

1 **Seed morphological traits as a tool to quantify variation maintained in *ex situ* collections: a**  
2 **case study in *Pinus torreyana* (Parry)**

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11

12 **Abstract**

13 Background: Understanding the within- and among-population distribution of trait variation  
14 within seed collections may provide a means to approximate standing genetic variation and  
15 inform plant conservation.

16 Aims: This study aimed to estimate population- and family-level seed trait variability for existing  
17 seed collections of Torrey pine (*Pinus torreyana*), and to use these data to guide sampling of  
18 future collections.

19 Methods: We quantified variation in 14 seed morphological traits and seedling emergence within  
20 and among Torrey pine populations. Using a simulation-based approach, we used estimates of  
21 within-population variance to assess the number of maternal families required to capture 95% of  
22 trait variation within each existing seed collection.

23 Results: Substantial structure was observed both within and among Torrey pine populations, with  
24 island and mainland seeds varying in seed size and seed coat thickness. Despite morphological  
25 differences, seedling emergence was similar across populations. Simulations revealed that 83%  
26 and 71% of all maternal families within island and mainland seed collections respectively needed  
27 to be resampled to capture 95% of seed trait variation within existing collections.

28 Conclusions: From a conservation perspective, our results indicate that to optimize genetic  
29 diversity captured in Torrey pine seed collections, maximizing the number of maternal families  
30 sampled within each population will be necessary.

31

32 **Keywords**

33 ex situ conservation; island-mainland phenotypic differentiation; seed trait variation; seedling  
34 emergence; Torrey pine; variance partitioning.

## 35 **Introduction**

36 Ex situ seed collections preserve species genetic diversity outside of their native range, providing  
37 the raw material for species reintroductions and germplasm to augment restoration (Guerrant Jr  
38 *et al.* 2014; Potter *et al.* 2017). Ensuring ex situ collections represent genetic variation found in  
39 natural populations is critical to both contemporary conservation and potential future restoration  
40 efforts (Schaal and Leverich 2004; Basey *et al.* 2015). An invaluable conservation resource,  
41 particularly for rare species, ex situ collections protect against biodiversity loss in the wild, while  
42 preserving species' evolutionary potential. However, the cost and logistical constraints  
43 associated with seed collection pose a significant challenge. Given this challenge, means are  
44 needed to optimize ex situ sampling efforts (Hoban and Schlarbaum 2014; Di Santo and  
45 Hamilton 2020).

46  
47 One approach may be to use the distribution of trait variation existing within contemporary *ex*  
48 *situ* seed collections as a proxy for quantifying standing genetic variation within and among  
49 populations. Although multiple factors influence plant phenotypes (Monty *et al.* 2013; Villellas  
50 *et al.* 2014), seed morphological variation is often considered highly heritable. For example, seed  
51 length, seed width, and seed mass have a heritability (or repeatability) estimated between 0.33  
52 and 0.98 in conifers, including maritime pine (*Pinus pinaster*), chir pine (*Pinus roxburghii*), and  
53 white spruce (*Picea glauca*) (Roy *et al.* 2004; Carles *et al.* 2009; Zas and Sampedro 2015). In  
54 addition, traits such as seed shape, seed coat thickness, or embryo length also exhibit high  
55 heritability, with values estimated between 0.59 and 0.96 for agronomic species, including  
56 soybean (*Glycine max*), narrow-leafed lupin (*Lupinus angustifolius*) and rice (*Oryza sativa*)  
57 (Pandey *et al.* 1994; Cober *et al.* 1997; Mera *et al.* 2004; Hakim and Suyamto 2017). Given these

58 observations, variation in seed morphological traits likely has a genetic basis and may reflect  
59 standing genetic variation within and among populations. In addition, morphological variation of  
60 seeds stored *ex situ* may reflect variation attributable to the maternal environment (Platenkamp  
61 and Shaw 1993; Singh *et al.* 2017). However, for rare species where existing genetic data are  
62 limited, quantifying within and between population variation for traits largely considered  
63 heritable within existing seed collections may be invaluable to optimizing future collections,  
64 even if estimates of genetic variation do not control for maternal environment.

65

66 The distribution of heritable genetic variation estimated via common garden experiments –  
67 experimental approaches used to understand the genetic contribution to phenotypic variation  
68 under common environmental conditions – (Weber and Kolb 2014; Hamilton *et al.* 2017; Yoko  
69 *et al.* 2020) or molecular genetic data (Zhang and Zhou 2013; Hausman *et al.* 2014; Tamaki *et*  
70 *al.* 2018) can be used to quantify the distribution of standing genetic variation. However, when  
71 common garden experiments or molecular genetic data are unavailable, quantifying trait  
72 variation within and among *ex situ* seed population collections may provide a reasonable proxy  
73 for the distribution of genetic variation. Millions of seed accessions have been stored in gene  
74 banks internationally (FAO 2010), representing a large conservation and research resource.

75 Although common garden experiments are preferred when available, heritability of seed  
76 morphological traits and ease of access to seeds through existing *ex situ* collections suggests that  
77 quantifying seed morphological variation may provide a timely approach to estimating variation  
78 preserved in collections. In addition, where the goal is to limit the loss of biodiversity and  
79 preserve evolutionary potential for rare species, existing seed morphological data may be  
80 leveraged to optimize supplemental conservation collections.

81  
82 *Pinus torreyana* Parry (Torrey pine), is one of the rarest pines in the world (Critchfield and Little  
83 1966; Dusek 1985), endemic to two discrete natural populations in California. Torrey pine  
84 occupies one mainland population (*Pinus torreyana* subsp. *torreyana*) of approximately 6,000  
85 trees at the Torrey Pine State Reserve in La Jolla, CA, and an island population (*Pinus torreyana*  
86 subsp. *insularis*) of approximately 3,000 reproductively mature trees on Santa Rosa Island, CA,  
87 one of the Channel Islands (Ledig and Conkle 1983; Haller 1986; Hamilton *et al.* 2017) (Fig. 1).  
88 Listed as critically endangered by the IUCN (2020), Torrey pine is of critical conservation  
89 concern due to multiple factors, including low population size (Franklin and Santos 2011; Hall  
90 and Brinkman 2015), low genetic diversity (Ledig and Conkle 1983; Whittall *et al.* 2010),  
91 climate change, and environmental and human-mediated disturbances (Franklin and Santos  
92 2011; Hamilton *et al.* 2017). While in situ conservation has preserved the whole of the species'  
93 range, with fewer than 10,000 reproductively mature individuals in native populations, there are  
94 substantial risks for population-level extirpation events. To mitigate potential losses in the wild,  
95 conservation efforts have focused on preservation of seed ex situ. While ex situ seed collections  
96 provide an invaluable conservation resource, they may also be used to quantify species' trait  
97 variation needed to inform future conservation efforts.

98  
99 In this study, we evaluate morphological trait variation in a large ex situ conservation collection  
100 of Torrey pine seed sourced from the two native extant populations. Specifically, we quantify the  
101 distribution of variation for 14 seed morphology traits and assess differences in emergence  
102 between island and mainland seedlings. In addition, we use existing ex situ collection data to  
103 provide supplemental population sampling guidance for future Torrey pine collections. For this

104 latter objective, we use simulations to estimate the number of maternal families required to  
105 capture 95% of seed morphological variation existing in contemporary ex situ collections, for  
106 both island and mainland population independently. This study evaluates the distribution of seed  
107 morphological variation in ex situ collections as a proxy for standing genetic diversity,  
108 quantifying variation attributable to within and between population differences. These data are  
109 then used to inform population sampling necessary to meet conservation objectives in future seed  
110 collections. Although presented here using Torrey pine, our approach is broadly applicable for ex  
111 situ collections within species with largely heritable seed trait variation.

