

1 **Limited reciprocal surrogacy of bird and habitat diversity and**
2 **inconsistencies in their representation in Romanian protected areas**

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23 **Abstract**

24 Because it is impossible to comprehensively characterize biodiversity at all levels of organization,
25 conservation prioritization efforts need to rely on surrogates. As species distribution maps of relict
26 groups as well as high-resolution remotely sensed data increasingly become available, both types of
27 surrogates are commonly used. A good surrogate should represent as much of biodiversity as possible,
28 but it often remains unclear to what extent this is the case. Here, we aimed to address this question by
29 assessing how well bird species and habitat diversity represent one another. We conducted our study in
30 Romania, a species-rich country with high landscape heterogeneity where bird species distribution
31 data have only recently started to become available. First, we prioritized areas for conservation based
32 on either 137 breeding bird species or 36 habitat classes, and then evaluated their reciprocal surrogacy
33 performance. Second, we examined how well these features are represented in already existing
34 protected areas. Finally, we identified target regions of high conservation value for the potential
35 expansion of the current network of reserves (as planned under the new EU Biodiversity Strategy for
36 2030). We found that bird species were a better surrogate for habitat diversity than vice versa. Highly
37 ranked areas based on habitat diversity were represented better than areas based on bird species, which
38 varied considerably between species. Our results highlight that taxonomic and environmental (i.e.,
39 habitat types) data may perform rather poorly as reciprocal surrogates, and multiple sources of data are
40 required for a full evaluation of protected areas expansion.

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42

43 **Introduction**

44

45 The ultimate goal of conservation prioritization is the protection of biodiversity at all levels of
46 organization [1]. However, limited financial resources and competing stakeholder interests constrain
47 the area that can reasonably be protected. The process of identifying potential regions for designation
48 as protected area (PA) should therefore be undertaken thoroughly and strategically [2, 3, see 4 for a
49 review]. The striking obstacle is however that biodiversity is very complex and difficult to characterize
50 [5], and surveying biodiversity in its entirety is nearly impossible. Shortcuts necessarily need to be
51 taken to quicken the prioritization process and ensure its feasibility [6]. One of these shortcuts is using
52 a biodiversity or environmental indicator as a conservation surrogate [see 4 for a review, 7], which is:
53 “An ecological process or element (e.g., species, ecosystem, or abiotic factor) that [should] [...]”
54 represent (i.e., serve as a proxy for) another aspect of an ecological system” [8]. The efficacy and
55 efficiency of surrogates for overall biodiversity (known and unknown) have progressively been
56 evaluated [7, 9-13], and appear to be influenced by factors such as the size of the study area, type of
57 surrogate, and the spatial resolution of surrogate data [e.g. 13, 14, 15]. Nevertheless, it often remains
58 ambiguous to what extent a surrogate represents other levels of biodiversity, in particular across
59 different levels of organization.

60

61 Biodiversity surrogates are usually subdivided into taxonomic and environmental surrogates [7, 10,
62 15]. Many studies have evaluated the efficacy of taxonomic surrogates for other taxonomic groups
63 [e.g. see 6, 16 for a review, 17-21]. The general consensus is that one taxonomic group alone might
64 not be an adequate surrogate for others [13, 14, 22, see 23 for a review, 24, 25], and the identification
65 of PAs should include more than one species or taxonomic group [14, 26]. Yet again, for many areas
66 in the world accurate species distribution data is scarce. One of the taxonomic groups for which rich
67 datasets are available are birds, because they are of interest to many people and are therefore one of
68 the best surveyed taxa in the world [26-28]. As such, birds are frequently used as biodiversity
69 indicators and conservation surrogates, and their surrogacy effectiveness varies from representing

70 overall species diversity well (other taxa than birds) [9, 23, 26, 29], or threatened birds being adequate
71 surrogates for non-threatened bird species [14], to being poor surrogates for other taxa [20, 30, 31].
72 Adding more taxa [26] or even different biodiversity features, such as environmental diversity [13,
73 32], increased the overall surrogacy of birds for other levels of biodiversity.

74

75 Environmental diversity (ED), in particular habitat diversity, has the potential to be a powerful
76 surrogate and represent other levels of biological organization, because habitat data can be generated
77 quickly and relatively inexpensively from remotely sensed or extrapolated ground data [7, 15, 23, 33-
78 36]. By using ED in conservation prioritization, it is assumed that selected areas do not only cover a
79 wide range of different environmental conditions, but also areas, rich in other biodiversity features,
80 such as species [35]. Furthermore, environmental surrogates may capture interactions between species
81 and their environment [32], and compensate for a potential lack of congruence between taxonomic
82 surrogates [22]. However, compared to taxonomic surrogates, the application of environmental
83 surrogates received less attention. One potential reason may be that so far no clear and explicit
84 recommendation about the efficiency of ED as a surrogate for biodiversity in general exists [e.g. 37,
85 38]. Moreover, there is no consensus on the way that ED is measured and included in a surrogacy
86 approach, e.g. continuous versus discrete measures of ED [e.g. 37]. Some studies suggested that
87 continuously distributed environmental variables (e.g. climate variables such as temperature and
88 precipitation, or vegetation characteristics such as percent tree cover) may outperform discrete
89 (classified) environmental data [e.g. 39, 40], but also may be inadequate [13, 23, 38] or at most better
90 than random surrogates for species occurrence [7, 36]. Other studies, however, found that categorical
91 environmental data in the form of pre-classified information (e.g. land classes, ecological vegetation
92 classes or habitat types), may nevertheless be better surrogates than continuous environmental data
93 [e.g. 15, 32, 37, 38]. The representation of habitat or land cover categories for other levels of
94 biodiversity may vary considerably, for instance being weak for plant species [32, 41], but better for
95 plants than for vertebrates [10, 13, 15, 42, 43]. Yet, such contrasting results could also result from
96 differences in the spatial extent and resolution of the study area, as well as the type of environmental

97 data used and the amount of different environmental features included as a surrogate (vegetation or
98 climate-based) [4, 15, 36, 44, 45].

