18 diverse than grasses (Fig. S1). The evolution of grasslands marked the emergence of a new landscape and provided the substrate
19 for the adaptive radiation of other life forms that coevolved along with this biome, including grazing mammals³ such as horses,
20 wombats, and capybaras.

21 The age of a given biome is often estimated by detecting when particular representative taxonomic groups first appear in the
22 fossil record. For example, the early evolution of the grassland biome—and open-habitat biomes in general—has been estimated
23 from the fossil record of grass phytoliths (plant silica)⁴ or from the record of fossil pollen of daisies, grasses and amaranths^{5,6}.
24 Phylogenetic trees based on DNA sequence data calibrated with fossils provide a powerful new perspective on the history of
25 biomes⁷. This approach has been used to estimate the timing of tropical-rainforest evolution based on phylogenetic trees of
26 plant groups that are characteristic of this biome (e.g., Malpighiales⁸, Arecaceae⁹, and the legume genus *Inga*¹⁰). Nevertheless,
27 phylogenetic approaches have barely been used to study the evolutionary history of grassy biomes; most previous studies of
28 grassland evolution have focused on the origins of C₄ grasslands¹¹. Here we estimate when grasslands first expanded using
29 phylogenetic trees of its two primary plant families, Asteraceae and Poaceae. We assembled a large calibrated phylogenetic tree
30 for daisies and used the largest tree yet inferred for grasses¹¹ to explore temporal shifts in rates of lineage diversification, and to
31 test correlations between diversification-rate shifts and past climatic fluctuations.

32 A major limitation when analyzing hyper-diverse groups—in our case Asteraceae with ~23,000 species and Poaceae with
33 ~11,000 species—is the inevitable sparse species sampling (Figs. 1, 2). Although existing approaches for inferring rates

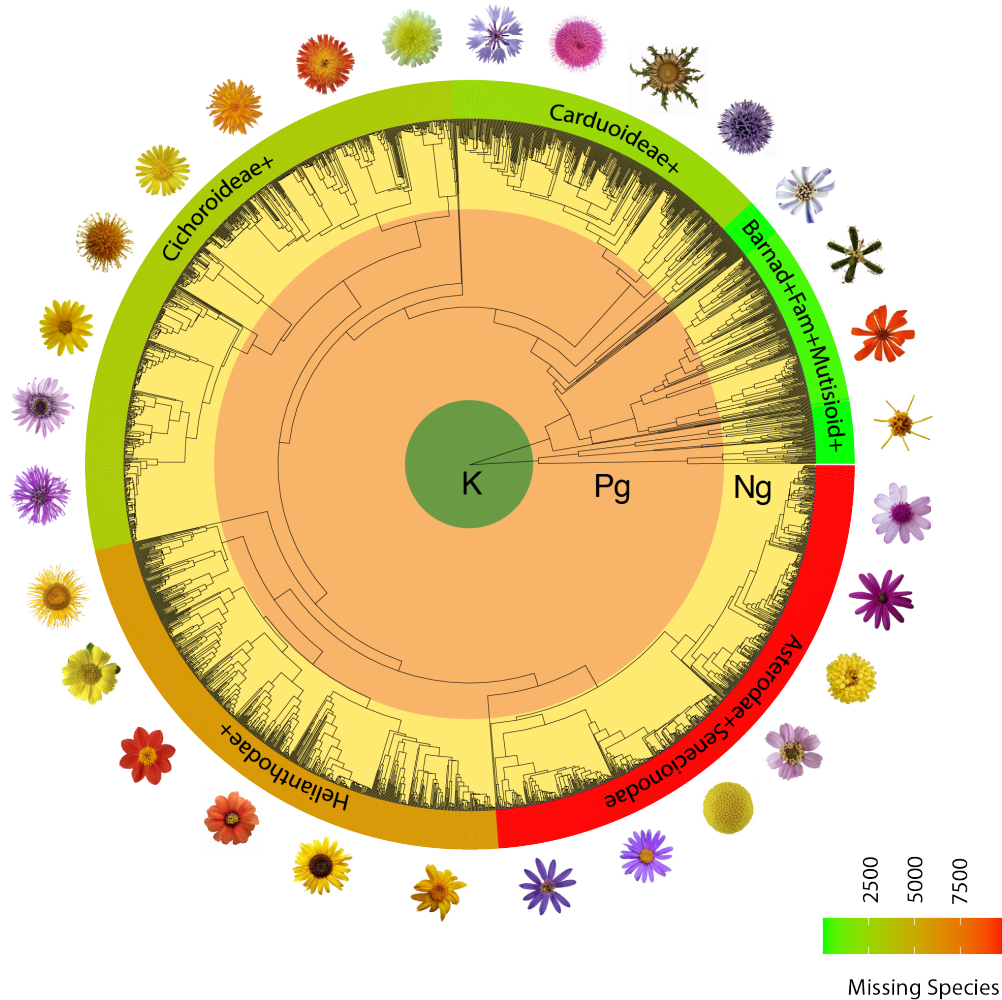


Figure 1. Phylogenetic tree scaled to geological time of Asteraceae with 2723 sampled tips. Asteraceae is one of the most species-rich families of flowering plants with more than 23,000 species. The number of non-sampled (missing) species increased enormously towards the more derived and specious lineages. For this reason, the sampling among clades is severely biased. Note that these rich and derived lineages evolved during the late Paleogene or early Neogene. K=Cretaceous, Pg=Paleogene, Ng=Neogene.

34 of lineage diversification (speciation and extinction) can accommodate incomplete species sampling^{12,13}, the distribution of
35 missing species on the tree in these approaches is modeled in a simplistic and somewhat unrealistic manner. Previous work has
36 shown that biased species sampling has a strong impact on diversification-rate estimates^{14–16}. We develop a novel Bayesian
37 approach for detecting diversification-rate shifts that incorporates a more realistic (non-uniform) model of species sampling
38 and implemented it in the open-source software RevBayes¹⁷. Our model builds on the episodic birth-death process, where
39 speciation and extinction rates are constant within an interval but may shift instantly to new rates at a rate-shift episode^{18–21}.
40 Furthermore, we tested for a correlation between diversification rate and two environmental variables —atmospheric CO₂
41 concentration and average global paleo-temperature— using one existing^{22–27} and three new environmentally-dependent
42 diversification models. We used an empirically informed and biologically realistic model to accommodate missing species that
43 assigns unsampled species to their corresponding clades using taxonomic information.