112

## 113 **Materials and methods**

### 114 ***Cone collection and seed processing***

115 Mature, open-pollinated Torrey pine (*Pinus torreyana* Parry) cones were collected from native  
116 extant populations as part of a large ex situ conservation collection between June and July of  
117 2017. Cones were collected from 157 trees on Santa Rosa Island (Channel Islands National  
118 Park), CA (island population) and 201 trees at the Torrey Pine State Reserve in La Jolla, CA  
119 (mainland population), representing the species' entire natural distribution (Fig. 1; See  
120 Supporting Information Figure S1). Where possible, we collected between five to ten cones per  
121 maternal tree at each location. Sampling of reproductive maternal trees was evenly spaced;  
122 however, adjacent maternal trees were occasionally sampled to ensure enough cones were  
123 collected. On average, mainland and island trees sampled were separated by approximately 714  
124 (range = 0 – 2,092) and 397 (range = 1 – 1,131) meters, respectively. Seeds, organized by  
125 individual maternal tree, were then extracted from cones using a combination of mallet and pliers  
126 and processed for inclusion in a long-term ex situ conservation collection (see below).

127

128 ***Seed viability tests***

129 Estimating viability of seeds preserved *ex situ* is necessary given their potential role in  
130 restoration, reforestation, or reintroduction. Given this, the potential viability of Torrey pine  
131 seeds was tested using two complementary approaches prior to inclusion in the final *ex situ*  
132 collection. A float test was first used as a rapid, low-cost approach to assess seed viability.  
133 Floating seeds were presumed to lack an endosperm or embryo, while seeds that sunk were  
134 presumed filled. Seeds were dropped into water for approximately 15 seconds to differentiate  
135 presumed non-viable, floating seeds from presumed viable, sinking seeds (Gribko and Jones  
136 1995; Morina *et al.* 2017). Those seeds classed as likely viable were organized by maternal tree  
137 using paper bags, and then placed in a Blue M drying oven (Thermal Product Solutions, White  
138 Deer, Pennsylvania, USA) maintained at 37°C for 24 hours to remove potential surface moisture.  
139 Following this, seeds from a haphazard sample of maternal families were x-rayed at the  
140 Placerville Nursery, CA. In addition to visualizing seed morphological variation, x-ray  
141 photographs were used to verify viability based on float tests. Acrylic seed trays [20.3 cm x 25.4  
142 cm x 0.48 cm], with a 9 x 11 array of wells, were used to separate and position each Torrey pine  
143 seed over the x-ray film. Kodak x-OMAT HBT film (20.3 cm x 25.4 cm) was placed in a  
144 lightproof x-ray film cassette which was positioned in the x-ray machine with the seed tray  
145 centered on top of the film, with a shelf height of 55.9 cm. The x-ray was taken using a 17 kVP  
146 exposure for a total of two minutes, based on standardized conditions established previously for  
147 *Pinus coulteri* (Sara Wilson, USDA Forest Service, pers. comm.). X-ray images were digitized  
148 using a Nikon D40 digital camera mounted on a tripod over a light box.

149

150 ***Morphological measurement of seed traits***

151 Using ImageJ (Abràmoff *et al.* 2004), eight seed morphological traits were measured across 80  
152 mainland maternal families and 30 island maternal families, representing a haphazard subset of  
153 the complete collection (Fig. 2; Table 1). Each x-ray picture was scaled using the diameter of a  
154 seed tray well (1.87 cm) to express pixels as trait values in centimeters. Directly measured seed  
155 traits included seed length (SL, cm), seed width (SW, cm), embryo length (EL, cm), embryo  
156 width (EW, cm), seed coat width (SCW, cm), seed area (SA, cm<sup>2</sup>), endosperm area (ESA, cm<sup>2</sup>),  
157 and embryo area (EA, cm<sup>2</sup>). We selected these traits as they can readily be measured from x-ray  
158 pictures of seeds and provide a ubiquitous means to evaluate morphological variation for plants  
159 preserved *ex situ*. Using measured seed traits, six additional traits were derived (Table 1),  
160 including seed length/width ratio (SLW), embryo length/width ratio (ELW), relative embryo size  
161 (RES), relative endosperm size (REndS), seed coat area (SCA, cm<sup>2</sup>), and relative seed coat size  
162 (RSCS). These traits were derived as they provide a means to relate different morphological  
163 traits to each other and can provide a fine-scale estimate of the relative contribution of growth  
164 and size traits within individual seeds. We measured five randomly selected seeds per maternal  
165 tree, including three technical replicates per seed for each trait (the same seed was measured  
166 three times for any given morphological trait). Measurements were averaged across technical  
167 replicates to summarize the mean trait value per seed. In total, 550 seeds were measured from  
168 across 110 maternal trees spanning the two Torrey pine populations.

169

170 ***Seedling emergence test***

171 Within a restoration or reintroduction context, concurrent seedling emergence is often preferred  
172 for nursery plantings. To evaluate the timing and probability of seedling emergence within



173 Torrey pine populations, a trial was conducted in January 2018 using a random subset of seeds  
174 from the ex situ collection, including seeds sourced from Torrey Pine State Reserve and Santa  
175 Rosa Island, CA. Following x-ray, seeds were stored at 4°C in sealed mylar bags (USA  
176 emergency supply, Beaverton, Oregon, USA) placed in plastic boxes; each box contained  
177 desiccant crystals to decrease ambient moisture and reduce likelihood of mold. Seeds from eight  
178 maternal families per population were selected for the emergence trial. Between eight to ten  
179 seeds per maternal tree were weighed and then stratified under cold, moist conditions for 30 days  
180 (placed in plastic boxes on a moist paper towel at 4°C). Seeds were sown directly into a 164 mL  
181 Ray Leach “Cone-tainer”™ (Stuewe & Sons, Tangent, Oregon, USA) filled with Sunshine®  
182 Mix #4 (Sungro horticulture, Agawam, Massachusetts, USA), pressed halfway into the soil, and  
183 then covered with a thin layer of gravel. For approximately one month following planting, seeds  
184 were misted for one minute at hourly intervals over a daily eight-hour period (9am – 4pm).  
185 Following emergence, seedlings were hand watered to saturation weekly to biweekly as needed.  
186 Emergence was quantified across three separate timepoints (Feb 06/2018, Feb 16/2018, and Mar  
187 07/2018) per maternal family as the proportion of seeds that successfully developed into living  
188 seedlings from the total initially planted.