99

100 Given uncertainties surrounding the potential for categorical habitat data to serve as a surrogate for
101 biodiversity, the goal of our study was to evaluate its representation of one of the most frequently used
102 biodiversity surrogates, bird species distributions, and vice versa. We implemented this analysis for
103 Romania, a country within the European Union exhibiting high bird species and habitat diversity,
104 likely caused by the variety of biogeographic regions it comprises [46, 47]. While 23% of Romania is
105 protected, either under the pan-European Natura 2000 network or as natural or national parks or
106 biosphere reserves [48], and despite its high levels of biodiversity, efforts to identify conservation
107 priorities and evaluate the efficacy of the network of reserves to protect biodiversity have been sparse
108 [mentioned by 11 but not examined, 21, 48-51]. One reason for this disparity is that species
109 distribution data have only recently become widely available. As such, PA management could greatly
110 benefit from prioritization efforts using systematic conservation planning principles and the latest
111 available data, particularly when establishing new PAs [48, 49]. The implementation of such scientific
112 research in the establishment and governance process of PAs is, however, often limited [52, 53]. This
113 is not a unique situation, as for instance Natura 2000 sites consist of a diverse array of reserves
114 designed for particular species, but not to protect biodiversity as a whole, so they often represent
115 species and habitat diversity only to a limited extent [38, 49, 54, 55]. Furthermore, the European
116 Commission decided to set new targets for 2030 and increase the percentage of protected areas in EU
117 member states to 30% [56]. Hence, there is a need to identify additional areas for protection, which is
118 best done using the principles of systematic conservation planning [57].

119

120

121 In this study, we first evaluated whether breeding bird species and habitat diversity based on remote-
122 sensing data are adequate surrogates for one another. We assessed surrogacy of the two datasets using
123 high-resolution data (1km) of (a) 137 modelled breeding bird species distributions and (b) 36 classes
124 of mapped habitat types from the “Ecosystem Types of Europe” (ETE) data set [58]. Second, we

125 evaluated whether existing protected areas (national and natural parks, biosphere reserves, wetland
126 reserves and SPAs (as part of the Natura 2000 network)) in Romania are effective in representing
127 areas of conservation concern for both birds and habitats. Finally, we identified additional areas that
128 could be prioritized in an effort to expand the current PA network to more comprehensively protect
129 bird and habitat diversity.

130

131

132 **Methods**

133

134 **Study region**

135 Romania is located in Eastern Europe, at the western shores of the Black Sea. It covers 238 397 km²
136 and natural landmarks and borders are dominated by the Carpathian Mountains and the Danube River.
137 Five biogeographical regions have been characterized across Romania: Pannonian, Continental,
138 Alpine, Steppic, and Black Sea. The heterogeneous landscape consists of an alternation between
139 intensive and extensive agricultural areas and (semi-) natural areas, such as forest, open woodland, and
140 grassland. As a member of the European Union, Romania is bound to the directives of the Natura 2000
141 network, and dedicated about 23% of its total landscape to conservation. The Natura 2000 network is
142 an important biodiversity conservation measure [11], and consists of different types of protected areas:
143 the terrestrial Special Protection Areas (SPA, for bird protection only), the terrestrial Sites of
144 Community Importance, and Special Areas of Conservation (SCI and SAC, for habitats and/or
145 species) [59]). In addition to, but partly overlapping with the Natura 2000 network, Romania also
146 implemented protected areas designated as natural and national parks as well as biosphere reserves
147 [48].

148

149 **Biodiversity features**

150

151 **Bird species distributions**

152 Bird species occurrence data (from the years 2006-2018) were obtained from the forthcoming Romanian
153 Breeding Bird Atlas ([60], in preparation), a scheme run by Milvus Group Association and the Romanian
154 Ornithological Society. Because data based on species atlases often entail a geographic sampling bias
155 [61], we first modelled the distributions of 137 breeding bird species using MaxEnt 3.4.1. [62] at two
156 different resolutions (1x1km and 2x2km), depending on the species' ecology or in some cases by the
157 available data (S1 Table). In-depth details on the species distribution modeling approach are provided
158 in S1 Supporting information.

159

160

161 **Habitat types**

162 We used the published maps of habitat types classified in the “Ecosystem Types of Europe” (ETE)
163 data set (version 3.1)[58]. ETE is a combination of the non-spatially referenced EUNIS (European
164 Nature Information System) habitat classification scheme and a spatially explicit habitat data set, the
165 Corine-based “Mapping and Assessment of Ecosystem and their services (MAES)” ecosystem classes
166 [63]. In Romania, 42 ETE habitat classes are mapped (level 2 classification) at 100 m resolution (S2
167 Table). Habitat classes including highly built-up areas (six classes) were excluded in the subsequent
168 spatial conservation mapping. These built-up areas were selected according to the ETE classification
169 category "J" (J1-J6, see S2 Table), which include buildings in cities and villages, industrial sites,
170 transport networks, artificial water structures and waste deposits.

171 To produce maps of habitat types that match the spatial resolution of those for the bird species, we
172 split the pre-defined ETE data set into single data layers per class (36 in total) and calculated the
173 proportion of each habitat type within 1 km² grid cells.

174

175 **Data handling**

176 All spatial data layers were re-projected to the Dealul Piscului 1970/ Stereo 70 projection and
177 processed at a 1 km resolution containing a total number of 381 248 grid cells. Species distribution

178 models at 2 km resolution were resampled to 1 km grid cell size. Preparation of input maps and post-
179 processing of results was done in R (version 3.6.1), using the packages (zonator, raster, rgdal, rgeos,
180 sp, maptools, tiff, data.table, plyr, dplyr, ggplot2, zoo). Maps were visualized in QGIS (version 3.10.6
181 'A Coruña').

182

183 **Spatial conservation prioritization**

184 We prioritized areas for conservation using the software Zonation 4.0 [64]. Zonation can handle large
185 data sets [65] and provides a priority ranking over the entire landscape rather than satisfying a specific
186 target. The ranking is produced by iteratively discarding locations (grid cells) with the lowest
187 conservation values, retaining the ones with the highest conservation value throughout the process [66,
188 67].

189

190 We used the additive-benefit function (ABF), which directly sums up the conservation value across
191 features [68] and results in a reserve network with high average performance across all features [69].
192 The ABF algorithm is appropriate for our study since we aim to identify areas representing overall
193 richness rather than core areas that lead to the equal representation of both common and rare species or
194 habitats. The algorithm accounts for the total and remaining distributions of features, and optional
195 feature-weights can be implemented [13]. We equally weighted habitat types and bird species
196 distributions at the aggregate level to avoid prioritization biases due to the different numbers of
197 features contained within (e.g., combined weights for 137 bird species or for 36 habitat types summed
198 to 1). To exclude land uses that for administrative or ecological reasons did not contribute to either
199 overall conservation value or to the expansion of protected areas (six classes of built-up area), we
200 applied a cumulative negative weight of -1 to these layers [68].

201

202 Performance curves were produced with the R package 'zonator' [70]. These curves show the
203 proportion of the original occurrence of features remaining in the landscape as a function of the
204 proportion of the landscape that is lost [71]. The curves start at 1.0, where the entire distribution of
205 features is represented in the full landscape, and end at 0.0, where the entire landscape is lost.

206 Because we observed a wide spread in the performance curves of the bird species, we explored
207 potential underlying patterns related to their broad habitat requirements. We grouped species into their
208 preferred breeding habitat types (S1 Table) to assess differences between groups and their
209 performance when the prioritization is accounting for all bird species.

