

1 **Freshwater diatom biomonitoring through benthic kick-net metabarcoding**

2 **Short title: Freshwater diatom biomonitoring through metabarcoding**

3 Victoria **Carley** Maitland<sup>1</sup>, Chloe Victoria Robinson<sup>1</sup>, Teresita M. Porter<sup>1,2</sup>, Mehrdad  
4 Hajibabaei<sup>1\*</sup>.

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6 <sup>1</sup>Centre for Biodiversity Genomics & Department of Integrative Biology, University of  
7 Guelph, Guelph, Ontario, N1G 2W1

8 <sup>2</sup>Great Lakes Forestry Centre, Natural Resources Canada, 1219 Queen Street East,  
9 Sault Ste. Marie, ON Canada

10

11 \* Corresponding author

12 Email: [mhajibab@uoguelph.ca](mailto:mhajibab@uoguelph.ca)

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## 22 **Abstract**

23           Biomonitoring is an essential tool for assessing ecological conditions and  
24 informing management strategies. The application of DNA metabarcoding and high  
25 throughput sequencing has improved data quantity and resolution for biomonitoring of  
26 taxa such as macroinvertebrates, yet, there remains the need to optimise these  
27 methods for other taxonomic groups. Diatoms have a longstanding history in freshwater  
28 biomonitoring as bioindicators of water quality status. However, periphyton scraping, a  
29 common diatom sampling practice, is time-consuming and thus costly in terms of  
30 labour. This study examined whether the benthic kick-net technique used for  
31 macroinvertebrate biomonitoring could be applied to bulk-sample diatoms for  
32 metabarcoding. To test this approach, we collected samples using both conventional  
33 microhabitat periphyton scraping and bulk-tissue kick-net methodologies in parallel from  
34 replicated sites with different habitat status (good/fair). We found there was no  
35 significant difference in community assemblages between conventional periphyton  
36 scraping and kick-net methodologies, but there was significant difference between  
37 diatom communities depending on site quality ( $P = 0.029$ ). These results show the  
38 diatom taxonomic coverage achieved through DNA metabarcoding of kick-net is  
39 suitable for ecological biomonitoring applications. The shift to a more robust sampling  
40 approach and capturing diatoms and macroinvertebrates in a single sampling event has  
41 the potential to significantly improve efficiency of biomonitoring programmes.

42

43 **Key words:** Biomonitoring, metabarcoding, periphyton, diatom, benthos, biodiversity,  
44 bioindicator, water quality, rbcL cpDNA, kick-net

## 45 **Introduction**

46 As climate change and other anthropogenic impacts continue to alter the  
47 environment, there is an increasing need for comprehensive ecological assessment.  
48 Rapid and robust biomonitoring is essential for informing management plans and  
49 mitigating further environmental degradation [1–3]. Freshwater biomonitoring typically  
50 involves sampling a range of aquatic taxa, with particular focus on biological indicator  
51 taxa, to assess environmental conditions based on diversity, richness, structure and  
52 function of the existing communities [3–5].

53  
54 Traditionally, biomonitoring data is generated through morphological taxonomic  
55 classifications, however there has been a recent shift towards DNA-based identification  
56 using metabarcoding [6] coupled with high throughput sequencing [7]. In aquatic  
57 systems such as wadable streams, a combination of bulk-tissue benthic sampling using  
58 kick-net methodology with DNA metabarcoding, facilitates rapid data collection whilst  
59 maintaining data integrity [8–10]. The metabarcoding approach has been employed for  
60 numerous biomonitoring studies involving macroinvertebrates [11,12] for assessing  
61 freshwater health [5,10,13].

62  
63 In addition to benthic macroinvertebrates, diatoms (members of Bacillariophyta) are also  
64 ideal biomonitoring target taxa for assessing freshwater system conditions [14–16].

65 These single-celled algae have a short generation time which allows for rapid  
66 responses to physical, chemical and biological changes in the environment [14,15,17].

