1	
2	
3	Body size mediates latitudinal population differences in response to Bd infection in two
4	amphibian species.
5	
6	
7	Sara Meurling ¹ *, Maria Cortazar-Chinarro ^{1,4,5} , Mattias Siljestam ¹ , David Åhlen ² , Erik Ågren ³ ,
8	Jacob Höglund ¹ and Anssi Laurila ¹
9	
10	¹ Animal Ecology/ Department of Ecology and Genetics, Uppsala University, Sweden
11	² Department of Ecology, Environment and Plant Sciences, Stockholm University, Sweden
12	³ Department of Pathology and Wildlife Diseases, National Veterinary Institute, Uppsala,
13	Sweden
14	⁴ MEMEG/Department of Biology, Lund University, Lund, Sweden
15	⁵ Department of Earth, Ocean and Atmospheric Sciences, University of British Columbia,
16	Vancouver, Canada
17	

18 *Corresponding author: anssi.laurila@ebc.uu.se

19 Abstract

20 Populations of the same species may differ in their sensitivity to pathogens but the factors 21 behind this variation are poorly understood. Moreover, infections may cause sub-lethal fitness 22 effects even in species resistant or tolerant to disease. The chytrid fungus Batrachochytrium 23 *dendrobatidis (Bd)*, is a generalist pathogen which has caused amphibian population declines 24 worldwide. In many species, *Bd* infection causes the disease chytridiomycosis, often leading 25 to high mortality. We investigated how geographical origin affects tolerance to Bd by 26 exposing newly metamorphosed individuals of two North European amphibians (moor frog 27 Rana arvalis, common toad Bufo bufo) from two latitudinal regions to two different BdGPL strains. Bd exposure strongly lowered survival in B. bufo, and in both species survival was 28 29 lower in the northern region, this difference being much stronger in B. bufo. Northern 30 individuals were smaller in both species, and the survival difference between the regions was 31 size-mediated with smaller individuals being more sensitive to Bd. In both species, Bd 32 exposure led to sub-lethal effects in terms of reduced growth suggesting that even individuals 33 surviving the infection may have reduced fitness mediated by smaller body size. Bd strain 34 affected size-dependent mortality differently in the two regions. We discuss the possible 35 mechanisms how body size and geographical origin can contribute to the present results. 36

37

38

40 Introduction

41 Natural populations are increasingly affected by emerging infectious diseases (Daszak et al. 42 2000, Pennisi 2010, Fisher et al. 2012, Scheele et al. 2019). Many of the emerging diseases 43 are caused by generalist and opportunistic fungal pathogens which can infect a wide range of 44 host species (Wibbelt et al. 2010, Fisher et al. 2012, Lorch et al. 2016, More et al. 2018). The 45 virulence of a fungal pathogen often differs among host species, leading to population 46 declines in some hosts while having no apparent effect on others (Casadevall 2007, Herceg et 47 al. 2021). There is also evidence that populations of a host species may differ in their 48 susceptibility, but apart from plant systems few studies have addressed this question in detail 49 (Ebert 2008; Laine et al. 2011; Bradley et al. 2015; Martin-Torrijos et al. 2017). 50 The chytrid fungus Batrachochytrium dendrobatidis (Bd), causing the disease 51 chytridiomycosis in amphibians, is a generalist pathogen which has caused the decline of over 52 500 amphibian species, including the presumed extinction of 90 species (Berger et al. 1998, 53 Skerratt et al. 2007, Lips 2016, Scheele et al. 2019). Bd is endemic in East Asia where 54 itcoexists with the native fauna, but severe outbreaks of chytridiomycosis have been observed 55 in the Americas and Australia (Lips 2016, Scheele et al. 2019). There is considerable variation 56 in virulence among genetic strains of *Bd* (Farrer et al. 2011, Bataille et al. 2013, Greenspan et 57 al. 2018) and BdGPL, the global panzootic lineage originating in Eastern Asia, has caused 58 most of the chytridiomycosis outbreaks (O'Hanlon et al. 2018). While genetic variation 59 within BdGPL is relatively limited (O'Hanlon et al. 2018), there is evidence for virulence differences between *Bd*GPL strains (Becker et al. 2017; Burrow et al. 2017, Dang et al 2017). 60 61 As a generalist pathogen, Bd infects a wide range of amphibian species (Fisher et al. 2012, 62 Olson et al. 2013). However, all species do not develop chytridiomycosis; many are resistant 63 to the disease and can clear the infection, while others can tolerate high infection loads 64 without developing the disease (Fisher et al. 2009, Gahl et al. 2012, Ellison et al. 2014,

65	Scheele et al. 2017). Similarly, geographical populations of the same species can differ in
66	their susceptibility to <i>Bd</i> (Savage and Zamudio 2011, Bradley et al. 2015, Kosch et al. 2019).
67	These differences can be due to genetic differences in traits like immune response and
68	behavior (Richards-Zawacki 2010), and are in some cases linked with direct Bd-mediated
69	selection (Savage & Zamudio 2016, Savage et al. 2018). Although infection does not cause
70	direct mortality in the resistant and tolerant species, sub-lethal fitness effects such as
71	decreased growth have been detected (Bielby et al. 2015, Burrow et al. 2017, Campbell et al.
72	2019).

73 Climate-related latitudinal divergence is an important structuring force of intraspecific genetic 74 variation (e.g., Hewitt 2000, Conover et al. 2009), but its potential effects mediating host-75 pathogen interactions have received little attention. Two lines of evidence suggest that 76 amphibian populations living at high latitudes in the northern hemisphere may be especially 77 vulnerable to disease. Firstly, due to post-glacial colonization patterns northern populations 78 often harbor less genetic variation (Hewitt 2000). In many amphibians, this is true also for 79 immunogenetic variation in major histocompatibility (MHC) genes (Zeisset and Beebee 2014, 80 Cortázar-Chinarro et al. 2017), which is associated with Bd resistance (Savage and Zamudio 81 2011, Savage et al. 2018, Kosch et al. 2019). Furthermore, pathogen richness and abundance 82 are significant predictors of adaptive MHC variation (Wang et al. 2017). As pathogen richness 83 and abundance decrease towards colder climates (Schemske et al. 2009), populations at higher 84 latitudes may encounter lower diversity and a lower number number of pathogens which may 85 lead to increased drift and loss of adaptive immunogenetic variation in these populations 86 (Cortázar-Chinarro et al. 2017). Secondly, time-constrained high-latitude environments select 87 for high larval growth and development rates (Palo et al. 2003, Luquet et al. 2019), which in 88 amphibians can trade-off with disease resistance (Johnson et al. 2011, Woodhams et al. 2016) 89 and immune response (Gervasi and Foufopoulos 2007, Murillo-Rincon et al. 2017). While all

90 these factors may contribute to lower ability to withstand novel pathogens in high-latitude 91 populations, no studies on disease resistance between latitudinal populations have been made. 92 Here we conducted a laboratory common garden experiment to examine inter- and 93 intraspecific population differences in response to *Bd* infection. Our aims were three-fold: 1) 94 to investigate the response of two common north European amphibians (moor frog Rana 95 arvalis and common toad Bufo bufo) to Bd infection, 2) investigate if the responses differ 96 between southern and northern Scandinavian populations of these species and 3) evaluate if 97 these responses differ between two geographically separated Bd lineages. To this end, we 98 infected newly metamorphosed amphibians with two different *Bd*GPL lineages and measured 99 their survival and growth during a 30-day exposure period. 100 101 Methods

102 Animal rearing

Both *R. arvalis* (hereafter *Ra*) and *B. bufo* (hereafter *Bb*) are widespread amphibians in Europe occurring up to the polar circle in the north (Sillero et al. 2014). Both species are explosive breeders and mate in early spring. In southern Sweden, *Bd* prevalence in breeding adults is 15.3% (n = 288) and 3.4% (n = 941) in *Ra* and *Bb*, respectively (Meurling et al. 2020).