210

211 We also suspected that the range size of feature types, in particular within bird species, influences their
212 performance in the prioritization. Specifically, we assumed that range restricted species would perform
213 better, since these species might be retained throughout many prioritization iterations. Yet, this may
214 only be the case when range-restricted features largely overlap with more widespread features. To
215 explore this further, we calculated the AUC (area under the curve) of each feature performance curve,
216 and plotted these as a function of range size (Fig 1). For bird species we calculated range sizes by
217 summing the Maxent probabilities, and for habitat types we summed the area in km².

218

219 **Figure 1** The Zonation performance of individual features (AUC) as a function of its corresponding
220 range size. (a) Individuals bird species, belonging to one of the four breeding habitat groups. Green
221 triangle = forest to (dense) woodland; grey cross = generalist and close to humans; yellow square =
222 arable land, open woodland to grassland; blue square = wetland and shores. The values for the range
223 sizes of bird species were computed by adding up Maxent species distribution values. (b) Individual
224 habitat types

225

226 **Surrogacy analyses**

227 We evaluated the reciprocal surrogacy of bird species and habitat types, and assessed the efficacy of
228 the existing network of protected areas to protect these biodiversity features. To test the surrogacy of
229 the two feature types, we ran separate analyses using one feature type as the surrogate and the other as
230 the target. To do so, bird species and habitat types were both included in each run, but positive weights
231 (=1) were only assigned to the surrogate, while the target was assigned a weight of 0.

232

233 We evaluated the surrogacy power of each feature type using the performance curves. A performance
234 curve by itself provides, however, little information, and for correct interpretation it should be
235 compared to an optimal and a random curve [23]. For instance, when testing whether habitat types are
236 a good surrogate for bird species, the optimal curve is equivalent to the surrogacy of bird species for
237 themselves. The random curve in this scenario reflects the representation of bird species expected in
238 the absence of biological data, when ‘area’ is used as a surrogate [72]. Qualitatively the surrogacy
239 value can be assessed visually by comparing the three curves. The closer the target curve is to the
240 optimal curve, the higher the surrogacy value. To quantify the surrogacy power, we calculated an
241 equivalent to the species accumulation index (SAI; Ferrier (73)):

$$242 \quad SAI = (S - R) / (O - R),$$

243 where S is the area under the target curve, R is the area under the random curve, and O is the area
244 under the optimal curve. The optimal curve was extracted from the runs when targets were used as a
245 surrogate themselves. To create the random curve, we executed 100 surrogacy runs with randomly,
246 uniformly distributed data as a surrogate and bird species and habitat types as targets. We used the
247 mean of the corresponding target curves to calculate SAI.

248

249 **Evaluation and potential expansion of the protected area network**

250 To evaluate the representation of habitat and birds in existing reserves, we specifically focussed on
251 SPAs, national and natural parks, and biosphere reserves. We thus excluded the SCI and SAC areas
252 (Natura 2000 sites), since they are designed to protect specific species or habitats, but do not
253 necessarily protect others - or even biodiversity as a whole. To evaluate the effectiveness of the current
254 network in Romania, we tested 1) how well current PAs represent areas of conservation concern for
255 bird species and habitat types, and 2) how much of the individual feature type’s distributions are
256 represented within the current network. Furthermore, we 3) assessed which areas should be prioritized
257 when expanding the current conservation network.

258

259 The analyses for 1) and 3) were based on a Zonation prioritization outputs, where both bird species
260 and habitat types had been considered simultaneously. We did not differentiate between protection

261 levels of the existing PAs. If PAs had been selected indiscriminately, we expected that Zonation
262 values within PAs would be uniformly distributed, as they are across the entire study region. We thus
263 tested the frequency of Zonation values within PAs against a uniform distribution using a Chi-square
264 test. For 2) we summarized the distribution of bird species and habitat types within current PAs as a
265 proportion of their total distribution via boxplots (S1 Fig b and c).

266

267 To identify potential areas that should be prioritized when expanding the current network of PAs, we
268 performed a mask analysis [74]. In this analysis, current PAs are included as a mask layer, and are
269 assigned a high rank (=1) in the final prioritization map. As such, the next highly ranked areas outside
270 protected areas can be identified as potential expansion areas that represent bird and habitat diversity
271 well.

272

273 **Results**

274

275 **Spatial conservation prioritization**

276 Both the separate and combined prioritization using bird species and habitat types resulted in broadly
277 similar patterns, with highly ranked areas in the Carpathian Mountains, river valleys and parts of the
278 Danube Delta. However smaller-scale differences are apparent, in particular with respect to the size
279 and clustering of those areas (Fig 2a and S2 Fig).

280

281 **Figure 2** Study region with Zonation ranking based on bird species and habitat type data without (a)
282 and with (b) considering currently protected areas (mask analysis). Colors indicate importance ranking
283 scores for conservation, with 0 meaning lowest importance and 1 meaning highest importance. Built-
284 up areas are indicated in white and were excluded from prioritization. Purple in panel (b) indicates
285 current protected areas

286

287 The overall performance of bird species for themselves was rather low (AUC=0.65, area under the
288 bird performance curve) (Fig 3a and S3 Table), but we observed considerable differences between
289 groups based on breeding habitat (S3 Table).

290

291 **Figure 3** Performance and surrogacy curves quantifying the average proportion of original feature
292 distributions represented as landscape is lost. Built-up areas were negatively weighted and hence
293 excluded from the prioritization (lower dashed line). The area between the target curve and the random
294 curve divided by the area between the optimal curve and random curve represents the efficacy of the
295 surrogate (SAI; Species accumulation index). In panel (a) bird species were used as a surrogate for
296 habitat types and in (b) habitats were used as a surrogate for birds

297

298 Wetland and shore-breeders were best retained through the ranking process, followed by those
299 breeding in “forest to (dense) woodland” areas (Fig 4). In contrast, birds breeding in “arable land, open
300 woodland to grassland” or being “generalist and close to humans” were lost much more quickly (Fig
301 4).

302

303 **Figure 4** Performance curves for bird species split by breeding habitats. The solid line is the average
304 performance curve of all bird species used in the surrogacy approach. Built-up areas were negatively
305 weighted and hence excluded from the prioritization (lower dashed line)

306

307 To explore this further, we plotted each species’ performance as a function of its range size (Fig 1 and
308 S1 Table), and found a clear negative trend. “Wetland and shore” breeders include more range-
309 restricted species compared to other groups and at the same time performed best in the prioritization,
310 whereas forest, generalist and grassland birds overall have larger ranges, and performed worst in the
311 prioritization. In addition, the distributions of wetland and shore breeders often overlap with those of
312 other groups, those resulting in areas of high species richness that are preferentially prioritized by the
313 ABF algorithm (S3 Fig).