44 Results and Discussions

45 Our analyses demonstrate that the most dramatic increase in diversification rates in both Asteraceae and Poaceae (calibration
46 scenario #1, see Methods) occurred from the late Oligocene (~28 Mya) to the early Miocene (~20 Mya) (Fig. 3 and Fig. S5).
47 This diversification rate shift are robust to several model assumptions. We recovered the same diversification rate shifts
48 regardless of the assumed number of time intervals (Fig. S6). Both autocorrelated diversification rate prior models qualitatively
49 agree on the overall pattern of diversification rates (Gaussian Markov random field (GMRF) or Horseshoe Markov random
50 field (HSRMF), Fig. S5 and S6). Only the uncorrelated diversification rate prior model differed in the inferred pattern (UCLN,

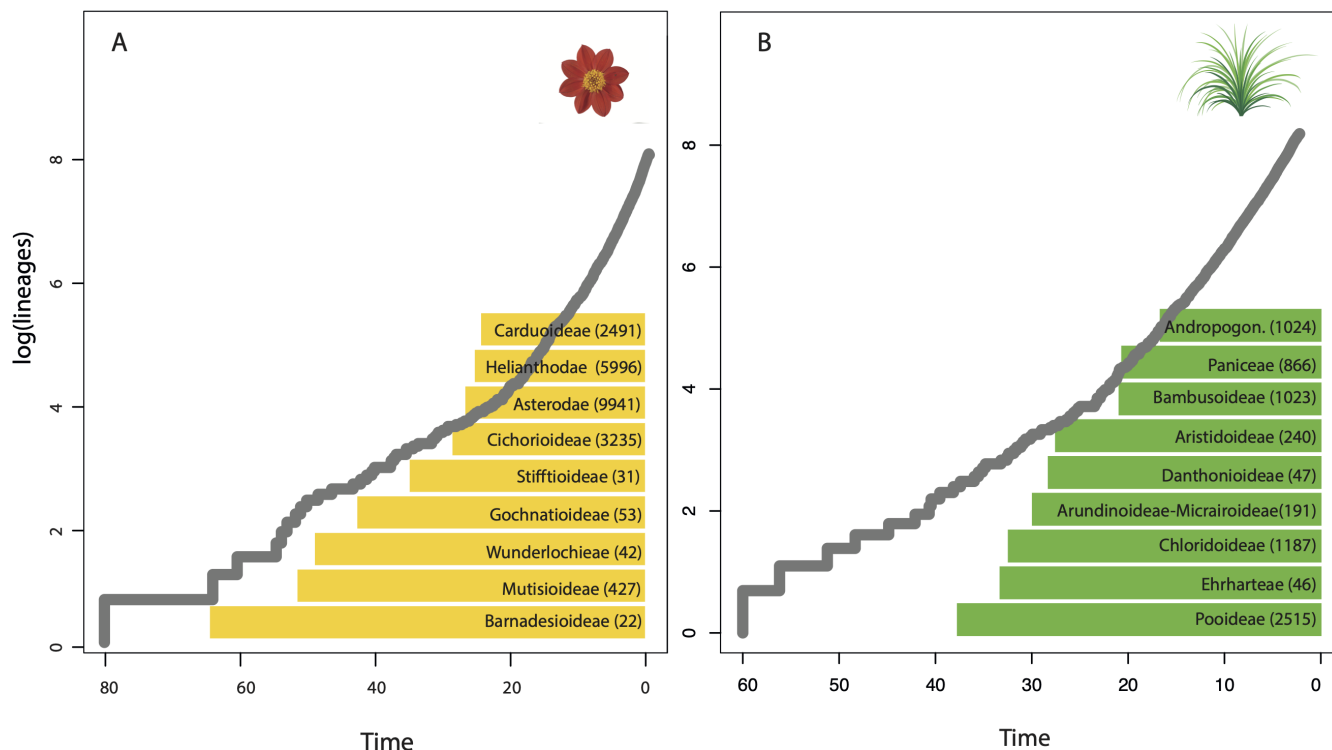


Figure 2. Lineage-Through Time (LTT) plots of Asteraceae (A) and Poaceae (B). Solid grey lines represent LTT curves derived from time-calibrated phylogenetic trees of Asteraceae and Poaceae (calibration scenario #1). Colored boxes depict the name and number of non-sampled (missing) species per clade that we integrated in our novel empirical taxon sampling. The shape of LTT curves have demonstrated to be a convenient summary metric for diversification diagnostics, particularly when diversification deviates from the expectation of constant rates^{28,29}. However, the distribution of missing species might not be uniform—as it is the case of these angiosperm families—and can severely impact on diversification-rate estimates. Our work shows that the most important increase in diversification rate for both Asteraceae and Poaceae is completely unnoticeable using the LTT analysis, even when calibrated phylogenetic trees include a large number of species.

51 Fig. S5 and S6). However, the autocorrelated diversification rate prior models were significantly favored according to our Bayes
 52 factor analyses (GMRP for the daisy phylogenetic tree and HSMRP for the grass phylogenetic tree, Fig. S7). The diversification
 53 rate patterns were strongly influenced by the assumed incomplete taxon sampling (Fig. S8). In our simulation study we show
 54 that incorrectly assuming uniform taxon sampling and thus disregarding taxonomic information about the distribution of missing
 55 species strongly biases diversification rates (Fig. S22). Conversely, our empirical taxon sampling informed by a more accurate
 56 distribution of missing species has good power to detect the correct time-varying diversification rates and low false-positive rate
 57 when diversification rates are in reality constant (Fig. S22). Thus, we recommend to include as much information as possible
 58 regarding the distribution of missing species.

59 The respective diversification rates of Asteraceae and Poaceae (calibration scenario #1, see Methods) peak between 20
 60 Mya and 15 Mya, and subsequently decreases for a brief period of time before increasing again from the late Miocene (~10
 61 Mya, Fig. 3 and Fig. S5). Our second analysis using the Poaceae phylogeny calibrated with a Cretaceous phytolith (calibration
 62 scenario #2) detects an earlier peak for Poaceae at about 30 Mya (Fig. S4). The phylogenetic placement of this fossil phytolith
 63 has been debated³⁰, thus this last result should be considered with caution. Our estimates of low diversification rates prior to
 64 the Oligocene is consistent with the scarcity of fossil forms assigned to both daisies (Table S1) and grasses^{4,31} known from
 65 this period. Similarly, our estimates of a high diversification rate in the late Oligocene and early Miocene is in line with the
 66 high diversity of fossil remains assigned to these groups^{4,32}. The Cenozoic 'temporal hotspot' of grassland diversification
 67 (~30 Mya to ~15 Mya) –based on daisies and grasses (calibration #1 and #2) phylogenetic trees– coincides with one of the
 68 most fundamental changes in global climate in the geologic record; a marked decline of atmospheric CO₂ occurred during the
 69 Oligocene (~34 Mya), reaching modern levels by the latest Oligocene^{33,34}. This scenario marks the onset of a cooler and more
 70 modern world (Coolhouse state), identified by the earliest Cenozoic glaciations in Antarctica, and the consequent drop in global
 71 paleotemperatures³⁵.