108 Eggs of both species were collected in April 2016 at two sites in Skåne county in

109 southernmost Sweden and May 2016 at two sites in in Norrbotten county in northern Sweden

110 (Fig. 1; Table S1). We collected approximately ten eggs from each of ten different clutches at

111 each site. The eggs and tadpoles were reared in walk-in climate-controlled rooms at Uppsala

- 112 University in plastic tanks filled with 20l reconstituted soft water (RSW; NaHCO₃, CaSO₄,
- 113 MgSO₄ and KCl added to deionized water; APHA 1985) until metamorphosis. Each clutch

114 was kept in a separate tank under 18:6 h light/dark regime at 19°C. The tadpoles were fed ad 115 *libitum* spinach and fish flakes and water was changed every third day. At metamorphic 116 climax (stage 42; Gosner 1960), the animals were moved to another tank of the same size with 117 access to aquatic and terrestrial (aquarium sand) habitat and a shelter. Four days after 118 completion of tail absorption (stage 46), the animals were transported to the sealed 119 experimental facilities at the Swedish Institute for Veterinary Science, Uppsala, where they 120 were kept individually in 1.2 l plastic tanks lined with moist paper towels and a lid of a plastic 121 bottle as a shelter. The metamorphs were kept in these tanks until the end of the experiment 122 and fed fruit flies and crickets *ad libitum* under 18/6h light/dark regime at 19°C. The 123 condition of each animal was checked daily and the tanks were cleaned every third day.

124 Infection experiment

125 The infection treatments were conducted after one week of acclimatization at the 126 experimental facility. The experimental animals were exposed to one of two isolates of Bd-127 GPL (SWE or UK) or a sham infection consisting of culture medium (Table 1). The UK 128 isolate (UKMal 01) was isolated from a wild alpine newt (Ichthyosaura alpestris) in the UK 129 in 2008. The Swedish isolate (SWED-40-5) originated from a wild green toad (Bufotes 130 viridis) in Malmö municipality in southern Sweden in 2015. The animals were exposed 131 individually for 5 h to 200µl culture media containing a dosage of 60 000 zoospores from one 132 of the Bd strains in 30 ml of RSW. The control group (C) was exposed for 5 h to an 133 equivalent volume of RSW and culture media without *Bd* spores. Altogether, we treated 74 134 (25 in SWE, 24 in UK and 25 in C treatment) southern and 46 (16 SWE, 14 UK, 16 C) 135 northern Ra. The corresponding numbers for Bb were 64 (21, 19, 24) southern and 90 (31, 31, 136 28) northern individuals.

After exposure the animals were monitored for 30 days, or until death. Animals showing
irreversible signs of chytridiomycosis (loss of righting function) were euthanized with an
overdose of MS222. Body mass of the animals was measured at the start and end of the
experiment (or at death). At the end of the experiment, the surviving animals were euthanized
and stored in 96% ethanol at 4 °C.

142 DNA extraction and qPCR analyses

143 To confirm infection status, we assessed the presence of *Bd* by using qPCR. DNA was 144 extracted from a hind leg using a Prepman Ultra method described in Boyle et al. (2004). 145 Presence of Bd was assessed by amplifying the internal transcribed spacer (ITS)-5.8S rRNA 146 region (Boyle et al 2004). 25µl reactions containing 12.5µl 2X Taqman Master Mix (Applied 147 Biosystem, ref. 4318157), 2.25 µl 10µM each of forward and reverse primers, 0.625 µl 10µM 148 MGB probe and 5µl of DNA (diluted x10 in water) were run. Each sample was run in 149 triplicate. An exogenous internal positive control (IPC; (Hyatt et al. 2007)) was added to one 150 well in each triplicate (1µl 10XExo IPC master mix and 0.5µl 50XExo IPC DNA to each 151 sample) (VICTIM dye, Applied Biosystems ref. 4304662) to avoid false negatives due to 152 inhibitors. The qPCR assays were run on a Biorad CFX96 Real Time System machine using 153 amplification conditions described in Boyle et al. (2004) with standards of 0.1, 1, 10 and 100 154 genomic equivalents (GE). An individual was recorded as positive if at least one of the 155 triplicate samples exhibited a positive signal (i.e. an exponential amplification curve). If the 156 IPC showed signs of inhibition, negative samples were rerun once before the samples were 157 assigned as not scoreable (NA) and removed from the data set. The above-mentioned 158 standards were used to create a standard curve which was then used to calculate the infection 159 intensity for each individual expressed in genome equivalents (GE).

For the statistical analysis of the infection, we used the logarithm (base 10, zero values were replaced by 0.001, one tenth of the lowest measured non-zero value) of GE and refer to this as infection load (IL).

163 Statistical analyses

All analyses were conducted in R 3.5.2 (R Core Team 2018). Survival was analysed using Generalised Linear Models with a binomial error distribution and a logit link function while data on growth and IL were analysed using linear models. The model assumptions of the linear models were checked using the model diagnostic plots in R. In *Bb*, the growth data were log-transformed due to heteroscedasticity. Consequently, growth in was defined as proportional growth per day: log(mass at death /mass at exposure)/lifespan) for both species.

170 The models that best explained differences in IL, survival and growth were selected using

171 bidirectional elimination (*stepAIC* function in R package *MASS*) starting from the full model

172 with Response ~ Region + *Bd*-strain + Size at infection + Infection load + Interactions. For

173 IL, the full model was IL ~ Region + Bd-strain + Size at infection + Interactions. The two

174 populations within each region were pooled in the analyses as our interest was increasing the

sampled genetic variation within each region rather than studying population differences, and

176 our experimental design did not allow for effective tests of population effects.

To separate the effect of each *Bd*-treatment we also ran the models without the control treatment following the same general structure. Differences in size at infection between regions and treatment were analysed with factorial ANOVAs. This was also done for differences in IL between the species.

181

175

182 **Results**

The qPCR analyses showed high infection success: only two of the successfully analysed individuals from the exposed groups were negative to *Bd* infection at the end of the experiment (one *Bb* and one *Ra*; Table 1) and were removed from the analyses. Thirteen control individuals were also found *Bd* positive (Table 1). However, the infection intensity was very low and we find it likely that these samples were contaminated during sample processing at the end of the experiment. For the statistical analysis, their infection intensity was therefore considered to be 0 GE.