67 Similar to macroinvertebrates, the high diversity and ubiquity of diatoms is used to

68 create biotic indices that can accurately report freshwater quality [16,18,19]. Studies  
69 have shown that diatoms respond more readily to the presence of heavy metal  
70 pollutants compared to macroinvertebrates, which are generally more sensitive to shifts  
71 in hydrological conditions [17,20–22]. Monitoring only one of these taxonomic groups to  
72 assess overall ecosystem health could potentially cause gaps in knowledge that could  
73 subvert subsequent management strategies. Hence, diatoms are being used in a  
74 number of national and regional biomonitoring programmes.

75  
76 Current methods for diatom sampling are time-consuming and laborious, which could  
77 hamper widespread use of diatoms for extensive freshwater biomonitoring [23,24]. The  
78 conventional diatom collection method involves the scraping of periphyton (a  
79 combination of algae, cyanobacteria, microbes, and detritus) from numerous substrates  
80 within littoral habitats [23–26]. These samples are then fixed and visualised using light  
81 microscopy [27–30]. From here, microscopy standards and keys are followed [29–31] to  
82 enable identification of diatoms to different taxonomic ranks. Within recent years, there  
83 has been the shift towards DNA metabarcoding-based identification of diatoms  
84 [15,16,32,33]. This involves the manual homogenized of periphyton scrapings into  
85 single samples, which are then processed via standard diatom metabarcoding  
86 procedures [34,35]. Alternative sampling methods, such as collection through the  
87 benthic kick-net technique, have not been tested for diatom biomonitoring applicability,  
88 however it is expected that this technique would drastically reduce time spent collecting  
89 samples. The ability to study diatom and macroinvertebrate assemblages from a single  
90 sample would allow biomonitoring programs to achieve an intensive appraisal of

91 freshwater conditions. In a rapidly changing world, streamlining current methodology to  
92 obtain as much data in as little time as possible is crucial.

93  
94 Because DNA-based analysis of environmental samples such as contents of a kick-net  
95 sample can provide a broad spectrum of organisms in the habitat sampled, we  
96 hypothesized that kick-net metabarcoding will provide diatom biodiversity comparable to  
97 commonly used scraping method. Specifically, we aimed to 1) investigate the feasibility  
98 of kick-net sampling for capturing community assemblages of freshwater diatoms  
99 versus conventional periphyton scraping using a high throughput sequencing coupled  
100 metabarcoding approach and 2) compare diatom community assemblages across a  
101 known habitat quality scale (Good and Fair) using both conventional and kick-net  
102 sampling to investigate presence of diatom indicator groups.

103

## 104 **Methods**

### 105 **Field Sampling**

106 Samples were collected in November 2019 from Grand River tributaries across  
107 four study sites in Waterloo, Ontario (Fig. 1). Status and location data were provided by  
108 Dougan & Associates based on a 2018 benthos biomonitoring project for the City of  
109 Waterloo (S1 Table). The four selected sites were a subset of the sites from this project  
110 and were chosen based on accessibility and habitat quality. Hilsenhoff Biotic Index  
111 ranges (weighted by species) informed the habitat quality scale [36] which categorized  
112 sites into 'Good' (4.51-5.50) and 'Fair' (5.51-6.50).

113

114 Collection occurred in riffles, starting with a benthic kick-net sample, followed by  
115 subsequent periphyton scrapings of microhabitats representative of the reach (S2  
116 Table). Periphyton scraping refers to the sampling of sediment, rock, macrophytes and  
117 leaf litter. Three replicates of each sampling type were collected at each site. Kick-net  
118 collection followed the Canadian Aquatic Biomonitoring Network [CABIN] protocol [37].  
119 Effort was standardized to three minutes. The sampler moved up stream in a zig-zag  
120 pattern to encompass all microhabitats within the reach. Periphyton scraping samples  
121 were comprised of five specimens per microhabitat type to account for variability within  
122 the microhabitat [23]. Negative controls, consisting of molecular grade water, were  
123 collected prior to the collection of each rock sample (n= 9) to ensure the toothbrushes  
124 used for scraping biofilms from rocks had been adequately sterilised (S3 Table). All  
125 other samples were collected using manufacture-sealed sterile equipment. All samples  
126 were collected in 1L sample jars and placed in a cooler to transport back to the lab.  
127 Upon arrival at the lab, samples (n=45) were preserved using 100% ethanol and stored  
128 in a -20°C freezer until processing.