314 Habitat types were generally retained well throughout the prioritization process (AUC=0.9, area under
315 the habitat surrogate curve) (Fig 3b and S3 Table). We observed that features with smaller ranges were
316 retained the longest (Fig 1 and S2 Table).

317

318 **Surrogacy analyses**

319 Birds were a moderately good surrogate for habitats (SAI = 0.60). Interestingly, birds represented
320 habitats better than themselves (Fig 3a), although as shown above this is only true for the
321 representation of all birds combined, and there are large differences between bird groups (Fig 4). The
322 representation of habitats for birds on the other hand, was less effective (SAI = 0.44; Fig 3b).

323

324 **Evaluation of protected areas and identification of expansion**

325 **regions**

326 We found that the Zonation values within current PAs, when both habitat types and bird species were
327 considered, differ significantly from a uniform distribution, with an overrepresentation of higher
328 values (Chi² test, Chi² = 29289, df = 9, p-value < 2.2e-16) (S1 Fig a). These results suggest that current
329 PAs generally comprise areas of high conservation value better than would be expected based on a
330 random assignment of areas for conservation. However, current PAs also comprise a considerable
331 amount of land surface area with relatively low conservation values based on bird and habitat
332 diversity, suggesting that improvements could be made.

333

334 Habitat types are relatively well represented in the current protected areas network (S1 Fig c), with the
335 exception of grassland, heathland and woodland habitats. Among the breeding groups, generalist and
336 grassland breeders are on average represented less well than expected under a random assignment,
337 although in the grassland breeding group much variation between the species can be observed (S1 Fig
338 b).

339

340 The mask analysis highlighted transition areas from highland to lowland regions, such as along the
341 northern Carpathian Mountains, the eastern foothills of the Carpathian Mountains, and the eastern part
342 of the Apuseni Mountains (Fig 2b) as particularly important expansion sites for bird and habitat
343 conservation.

344

345 **Discussion**

346

347 The necessity to rely on surrogates for conservation prioritization raises the question of how effective
348 they are. Here we evaluated the mutual surrogacy power of bird species and habitat types in Romania,
349 an area in Europe with high biodiversity, and demonstrated that neither birds nor habitat types are
350 effective surrogates of one another. Birds represented 60% of habitat conservation priorities, while
351 habitats were less effective at representing bird conservation priorities (44%). These results are not
352 only concordant with studies in other regions suggesting to use more than one type of surrogate for
353 conservation prioritization [13, 32, 42], but also point out that environmental data as conservation
354 surrogates for species should be carefully evaluated prior to applications of protected area expansions.
355 We also found that existing protected areas in Romania capture areas of high conservation value for
356 both biodiversity features better than expected at random, but could potentially be designed more
357 effectively and more efficiently. Finally, we identified additional areas that should be prioritized in
358 case the existing network were to be expanded under the European Union Biodiversity Strategy to
359 2030, or where conservation strategies for conserving avian and habitat diversity on private lands
360 could be incentivized.

361 **Bird species as a surrogate**

362 The effectiveness of 137 breeding bird species as a surrogate for habitats was ~60% of that of habitats
363 for themselves (Fig 3a). Thus, in the absence of other data, birds could represent habitat types better
364 than random, but only to a limited extent. These results appear robust because we included many bird
365 species, breeding in a wide variety of habitats (S1 Table), thus covering the existing habitat diversity

366 quite well. Our results corroborate other studies that found that taxonomic groups are poor surrogates
367 for one another [for other taxonomic groups, e.g. 13, 14, 25, 26], and should be used cautiously as
368 surrogates for habitat diversity.

369

370 Interestingly, when prioritizing bird species only (Fig 4) we found that wetland and shore birds were
371 much better represented than forest, grassland, and generalist species. This unexpected result
372 corroborates the focal areas of the Bird Directive, which demands particular attention to wetland
373 species [75, Art 4 (2)]. A potential explanation for this representation bias is the emphasis of the
374 additive-benefit function (ABF) on high average performance across all features - in the case of bird
375 species, areas with high species richness [69] – combined with differences in range sizes between the
376 bird groups [14, 67]. We found that species richness was highest in areas where the distributions of
377 wetland-breeding species overlapped with those of species breeding in other types of habitat (S3 Fig).
378 Because wetland birds generally have small ranges due to the limited availability of suitable habitat
379 [76], Zonation prioritized the species-rich wetlands over areas with fewer species, where more widely
380 distributed species occur (Fig 1). These results are in line with similar patterns in small versus large-
381 range moths [20], butterflies, reptiles, and amphibians [14]. The representation bias in our study may
382 be exacerbated by associations of generalist species to human-dominated landscapes. Because we
383 negatively weighted and hence excluded built-up areas from the prioritization, species occurring in
384 those areas may be underrepresented in the final results.

385

386 **Habitats as a surrogate**

387 Habitats as a surrogate for birds were only 44 % as effective as the maximum possible. This result is
388 consistent with other studies showing that environmental diversity may have limited suitability as a
389 proxy for the diversity of small vertebrates (including bird species) [32, 77]. Yet, habitats represented
390 birds better than random (Fig 3b), potentially due to the influence of habitat structure on bird species
391 occurrence and distributions [43]. It remains unclear whether higher spatial and thematic resolutions –
392 in particular more detailed habitat classifications – could improve the mutual representation.

393

394 The underlying concept of using environmental diversity (ED) as a conservation surrogate is that by
395 selecting areas that cover a wide range of environmental conditions, other levels of biodiversity should
396 be covered equally well [35]. The surrogacy power of ED may, however, depend on how it is tested
397 and implemented in prioritization schemes, and whether it covers multiple taxonomic groups. Some
398 studies suggested that pre-classified environmental data such as the ETE dataset [58] perform equally
399 well or better than continuous environmental variables as a surrogate for species diversity [e.g. 15, 32,
400 37, 42, 78]. However, inconsistencies in the application of ED as a conservation surrogate and in what
401 form it should be implemented (e.g. as discrete classes or continuous variation), are still unresolved,
402 and also likely depend on the types of species data available. For instance, implementing ED as a
403 distance matrix across as many environmental variables as possible, in combination with a different
404 optimization procedure may in some cases help improve its surrogacy power [e.g. 36]. Yet, continuous
405 environmental data such as climate or vegetation characteristics such as density or cover may fail to
406 capture potentially important attributes summarized in classified habitat data like the ETE. To this end,
407 more empirical evidence is needed from direct comparisons between methods as well as the
408 performance of different measures of ED to provide a solid basis for recommendations of best
409 practice.