- 190 At the start of the experiment, northern animals were significantly smaller than southern
- 191 animals both in *Ra* ($F_{1, 115} = 25.05$, p < 0.001) and in *Bb* ($F_{1, 149} = 156.94$ p < 0.001; Fig. 2).
- 192 There was no difference in size at exposure between treatments (*Ra*: $F_{2, 115} = 0.15$, p = 0.86,
- 193 $Bb: F_{2, 149} = 0.9, p = 0.41$).
- 194 Infection load
- 195 IL differed between species ($F_{1, 163} = 15.09$, p < 0.001, Fig S1), *Bb* having higher loads. The
- 196 selected model for Ra was IL ~ Size + Region + Size × Region (Table S2a). We found a
- 197 significant effect of the interaction between size and region on IL ($F_{1, 69} = 6.5$, p = 0.013).,
- 198 Size at infection had a negative effect on IL in the northern region, while in the southern
- 199 region size had no effect (Fig. 3a).
- For *Bb*, the selected model was IL ~ *Bd*-strain + Size + *Bd*-strain × Size (Table S2b). Size had
- a significant negative effect on infection load ($F_{1, 87} = 64.45$, p < 0.001, Fig. 3b). The
- interaction between *Bd*-strain and size was close to significant ($F_{1, 87} = 3.66$, p = 0.059), large
- 203 toadlets infected with SWE strain having somewhat higher loads than large individuals
- 204 infected with UK strain.

205 Survival

All southern *Ra* survived the experiment, whereas three infected individuals from the northern region died during the experiment, resulting in survival of 92.9 % in the UK and 87.5 % in the SWE treatment. Survival was complete in the control treatment and in the southern region, and these were excluded from the model. The selected model was: Survival ~ Size (Table S3a) indicating poorer survival of smaller individuals in the two infection treatments in the northern region ($\chi^2_{1,26} = 6.49 \text{ p} = 0.011$; Fig. 4a).

212 While all *Bb* in the control treatment survived the experiment, there was considerable 213 mortality in the infection treatments. Furthermore, survival was higher in southern (66.7 % in 214 SWE and 89.5 % in UK) than in the northern region (38.7 % in SWE and 12.9 % in UK). Due 215 to 100% survival in the control treatment we excluded it from the full model. We also 216 excluded the interaction term between region and size as well as region and infection load, as 217 the range of infection load and size from the northern region only covered a small subset of 218 the range of the southern region, and including these interactions may lead to problematic 219 extrapolations. Survival of *Bb* in the two infection treatments was best explained by the model 220 Survival ~ Bd-strain + Size + IL + Bd-strain × IL (Table S3b). Initial size had a strong 221 positive effect on survival ($F_{1, 86} = 13.10$, p < 0.001, Fig. 4b), whereas IL had a strong 222 negative effect ($F_{1, 86} = 27.61$, p < 0.001). In addition, the significant interaction between Bd 223 strain and infection load ($F_{1, 86} = 7.05$, p = 0.009) corresponded to SWE strain having a 224 steeper slope between survival and infection load compared to UK strain. This results in 225 higher mortality at high and lower mortality at low IL in SWE treatment (Fig. 4b). 226 Growth

Growth in *Ra* was best explained by the model Growth ~ Bd-strain + Region + Size + IL + Region × Size + Size × IL (Table S4a). IL had a negative effect on growth in both infected groups ($F_{1, 107} = 6.16$, p = 0.015; Fig. 5a). Size at infection had a negative effect on growth 230 $(F_{1,107} = 6.16, p = 0.015, Fig. 5b)$, the Size \times IL interaction $(F_{1,107} = 7.21, p = 0.008)$ was 231 caused by the stronger of IL on growth of small individuals. We also found a significant effect 232 of Bd strain on growth ($F_{2, 107} = 3.13$, p = 0.048) and an analysis using only the infected 233 individuals showed that froglets infected with the Swedish strain had an overall lower growth 234 than those infected with the UK strain ($F_{1,70} = 5.63$, p = 0.02; Fig. 5b). 235 Growth analyses in Bb revealed strong negative effects of infection load and a size \times IL 236 interaction when all three Bd-treatments were analysed together (Table S4b). To get more 237 insight on how Bd strain affected Bb growth we analysed the in the two Bd- infection 238 treatments. This best model was Growth ~ Bd-strain + Region + Size + IL + Bd-strain \times 239 Region + Region \times Size + Size \times IL (Table S4c). The effect of size depended on the region (Region × Size : $F_{1.83} = 17.61$, p < 0.001) and IL (Size × IL: $F_{1.83} = 9.00$, p = 0.004). 240 241 Therefore, we present the effect of size on growth separately for the less and more infected 242 half of the individuals (Fig 6a, b). In a similar manner, how infection load affects growth is 243 presented separately for smaller and larger half of the individuals (Fig 6c, d). 244 In small individuals, IL had strong negative effect on growth in both regions (Fig. 6a). In 245 larger toadlets, growth of southern individuals was unaffected by IL whereas larger northern 246 individuals were negatively affected by IL (Fig. 6b). This was also reflected on how initial 247 mass affected growth: in southern toadlets larger individuals grew proportionally faster, 248 whereas in northern toadlets proportional growth was negatively correlated with initial size 249 (Fig. 6c, d). This difference was especially clear in individuals with higher IL (Fig. 6b). There 250 was no significant difference between the two *Bd*-strains ($F_{1, 83} = 0.06$, p = 0.812) nor 251 significant interaction (*Bd*-strain × Region: $F_{1, 83} = 2.20$, p = 0.142).

252

253 Discussion

254 We found that Bd infection lowered survival in both Ra and Bb, but the effect of Bd was more 255 severe on the latter species, especially in the northern region. Our analyses suggest that the 256 survival differences between the regions were largely mediated by body size, smaller 257 individuals being more sensitive to Bd. Furthermore, we found that Bd infection led to sub-258 lethal effects in terms of reduced growth, suggesting that individuals surviving the infection 259 may have lower fitness mediated by their smaller body size. These results suggest that Bd 260 infection may have both direct and indirect effects on amphibian populations and that. high 261 latitude populations may run a higher risk of negative effects than their low-latitude 262 counterparts.

263 While both species became infected in our experiment, in Ra IL was lower and Bd-mediated 264 mortality was only a fraction of the mortality experienced by *Bb*. These results agree with 265 previous studies showing that brown frogs, such as *Ra* have higher tolerance to *Bd*, while *Bb*, 266 like many other bufonids, is more susceptible to Bd-infection (Bosch & Martínez-Solano 267 2006, Garner et al. 2011, Gahl et al. 2012, Balaz et al. 2014, Bielby et al. 2015). When 268 comparing susceptibility to infection and *Bd*-mediated mortality between two anuran species, 269 Bielby et al. (2015) found that R. temporaria, closely related to Ra, was resistant to infection 270 even at high doses, while *Bb* showed near complete infection and dose-dependent mortality. 271 Since Ra has higher infection prevalence in the wild (Meurling et al. 2020) and higher 272 infection tolerance (this study), we suggest that Ra may act as a reservoir species and a 273 possible vector for *Bd*-transmission to more sensitive species such as *Bb*. Indeed, Kärvemo et 274 al. (2019) showed that Bb populations coexisting with Ra had higher Bd-prevalence than 275 populations breeding in ponds without Ra.