189

### 190 ***Evaluating the distribution of seed trait variation***

191 We conducted a principal component analysis (PCA) using all 14 measured and derived seed  
192 traits averaged by maternal family to evaluate population-specific differentiation in seed  
193 morphology. Prior to performing the PCA, to account for differences in measurement units, all  
194 seed traits were standardized using the *scale()* function in R implementing the z-score  
195 standardization:  $z_{ij} = \frac{x_{ij} - \mu_j}{\sigma_j}$  (1), where  $x_{ij}$  is the non-transformed trait value,  $\mu_j$  is the mean of

196 a given seed trait across populations, and  $\sigma_j$  the standard deviation of the same seed trait across  
197 populations. Subsequently, to test for seed trait differences between population means, we used  
198 either Student's two-sample test or its non-parametric equivalent when normality was not met,  
199 Wilcoxon's two-sample test, within the R package "exactRankTests" (Hothorn and Hornik  
200 2019). Normality was assessed using Shapiro-Wilk's test of normality within each population. In  
201 total, four of the fourteen traits were distributed normally in both Torrey pine populations,  
202 including seed width (mainland:  $W = 0.97$ ,  $P = 0.06$ ; island:  $W = 0.97$ ,  $P = 0.52$ ), embryo length  
203 (mainland:  $W = 0.98$ ,  $P = 0.29$ ; island:  $W = 0.95$ ,  $P = 0.21$ ), embryo width (mainland:  $W = 0.97$ ,  
204  $P = 0.09$ ; island:  $W = 0.93$ ,  $P = 0.05$ ), and embryo area (mainland:  $W = 0.98$ ,  $P = 0.45$ ; island:  $W$   
205  $= 0.96$ ,  $P = 0.26$ ).

206

207 To evaluate the distribution of morphological trait variation within and between Torrey pine  
208 populations, we quantified the proportion of variation attributed to population and maternal tree  
209 families using measured and derived morphological traits summarized by seeds. For each trait,  
210 we fit a linear mixed model using the R package "lme4" (Bates *et al.* 2015) with population  
211 considered a fixed effect and maternal families within populations considered a nested random  
212 effect:  $Y_{ij} = \mu + \pi_i + r_{i/j} + e_{ij}$ , where  $Y_{ij}$  is the observed seed trait value,  $\mu$  is the seed trait  
213 overall mean,  $\pi_i$  is the effect of population origin on the seed trait mean,  $r_{i/j}$  is the effect of  
214 maternal family within populations on the observed seed trait value, and  $e_{ij}$  are the effects on the  
215 seed trait value of any other variables unaccounted for in the model (residual error). For each  
216 model, normality of residual errors was visually assessed and significance of fixed- and random-  
217 effect terms was tested using the functions *anova()* and *ranova()* respectively, implemented in  
218 the R package "stats" (R Core Team 2020) and "lmerTest" (Kuznetsova *et al.* 2017). Proportions

219 of seed morphological variance explained by populations (marginal  $R^2$ ,  $R^2_m$ ), both populations  
220 and maternal families (conditional  $R^2$ ,  $R^2_c$ ), and maternal families alone ( $R^2_c - R^2_m$ ) were  
221 quantified for each model independently using the function *r.squaredGLMM()* implemented in  
222 the R package “MuMIn” (Bartoń 2020).

223

### 224 *Assessing differences in seedling emergence across populations*

225 To test for differences in the probability and the timing of seedling emergence in Torrey pine, we  
226 evaluated the proportion of seeds that produced seedlings both within and between populations  
227 across timepoints. First, we used Friedman’s rank sum test (non-parametric repeated measures  
228 ANOVA) followed by Wilcoxon’ paired two-sample test, both implemented in the R package  
229 “rstatix” (Kassambara 2020), to assess differences in the proportion of emerged seedlings  
230 between timepoints within populations. We used a non-parametric approach for both Torrey pine  
231 populations because normality could not be assumed at select timepoints due to high frequency  
232 of zero values. We accounted for multiple testing using Benjamini and Hochberg (1995)’s False  
233 Discovery Rate (FDR) correction implemented in the *wilcox\_test()* function. Following this, we  
234 evaluated timepoint-specific population differences in seedling emergence. We used Shapiro-  
235 Wilk’s test to assess populations’ deviation from normality at each timepoint and either  
236 Student’s (for timepoints passing the normality test) or Wilcoxon’s two-sample test (for  
237 timepoints failing the normality test) to evaluate differences in population emergence.  
238 Timepoints Feb 16/2018 (mainland:  $W = 0.94$ ,  $P = 0.57$ ; island:  $W = 0.9$ ,  $P = 0.28$ ) and Mar  
239 07/2018 (mainland:  $W = 0.96$ ,  $P = 0.85$ ; island:  $W = 0.88$ ,  $P = 0.21$ ) passed the normality test,  
240 while timepoint Feb 06/2018 (mainland:  $W = 0.52$ ,  $P < 0.001$ ; island:  $W = 0.73$ ,  $P = 0.004$ ) failed

241 the normality test. All statistical analyses were performed using R version 4.0.2 and 4.0.5 (R  
242 Core Team 2020, 2021).

243

#### 244 ***Simulating variation captured in the ex situ collection using seed morphological traits***

245 For each of the 14 measured and derived seed traits, we conducted a separate simulation  
246 quantifying morphological variation captured when increasing the number of maternal families  
247 sampled from contemporary Torrey pine seed collections. Simulations were conducted in R  
248 version 3.6.3 (R Core Team 2020) using a customized script [See Supporting Information Figure  
249 S2]. Resampling of ex situ collections were performed for island and mainland Torrey pine  
250 populations independently, using between one and the total number of maternal families  
251 available within each ex situ population collection (mainland: 80 maternal families, island: 30  
252 maternal families) ( $N_{fam}$ ). Maternal trees were sampled randomly without replacement from the  
253 pool of available families. All seeds within each selected maternal family were sampled as part  
254 of this simulation, except those with missing values for the trait simulated. Overall, between two  
255 to five seeds per maternal family were sampled within each population.

256

257 To evaluate the number of maternal families needed to capture 95% of seed trait variation in both  
258 island and mainland populations, we estimated the number of unique seed trait values captured in  
259 a sample of  $N_{fam}$  maternal families ( $N_c$ ) relative to the total number of unique seed trait values  
260 present in a seed population ( $N_t$ ). Here, we define “unique seed trait values” as the number of  
261 non-redundant standardized measurements for the seed trait simulated rounded to the first digit.  
262 Seed morphological measurements were rounded to the first digits as we believe that seed trait  
263 variation estimated using additional digits is more likely to fail to capture meaningful biological

264 variation. Standardization of the data was performed so that all seed traits share the same unit  
265 (the number of standard deviations a value is from the overall trait mean across populations, see  
266 equation (1) above) and become comparable. Sampling of maternal families and estimation of  
267 the summary statistic, defined as the proportion of total seed trait variance captured ( $N_c/N_t$ ), were  
268 repeated 500 times for each seed morphological trait and Torrey pine population. In this way,  
269  $N_c/N_t$  accounts for potential variation in number of seeds sampled per maternal family or  
270 variation in sampled maternal families included.

271  
272 Finally, for each number of maternal trees sampled ( $N_{fam}$ ), we averaged the summary statistic  
273 across all 500 replicates. This process was repeated for each of the 14 seed morphological traits  
274 and performed for each Torrey pine population separately. Following this, the summary statistic  
275 was averaged across all seed traits and separated by populations (see Results below). Proportions  
276 of total seed trait variance captured ( $N_c/N_t$ ) are provided based on proportions of maternal  
277 families sampled (instead of the number of maternal families sampled) as sample sizes varied  
278 across Torrey pine populations.