72 In line with the reconstructed climatic scenario, our analyses of correlation between diversification rates and CO₂ or
 73 paleo-temperature show very interesting results (Fig. 3 and S8). Diversification rates inferred from both the daisy and grasses
 74 phylogeny support correlation to CO₂ over paleo-temperature (Fig. S10). Surprisingly, the best fitting environmentally-

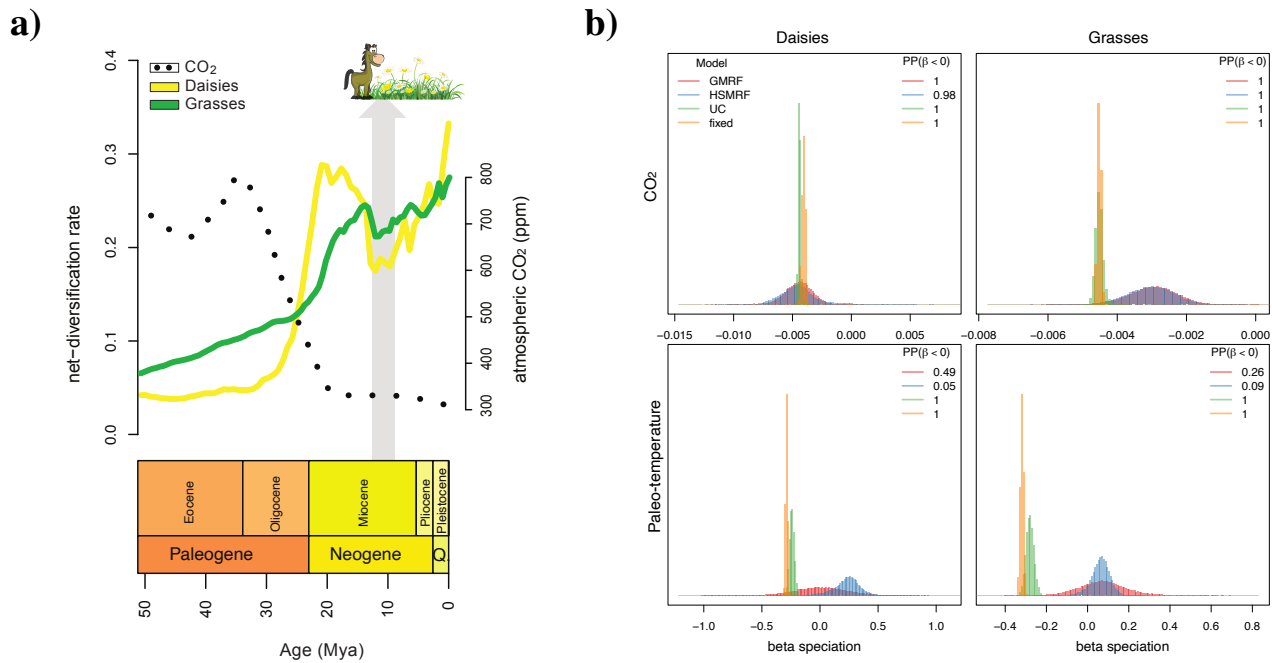


Figure 3. Estimating diversification rates and correlation to CO₂ and paleo-temperature. **a)** Diversification rates through time of daisies (yellow) and grasses (green) –calibration scenario #1– for the last 50 Mya. Dotted line represents atmospheric CO₂ fluctuations (22); note that the Oligocene steep decrease mirrors the onset of the increase in diversification of daisies and grasses. Grey arrow indicates a period (13-10 Mya) of lower diversification rates probably linked either with a brief increase in CO₂ (not represented in the dotted smoothed curve) or the explosive radiation of hypsodont grazers (*e.g.*, horses) and other mixed feeder grazers³ who may have had a tremendous impact on grasslands through their effects on plant populations and community composition. **b)** The correlation coefficient (β) between diversification rates and CO₂ concentrations for daisies and grasses is significantly negative (posterior probability of 1.0 for all models except the HSMRF, which has a posterior probability of 0.98; Bayes factors of 37,501 and 49 respectively). The support for a negative correlation between paleo-temperature and diversification rates is ambiguous; the UC and *fixed* models show significant support (posterior probability of 1.0, Bayes factor of 37,501) while the autocorrelated models show no support (posterior probability between 0.05 and 0.095, Bayes factors supporting a *positive* correlation of 1.04 and 17.86 for the daisy dataset and 2.85 and 9.75 for the grasses dataset for the GMRF and HSMRF models respectively).

75 dependent diversification model for the daisy phylogeny was the uncorrelated lognormal (UC) variation model and for the
 76 grasses phylogeny the *fixed* rate model without additional variation. The support of the uncorrelated model over the two
 77 autocorrelated models (GMRF and HSMRF), although the autocorrelated models were favored when using time-varying
 78 diversification rates without environmental variables (Fig. S7), could stem from the use of vague prior distribution which
 79 allows for more rate variation in autocorrelated models²¹. However, regardless of the specific environmentally-dependent
 80 diversification model, we inferred a negative correlation between diversification rates and environmental CO₂ (Fig. S9). The
 81 resulting Bayes factors for a negative correlation were decisive with values of 37,501 for the *fixed*, UC and GMRF models and
 82 49 for the HSMRF model (Fig. 3). We also see the same agreement between the four environmentally-dependent diversification
 83 models in our simulation study (Fig. S19 and S20). Thus, if there is a clear signal of correlation between the environmental
 84 variable and diversification rates, then our analyses appear robust to modeling of the additional component of time-varying
 85 diversification rates. This agreement can also be seen when all four environmentally-dependent diversification models show the
 86 same estimated diversification rates (Fig. S13-14). When the signal is less clear, as for the paleo-temperature analyses, the four
 87 models disagree and range from significant positive to significant negative correlation and the estimated diversification rates
 88 of the environmentally-dependent diversification models also differ (Fig. S13-14). Finally, our results of correlation between
 89 environmental CO₂ and diversification rates are also robust to the chosen epoch size (Fig. S11).

90 The negative correlation between diversification rates of these selected grassland families and atmospheric CO₂ might not
 91 be surprising; atmospheric CO₂—the main source of carbon for photosynthesis—serves as a fundamental substrate for plant
 92 growth. The available experimental evidence shows that low atmospheric CO₂ limits plant performance³⁶, although responses
 93 vary significantly between species. At a landscape scale, carbon limitation and water stress due to lower atmospheric CO₂

concentrations ('ecophysiological drought'), rather than water stress due to lower precipitation ('climatic drought'), cause changes in vegetation structure³⁷. During the Last Glacial Maximum (LGM; ~21,000 years ago), for example, atmospheric CO₂ was at its lowest concentration in the history of land plants (~180–200 ppm)³⁸. Models have predicted that the direct physiological impact of the of low CO₂ concentrations during the LGM drove the expansion of grasslands and dry shrublands at the expense of forest³⁹ (Fig. S2). Other modeling experiments indicate that low atmospheric CO₂, in combination with increased aridity and decreased temperatures, causes new xeric biomes to develop³⁸.