410

411 Given the still existing uncertainties surrounding the use of ED, we set out with pre-classified habitat
412 types, which are themselves based on a wide range of climatic and environmental conditions [58].
413 Within the methodological and geographical scope of our study, results suggest that habitat classes
414 performed relatively poorly at representing bird biodiversity in Romania, and ideally should not be
415 used on their own in prioritization efforts. Instead, combining taxonomic and environmental surrogates
416 could increase the surrogacy power for the protection of overall biodiversity [13, 42], but a single
417 taxonomic group may not suffice. For instance, habitats and birds did not perform well in representing
418 amphibians and reptiles in other areas [15, 38, 43]. Thus, we recommend to combine environmental
419 and taxonomic surrogates, preferentially from multiple taxonomic groups.

420

421 **Representation in existing protected areas and conservation**

422 **implications**

423

424 We found that a considerable fraction of PAs is located in areas with high conservation values. It is
425 important to stress, however, that our evaluations by no means suggest that the current network of PAs
426 is sufficient. Around 23% of Romania's land surface area is currently under protection, and
427 improvements to the protected area network may be necessary [48, 49]. Large ecoregions and several
428 widespread bird and mammal species may be protected sufficiently well, but smaller ecoregions, as
429 well as invertebrate and plant species are for example underrepresented in the existing Natura 2000
430 network [49]. The current network of PAs consists of reserves designed for various purposes. In our
431 evaluation, we specifically focused on those that have been designed to protect birds, habitats, or
432 biodiversity as a whole, i.e. SPAs, national and natural parks, and biosphere reserves. We found that
433 these PAs represent areas of high bird or habitat conservation value better than a random assignment
434 of areas for protection. However, habitats were better represented than birds (S1 Fig b and c). We also
435 found that rare habitats are well represented, which is consistent with results for the Czech Republic
436 [54]. These habitats typically are wetlands and shores, large areas of which are protected in the
437 Danube Delta. Surprisingly, the representation of grassland and woodland habitats was rather poor. A
438 likely reason for this result is the large area of wood- and grassland habitats in Romania, only part of
439 which can be represented in PAs (S4 Fig). In contrast, rare habitats such as littoral areas are
440 represented at high percentages, because they can be entirely contained within a fraction of the total
441 land surface area. Despite the fact that current PAs capture important areas for conservation relatively
442 well, a tail of areas with low conservation value can also be observed (S1 Fig). It remains unclear
443 whether these areas may be important for other reasons, such as for other taxonomic groups, or as
444 corridors between areas of high conservation value. Yet, the presence of areas with low conservation
445 value also suggests that improvements in both the efficiency and efficacy of the network may be
446 possible. To this end, we identified areas that should be prioritized based on bird and habitat diversity
447 in a scenario of future expansions of the current network. A recent study suggests that such

448 improvements may best be developed at the level of Biogeographical Regions rather than at the
449 national level [57].

450

451 Protected areas are a crucial component of conservation, but the identification and designation of PAs
452 is often a lengthy and difficult process. In addition, even when the new targets for the EU Biodiversity
453 Strategy are met, 70% of the land surface area will remain unprotected. Hence, effective conservation
454 also depends on the protection of biodiversity outside of PAs. To do so, the development of incentives
455 for targeted management practices to retain high diversity of species and habitats should be prioritized
456 [79], yet scientific research that can support management decision is largely lacking [80].

457

458 We conducted this study in Romania, a heterogeneous and species-rich country, comprising five
459 biogeographical regions. We exploited an extensive data set on breeding bird species that has only
460 recently become available for the country. Our study adds to the body of evidence that taxonomic and
461 environmental surrogates represent one another only to a limited extent. Hence, the use of just one
462 type of surrogate may not capture the broad patterns of biodiversity sufficiently well. This situation is
463 less than ideal, as conservation measures respond to the biodiversity crisis, with little time to collect
464 data on the distribution of species or habitats. Although these data are becoming increasingly
465 available, our results highlight the need for investing in survey and monitoring schemes in countries
466 such as Romania, where data still remains relatively scarce. Our study also presents an example of the
467 importance of scientific research in informing conservation strategies as a stakeholder, which often
468 remains underrated [52, 53].

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474 **References**

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669

670 **Supporting information - Figures captions**

671 **S1 Figure** (a) Barplot of conservation values of areas in current reserves. The horizontal dashed line
672 indicates the expected frequency of each conservation value (freq = 3338.6), had the current PAs be
673 selected at random. The high frequencies of high conservation values, combined with the low
674 frequencies of low conservation values suggest that current PAs were selected efficiently. (b-c) Box-
675 and-whisker plots for birds (b) and habitats (c) showing the proportion of the total distribution of each
676 group of feature types that is represented in the existing protected area network. A dotted line indicates
677 the random expectation for the representation of each feature class based on the amount of protected
678 area in Romania (~ 23% of land surface area)

679

680 **S2 Figure** Study region with Zonation ranking based on (a) Bird species and (b) habitat types. Colors
681 indicate importance ranking scores for conservation, with 0 meaning lowest importance and 1 meaning
682 highest importance. Built-up areas are indicated in white and were excluded from prioritization

683

684 **S3 Figure** Overlapping bird species occurrences per breeding habitat group: (a) forests to (dense)
685 woodland, (b) generalist and close to humans, (c) arable land, open woodland to grasslands, and (d)
686 wetlands and shores. Red indicates species-rich areas; white to grey indicate no or low overlap of
687 species occurrences

688

689 **S4 Figure** Study region with (a) forest habitats and (b) grassland habitats highlighted. The used
690 protected area network is highlighted in grey