We found a clear difference in survival between northern and southern populations especially in *Bb*. Northern individuals were smaller at the time of infection than southern individuals, and our analyses suggest that the survival difference was mainly mediated by body size. This is in accordance with previous studies (Bradley et al. 2015, Burrow et al. 2017) showing that smaller individuals were more vulnerable to *Bd* infection. Smaller individuals may have less developed immune system which may render them more vulnerable to disease (Møller et al. 1998, Burrow et al. 2017). Smaller individuals may also be more vulnerable to *Bd*-mediated water loss as they have larger surface area to body mass ratio. Increased water loss via sloughing is an important symptom in chytridiomycosis, which may render smaller individuals more sensitive to *Bd* infection (Russo et al. 2018, Wu et al. 2019).

286 Our results suggest that much of the differences in *Bd*-mediated mortality can be explained by 287 size differences between the regions. As we raised the tadpoles under common garden conditions, the differences in body mass most likely have a genetic origin. Two additional, not 288 289 mutually exclusive, explanations may further explain higher mortality in the northern 290 populations. Firstly, northern populations may have less effective immune systems because of 291 reduced genetic variation due to postglacial colonization processes (Hewitt 2000, see 292 Cortazar-Chinarro et al. 2017, Rödin-Mörch et al. 2019 for Ra, Thörn et al. 2021 for Bb), or 293 lower pathogen abundance at higher latitudes (Schemske et al. 2009). This hypothesis gains 294 support from the fact that MHC variation in both our study species is lower at higher latitudes 295 (Cortázar-Chinarro et al. 2017, Meurling 2019). Moreover, Bd-mediated survival in Bb seems 296 to be linked with certain MHC alleles (Meurling 2019), as also found in other species (Savage 297 & Zamudio 2011, Savage et al. 2018, Kosch et al. 2019). Secondly, higher larval development 298 rates in the northern populations may trade off with disease resistance (Johnson et al. 2011, 299 Woodhams et al. 2016). Also this hypothesis is indirectly supported by the facts that more 300 time-constrained populations have higher development rates in both our study species (Luquet 301 et al. 2015, 2019) and that *Ra* tadpoles experimentally induced to develop faster have weaker 302 immune response (Murillo-Rincon et al. 2017). Additional studies focusing on Bd resistance 303 in known MHC and developmental genotypes would be highly interesting.

304 *Bd* infection had clear negative effects on growth in both species. As body size is positively 305 related to fitness in juvenile amphibians (Earl and Whiteman 2015), these results suggest that 306 Bd may have sublethal fitness effects. For example, hibernation success is often positively 307 related to body size and failing to reach a sufficient size before hibernation can greatly reduce 308 overwinter survival (Altwegg and Reyer 2003). This can be especially detrimental at higher 309 latitudes where the hibernation period can reach eight months. Small body size may also lead 310 to higher risk of predation, delayed maturation and lower ability to compete for resources and 311 mates (reviewed in Earl and Whiteman 2015). In the long run, these effects may decrease 312 population growth rate and ability to cope with environmental changes such as higher 313 temperature due to climate change. In our case, even if survival of *Ra* was not strongly 314 affected by *Bd* infection, the results suggest that sublethal effects of infection mediated by 315 body size may still lower individual and population fitness.

316 We found relatively little evidence for differences in pathogenicity between the two Bd 317 isolates. However, we found significant treatment \times size interaction in survival of *Bb* where 318 survival was more strongly size-dependent when toadlets were infected with the UK strain. 319 These results suggest that individuals infected with the UK strain may relatively quickly reach 320 a size where the lethality of Bd is reduced, while Bd-mediated mortality induced by the 321 Swedish strain is less size-dependent. This is especially the case in southern individuals which 322 are larger at metamorphosis, while the smaller northern individuals stay longer in the 323 vulnerable size classes.

A potential caveat in our study is that as we used laboratory-raised (but wild-collected) individuals which may not have developed as diverse community of skin microbiota as wild individuals. Indeed, captive amphibians often have a reduced and less varied bacterial community than wild populations of the same species (Antwis et al. 2014, Bataille et al. 2016). As skin microbiome plays an important role in defending against fungal and other pathogens, this could impact the ability of amphibians reared in captivity to respond to *Bd*infection (Harris et al. 2009, Walke et al. 2015, Madison et al. 2017, Woodhams et al. 2018).
We currently lack knowledge on the skin microbiomes of our study species and if these differ
between geographical regions. Microbiome studies are needed for additional insight on
factors behind the high mortality found in this study.

334 Bd is widespread in the southern parts of Sweden (Kärvemo et al. 2018, 2019, Meurling et al. 335 2020). However, in a pattern similar to much of Europe (Lips 2016, Scheele et al. 2019), no 336 cases of chytridiomycosis or unusual die-offs have been found in Sweden. Our experimental 337 results suggest that even though no negative effects of the infection have been seen in the 338 wild, this might not be the complete picture. It is currently unclear how well the present 339 results translate to natural conditions, but we note that *Bd* causes sublethal effects in terms of 340 reduced movements and body condition in wild Scandinavian amphibians (Kärvemo et al. 341 2019, 2020). Furthermore, the lethality of Bd is highly dependent on environmental 342 conditions, including temperature (e.g., Novakowski et al. 2106, Mosher et al. 2018, Cohen et 343 al. 2019), and relatively minor elevations in mortality may risk long-term survival of Bd-344 infected amphibian populations (Muths et al. 2011; Spitzen-van der Sluijs et al. 2017). Two 345 important conclusions can be drawn. Firstly, very few surveys have been conducted in 346 northern Scandinavia (Meurling et al. 2020). As populations at higher latitudes can be more 347 vulnerable to infection, it is important to investigate the occurrence of Bd in these areas and, if 348 still possible, prevent or limit the northward spread of the fungus. Secondly, we showed that 349 infection leads to higher mortality and reduced body size. These, in turn, can lead to reduced 350 population growth rates in the long-term even in the absence of major mortality effects. As 351 the potential negative effects of *Bd* on population growth can be relatively subtle and difficult 352 to detect (Doddington et al. 2013, Spitzen-van der Sluijs et al. 2017, Mosher et al. 2018), 353 long-term monitoring of amphibian populations is of high importance.

354

355 Acknowledgements

356 We thank Lola Brooks and Trent Garner for discussions and providing the *Bd*-strains.

357

358 **Declarations**

- **Funding:** Funding from the Swedish Research Council Formas (215-2014-294), Stiftelsen
- 360 Oscar och Lili Lamms Minne and Stiftelsen för zoologisk forskning is acknowledged.
- 361 **Conflict of interest:** The authors declare no conflicts of interest.
- 362 **Ethics approval:** The study was conducted with a permit (C28/15) from Uppsala ethical
- 363 committee for animal experiments and collection permits from the county administrative
- boards in Skåne and Norrbotten.
- 365 Availability of data and materials: The data will be deposited in DRYAD upon acceptance.
- 366 Authors' contributions: SM, MCC, JH and AL conceived and designed the experiments,
- 367 SM, MCC and DÅ performed the experiments and EÅ provided advice and logistic help, SM
- and MLS analysed the data, SM and AL wrote the paper with input from all the authors.