Although our primary hypothesis is that a CO₂-depleted atmosphere played a role in the geographic expansion and diversification of grassland families since the Oligocene, other environmental and biological variables could have also been involved. In particular, the decreasing temperatures, increasing aridity, and increasing seasonality of temperature and/or precipitation of the late Cenozoic have been traditionally linked to the early radiation of grasslands^{40,41}. The role of cooling in the emergence of open-habitat grasses has been debated as the adaptation to low temperatures became prominent in the more derived groups of grasses^{4,42}. Grazing mammals have been also important components in the evolution of grasslands; grazers and grassland ecosystems probably coevolved over millions of years⁴³. Grazing increased species diversity according to experimental studies, as grazers prevent dominant plant species from monopolizing resources. Without grazing, tall, vegetatively reproducing plant species increase in cover and shade out short and sexually reproducing species⁴⁴. Grazing also affects the flux of nutrients by accelerating the conversion of plant nutrients from forms that are unavailable for plant uptake to forms that can be readily used. Overall, grazing mammals have an important role in the diversity of present-day natural grasslands and we assume they might have done so during their early radiation. However, the explosive radiation of true hypsodonts may have negatively impacted grasslands' distribution and diversity (see below). Sorting out the relative importance of all these environmental and biological competing forces from the hypothesized CO₂-induced shift is challenging.

We detected a short decrease in diversification rates for daisy and grass plant groups during the mid-Miocene, about 13-10 Mya (Fig. 3). The causal mechanism underlying remains to be elucidated. However, we suspect that the dramatic radiation of hypsodont grazers –such as horses– and other mixed feeder grazers may have had an impact on grasslands^{3,45}. Since the late Miocene (~10 Mya), however, the more recent expansion of C₄ grass lineages¹¹ may have contributed to the increased diversification rates in these groups. Plants using the C₄ photosynthetic pathway have anatomical and biochemical adaptations for concentrating CO₂ within leaf cells prior to photosynthesis, which may lead to a selective advantage over C₃ plants under conditions of low atmospheric CO₂. Although the evolutionary origin of C₄ photosynthesis in grasses most likely occurred early in the Cenozoic³⁰, their expansion and ecological dominance may have taken place during the last 10 Mya, by the late Miocene in warmer and fire-prone landscapes of the world⁴⁶. Likewise, the evolution of hyper-diverse Asteraceae lineages (e.g. *Senecio*)⁴⁷ have also contributed to the increasing rates of diversification since the last 10 Mya. Our evidence also supports the notion that the ongoing rise of atmospheric CO₂ will likely altered vegetation distributions through differential effects on C₃ and C₄ plant types. In fact, modelling future distributions predicts the near-complete eradication of C₄ species across the globe for the next 50 year⁴⁸; this implies that about half of the species in the grass family will be extinct. In summary, our study reveals episodic shifts in diversification rates of grasses and daises which are correlated with changes in atmospheric CO₂ (Fig. 3); these insights are made possible by the development of our new Bayesian phylogenetic approach which combines the episodic birth-death process^{18–21} with environmentally-dependent diversification rates^{22–24,26} and empirical taxon sampling^{15,16,49}. Our environmentally-dependent and episodic birth-death diversification model provides a novel approach for exploring the evolution of hyper-diverse groups of plants and animals in the context of historical environmental changes.

Material and Methods

Grassland diversity. To quantify the taxonomic representativeness of vascular-plant families found in open-habitat landscapes (Fig. S1), we selected seven distantly distributed eco-regions dominated by grasslands from the World Wide Fund for Nature (WWF)⁵⁰. Using the coordinate boundaries of each of the selected eco-regions, we extracted the vascular plant taxa (=Tracheophyta) from the Global Biodiversity Information Facility (GBIF), using the R 'RGBIF' package⁵¹ with the option "hasGeospatialIssue=FALSE", that includes only records without spatial issues. Plant families were sorted according to the number of species, removing duplicated species.

Palaeobotanical analysis. Asteraceae and Poaceae have a fairly similar fossil record; their oldest findings are known from the Late Cretaceous —which mainly comprise microscopic remains (that is, phytoliths⁵² or pollen grains⁵³)— whereas the first indisputable macroscopic Asteraceae and Poaceae fossils are first known from the Eocene^{54,55}, with a substantial increase of diversity since since the Oligocene/Miocene. While the fossil record of Poaceae has been fully revised^{4,56,57}, the fossil record of Asteraceae has not been as carefully reviewed. We compiled published pollen and macroscopic fossil data for Asteraceae including all fossil species assigned to Asteraceae (Table S1). The earliest record of the Asteroideae (the clade that includes the most common open-habitat daisy tribes) occurs since the Late Oligocene of New Zealand but in very low frequencies. Fossils refer to this subfamily increased in abundance and diversity during the Miocene and Pliocene. Pollen referred to *Artemisia*, in

147 particular, did not become abundant until the Middle-Late Miocene with several reports from central Europe, Asia and North
148 America. Pre-Miocene findings need further verification. Overall, the Late Oligocene and in particular the Miocene witnessed
149 the major step in the diversification of Asteraceae; ca. 80% of the fossil species recorded have been assigned to this time
150 interval.