279

## 280 **Results**

### 281 *Island-mainland differentiation in seed morphology*

282 A principal component analysis (PCA) using all 14 measured and derived seed traits averaged by  
283 maternal family revealed substantial differences in seed morphology between island and  
284 mainland populations of Torrey pine (Fig. 3). The first PC axis explained 57.8% of variation in  
285 seed morphological traits, primarily separating the island from the mainland population. Seed  
286 length, seed width, seed area, endosperm area, and seed coat area exhibited the five highest

287 loadings (absolute values) on PC1 [See Supporting Information Table S1], indicating that seed  
288 size and seed coat thickness can largely discriminate island from mainland individuals. On  
289 average, seeds collected on island trees were longer, wider, larger, and thicker than seeds  
290 collected on mainland trees (Table 1). The second PC axis explained 15.9% of seed trait  
291 variation and summarizes within population variability in seed morphology (Fig. 3). Relative  
292 seed coat size, relative endosperm size, and relative embryo size had the three highest loadings  
293 (absolute values) on PC2 [See Supporting Information Table S1]. This suggests that once  
294 corrected for seed size, seed coat thickness, endosperm size, and embryo size are traits  
295 contributing to within-population variation.

296

### 297 *Contribution of population origin and maternal family to seed trait variation*

298 Consistent with our principal component analysis, linear mixed models constructed for each of  
299 the 14 measured and derived seed traits demonstrated that considerable variation in seed  
300 morphology in Torrey pine is explained by population origin (Fig. 4). On average, population  
301 origin explained 23% (range = 0.02–0.57) of variation across the species' distribution [See  
302 Supporting Information Table S2]. Traits associated with seed size and seed coat thickness  
303 exhibited the highest proportion of variance explained by population origin. These include seed  
304 coat area (0.57;  $F_{1,107.60} = 221.91$ ,  $P < 0.001$ ), seed area (0.49;  $F_{1,107.56} = 156.45$ ,  $P < 0.001$ ),  
305 endosperm area (0.37;  $F_{1,108.07} = 100.58$ ,  $P < 0.001$ ), seed width (0.36;  $F_{1,108} = 126.04$ ,  $P <$   
306  $0.001$ ), seed coat width (0.32;  $F_{1,108} = 96.04$ ,  $P < 0.001$ ), and seed length (0.30;  $F_{1,108.50} = 78.92$ ,  
307  $P < 0.001$ ). Overall, this suggests seed size and seed coat thickness are major discriminants of  
308 island and mainland Torrey pine seeds.

309

310 While population origin explained substantial variation across populations, assessment of  
311 maternal seed families within populations indicated substantial family structure to seed trait  
312 variation (Fig. 4). On average, maternal seed family explained 24% (range = 0.07–0.37) of  
313 variation within populations [See Supporting Information Table S2]. Embryo length (0.37;  $\chi^2 =$   
314 124.82, df = 1,  $P < 0.001$ ), seed length (0.35;  $\chi^2 = 180.72$ , df = 1,  $P < 0.001$ ), endosperm area  
315 (0.34;  $\chi^2 = 211.87$ , df = 1,  $P < 0.001$ ), seed area (0.31;  $\chi^2 = 256.71$ , df = 1,  $P < 0.001$ ), embryo  
316 area (0.29;  $\chi^2 = 100.16$ , df = 1,  $P < 0.001$ ), and seed coat width (0.28;  $\chi^2 = 126.93$ , df = 1,  $P <$   
317 0.001) exhibited the highest proportion of seed trait variation explained by within-population  
318 maternal families. This suggests that there is substantial family-level structure to seed size,  
319 endosperm size, embryo size, and seed coat thickness within Torrey pine populations.

320

### 321 *Impact of population seed trait differentiation on seedling emergence*

322 The proportion of emerged seedlings increased over time for both island ( $Q = 15.5$ , df = 2,  $P <$   
323 0.001) and mainland ( $Q = 15.2$ , df = 2,  $P < 0.001$ ) populations (Fig. 5). However, we found no  
324 significant differences in the proportion of individuals emerging between populations across  
325 observed time points. On average, 7% and 9% of mainland and island seedlings emerged a  
326 month after sowing (Feb 06/2018;  $W = 28$ ,  $P = 0.64$ ), 63% and 53% of mainland and island  
327 seedlings emerged a month and a half after sowing (Feb 16/2018;  $t = 0.81$ , df = 14,  $P = 0.43$ ),  
328 and 78% of mainland and island seedlings emerged two months after sowing (Mar 07/2018;  $t =$   
329  $-0.06$ , df = 14,  $P = 0.95$ ). Overall, this indicates that under controlled conditions, timing and  
330 probability of emergence may not be impacted by population differences in seed morphology for  
331 Torrey pine seedlings.

332

333 ***Morphological variation captured in simulated seed collections***

334 Simulations revealed that to capture 95% of seed trait variation present in our existing ex situ  
335 collections, on average 83% (25 out of 30) and 71% (57 out of 80) of all island and mainland  
336 families would need to be resampled, respectively (Fig. 6). This indicates that both island and  
337 mainland populations harbor considerable within-population structure for seed morphological  
338 traits. Interestingly, capturing equal morphological variation across seed collections always  
339 required a higher proportion of island maternal families to be collected relative to the mainland  
340 population.

341

342 **Discussion**

343 Evaluating between- and within-population morphological variation in contemporary ex situ seed  
344 collections may aid in understanding the distribution of variation needed to guide future  
345 conservation efforts. Here, we quantified the distribution of trait variation within an ex situ  
346 collection of Torrey pine seeds, with an aim to optimize future supplemental collections.  
347 Morphologically, island and mainland seeds were significantly different from each other. Island  
348 seeds were larger on average with thicker seed coats relative to their mainland counterparts.  
349 These morphological differences may be explained by a combination of stochastic and  
350 deterministic factors associated with population origin, including genetic bottlenecks following  
351 island colonization, genetic drift, and selection associated with unique biotic pressures.  
352 Interestingly, despite substantial morphological differentiation, seedling emergence did not vary  
353 among populations, suggesting that either the probability and timing of emergence under  
354 controlled conditions is not impacted by differences in seed morphology or that island and  
355 mainland seeds respond similarly to an artificial germination protocol. In addition to population



356 origin, a considerable proportion of seed trait variation within each population was explained by  
357 maternal family. This suggests that there is substantial within-population variation that will be  
358 important to conserve and maintain populations' evolutionary potential. Finally, simulations  
359 demonstrated that 83% and 71% of all maternal families within island and mainland seed  
360 collection respectively were necessary to capture 95% of morphological variation within the  
361 existing collections. To ensure we maintain representative levels of trait variation in future seed  
362 collections, our results suggest that the number of maternal families sampled within natural  
363 populations should be maximized, with the island population potentially requiring additional  
364 sampling efforts.

365

366 Species distributed between island and mainland origins often exhibit marked among-population  
367 phenotypic differentiation, including differentiation in seed morphology (Burns *et al.* 2012; Lens  
368 *et al.* 2013; Kavanagh and Burns 2014; Burns 2016). Our results revealed considerable seed  
369 morphological differences between Torrey pine populations (Fig. 3), primarily in seed size and  
370 seed coat thickness (Fig. 4; See Supporting Information Table S1). On average, islands seeds  
371 were larger and had thicker seed coats than seeds collected on the mainland (Table 1). These  
372 results are consistent with previous studies of island-mainland systems that noted island  
373 populations exhibited larger seeds relative to mainland populations (Kavanagh and Burns 2014;  
374 Burns 2016; Biddick *et al.* 2019). A combination of different factors could contribute to  
375 morphological variation among seed populations, including both stochastic and deterministic  
376 forces.