369

370 References

- Altwegg R, Reyer HU (2003) Patterns of natural selection on size at metamorphosis in water
 frogs. Evolution 57:872-882
- Antwis RE, Haworth RL, Engelmoer DJ et al.(2014) Ex situ diet influences the bacterial
 community associated with the skin of red-eyed tree frogs (*Agalychnis callidryas*).
 PLoS One 9:e85563
- 376 APHA. 1985. Standard methods for the examination of water and wastewater. 16th ed.
- 377 American Public Health Association, Washington, DC.

378	Balaz V, Voros J, Civis P et al. (2014) Assessing risk and guidance on monitoring of
379	Batrachochytrium dendrobatidis in Europe through identification of taxonomic
380	selectivity of infection. Conserv Biol 28:213-223
381	Bataille A, Fong JJ, Cha M et al. (2013) Genetic evidence for a high diversity and wide
382	distribution of endemic strains of the pathogenic chytrid fungus Batrachochytrium
383	dendrobatidis in wild Asian amphibians. Mol Ecol 22:4196-4209.
384	Bataille A, Lee-Cruz L, Tripathi B et al. (2016) Microbiome variation across amphibian skin
385	regions: Implications for chytridiomycosis mitigation efforts. Microb Ecol 71:221-232
386	Berger L, Speare R, Daszak P et al. (1998) Chytridiomycosis causes amphibian mortality
387	associated with population declines in the rain forests of Australia and Central
388	America. Proc Natl Acad Sci USA 95:9031-9036
389	Becker CG, Bletz MC, Greenspan SE et al. (2019) Low-load pathogen spillover predicts
390	shifts in skin microbiome and survival of a terrestrial-breeding amphibian. Proc R Soc
391	B 286:20191114
392	Bielby J, Fisher MC, Clare FC et al. (2015) Host species vary in infection probability, sub-
393	lethal effects, and costs of immune response when exposed to an amphibian parasite.
394	Sci Rep 5:10828
395	Bosch J, Martínez-Solano I (2006) Chytrid fungus infection related to unusual mortalities of
396	Salamandra salamandra and Bufo bufo in the Peñalara Natural Park, Spain. Oryx
397	40:84-89
398	Boyle DG, Boyle DB, Olsen V et al. (2004) Rapid quantitative detection of chytridiomycosis
399	(Batrachochytrium dendrobatidis) in amphibian samples using real-time Taqman PCR
400	assay. Dis Aquat Org 60:141-148
401	Bradley PW, Gervasi SS, Hua J et al. (2015) Differences in sensitivity to the fungal pathogen
402	Batrachochytrium dendrobatidis among amphibian populations. Conserv Biol 29:134-
403	1356
404	Burrow AK, Rumschlag SL, Boone MD (2017) Host size influences the effects of four
405	isolates of an amphibian chytrid fungus. Ecol Evol 7:9196-9202
406	Casadevall A (2007) Determinants of virulence in the pathogenic fungi. Fungal Biol Rev
407	21:130-132
408	Campbell L, Bower DS, Clulow S et al. (2019) Interaction between temperature and sublethal
409	infection with the amphibian chytrid fungus impacts a susceptible frog species. Sci
410	Rep 9:83

411 Cohen JM, McMahon TA, Ramsay C et al. (2019) Impacts of thermal mismatches on chytrid 412 fungus Batrachochytrium dendrobatidis prevalence are moderated by life stage, body 413 size, elevation and latitude. Ecol Lett 22:817-825 414 Cortázar-Chinarro M, Lattenkamp EZ, Meyer-Lucht Y et al. (2017) Drift, selection, or 415 migration? Processes affecting genetic differentiation and variation along a latitudinal 416 gradient in an amphibian. BMC Evol Biol 17:189 417 Dang TD, Searle CL, Blaustein AR (2017) Virulence variation among strains of the emerging 418 infectious fungus Batrachochytrium dendrobatidis in multiple amphibian host species. 419 Dis Aquat Org 124:233-239 420 Daszak P, Cunningham AA, Hyatt AD (2000) Emerging infectious diseases of wildlife--421 Threats to biodiversity and human health. Science 287:443-449 422 Doddington BJ, Bosch J, Oliver JA et al. (2013) Context-dependent amphibian host 423 population response to an invading pathogen. Ecology 94:1795-1804 424 Earl JE, Whiteman HH (2015) Are commonly used fitness predictors aAccurate? A meta-425 analysis of amphibian size and age at metamorphosis. Copeia 103:297-309 426 Ebert D (2008) Host-parasite coevolution. insights from the Daphnia-parasite model system. 427 Curr Opin Microbiol 11:290-301 428 Ellison AR, Tunstall T, DiRenzo GV et al. (2014) More than skin deep: functional genomic 429 basis for resistance to amphibian chytridiomycosis. Genome Biol Evol 7:286-298 430 Farrer RA, Weinert LA, Bielby J et al. (2011) Multiple emergences of genetically diverse 431 amphibian-infecting chytrids include a globalized hypervirulent recombinant lineage. 432 Proc Natl Acad Sci USA 108:18732-18736 433 Fisher MC, Garner TW, Walker SF (2009) Global emergence of Batrachochytrium 434 dendrobatidis and amphibian chytridiomycosis in space, time, and host. Annu Rev 435 Microbiol 63:291-310 436 Fisher MC, Henk DA, Briggs CJ et al. (2012) Emerging fungal threats to animal, plant and 437 ecosystem health. Nature 484:186 438 Gahl MK, Longcore JE, Houlahan JE (2012) Varying responses of rortheastern North 439 American amphibians to the chytrid pathogen *Batrachochytrium dendrobatidis*. 440 Conserv Biol 26:135-141 441 Garner TWJ, Rowcliffe JM, Fisher MC (2011) Climate change, chytridiomycosis or 442 condition: an experimental test of amphibian survival. Glob Change Biol 17:667-675 443 Gervasi S, Foufopoulos J (2007) Costs of plasticity: Responses to desiccation decrease post-444 metamorphic immune function in a pond-breeding amphibian. Funct Ecol 22:100-108