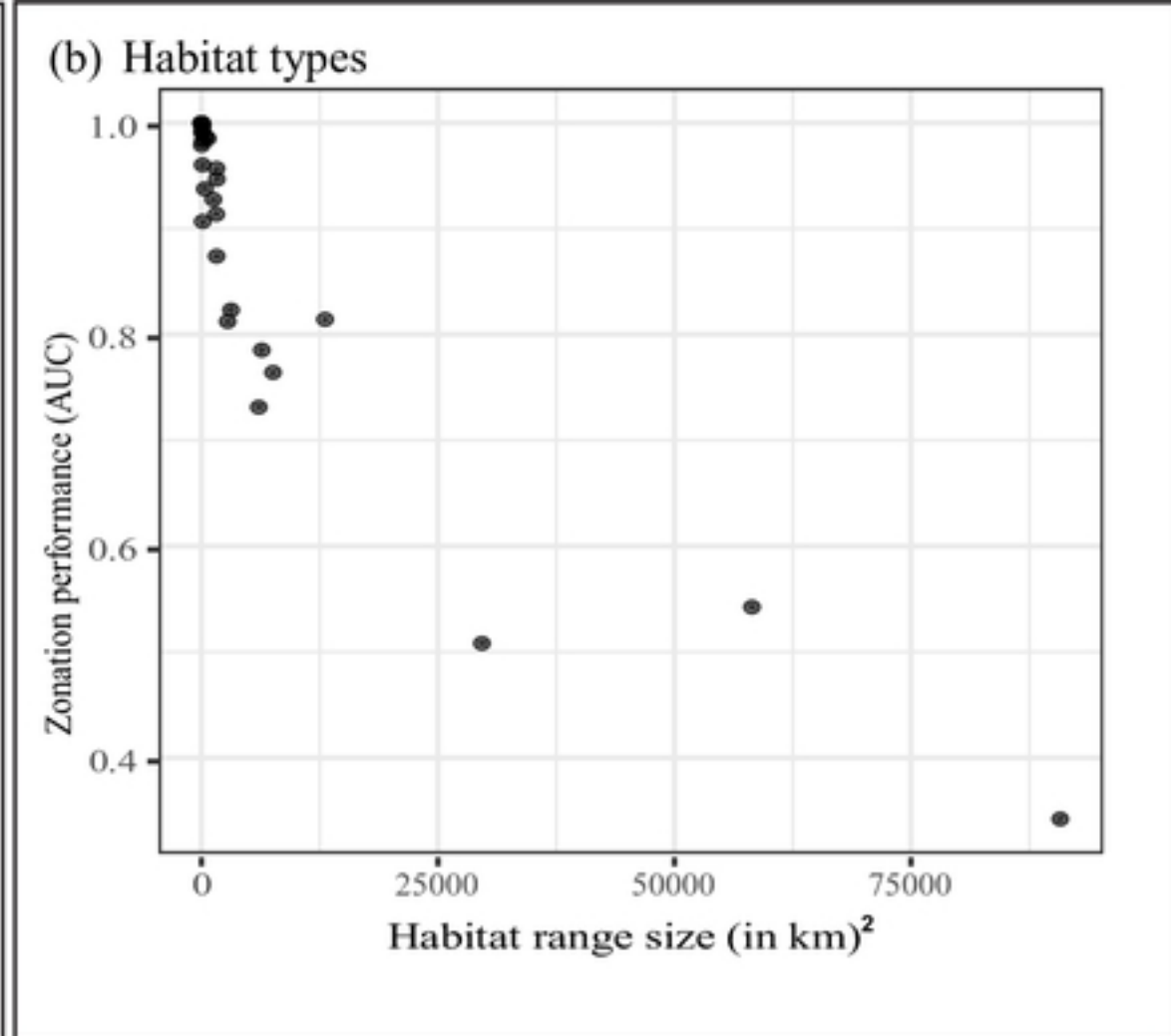
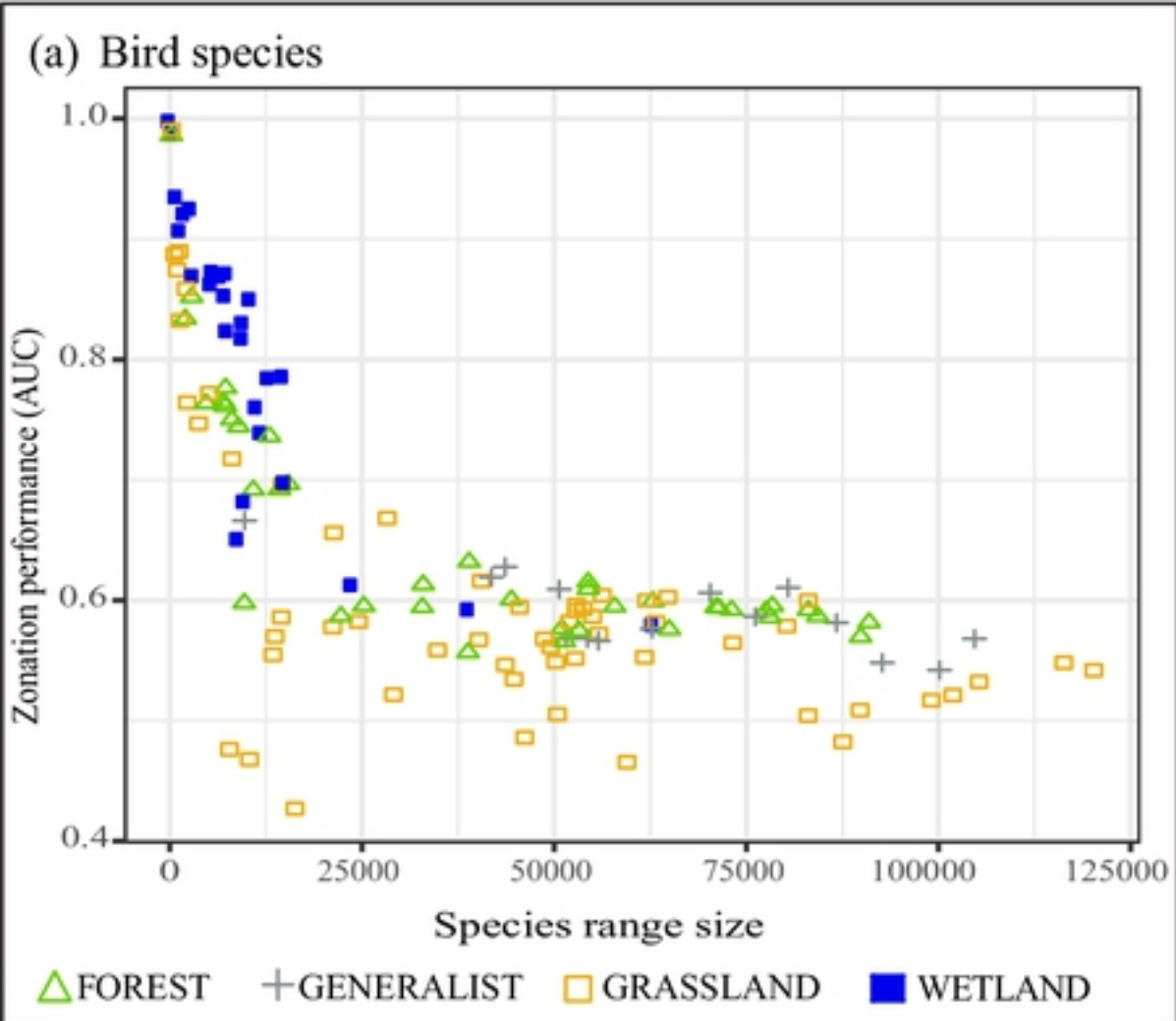