377

378 On islands, seeds traits associated with long-distance dispersal may be selected against as they  
379 can increase the probability an individual would disperse beyond an island's limits (Cody and  
380 Overton 1996; Kavanagh and Burns 2014; Ottaviani *et al.* 2020, but see Burns 2018). For Torrey  
381 pine, increased seed size on the island may have evolved to limit potential seed losses via wind-  
382 dispersal, as seed mass negatively correlates with dispersal distance in pines (Greene and  
383 Johnson 1993; Debain *et al.* 2003, but see Wyse and Hulme 2020). Nonetheless, Torrey pine  
384 seeds possess degenerated wings (Ledig and Conkle 1983), suggesting that other mechanisms  
385 likely contribute to seed dispersal in this species. Rodents and birds both feed on Torrey pine,  
386 suggesting that seeds may undergo animal-mediated dispersal (Johnson *et al.* 2003). Thus, seed  
387 predation may contribute to differences in seed size observed between populations. On the  
388 island, *Peromyscus maniculatus* (Deer mouse) is the only rodent present to predate on Torrey  
389 pine seeds (Johnson *et al.* 2003). This contrasts with the mainland, where multiple seed predators  
390 have been documented; including *Peromyscus boylei* (Brush mice), *Peromyscus maniculatus*  
391 (Deer mice), *Peromyscus eremicus* (Cactus mice), *Chaetodipus californicus* (California pocket  
392 mice), *Spermophilus beecheyi* (California ground squirrels), or *Aphelocoma californica* (Scrub  
393 jays) (Johnson *et al.* 2003). If large seeds are preferentially targeted by seed predators (Reader  
394 1993; Gómez 2004), reduced seed size on the mainland may have evolved as a consequence of  
395 the trade-off between attracting predators to promote seed dispersal and mitigating fitness loss  
396 due to seed consumption.

397

398 While selection may contribute to population differences, differentiation in seed morphology  
399 may result from stochastic evolutionary forces. Founder effects associated with the colonization  
400 of Santa Rosa island by mainland individuals, and genetic drift in the face of limited gene flow,

401 may have led to morphological differentiation between Torrey pine populations (Ledig and  
402 Conkle 1983). Alternatively, a more complex demographic history of the two populations,  
403 including colonization, extinction, and recolonization events may have led to the differences  
404 observed between populations (Haller 1986). While both stochastic and deterministic factors  
405 may contribute to population differences in seed morphology, additional experiments are  
406 required to test mechanistic hypotheses. Seeds evaluated in this manuscript were collected from  
407 natural populations. To tease apart the contribution of environment and genetics to seed trait  
408 differences observed among populations, a common garden experiment is required. Indeed, a  
409 reciprocal transplant experiment would be the most effective test of the action of natural  
410 selection in shaping morphological differences between island and mainland seeds.

411  
412 Despite significant differences in seed morphology between populations, timing and probability  
413 of emergence was similar across populations (Fig. 5). Emergence rates were high throughout the  
414 trial, with 78% of island and mainland seedlings emerging two months after sowing. The absence  
415 of differences in seedling emergence between populations was surprising, as seed size often  
416 negatively correlates with time to germination (Daws *et al.* 2005; Tanveer *et al.* 2013). However,  
417 seed coat thickness can also influence rates of emergence. Seeds with thick seed coats relative to  
418 their mass often germinate later than seeds with thinner seed coats (Daws *et al.* 2005; Hamilton  
419 *et al.* 2013). For Torrey pine, Hamilton *et al.* (2017) found that island seeds germinate on  
420 average two days after mainland seeds. Interestingly, island seeds were not only larger, but also  
421 had thicker seed coats relative to mainland seeds (Table 1). Even after correcting for differences  
422 in seed size, seed coat thickness (relative seed coat size; RSCS) remained moderately higher in  
423 island seeds. Together, these results predict that island seedlings should emerge at similar or later

424 timepoints relative to mainland seedlings, which is consistent with current and previous  
425 observations.

426

427 Similar emergence rates may also result from our experimental design. Abe and Matsunaga  
428 (2011), in a mainland-island comparison study, observed that cold stratification attenuates  
429 differences in germination rates between populations of *Rhaphiolepis umbellata*. Additionally,  
430 complete and rapid germination of pine seeds is generally observed when pretreated under cold  
431 and moist conditions (Krugman and Jenkinson 2008). Overall, this suggests that cold  
432 stratification may mask population-specific differences in seedling emergence. Concurrent  
433 seedling emergence from both Torrey pine populations coupled with high emergence success  
434 suggests a cold stratification protocol is valuable for Torrey pine, particularly where  
435 simultaneous emergence for nursery-grown seedlings is desired. Note, however, that variation in  
436 the proportion of emerged seedlings within populations across timepoints may have concealed  
437 population-specific differences in emergence rates. Consequently, weak differences in the timing  
438 and probability of seedling emergence observed between island and mainland populations may  
439 be an artifact of small numbers of seeds and maternal families used during emergence trials.

440

441 Although population origin explained a substantial proportion of seed trait variation, linear  
442 mixed models demonstrated that maternal seed families within populations explained as much  
443 variation (Fig. 4; See Supporting Information Table S2). Given generally high heritability for  
444 seed morphological traits and the half-sib design of our collection (Pandey *et al.* 1994; Cober *et*  
445 *al.* 1997; Mera *et al.* 2004; Roy *et al.* 2004; Carles *et al.* 2009; Zas and Sampedro 2015; Hakim  
446 and Suyamto 2017), family-level seed trait variation likely provides a useful proxy for assessing

447 within-population genetic diversity. With nearly 25% of variation explained on average by  
448 maternal families [See Supporting Information Table S2], this suggest there is substantial genetic  
449 structure within Torrey pine populations. These results were notable as previous studies using  
450 allozymes and chloroplast DNA suggested that the species exhibits little to no within-population  
451 genetic variability (Ledig and Conkle 1983; Waters and Schaal 1991; Whittall *et al.* 2010).  
452 However, the common garden experiment conducted by Hamilton and colleagues (2017)  
453 indicated substantial family-level variation in tree height within both island and mainland  
454 populations. Overall, these results indicate that Torrey pine populations may possess within-  
455 population genetic variation necessary for natural selection to act upon. From a conservation  
456 perspective, these findings suggest that a strategy maximizing the number of maternal families  
457 sampled would optimize genetic diversity preserved in future ex situ seed collections and  
458 increased distance among individuals may limit relatedness among maternal trees.

459

460 Generally, ex situ seed collections aim to capture 95% of genetic diversity present throughout a  
461 species' distribution (Marshall and Brown 1975; Brown and Marshall 1995; Li *et al.* 2002;  
462 Gapare *et al.* 2008). Simulations revealed that, in order to capture 95% of morphological  
463 variation currently maintained ex situ, 25 (83% of island collection) and 57 (71% of mainland  
464 collection) maternal families within each seed collection would need to be sampled (Fig. 6).  
465 These data indicate that sampling more maternal families from the island population may be  
466 necessary to achieve the same level of representation of morphological variation. Assuming  
467 increased phenotypic variation observed on the island results from higher allelic diversity,  
468 capturing 95% of genetic variation within the island population will always require more  
469 maternal families relative to the mainland population. For these simulations, we assumed that

470 contemporary ex situ collections captured all morphological variation both within and between  
471 populations, including seed phenotype frequencies. However, if this is not the case, these  
472 recommendations may result in suboptimal sampling of standing variation within targeted  
473 populations. This caveat is important because the number of x-rayed maternal families differed  
474 between island (30 maternal families) and mainland (80 maternal families) seed collections. To  
475 address this caveat, it will be important to have a general understanding of the fraction of natural  
476 morphological variation captured across ex situ seed populations and adapt sampling efforts  
477 accordingly.