445 Gosner KL (1960) A Simplified table for staging anuran embryos and larvae with notes on 446 identification. Herpetologica 16:183-190 447 Greenspan SE, Lambertini C, Carvalo T et al. (2018) Hybrids of amphian chytrid show high 448 virulence in native hosts. Sci Rep 8:9600 449 Rowley-Harris RN, Brucker RM, Walke JB et al. (2009) Skin microbes on frogs prevent 450 morbidity and mortality caused by a lethal skin fungus. ISME J 3:818-824 451 Herczeg D, Ujszegi J, Kasler A et al. 2021. Host-multiparasite interactions in amphibians: a 452 review. Parasit Vectors 14:296 453 Hewitt G (2000) The genetic legacy of the Quaternary ice ages. Nature 405:907-913 454 Hyatt AD, Boyle DG, Olsen V et al. (2007) Diagnostic assays and sampling protocols for the 455 detection of Batrachochytrium dendrobatidis. Dis Aquat Org 73:175-192 456 Johnson PTJ, Kellermanns E, Bowerman J (2011) Critical windows of disease risk: 457 amphibian pathology driven by developmental changes in host resistance and 458 tolerance. Funct Ecol 25:726-734 459 Kärvemo S, Laurila A, Höglund J (2019) Urban environment and reservoir host species are 460 associated with Batrachochytrium dendrobatidis infection prevalence in the common 461 toad. Dis Aquat Org 134:33-42 462 Kärvemo S, Meurling S, Berger D et al. (2018) Effects of host species and environmental 463 factors on the prevalence of *Batrachochytrium dendrobatidis* in northern Europe. 464 PLoS ONE 13:e0199852 465 Kärvemo S, Wikström G, Widenfalk LA et al. (2020) Chytrid fungus dynamics and infections 466 associated with movement distances in a red-listed amphibian. J Zool 311:164-174 467 Kosch TA, Silva CNS, Brannelly LA et al. (2019) Genetic potential for disease resistance in 468 critically endangered amphibians decimated by chytridiomycosis. Anim Conserv 469 22:238-250 470 Laine A-L, Burdon JJ, Dodds PN et al. (2011) Spatial variation in disease resistance: from 471 molecuales to metapopulations. J Ecol 99:96-112. 472 Lips KR (2016) Overview of chytrid emergence and impacts on amphibians. Phil Trans R Soc 473 B 371:20150465 474 Lorch JM, Knowles S, Lankton JS et al. (2016) Snake fungal disease: an emerging threat to 475 wild snakes. Phil Trans R Soc B 371:20150457 476 Luquet E, Léna JP, Miaud C et al. (2015) Phenotypic divergence of the common toad (Bufo 477 bufo) along an altitudinal gradient: evidence for local adaptation. Heredity 114:69-79

- 478 Luquet E, Rödin Mörch P, Cortázar-Chinarro M et al. (2019) Post-glacial colonization routes 479 coincide with a life-history breakpoint along a latitudinal gradient. J Evol Biol 32:356-480 368 481 Martin-Torrijos L, Campos Llach M, Pou Rovira Q et al. (2017) Resistance to crayfish 482 plague, Aphanomyces astaci (Oomycota) in the endangered freshwater crayfish 483 species Austropotamobius pallipes. PLoS ONE 12:e0181226 484 Madison JD, Berg EA, Abarca JG et al. (2017) Characterization of Batrachochytrium 485 dendrobatidis inhibiting bacteria from amphibian populations in Costa Rica. Front 486 Microbiol 8:290 487 Meurling S (2019) The response in native wildlife to an invading pathogen: Swedish 488 amphibians and Batrachochytrium dendrobatidis. PhD thesis, Uppsala University. 489 Meurling S, Kärvemo S, Chondrelli N et al. (2020) Occurrence of Batrachochytrium 490 dendrobatidis in Sweden: higher infection prevalence in southern species. Dis Aquat 491 Org 140:209-218 492 More S, Angel Miranda M, Bicout D et al. (2018) Risk of survival, establishment and spread 493 of Batrachochytrium salamandrivorans (Bsal) in the EU. EFSA J 16:e05259 494 M1ller AP, Christe P, Erritzoe J et al. (1998). Condition, disease and immune defence. Oikos 495 83:3010306 496 Mosher BA, Bailey LL, Muths E et al. (2018) Host-pathogen metapopualtion suggest high 497 elevation refugia for boreal toads. Ecol Appl 28:926-937 498 Murillo-Rincon AP, Laurila A, Orizaola G (2017) Compensating for delayed hatching reduces 499 offspring immune response and increases life-history costs. Oikos 126:565-571 500 Muths E, Scherer RD, Pilliod DS (2011) Compensatroy effects of recruitment and survival 501 when amphibians are perturbed by disease. Ecol Appl 48:873-879 502 Nowakowski AJ, Whitfield SM, Eskew EA et al. (2016) Infection risk decreases with 503 increasing mismatch in host and pathogen environmental tolerances. Ecol Lett 504 19:1051-1061 505 O'Hanlon SJ, Rieux A, Farrer RA et al. (2018) Recent Asian origin of chytrid fungi causing 506 global amphibian declines. Science 360:621-627 507 Olson DH, Aanensen DM, Ronnenberg KL et al. (2013) Mapping the global emergence of
- 508 *Batrachochytrium dendrobatidis*, the amphibian chytrid fungus. PLoS ONE 8:e56802
- 509 Palo JU, O'Hara RB, Laugen AT et al. (2003) Latitudinal divergence of common frog (*Rana* 510 *temporaria*) life history traits by natural selection: evidence from a comparison of
- 511 molecular and quantitative genetic data. Mol Ecol 12:1963-1978

512 Pennisi E (2010) Armed and dangerous. Science 327:804-805

- 513 R Core Team (2018) R: A Language and Environment for Statistical Computing. R
- 514 Foundation for Statistical Computing, Vienna, Austria
- 515 Richards-Zawacki CL (2010) Thermoregulatory behaviour affects prevalence of chytrid
 516 fungal infection in a wild population of Panamanian golden frogs. Proc R Soc B
 517 277:519-528
- Rödin-Mörch P, Luquet E, Meyer-Lucht Yet al. (2019) Latitudinal divergence in a
 widespread amphibian: Contrasting patterns of neutral and adpative genomic
 variation. Mol Ecol 28:2996-3011
- Russo CJM, Ohmer MEB, Cramp RL et al. (2018) A pathogenic skin fungus and sloughing
 exacerbate cutaneous water loss in amphibians. J Exp Biol 221
- 523 Savage AE, Zamudio KR (2011) MHC genotypes associate with resistance to a frog-killing
 524 fungus. Proc Natl Acad Sci USA 108:16705-16710
- 525 Savage AE, Mulder KP, Torret T et al. (2018) Lost but nmot forgotten: MHC genotypes
 526 predict overwinter survival despite depauparate MHC diversity in a declining frog.
- 527 Conserv Genet 19:309-322
- Scheele BC, Hunter DA, Brannelly LA et al. (2017) Reservoir-host amplification of disease
 impact in an endangered amphibian. Conserv Biol 31:592-600
- Scheele BC, Pasmans F, Skerratt LF et al. (2019) Amphibian fungal panzootic causes
 catastrophic and ongoing loss of biodiversity. Science 363:1459-1463
- Schemske DW, Mittelbach GG, Cornell HV et al. (2009) Is there a latitudinal gradient in the
 importance of biotic interactions? Annu Rev Ecol Evol Syst 40:245-269
- Sillero N, Campos J, Bonardi A et al. (2014) Updated distribution and biogeography of
 amphibians and reptiles of Europe. Amph-Reptil 35:1-31
- 536 Skerratt LF, Berger L, Speare R et al. (2007) Spread of chytridiomycosis has caused the rapid
 537 global decline and extinction of frogs. EcoHealth 4:125-134
- Spitzen van der Sluijs A, Canessa S, Martel A 2017 Fragile coexistence of a global chytrid
 pathogen with amphibian populations is medaited by environment and demography.
 Proc R Soc B 284:20171444
- 541 Thörn F, Rödin-Mörch P, Cortazar-Chinarro M et al. (2021) The effects of drift and selection
 542 on latitudinal genetic variation in Scandinavian common toads (*Bufo bufo*) following
- 543 postglacial recolonization. Heredity 126:656-667
- Voyles J, Woodhams DC, Saenz V et al. (2018) Shifts in disease dynamics in a tropical
 amphibian assemblage are not due to pathogen attenuation. Science 359:1517