Figure 1

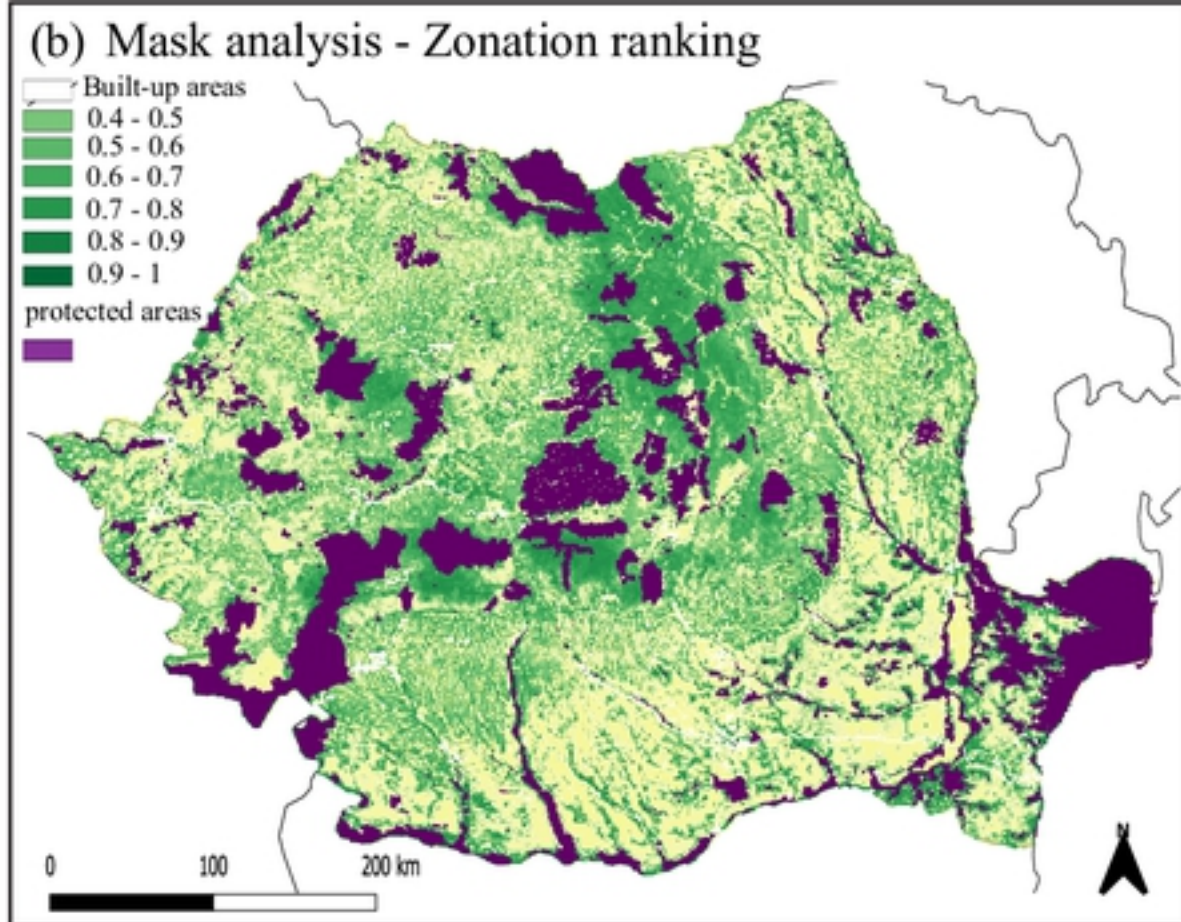
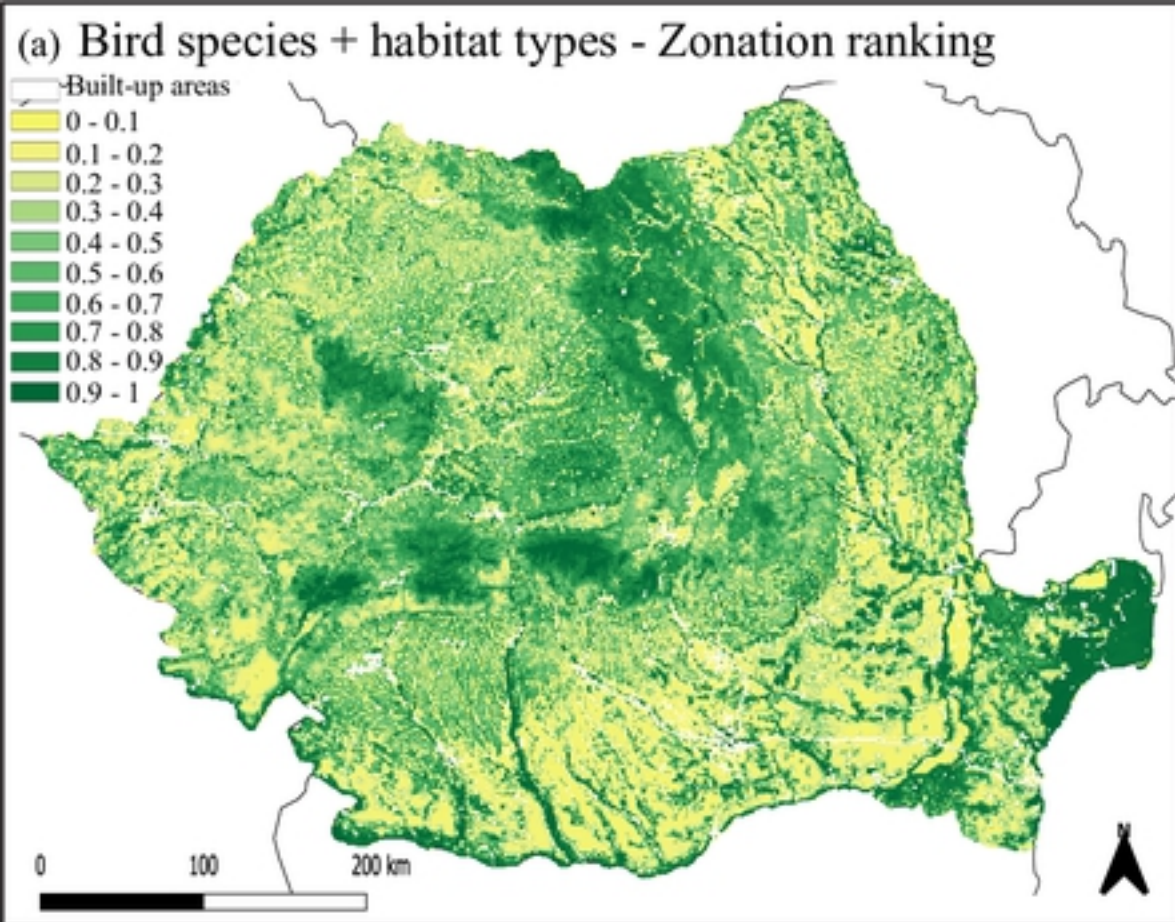