478  
479 Practical and cost-effective, long-term storage of seeds ex situ is widely used to capture and  
480 maintain rare species genetic diversity. These seed collections represent an invaluable resource  
481 to quantify within and between population trait variation that may be used to guide future ex situ  
482 sampling efforts. Using Torrey pine as a model, we demonstrate that incorporating existing  
483 information from ex situ collections offers a unique opportunity to monitor and optimize  
484 conservation objectives, particularly important for rare species. While our results and  
485 recommendations may be specific to Torrey pine, the empirical, statistical, and simulation-based  
486 approaches presented here are broadly applicable to heritable traits across ex situ seed  
487 collections.

488

#### 489 **Data**

490 The data for this article, including seed morphological measurements and R scripts used are  
491 available from GitHub: <https://github.com/lnds-anonymous/AoBP2021>.

492

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507

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663

664 **Tables**

665 **Table 1.** Morphological measurements of Torrey pine seeds sourced from Santa Rosa Island (Island, n=30) and Torrey Pine State  
 666 Reserve (Mainland, n=80), CA. Listed are population mean estimates ( $\pm$ SE) of measured (A) and derived (B) seed traits summarized  
 667 by maternal families. Measurable traits: seed length (SL), seed width (SW), embryo length (EL), embryo width (EW), seed coat width  
 668 (SCW), seed area (SA), endosperm area (ESA), and embryo area (EA). Derived traits: seed length/width ratio (SLW), embryo  
 669 length/width ratio (ELW), relative embryo size (RES), relative endosperm size (REndS), seed coat area (SCA), and relative seed coat  
 670 size (RSCS). Differences in seed morphology between mainland and island populations were significant ( $\alpha=0.05$ ) for all 14 seed traits.

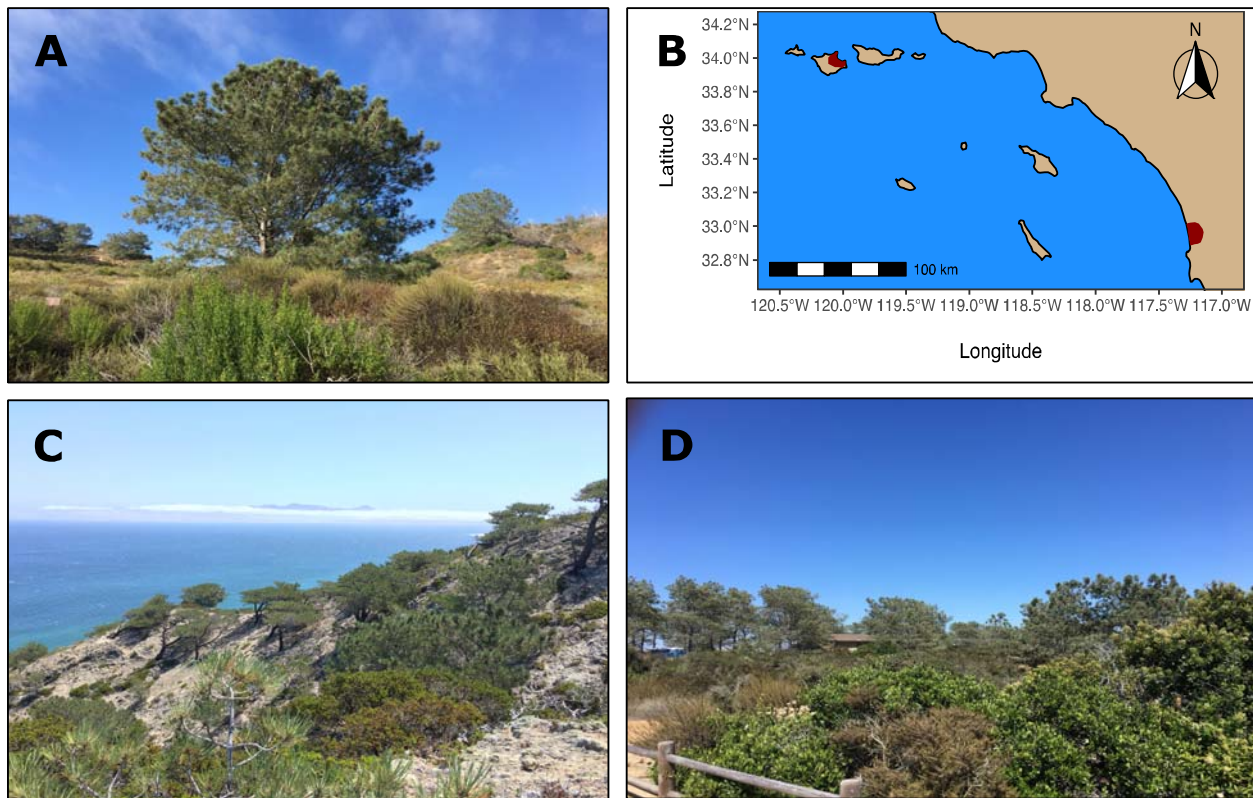
671 **A – Measured seed traits**

Population	SL (cm)	SW (cm)	SCW (cm)	EL (cm)	EW (cm)	SA (cm <sup>2</sup> )	ESA (cm <sup>2</sup> )	EA (cm <sup>2</sup> )
Mainland	1.41 $\pm$ 0.016	0.77 $\pm$ 0.009	0.09 $\pm$ 0.002	1.08 $\pm$ 0.012	0.15 $\pm$ 0.003	0.89 $\pm$ 0.023	0.55 $\pm$ 0.014	0.17 $\pm$ 0.004
Island	1.67 $\pm$ 0.02	0.97 $\pm$ 0.015	0.12 $\pm$ 0.002	1.2 $\pm$ 0.02	0.19 $\pm$ 0.004	1.41 $\pm$ 0.031	0.81 $\pm$ 0.021	0.23 $\pm$ 0.007

672 **B – Derived seed traits**

Population	SLW (SL/SW)	ELW (EL/EW)	SCA (cm <sup>2</sup> ) (SA-ESA)	RES (EA/ESA)	REndS (ESA/SA)	RSCS (SCA/SA)
Mainland	1.86 $\pm$ 0.013	7.28 $\pm$ 0.094	0.34 $\pm$ 0.01	0.31 $\pm$ 0.005	0.62 $\pm$ 0.003	0.38 $\pm$ 0.003
Island	1.77 $\pm$ 0.037	6.64 $\pm$ 0.136	0.6 $\pm$ 0.013	0.29 $\pm$ 0.004	0.57 $\pm$ 0.004	0.43 $\pm$ 0.004