546	Walke JB, Becker MH, Loftus SC et al. (2015) Community structure and function of
547	amphibian skin microbes: an experiment with bullfrogs exposed to a chytrid fungus.
548	PLoS ONE 10:e0139848
549	Wang SP, Liu CH, Wilson AB et al. (2017) Pathogen richness and abundance predict patterns
550	of adaptive major histocompatibility complex variation in insular amphibians. Mol
551	Ecol 26:4671-4685
552	Wibbelt G, Kurth A, Hellmann D et al. (2010) White-nose syndrome fungus (Geomyces
553	destructans) in bats, Europe. Emerg Infect Dis 16:1237-1243
554	Woodhams DC, Bletz M, Kueneman J et al. (2016) Managing amphibian disease with skin
555	microbiota. Trends Microbiol 24:161-164
556	Woodhams DC, LaBumbard BC, Barnhart KL et al. (2018) Prodigiosin, violacein, and
557	volatile organic compounds produced by widespread cutaneous bacteria of amphibians
558	can inhibit two Batrachochytrium fungal pathogens. Microb Ecol 75:1049-1062
559	Wu NC, McKercher C, Cramp RL et al. (2019) Mechanistic basis for loss of water balance in
560	green tree frogs infected with a fungal pathogen. Am J Physiol Regul Integr Comp
561	Physiol 317:R301-R311
562	Zeisset I, Beebee TJC (2014) Drift rather than selection dominates MHC Class II Allelic
563	diversity patterns at the biogeographical range scale in natterjack toads Bufo calamita.
564	PLoS ONE 9:e100176
565	

Table 1 Number of experimental and infected individuals (as determined by qPCR) in

568 different infection treatments.

Species	Region	Treatment	Total Nb.	Positive	%	Negative	%	Failed	%
Bufo bufo	North	Control	28	1	3,6	16	57,1	11	39,3
		<i>Bd</i> SWE	31	31	100,0	0	0,0	3	9,7
		<i>Bd</i> UK	31	31	100,0	0	0,0	0	0,0
	South	Control	24	9	37,5	13	54,2	2	8,3
		<i>Bd</i> SWE	21	21	100,0	0	0,0	3	14,3
		<i>Bd</i> UK	19	14	73,7	1	5,3	4	21,1
Rana arvalis	North	Control	16	1	6,3	6	37,5	9	56,3
		<i>Bd</i> SWE	16	16	100,0	0	0,0	1	6,3
		<i>Bd</i> UK	14	13	92,9	0	0,0	1	7,1
	South	Control	25	2	8,0	18	72,0	5	20,0
		<i>Bd</i> SWE	25	25	100,0	0	0,0	0	0,0
		<i>Bd</i> UK	24	23	95,8	1	4,2	0	0,0

571 Figure legends

- Fig. 1. Map of Sweden showing the sites of egg collection. Blue and green circles show *R*. *arvalis* and green *B. bufo* sites, respectively.
- 574 **Fig. 2.** Mass at exposure $(g \pm SE)$ per region for *R. arvalis* and *B. bufo*

575 **Fig. 3.** Infection load (log10 average GE, at the end of treatment) as the function of size at

576 infection with either the Swedish or UK strain in **a**. *R. arvalis* and **b**. *B. bufo*. Lines give the

577 predictions of the model. Filled dots and solid lines represent the northern region, while open

- 578 dots and dashed lines represent the southern region.
- Fig. 4. Survival as a function of size at infection with either the Swedish or UK strain for a: *R*. *arvalis* from the northern region. b: Survival in *B. bufo* as function of infection load and size
 at infection, where dark and pale dots represent the northern and southern regions,

respectively. Line (a) and surfaces (b) give the predictions of the model. Blue dots refer to the

583 Swedish *Bd* strain and red dots to the UK strain. In **b** some pale dots are hidden among the

584 dark dots in the lower corner in the front (high infection load and low size).

Fig. 5. Growth per day (from infection to death or end of experiment) as a function of **a:** infection load (log10 average GE, at the end of treatment) and, **b:** size (at start of treatment) in different infection treatments for *R. arvalis*. Filled and open dots represent the northern and southern regions, respectively. Lines gives the predictions of the model and are evaluated for average infection load in **a** and average size in **b**. Separate models were run for the control group and infected individuals.

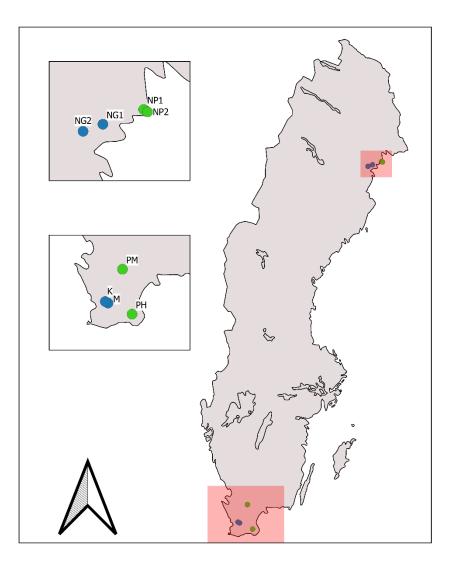
591 Fig. 6. Growth per day as a function of a-b: infection load (log10 average GE, at the end of

treatment) and, **c-d:** size at infection with either the Swedish or UK strain or no infection

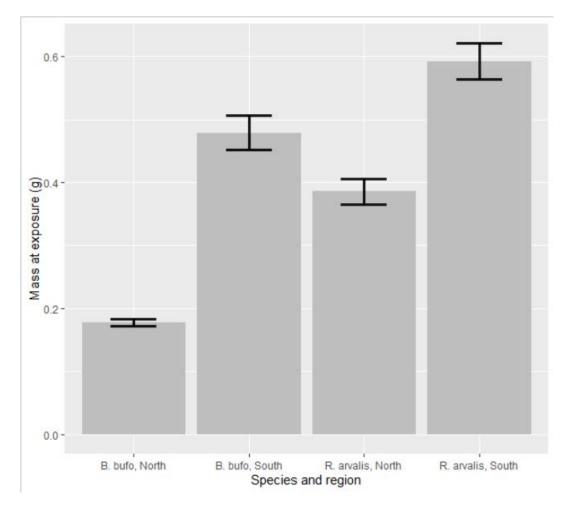
593 (control) for *B. bufo*. Smaller individuals are represented by squares in **a** (smaller than the

- 594 median of their respective region) and the larger individuals as triangles in **b** (larger or equal
- 595 to the median of their respective region). The same holds for **c-d** but with respect to infection
- 596 load. Lines shows model prediction. Filled markers and solid lines represent the northern and
- 597 open markers and dashed lines represent southern region. Separate models were run for
- 598 control treated and infected individuals.