129

## 130 **Sample Validation and Extraction**

131 To account for potential false negatives [38], diatom presence in the samples  
132 was confirmed using microscopy. A small amount of ethanol used to preserve the  
133 samples was placed on a slide and observed under a compound microscope at 100X  
134 magnification. Visual inspection confirmed the presence of diatoms in each sample type  
135 (S1 Fig.), however no taxonomic information was taken as morphological identification  
136 was beyond the scope of this study.

137  
138 Once diatom presence was validated, samples were homogenized using standard  
139 blenders decontaminated by washing with ELIMINase® (VWR, Canada) then rinsing  
140 with deionized water before treating with UV light for 30 minutes. Homogenate was  
141 subsequently transferred to 50 mL Falcon tubes, where one tube was set aside and  
142 centrifuged at 2400 rpm for two minutes. Supernatant was removed and residual pellets  
143 were incubated at 70 °C until fully dried. Next, approximately 300 mg dried tissue was  
144 subsampled into PowerBead tubes and DNA extractions were completed using the  
145 DNeasy Power Soil kit (Qiagen, CA) following the manufacturer's protocol. The only  
146 exception being that 50 µL of buffer C6 (TE) was used for final elution. Negative  
147 controls containing no tissue were also included with each batch of extractions. All  
148 negative controls failed to amplify and therefore were not sequenced.

149

## 150 **DNA Amplification, Library Preparation and Sequencing**

151 Amplification targeted the 312 base pair long region of the chloroplast gene  
152 ribulose biphosphate carboxylase large chain (rbcL) using five diatom specific primers.  
153 Following the methods of Rivera et al. [39], forward primers Diat\_rbcL\_708F\_1 (5'-  
154 AGGTGAAG- TAAAAGGTTTCWTTACTTAAA-3'), Diat\_rbcL\_708F\_2 (5'-AGGT-  
155 GAAGTTAAAGGTTTCWTAYTTAAA-3') and Diat\_rbcL\_708F\_3 (5'-AGGTGAAAC-  
156 TAAAGGTTTCWTTACTTAAA-3') were combined in an equimolar mix. Two reverse  
157 primers, Diat\_rbcL\_R3\_1 (5'-CCTTCTAATTTACC- WACWACTG-3') and  
158 Diat\_rbcL\_R3\_2 (5'-CCTTCTAATTTACCWA-CAACAG-3'), were also combined and  
159 used for amplification. Each reaction used the following reagents: 17.5 µL HyPure™

160 molecular biology grade water, 2.5  $\mu$ L 10X reaction buffer (200 mM Tris-HCl, 500 mM  
161 KCl, pH 8.4), 1  $\mu$ L  $MgCl_2$  (50 mM), 05.  $\mu$ L dNTPs mix (10 mM), 0.5  $\mu$ L of both forward  
162 (10 mM) and reverse (10 mM) equimolar mixes, 0.5  $\mu$ L Invitrogen's Platinum Taq  
163 polymerase (5 U) and 2  $\mu$ L of DNA. Final reaction volume totaled 25  $\mu$ L.  
164  
165 PCR protocol largely followed Rivera et al. [39] with minor adjustments. Instead of thirty  
166 cycles of denaturation at 95°C for 45 seconds, annealing at 55°C for 45 seconds and  
167 extension at 72°C for 45 seconds [39], this study increased the number of cycles to  
168 thirty-five. PCR amplification was also performed in two-steps, with the second PCR  
169 using 2  $\mu$ L of amplicons from the first PCR instead of DNA, and Illumina-tailed primers.  
170 All PCRs were completed in Eppendorf Mastercycler ep gradient S thermal cycler.  
171 Successful amplification was confirmed using 1.5% agarose gel electrophoresis before  
172 purifying second PCR amplicons with the MinElute Purification kit (Qiagen). The next  
173 step was quantifying purified samples with a QuantIT PicoGreen daDNA assay kit and  
174 using these values to normalize all samples to 3 ng/ $\mu$ L. Samples were then indexed and  
175 pooled before purifying with AMPure magnetic beads. QuantIT PicoGreen daDNA assay  
176 kit was once again used to quantify the library and Bioanalyzer was used to determine  
177 fragment length. The library was diluted to 4 nM and 10% PhiX was added before being  
178 sequenced using Illumina MiSeq with a V3 MiSeq sequencing kit (300 X 2; MS-102-  
179 2003).