Figure 2

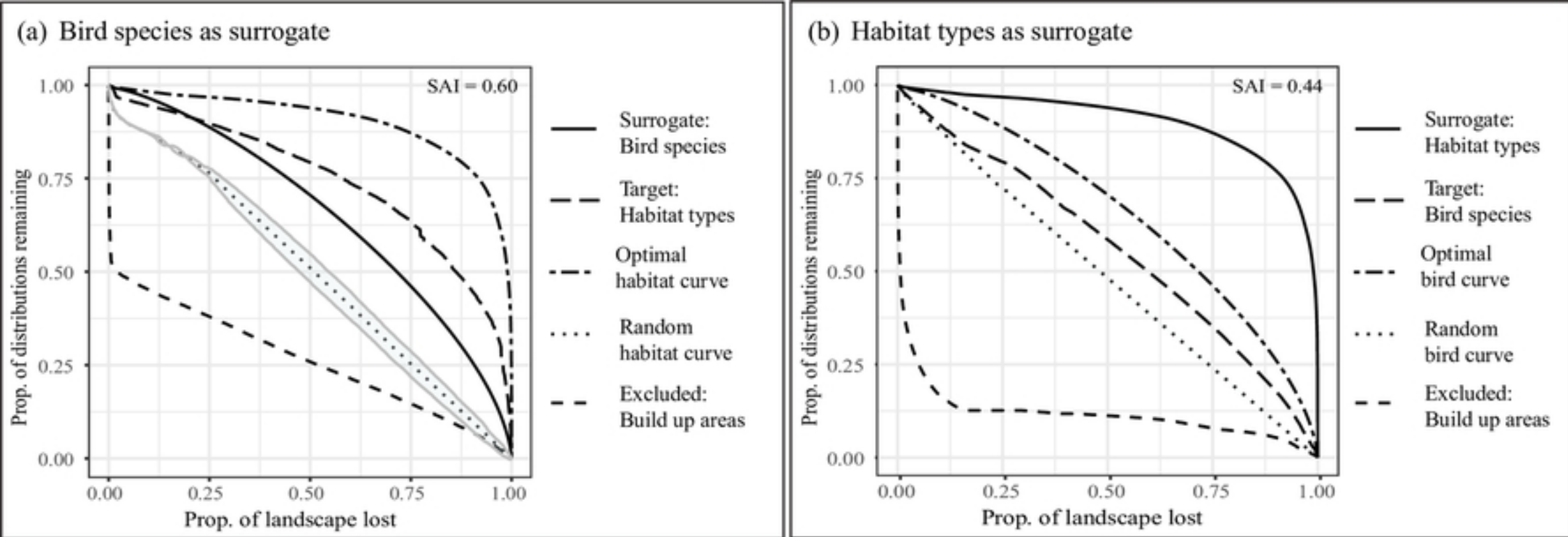


Figure 3

Bird species groups

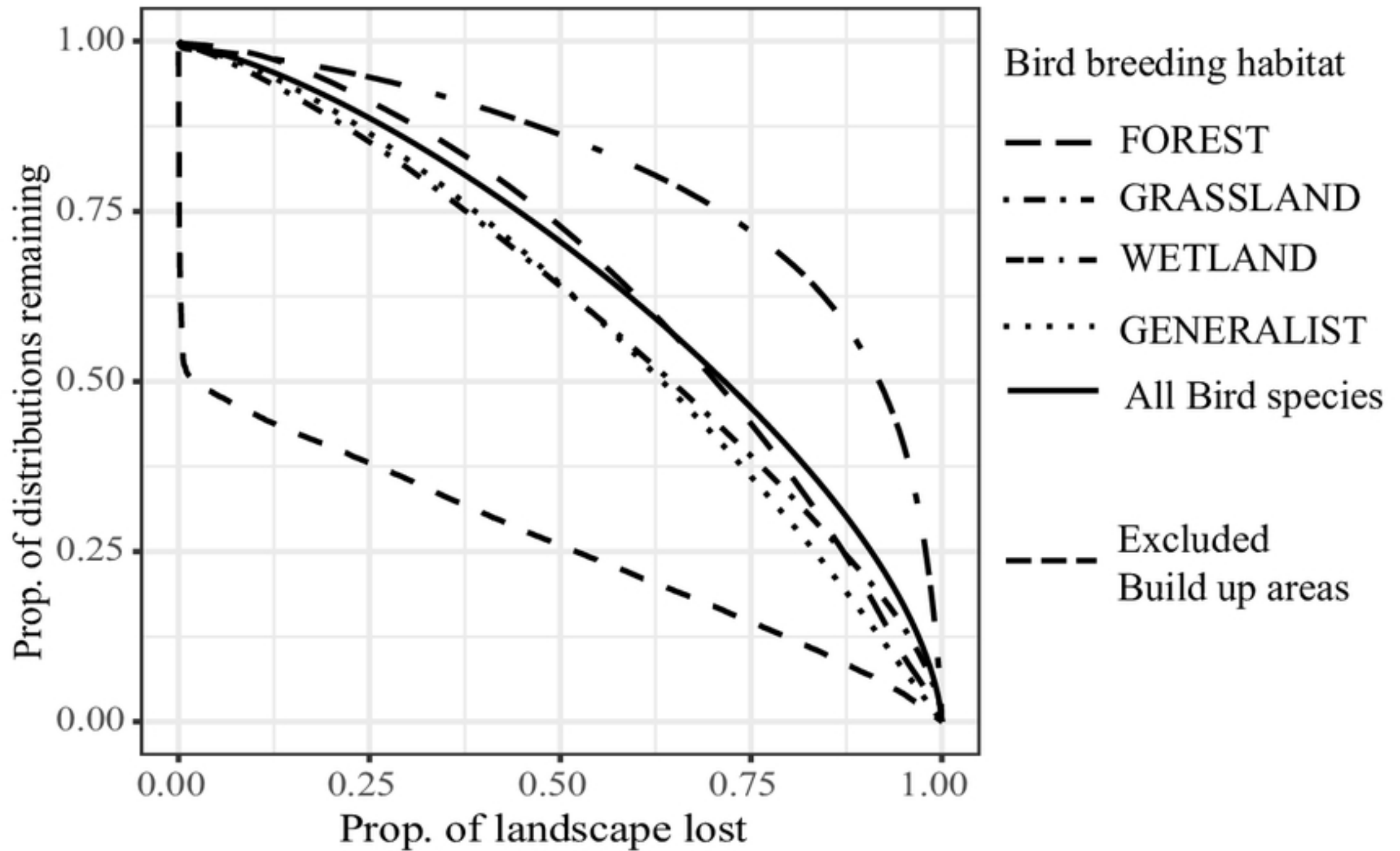


Figure 4