151 **Divergence-Time estimation.** To construct the Asteraceae supertree (2,723 tips), we first inferred a backbone chronogram
152 using 14 plastid DNA regions from 54 species, including representatives of all 13 subfamilies, with an additional four species
153 of Calyceraceae used as outgroup taxa (Table S2). Sequences were compiled from GenBank and each region was aligned
154 separately using MAFFT⁵⁸ with the options maxiterate 1000 and localpair. Two fossil constraints were applied: (i) a macrofossil
155 (capitulum) and associated pollen (*Raiguenrayun cura* + *Mutisiapollis telleriae*) from the Eocene (45.6 Mya) to calibrate
156 the non-Barnadesioideae Asteraceae clade⁵⁵ and; (ii) the fossil-pollen species *Tubulifloridites lilliei* type A from the late
157 Cretaceous (72.1 Mya)⁵³ to calibrate the crown Asteraceae (considering *T. lilliei* as a stem group, see Huang *et al.*⁵⁹ for further
158 discussion). Divergence-time estimates and phylogenetic relationships were inferred using RevBayes¹⁷. For the aligned
159 molecular sequences we assume a general-time reversible substitution model with gamma-distributed rate variation among
160 sites (GTR+ Γ), an uncorrelated log-normal prior on substitution-rate variation across branches (UCLN relaxed clock), and a
161 birth–death prior model on the distribution on node ages/tree topologies. A densely sampled phylogeny is crucial to identify
162 shifts in diversification rates. Therefore, we constructed a supertree by inserting eleven individual sub-trees —representing
163 all subfamilies of the Asteraceae except those less diverse or monotypic clades (that is, Gymnarrhenoideae, Corymbioideae,
164 Hecastocleidoideae, Pertyoideae)— into the calibrated backbone chronogram. This method follows a previous study that
165 constructed a supertree of grasses using the same approach¹¹. Each of the eleven clades of Asteraceae was built using their own
166 set of markers and the same phylogenetic approach as the one used to infer the backbone tree (Table S2). Sequence data for
167 each of the eleven trees and their respective outgroup taxa were collected from Genbank using the NCBIminer tool⁶⁰. The
168 estimated ages of the nodes given by the backbone analysis were used to constrain the age of each of the eleven sub-trees
169 (Table S2). Divergence-time estimates and phylogenetic relationships for each of the eleven sub-clades were estimated using
170 RevBayes as described above. The eleven trees were grafted onto the backbone tree using the function ‘paste.tree’ from the
171 `phytools` R package⁶¹. We used `GGTREE` R package⁶² to plot the circle phylogenetic tree of Figure 1 and `phytools`⁶¹
172 to include the concentric geological scale. The supertree of the grass family (3,595 taxa) was obtained from Spriggs *et al.*¹¹
173 (Table S3). They inferred two chronograms using two different calibration scenarios, that is, a younger scenario (#1) calibrated
174 using an Eocene megafossil⁵⁴ and an older scenario (#2) calibrated using Cretaceous phytoliths⁵². We run our diversification
175 analyses using these two chronograms.

176 **Inferring Changes in Diversification Rate Through Time.** Our species-diversification model is based on the *reconstructed*
177 *evolutionary process* described by Nee *et al.*¹² and more specifically on the episodic birth-death process^{18–21}. We assume that
178 each lineage gives birth to another species with rate λ (cladogenetic speciation events) and dies with rate μ (extinction event;
179 see Figure 4). We model diversification rates (i.e., speciation and extinction rates) as constant within an interval but independent
180 between intervals, where intervals are demarcated by instantaneous rate-shift events. We denote the vector of speciation rates
181 $\Lambda = \{\lambda_1, \dots, \lambda_k\}$ and extinction rates $\mathbf{M} = \{\mu_1, \dots, \mu_k\}$ where λ_i and μ_i are the (constant) speciation and extinction rates in
182 interval i . Additionally, we use the taxon sampling fraction at the present denoted by ρ ^{15,16}. Following the notation of May *et al.*²⁰,
183 we construct a unique vector, \mathbb{X} , that contains all divergence times and rate-shift event times sorted in increasing order. It
184 is convenient for notation to expand the vectors for all the other parameters so that they have the same number of elements
185 $k = |\mathbb{X}|$. Let Ψ denote an inferred tree relating n species, comprising a tree topology, τ , and the set of branching times, \mathbb{T} . We
186 use the notation $S(2, t_1=0, T)$ to represent the survival of two lineages in the interval $[t_1, T]$, which is the condition we enforce
187 on the reconstructed evolutionary process. Transforming Equation (A4) in May *et al.*²⁰ to our model yields the probability
188 density of a reconstructed tree as:

$$\begin{aligned}
 & f(\Psi|N(t_1=0)=2, S(2, t_1=0, T)) \\
 &= \frac{2^{n-1}}{n!} \\
 & \times \left(1 + \sum_{i=0}^k \left(\frac{\mu_i}{\mu_i - \lambda_i} \times e^{\sum_{j=0}^{i-1} (\mu_j - \lambda_j)(x_{j+1} - x_j)} \times \left(e^{(\mu_i - \lambda_i)(x_{i+1} - x_i)} - 1 \right) \right) - \frac{\rho - 1}{\rho} \times e^{\sum_{i=0}^k (\mu_i - \lambda_i)(x_{i+1} - x_i)} \right)^{-2} \times \left(e^{-\log(\rho) \sum_{j=0}^k (\mu_j - \lambda_j)(x_{j+1} - x_j)} \right)^2 \\
 & \times \prod_{i \in \mathbb{I}_T} \left[\lambda_i \times \left(1 + \sum_{l=1}^k \left(\frac{\mu_l}{\mu_l - \lambda_l} \times e^{\sum_{j=0}^{l-1} (\mu_j - \lambda_j)(x_{j+1} - x_j)} \times \left(e^{(\mu_l - \lambda_l)(x_{l+1} - x_l)} - 1 \right) \right) - \frac{\rho - 1}{\rho} \times e^{\sum_{i=1}^k (\mu_i - \lambda_i)(x_{i+1} - x_i)} \right) \right]^{-2} \quad \text{The first term, } \frac{2^{n-1}}{n!} \\
 & \times e^{-\log(\rho) \sum_{j=i}^k (\mu_j - \lambda_j)(x_{j+1} - x_j)}. \tag{1}
 \end{aligned}$$

190 corresponds to the combinatorial constant for the number of labelled histories¹⁸, the second term corresponds to the condition
 191 of two initial lineage at the root of the phylogeny surviving until the present, and the third term corresponds to the product of all
 192 speciation events and the new lineages surviving until the present.

193 **Empirical taxon-sampling model.** Here we develop an *empirical* taxon sampling model that uses taxonomic information
 194 on the membership of unsampled species to clades and speciation times of unsampled species, which is an extension to the
 195 work by Höhna *et al.*^{15, 16} and similar to the approach used by Stadler and Bokma⁴⁹. The main difference of our approach and
 196 the approach by Stadler and Bokma⁴⁹ is that their model uses a constant-rate birth-death process (compared to our episodic
 197 birth-death process). Additionally, Stadler and Bokma⁴⁹ derive the density of the missing species using a random probability s
 198 of an edge being sampled, which differs from our approach where we integrate over the time of the missing speciation event.
 199 Nevertheless, at least for the constant-rate birth-death process, both approaches arrive at the same final likelihood function.