180

## 181 **Bioinformatic Processing**



182           Illumina MiSeq paired-end reads were processed using the SCVURL rbcL  
183 metabarcode pipeline-1.0.2 pipeline available from  
184 [https://github.com/terrimporter/SCVURL\\_rbcL\\_metabarcode\\_pipeline](https://github.com/terrimporter/SCVURL_rbcL_metabarcode_pipeline)  
185 . SCVURL is an automated snakemake [40] bioinformatic pipeline that runs in a conda  
186 [41] environment. SeqPrep v1.3.2 [42] was used to pair raw reads requiring a minimum  
187 Phred score of 20 to ensure 99% base-calling accuracy. CUTADAPT v2.6 was used to  
188 trim primers from sequences, leaving a minimum fragment length of at least 150 base  
189 pairs [43]. Global exact sequence variant (ESV) [44] analysis was performed on the  
190 primer-trimmed reads. Reads were dereplicated using the ‘derep\_fulllength’ command  
191 with the ‘sizein’ and ‘sizeout’ options of VSEARCH v2.14.1 [45]. VSEARCH was also  
192 used to denoise the data using the unnoise3 algorithm [46]. These steps were taken to  
193 remove sequences with errors, chimeric sequences, PhiX carry-over and rare reads  
194 (singletons or doubletons) [47]. ESVs were classified using the rbcL diatom Classifier  
195 available from <https://github.com/terrimporter/rbcLdiatomClassifier>. Reference rbcL  
196 sequences were downloaded from the INRA diatom project [48] and reformatted to train  
197 the naive Bayesian classifier to make rapid, accurate taxonomic assignments [49]. This  
198 method makes assignments to the species rank and produces a statistical measure of  
199 confidence for each taxon up to the domain rank to help reduce false positive taxonomic  
200 assignments. We used 0.60 cutoff at the family rank (99% accuracy) and 0.20 cutoff at  
201 the genus rank (95% accuracy). The accuracy of the method assumes that target taxa  
202 are present in the reference database.

203

## 204 **Statistical Analysis**

205 RStudio was used to analyze the data [50]. To account for variable reads within  
206 the library each sample was normalized to the 15<sup>th</sup> percentile using the 'rrarefy' function  
207 in the vegan package [51,52].

208  
209 ESV richness across the various sampling and status categories was calculated to  
210 assess differences between the methods and sites. A non-metric multi-dimensional  
211 (NMDS) analysis on Sorensen dissimilarities (binary Bray-Curtis) was conducted using  
212 the vegan 'metaMDS' function to determine if sampling method or site status created  
213 variation in community structure [5]. A scree plot was run using the 'dimcheckMDS'  
214 command from the goeveg package to determine the number of dimensions (k=2) to  
215 use with vegan metaMDS function[53]. Shephard's curve and goodness of fit  
216 calculations were calculated using the vegan 'stressplot' and 'goodness' functions. The  
217 vegan 'vegdist' command was used to build a Sorensen dissimilarity matrix. We  
218 checked for heterogeneous distribution of dissimilarities using the 'betadisper' function.  
219 We used the 'adonis' function to perform a permutational analysis of variance  
220 (PERMANOVA). PERMANOVA was performed on conventional sampling methods  
221 (periphyton scraping) and kick-net methods, as well as site status to test for significant  
222 interactions between the categories [54].

223  
224 To maintain a balanced design during statistical testing, we pooled all periphyton  
225 sampling into one sample type (conventional) and maintained kick-net samples as a  
226 separate sample type. The Jaccard index was calculated to assess the overall  
227 similarities between the sites, collection methods and site status. Nestedness and

228 turnover of between kick-net and conventional samples were calculated using R  
229 package betapart function ‘beta-pair’ [55] followed by vegan function ‘betadisper’. The  
230 number of diatom family ESVs detected from kick-net or pooled conventional samples  
231 was also plotted. A dendrogram of diatom families detected was plotted using  
232 RAWGraphs ([app.rawgraphs.io](http://app.rawgraphs.io)) and color-coded to show the samples the families were  
233 detected in [56]. Lastly, the frequency of ESVs detected from diatom families was  
234 visualized using a heatmap generated using geom\_tile (ggplot) in R, plotting individual  
235 sample types for each site, split into two plots according to site status.