600 Meurling et al., Fig. 1

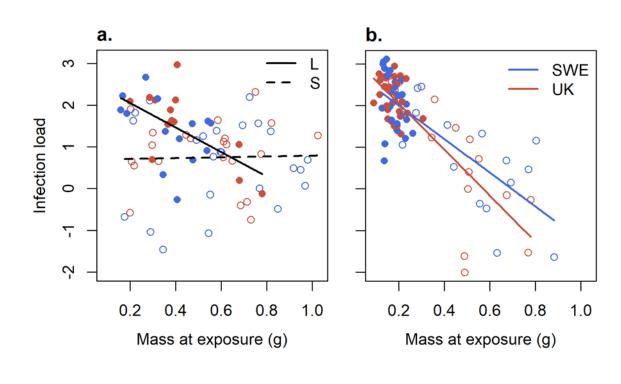


604 Meurling et al., Fig. 2

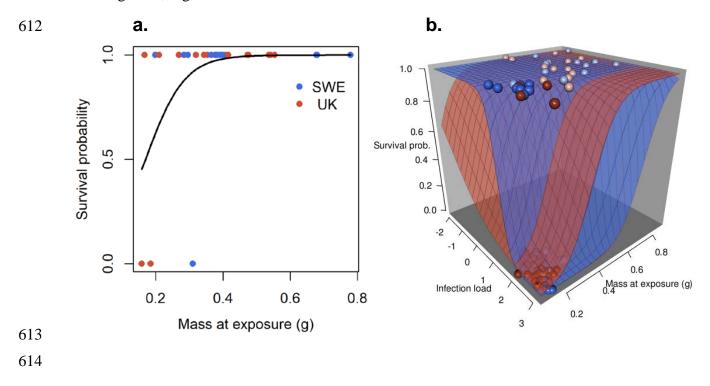


605

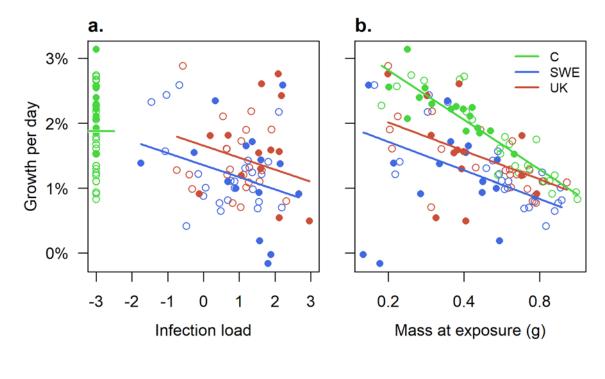
607 Meurling et al., Fig. 3



611 Meurling et al., Fig. 4



615 Meurling et al., Fig. 5



618 Meurling et al., Fig. 6

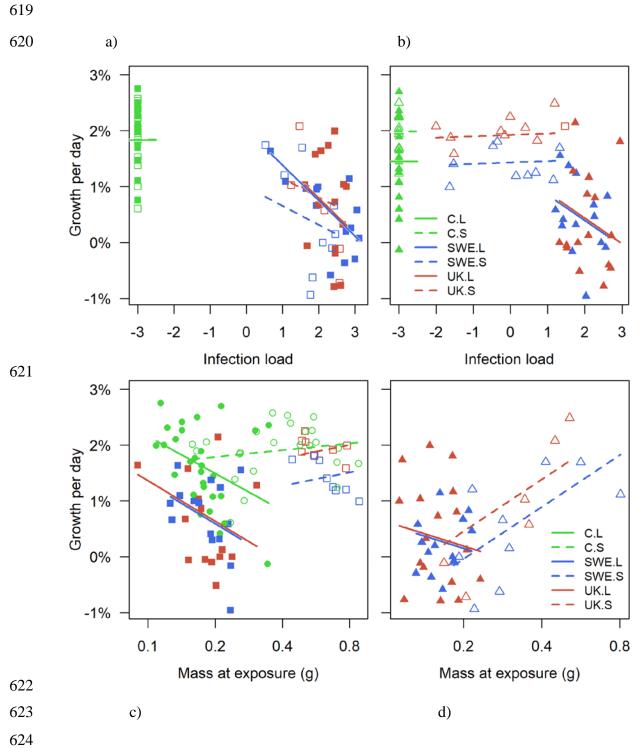


Table S1. Coordinates for the collection sites.

Species	Region	Population	Ν	Ε
Rana arvalis	North	NG1	65.519974	21.685978
		NG2	65.488998	21.378151
	South	М	55.699774	13.360416
		K	55.722114	13.284693
Bufo bufo	North	NP1	65.583139	22.319458
		NP2	65.56554	22.37404
	South	PM	56.217897	13.731548
		PH	55.539031	14.009702

630 Table S2. Results from final general linear models on infection load. a) *R. arvalis*, b) *B. bufo*.

a)

	Sum of squares	Df	F	Р
Size	5.5	1	7.54	0.008
Region	8.8	1	12.11	< 0.001
Size x Region	4.7	1	6.50	0.013
Residuals	50.3	69		

b)

	Sum of squares	Df	F	Р
Bd-strain	0.8	1	1.70	0.196
Size	31.9	1	64.45	< 0.001
Strain x Size	1.8	1	3.66	0.059
Residuals	43.0	87		

Table S3. Results from final generalized linear models on survival. a) *R. arvalis*: the analyses

only cover the northern population and the two *Bd*-treatments as survival in the southern

638 population and control treatment were complete in this species. b) *B. bufo:* only the two *Bd*-

treatments are included as survival was complete in the control treatment in this species.

640

641

a)

b)

	LR chisquared	Df	Р
Size	6.5	1	0.011
Residuals	12.6	26	

642

643

	Sum of squares	Df	F	Р
Bd-strain	0.2	1	0.24	0.624
Size	9.3	1	13.10	< 0.001
Infection load (IL)	19.7	1	27.61	< 0.001
Strain x IL	5.0	1	7.05	0.009
Error	61.3	86		

644

Table S4. Results from general linear models on growth. a) R. arvalis, b) B. bufo all

a)

	Sum of squares	Df	F	Р
<i>Bd</i> strain	0.000131	2	3.13	0.048
Region	0.000001	1	0.05	0.827
Size	0.000143	1	6.84	0.010
Infection load (IL)	0.000129	1	6.16	0.015
Region x Size	0.000053	1	2.54	0.114
Size x IL	0.000150	1	7.21	0.008
Residuals	0.0002233	107		

b)

	Sum of squares	Df	F	Р
Bd-strain	0.000209	2	2.43	0.092
Region	0.000117	1	2.72	0.101
Size	0.000006	1	0.14	0.712
Infection load (IL)	0.000336	1	7.82	0.006
Strain x Size	0.000252	2	2.94	0.056
Region x Size	0.000937	1	3.06	0.083
Region x IL	0.000131	1	3.66	0.059
Size x IL	0.000391	1	21.84	< 0.001
Residuals	0.005667	132		

c)

	Sum of squares	Df	F	Р
Bd-strain	0.000003	1	0.06	0.812
Region	0.000001	1	0.02	0.877
Size	0.000494	1	10.67	0.002
Infection load (IL)	0.000513	1	11.09	0.001
Strain x Region	0.000102	1	2.20	0.142
Region x Size	0.000814	1	17.61	< 0.001
Size x IL	0.000416	1	9.00	0.004
Residuals	0.003838	83		

individuals, c) *B. bufo*, *Bd*-infected individuals only