We include information on the missing speciation events by integrating over the known interval when these speciation events must have occurred (that is, between the stem age t_c of the MRCA of the clade and the present). This integral of the probability density of a speciation event is exactly the same as one minus the cumulative distribution function of a speciation event¹⁶,

$$F(t_c|N(t_1) = 1, t_1 \leq t \leq T) = 1 - \frac{1 - P(N(T) > 0|N(t_c) = 1) \exp(r(t_c, T))}{1 - P(N(T) > 0|N(t_1) = 1) \exp(r(t_1, T))}, \tag{2}$$

200 where t_1 is the age of the root. The probability of survival is given by:

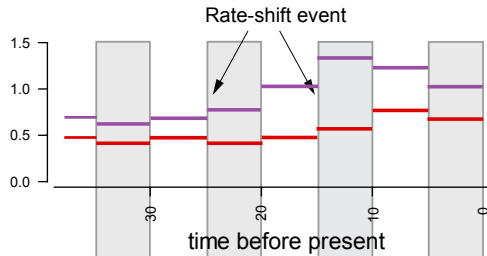
$$\begin{aligned}
 & P(N(T) > 0|N(t_c) = 1) \\
 &= \left(1 + \sum_{i=c}^k \left(\frac{\mu_i}{\mu_i - \lambda_i} \times e^{\sum_{j=c}^{i-1} (\mu_j - \lambda_j)(x_{j+1} - x_j)} \times \left(e^{(\mu_i - \lambda_i)(x_{i+1} - x_i)} - 1 \right) \right) - \frac{\rho - 1}{\rho} \times e^{\sum_{i=c}^k (\mu_i - \lambda_i)(x_{i+1} - x_i)} \right)^{-1} \tag{3}
 \end{aligned}$$

201 where $k = |\mathbb{X}|$. Let us define n as the number of sampled species, m as the total number of species in the study group, \mathbb{K} as the
 202 set of missing species per clade and $|\mathbb{K}|$ the number of clades with missing species. Additionally, we define c_i as the time of
 203 most recent common ancestor of the i^{th} clade. Then, the joint probability density of the sampled reconstructed tree and the
 204 empirically informed missing speciation times is

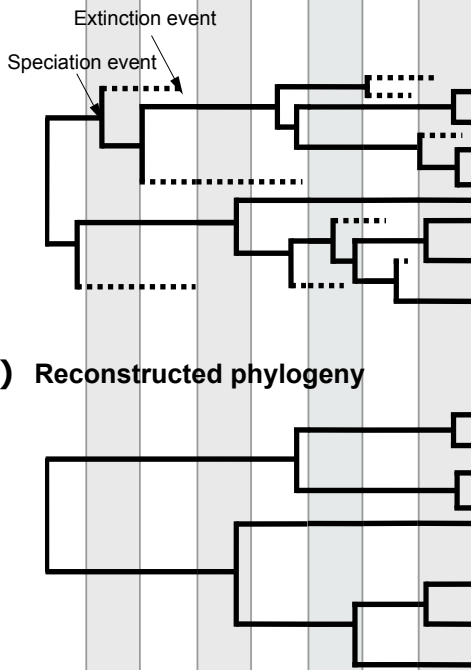
$$\begin{aligned}
 f(\Psi, \mathbb{K}|N(t_1=0)=2, S(2, t_1=0, T)) &= f(\Psi|N(t_1=0)=2, S(2, t_1=0, T)) \\
 & \times \frac{(m-1)!}{(n-1)!} \prod_{i=1}^{|\mathbb{K}|} \frac{1}{k_i!} (1 - F(t|N(c_i) = 1, c_i \leq t \leq T))^{k_i} \tag{4}
 \end{aligned}$$

205 **Prior models on diversification rates.** Our model assumes that speciation and extinction rates are piecewise constant but
 206 can be different for different time intervals (Figure 4). Thus, we divide time into equal-length intervals (e.g., $\Delta t=1$). Following
 207 Magee *et al.*²¹, we specify prior distributions on the log-transformed speciation rates ($\ln(\lambda_i)$) and extinction rates ($\ln(\mu_i)$)

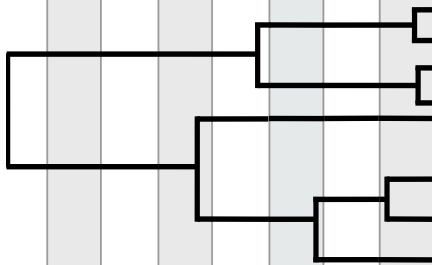
A) Diversification rates through times



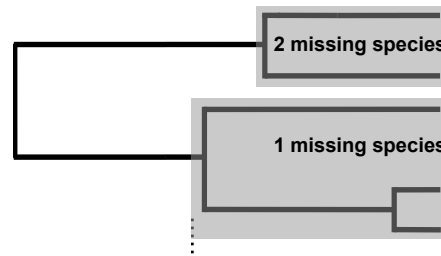
B) Complete phylogeny



C) Reconstructed phylogeny



D) Sampled phylogeny with missing species



E) Distribution function of missing speciation event

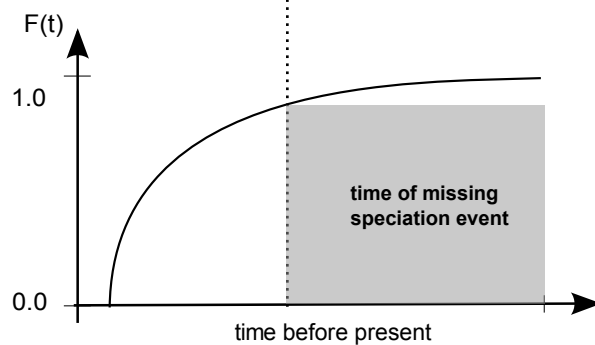


Figure 4. Cartoon of the birth-death process with rate-shift events and empirical taxon sampling. A) Depiction of the speciation (purple lines) and extinction (red lines) rates through time. Here we assume that speciation and extinction rates are episodically constant, that is, diversification rates shift instantly and only at the beginning of an episode. Each episode lasts 5 time units in this example. B) A realization (complete phylogeny) of the birth-death process. Lineages that have no extant or sampled descendant are shown as dashed lines and surviving lineages are shown as solid lines. C) Reconstructed phylogeny corresponding exactly to the one shown in B with the extinct lineages pruned away. Thus, plot C depicts the “observed” phylogeny from which the speciation times are retrieved. D) Sampled phylogeny with gray boxes depicting named clades with known number of missing species. The phylogeny is the same as in C with fewer taxa. E) Distribution function of the time of the missing speciation event. The missing speciation event could have occurred any time between the crown age of the named clade and the present time (gray box). The distribution function is integrated over and hence the uncertainty of the missing speciation event accounted for.

208 because the rates are only defined for positive numbers and our prior distributions are defined for all real numbers. We apply
209 and compare three different prior models: (i) an uncorrelated log-normal (UCLN) prior distribution, (ii) a Gaussian Markov
210 random field (GMRF) prior²¹, and (iii) a Horseshoe Markov random field (HMRF) prior²¹. The first prior distribution specifies
211 temporally uncorrelated speciation and extinction rates, whereas the second and third prior distributions are autocorrelated
212 prior models. The assumption of autocorrelated rates might make more sense biologically (an interval of high speciation
213 rates is likely to be followed by another interval with high speciation) but also improves our ability to estimate parameters²¹.
214 Nevertheless, our inclusion of both uncorrelated and autocorrelated prior distributions allows for testing whether an uncorrelated
215 or autocorrelated model is preferred.