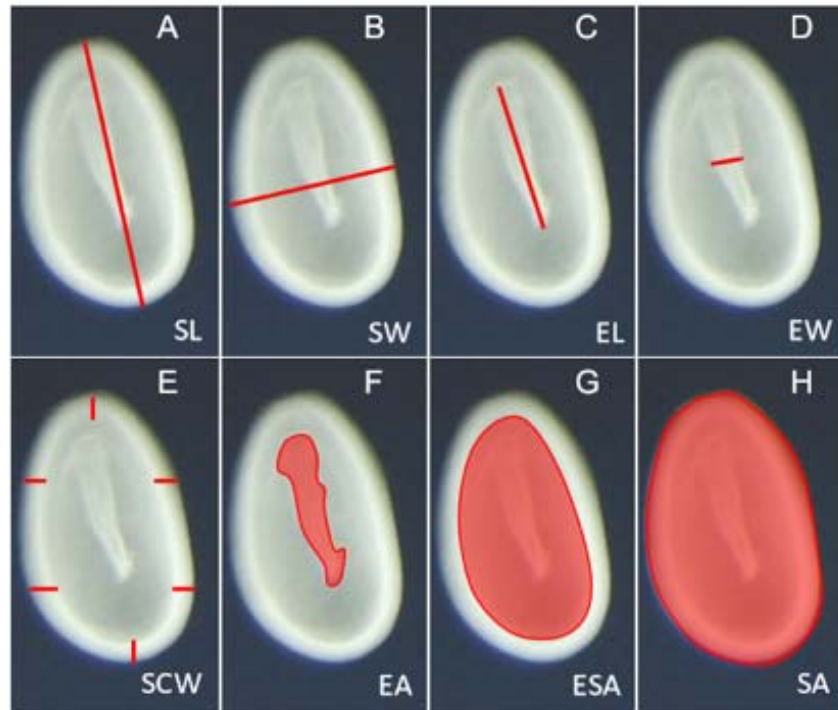
673 **Figures**



674

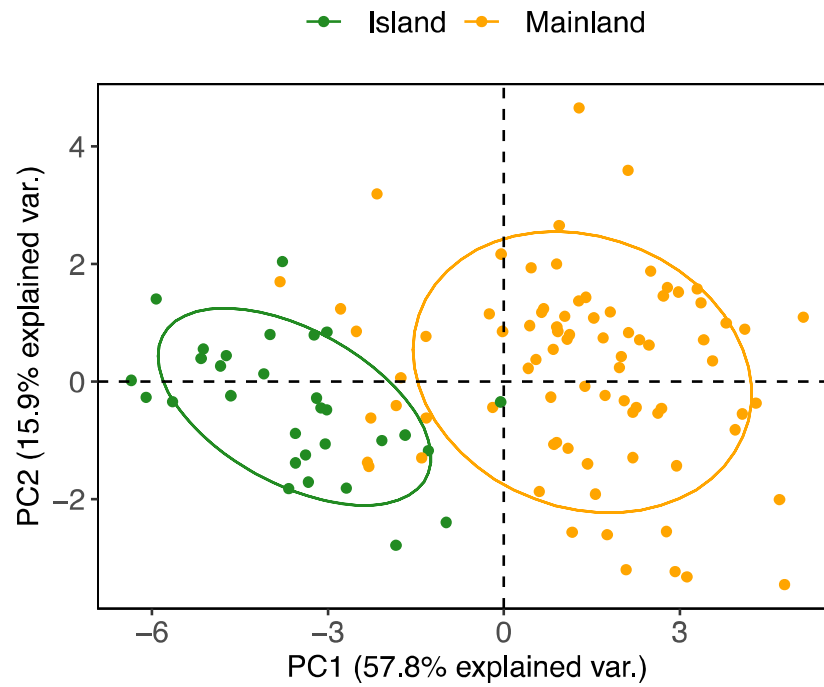
675 **Fig. 1.** (A) *Pinus torreyana* individual. (B) *Pinus torreyana* distribution map, including Torrey  
676 pine distribution on Santa Rosa Island, CA (*Pinus torreyana* subsp. *insularis*, top left red-shaded  
677 area) and at the Torrey Pine State Reserve, CA (*Pinus torreyana* subsp. *torreyana*, bottom right  
678 red-shaded area). (C) Torrey pine stand on Santa Rosa Island, CA. (D) Torrey pine stand at the  
679 Torrey Pine State Reserve, CA.





680

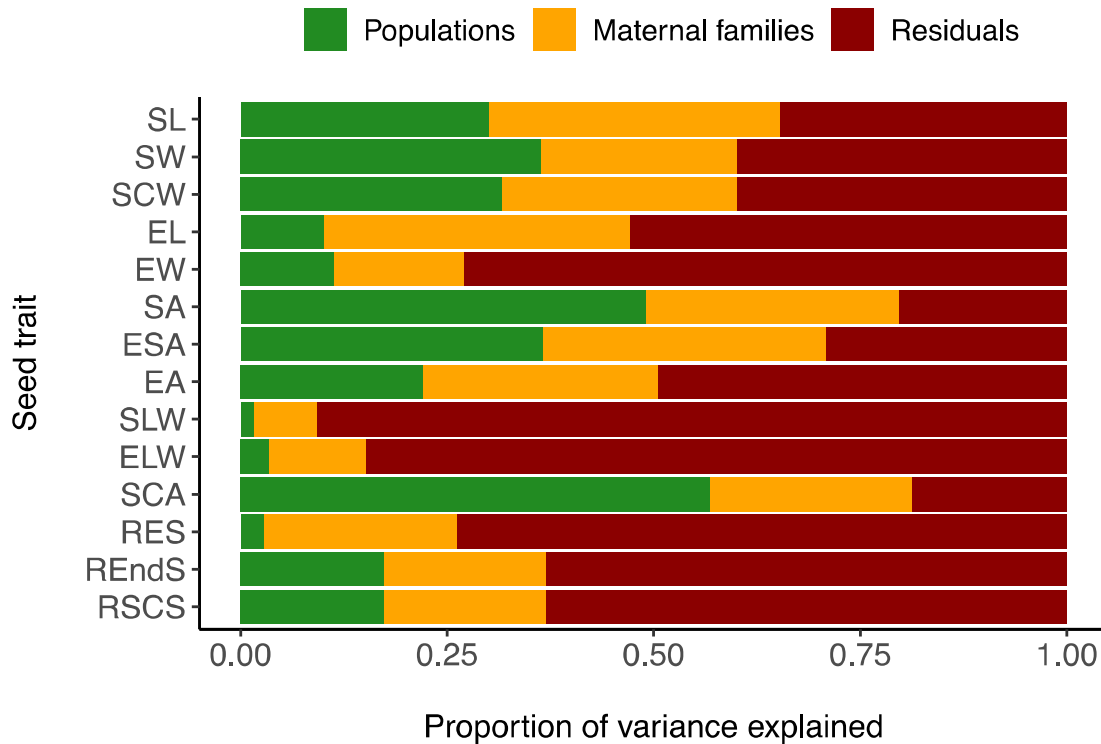
681 **Fig. 2.** Visual of morphological measurements taken using ImageJ for seeds collected on Santa  
682 Rosa Island and at the Torrey pine State Reserve. (A) Seed length [cm]. (B) Seed width [cm]. (C)  
683 Embryo length [cm]. (D) Embryo width [cm]. (E) Seed coat width [cm]. (F) Embryo area [cm<sup>2</sup>].  
684 (G) Endosperm area [cm<sup>2</sup>]. (H) Seed area [cm<sup>2</sup>].