216 The prior distribution on the speciation rates λ_i and extinction rates μ_i are set in exactly the same form in our models with
217 their respective hyperprior parameters. Thus, for the sake of simplicity, we omit the prior distribution on the extinction rates
218 here in the text. Our first prior distribution, the uncorrelated log-normal (UCLN) distributed prior, specifies the same prior
219 probability for each speciation rate λ_i ,

$$\ln(\lambda_i) \sim \text{Normal}(m, \sigma) \quad . \quad (5)$$

220 Thus, each speciation rate is independent and identically distributed.

221 Our second, prior distribution, the Gaussian Markov random field (GMRF) prior, models rates in an autocorrelated form
222 analogous to a discretized Brownian motion. That is, we assume that diversification rates $\lambda(t)$ and $\mu(t)$ are autocorrelated and
223 the rates in the next time interval will be centered at the rates in the current time interval,

$$\ln(\lambda_i) \sim \text{Normal}(\ln(\lambda_{i-1}), \sigma_\lambda) \quad . \quad (6)$$

224 The standard deviation σ regulates the amount of change between each time interval.

225 Our third prior distribution, Horseshoe Markov random field (HSMRF) prior, is very similar to the GMRF but additionally
226 allows for the variance to change between time intervals,

$$\gamma_i \sim \text{halfCauchy}(0, 1) \quad (7)$$

$$\ln(\lambda_i) \sim \text{Normal}(\ln(\lambda_{i-1}), \sigma\gamma_i) \quad . \quad (8)$$

227 The HSMRF prior model is more adaptive than the GMRF; it allows for more extreme jumps between intervals while
228 favoring/smoothing more constant rate trajectories if there is no evidence for rate changes.

229 These three prior models of diversification rates provide the null models of our analyses as it does not assume any
230 dependence to an environmental variable. We use this model first to estimate diversification rates through time before testing
231 for a correlation of the speciation or extinction rate to an environmental variable (e.g., atmospheric CO₂ or paleo-temperature).
232 Magee *et al.*²¹ found that 100 epochs perform well for autocorrelated models. Since we do not know how many bins (i.e.,
233 epochs) should be used for the episodic birth-death process, we test various numbers of equal-sized epochs (4, 10, 20, 50, 100
234 and 200). We show both the mean posterior diversification rates (Fig. S5) as well as select the best fitting model based on the
235 number of epochs (Fig. S7).

236 **Correlation between speciation and extinction rate to CO₂.** Previously, Condamine *et al.*²² introduced an environmentally-
237 dependent diversification model. In their model, diversification rates are correlated with an environmental variable^{22–27}.
238 For example, the speciation rate can be modeled as $\lambda(t) = \lambda_0 e^{\beta \times \text{CO}_2(t)}$ (see Box 1 in Condamine *et al.*²²), which is
239 equivalent to $\ln(\lambda(t)) = \ln(\lambda_0) + \beta \times \text{CO}_2(t)$. Since we are using the episodic birth-death process which has piecewise-
240 constant diversification rates, we modify the original continuous-time environmentally-dependent diversification model to
241 $\ln(\lambda_i) = \ln(\lambda_0) + \beta \times \text{CO}_{2,i}$, which is equivalent to and more conveniently written as $\ln(\lambda_i) = \ln(\lambda_{i-1}) + \beta \times \Delta\text{CO}_{2,i}$ where
242 $\Delta\text{CO}_{2,i} = \text{CO}_{2,i} - \text{CO}_{2,i-1}$. Note that we only use the so-called exponential dependency and not the linear dependency²⁴
243 because the linear dependency can result in negative rates which are mathematically and biologically impossible⁶³.

244 We applied this original environmentally-dependent diversification model and three new environmentally-dependent diversi-
245 fication models. The original environmentally-dependent diversification model of Condamine *et al.*²² does not accommodate
246 diversification-rate variation that is independent of the environmental variable. Instead, our three new environmentally-
247 dependent diversification models build on our diversification-rate prior models which allow for rate-variation through time (see
248 above). Thus, our environmentally-dependent diversification models will collapse to the episodic birth-death model if rates of
249 diversification and atmospheric CO₂ are uncorrelated and hence inherently allows for diversification rate variation. The linkage
250 of environmental variable and diversification rates without allowing for independent diversification rate variation might provide
251 spurious results, as has been noticed for trait evolution⁶⁴ and state-dependent diversification rates⁶⁵. We explore this potential
252 of misattribution of diversification rate variation to the environmental variable in our model selection procedure and simulation
253 study (see below).

254 As before, we omit the description of the extinction rates in the text for the sake of notational simplicity. Both speciation
 255 and extinction rates are model exactly in the same way with their corresponding set of hyperparameters (e.g., see the Tables
 256 S4-S7). Our first environmentally-dependent diversification model has a *fixed* linkage between the diversification rate variation
 257 and variation in the environmental variable;

$$\lambda_0 \sim \text{Uniform}(0, 100) \quad (9)$$

$$\ln(\lambda_i) = \ln(\lambda_{i-1}) + \beta_\lambda \times \Delta\text{CO}_2 \quad (10)$$

258 This model does not have a counterpart in the above diversification rate priors, but is included as a comparison to the work
 259 Condamine *et al*²².

260 Our second environmentally-dependent diversification model adds uncorrelated lognormal variation on top of the variation
 261 in the environmental variable;

$$\lambda_0 \sim \text{Uniform}(0, 100) \quad (11)$$

$$\ln(\hat{\lambda}_i) = \ln(\hat{\lambda}_{i-1}) + \beta_\lambda \times \Delta\text{CO}_2 \quad (12)$$

$$\varepsilon_i \sim \text{Normal}(0, \sigma) \quad (13)$$

$$\ln(\lambda_i) = \ln(\hat{\lambda}_i) + \varepsilon_i \quad (14)$$

262 Thus, this model collapses to the above UCLN model if there is no correlation between the environmental variable and
 263 diversification rates ($\beta = 0$). Importantly, the difference in the variation of the diversification rates and environmental variable
 264 is independent in each epoch, contributed by the variable ε_i . The environmental-dependent part of the diversification rates $\hat{\lambda}_i$ is
 265 equivalent to the *fixed* environmentally-dependent diversification model.

266 Our third environmentally-dependent diversification model adds correlated lognormal variation on top of the *fixed*
 267 environmentally-dependent diversification mode;

$$\lambda_0 \sim \text{Uniform}(0, 100) \quad (15)$$

$$\ln(\lambda_i) \sim \text{Normal}(\ln(\lambda_{i-1}) + \beta_\lambda \times \Delta\text{CO}_2), \sigma) \quad (16)$$

268 This model an extension of the above GMRF model and collapses to it if there is no correlation between the environmental
 269 variable and diversification rates ($\beta = 0$). As the GMRF model is a discretized Brownian motion model, this environmentally-
 270 dependent extension can be considered as a Brownian motion with trend model, where the trend is predicted by the environmental
 271 variable. Instead of writing this model with a separate environmentally-dependent part $\hat{\lambda}_i$ and autocorrelated part ε_i , we directly
 272 use the combined environmentally-dependent and independent rate variation as the mean for the next time interval. Nevertheless,
 273 we want to emphasize this equivalence to bridge the connection to the UCLN model above.

274 Finally, our fourth environmentally-dependent diversification model extends the above HSRMF to allow for diversification
 275 rates predicted by the environmental variable;

$$\lambda_0 \sim \text{Uniform}(0, 100) \quad (17)$$

$$\gamma_i \sim \text{halfCauchy}(0, 1) \quad (18)$$

$$\ln(\lambda_i) \sim \text{Normal}(\ln(\lambda_{i-1}) + \beta_\lambda \times \Delta\text{CO}_2), \sigma\gamma_i) \quad (19)$$

276 This model follows the same extension as the environmentally-dependent GMRF model with local adaptability of the rate
 277 variation through the parameter γ_i , as before for the HSMRF.

278 In all our four models, we denote the correlation coefficient by β . If $\beta > 0$ then there is a positive correlation between the
 279 speciation rate and CO_2 , that is, if the CO_2 increases then the speciation rate will also increase. By contrast, if $\beta < 0$ then there
 280 is a negative correlation between the speciation rate and CO_2 , that is, if the CO_2 concentration increases then the speciation rate
 281 will decrease. Finally, if $\beta = 0$ then there is no correlation and our environmentally-dependent diversification model collapses
 282 to corresponding episodic birth-death model.

283 All four models have the same parameter for the initial speciation rate λ_0 with a uniform prior distribution between 0
 284 and 100. The models are constructed in increasing complexity and all three new models can collapse either to the *fixed*
 285 environmentally-dependent diversification model or to their environmentally-independent episodic birth-death process.

286 **Environmental Data.** In our analyses we tested for correlation between two environmental factors: CO_2 and temperature.
 287 The concentration of atmospheric CO_2 throughout the Cenozoic were compiled by Beerling & Royer³³ using terrestrial and
 288 marine proxies. An updated dataset was provided by Dr. Dana Royer. Paleo-temperature fluctuations come from Zachos et al.
 289 (2001)⁶⁶. Raw data were extracted from <ftp://ftp.ncdc.noaa.gov/pub/data/paleo/>.

290 Analogous to our tests about the number of epochs for the diversification rate analyses, we computed the arithmetic mean for
 291 the environmental variable for 1-, 2- and 5-million year intervals. We both estimated the correlation between the environmental
 292 variable and diversification rates for each interval size and performed model selection using Bayes factors.

Model Selection. We performed three sets of empirical diversification rate analyses for each dataset. We estimated the diversification rates over time using three different models, we estimated the environmentally-dependent diversification rates using four different models, and we applied two different taxon sampling schemes. For the first two sets of analyses we performed standard model selection in a Bayesian framework using Bayes factors⁶⁷. Thus, we computed the marginal likelihood for each model using stepping-stone sampling⁶⁸ as implemented in `RevBayes`. We run 128 stepping stones with each stone comprising of its own MCMC run with 2,000 iteration and on average 1,374 moves per iteration (i.e., the runs being equivalent to standard single-move-per-iteration software with 2,748,000 iterations).

We tested the support for the environmental correlation using Bayes factors computed from the posterior odds. Our prior probability for the correlation coefficient β was symmetric and centered at zero, that is, we specified exactly a probability of 0.5 that $\beta < 0$ and $\beta > 0$. Thus, the prior probability ratio of $\frac{P(\beta < 0)}{P(\beta > 0)} = 1.0$. Then, to compute the Bayes factor for in support of a negative correlation is simply the number of MCMC samples with $\beta < 0$ divided over the total number of MCMC samples.

We did not compute marginal likelihoods for the two different sampling schemes; the uniform taxon sampling and the empirical taxon sampling. Empirical taxon sampling uses additional data, the age ranges of the missing speciation events, and two analyses with different data cannot be compared using traditional model selection. Instead, we performed a simulation study to show the robustness of our parameter estimates under empirical taxon sampling and the resulting bias if wrongly uniform taxon sampling was assumed.

Simulation Study. We performed two sets of simulations; focusing (a) on the environmentally-correlated diversification model, and (b) the incomplete taxon same scheme. First, we simulated phylogenies under the UCLN and GMRF environmentally-correlated diversification model using the R package `TESS`^{69,70}. We set the diversification rate variation to $\sigma = \{0, 0.02, 0.04\}$ and correlation coefficient to $\beta = \{0, -0.005, -0.01\}$. Thus, our simulations included the constant-rate birth-death process (when $\sigma = 0$ and $\beta = 0$), time-varying but environmentally independent diversification rates (when $\sigma > 0$ and $\beta = 0$), the *fixed* environmentally-dependent diversification model (when $\sigma = 0$ and $\beta \neq 0$), and the time-varying and environmentally-dependent diversification model (when $\sigma > 0$ and $\beta \neq 0$). For each setting, we simulated ten diversification rate trajectories (Figure S14 and S15) and trees (Figure S12 and S13). We analyzed each simulated tree under the same four environmentally-dependent diversification model as in our empirical analysis (see above).

Second, we simulated phylogenetic trees under empirical taxon sampling to validate the correctness of our model derivation. Unfortunately, simulation of empirical taxon sampling is not straight forward. We circumvented the problem by randomly adding the missing species to the daisy phylogenetic tree, then drawing new divergence times under (a) a constant-rate birth-death process, and (b) a time-varying episodic birth-death process with rates taken from the empirical estimates. Then, we pruned the additional species to mimic empirical taxon sampling. The simulations under the constant-rate birth-death process provide information about falsely inferring diversification rate variation (false positives) and the simulations under the time-varying episodic birth-death process provide information about the power to correctly inferring diversification rate variation (power analysis). We simulated 100 trees under each setting and analyzed each tree using the GMRF prior model with both empirical and uniform taxon sampling. The MCMC inference settings were identical to the empirical analyses.

Software Implementation and Availability. Both models, the episodic birth-death process and the environmentally-dependent diversification model, are implemented in the Bayesian phylogenetics software `RevBayes`¹⁷. Moreover, the implementation is not restricted to the models we introduce here because `RevBayes` is built on the principle of probabilistic graphical models⁷¹. The graphical model approach provides full flexibility to extend or modify the current analyses to other models and assumption, for example, testing for correlation to multiple environmental variables. `RevBayes` is open-source and freely available from <https://github.com/revbayes/revbayes>. The analysis from this paper are described in detail in several tutorials available at <http://revbayes.github.io/tutorials.html>.

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