685

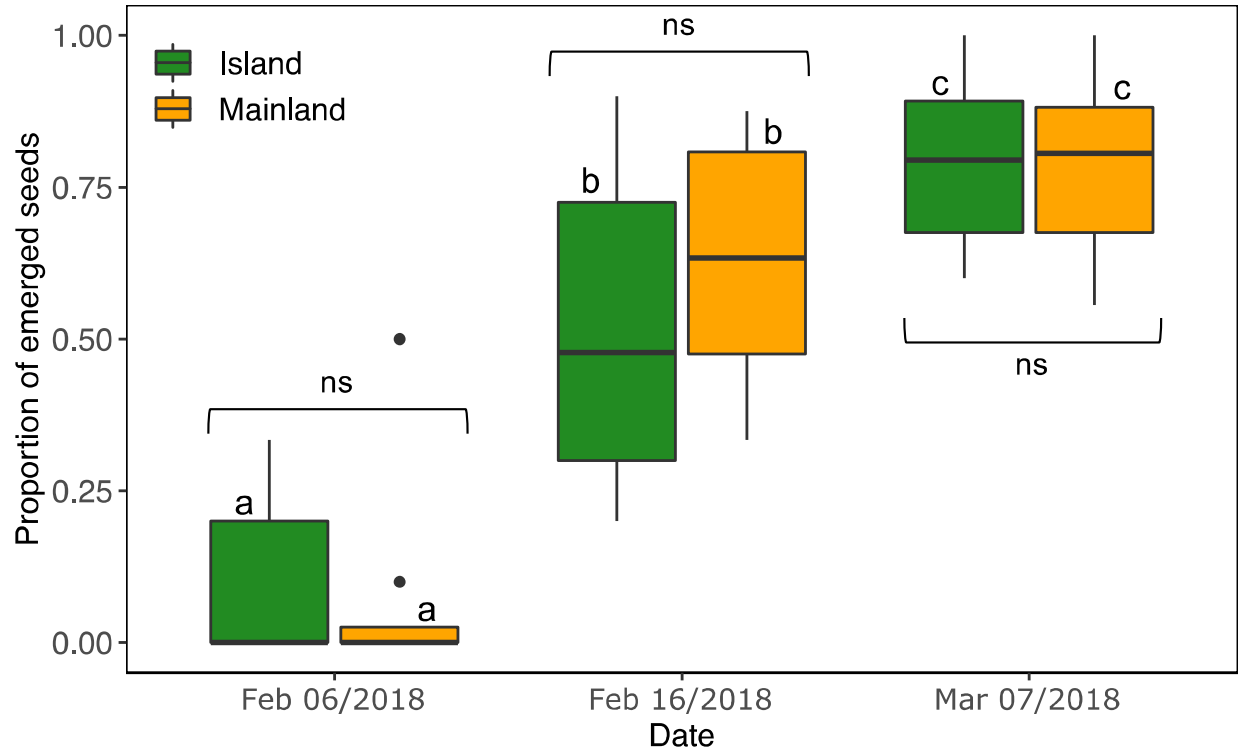
686 **Fig. 3.** Principal components analysis (PCA) using all 14 seed morphological traits measured and  
687 derived from maternal plants collected on Santa Rosa Island (green) and at the Torrey Pine State  
688 Reserve (orange).

689



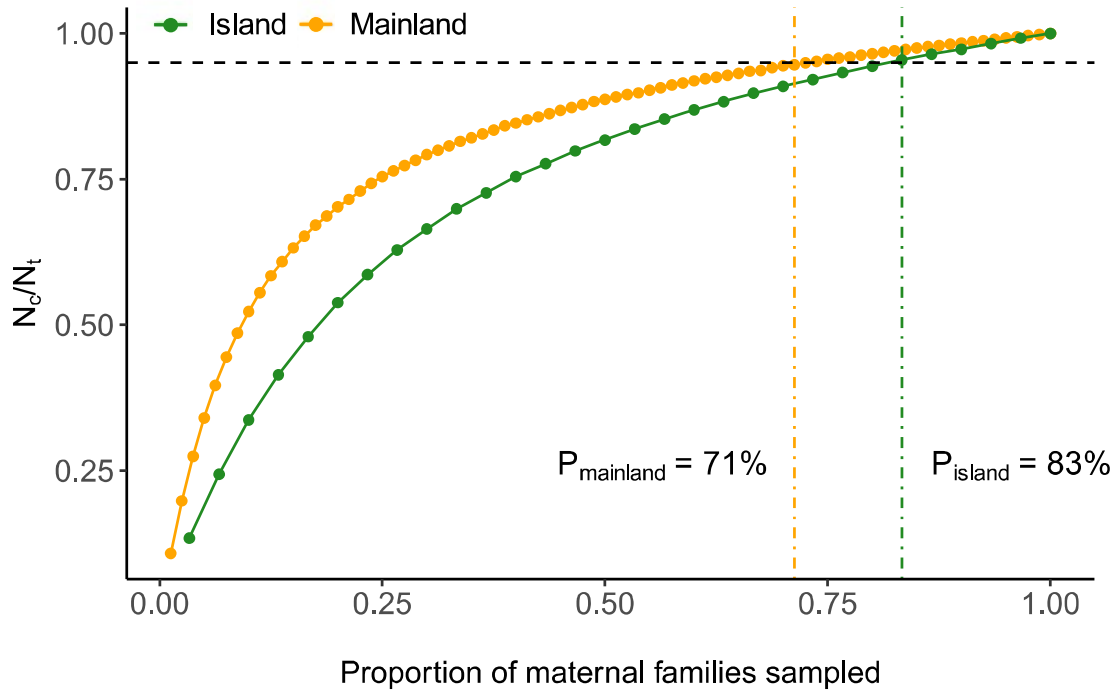
690

691 **Fig. 4.** Proportion of variance in seed morphology explained by populations (green), maternal  
692 families within populations (orange), and other variables not accounted for in the model  
693 (residuals; dark red) for each of the 14 measured and derived seed traits. SW, seed width (cm);  
694 SLW, seed length/width ratio; SL, seed length (cm); SCW, seed coat width (cm); SCA, seed coat  
695 area (cm<sup>2</sup>); SA, seed area (cm<sup>2</sup>); RSCS, relative seed coat size; RES, relative embryo size;  
696 REndS, relative endosperm size; EW, embryo width (cm); ESA, endosperm area (cm<sup>2</sup>); ELW,  
697 embryo length/width ratio; EL, embryo length (cm); EA, embryo area (cm<sup>2</sup>). See Supporting  
698 Information Table S2 for numerical estimates.



699

700 **Fig. 5.** Proportion of emerged seedlings (y axis) recorded at three different timepoints (x axis)  
701 for seeds sampled on Santa Rosa Island (green) and at the Torrey Pine State Reserve (orange).  
702 Significant differences in emergence time across timepoints within populations are indicated  
703 with different letters. Comparisons between populations at each timepoint is indicated with  
704 square brackets. ns, non-significant differences ( $\alpha=0.05$ ).



705

706 **Fig. 6.** Phenotypic variation captured across seed traits in simulated collections ( $N_c$ ) relative to  
707 total phenotypic variation present in seed populations ( $N_t$ ). Average proportion of phenotypic  
708 variation captured ( $N_c/N_t$ ) was estimated for various proportions of maternal families sampled.  
709  $P_{\text{island}}$  and  $P_{\text{mainland}}$  represent the proportion of maternal families required to capture 95% of  
710 morphological variation (horizontal dashed line) present in island (green) and mainland (orange)  
711 ex situ seed populations, respectively.