236

## 237 **Results**

238 After bioinformatic processing, we generated 4,272 ESVs (2,166,157 reads).  
239 After taxonomic filtering (removal of non-diatom phyla), a total of 3,940 diatom ESVs  
240 (2,125,984 reads) were retained for data analysis. Read coverage per sample after  
241 normalisation (15<sup>th</sup> percentile cut-off) was 37,735.

242

243 Since the rarefaction curves plateau, this indicated that the sequencing depth was  
244 sufficient to capture the ESV diversity in our PCRs (S2 Fig.). In terms of the top 10  
245 orders identified, the order Naviculales represented 30.6% of ESVs (30% of reads) and  
246 Bacillariales represented 18.6% of ESVs (15.4% of reads; S3 Fig.).

247

## 248 **Taxonomic Coverage**

249 In terms of taxonomic assignment, we identified a total of 1 phyla (Bacillariophyta), 4  
250 classes, 23 orders, 44 families and 77 genera at the 95% correct assignment level. ESV

251 richness varied across different sampling methods (Fig. 2). Mean overall ESV richness  
252 was used to calculate alpha diversity which displayed very similar values for all  
253 sampling methods across the four sites (S4 Table). Averaged across sites, kick-net  
254 samples produced the lowest mean ESV richness ( $225 \pm 85$ ), with sediment samples  
255 producing the highest ESV richness ( $317 \pm 92$ ).

256  
257 Through investigating diatom families, a majority of families detected were present in all  
258 microhabitats and kick-net samples (Fig. 3). Two families (Coscinodiscaceae and  
259 Orthoseriaceae) were solely present in leaf litter samples and two families  
260 (Entomoneidaceae and Diadesmidaceae) were present only in sediment samples (Fig.  
261 3).

262  
263 In terms of diatom genera, some of the confidently identified genera represented by  
264 more than 2 sequence variants, identified from kick-net and conventional samples,  
265 included: *Nitzschia* (Bacillariales), *Polypedilum* (Chironomidae), *Navicula* (Naviculales),  
266 *Amphora* (Thalassiophysales) and *Ulnaria* (Licmophorales; Fig. 4).

## 267 268 **Diatom Diversity by Method and Site Status**

269 NMDS plots showed that replicates clustered close together for site and status,  
270 with overlap observed between sampling methods and replicates (Fig. 5). When pooling  
271 conventional periphyton samples (i.e. macrophyte, leaf litter, rock, and sediment) at  
272 each site, there remained overlap between kick-net and conventional samples and  
273 samples also remained clustered by site and status (S4 Fig). PERMANOVA of the

274 pooled samples, shows that analyzing data from kick-net or conventional samples  
275 (method) explains 13% of the variation in Bray Curtis dissimilarities (p-value = 0.776),  
276 sampling site (site) explains 58% of the variation (p-value = 0.009) and habitat quality  
277 status (status) explains 22% of the variation observed (p-value = 0.029; S5 Table). The  
278 Jaccard index for kick-net compared with conventional samples is 0.53, indicating  
279 samples are 53% similar, whereas the Jaccard index for fair compared to good site  
280 quality status samples is 0.20, indicating samples are only 20% similar. In terms of beta  
281 diversities of communities aggregated by the treatments of "kick-net" and  
282 "conventional", there was no significant difference between turnover. For beta  
283 diversities of communities aggregated by site status, there was a significant difference  
284 between nestedness ( $P < 0.05$ ) but not for turnover ( $P = 0.06$ ). Fair samples appear to  
285 be significantly nested within good samples. These results further indicate that site  
286 status has a significant effect on the sampled community composition whereas  
287 conventional versus kick-net sampling methods do not.

288  
289 For individual sample types (i.e. kick-net, macrophyte, leaf litter, rock, and sediment),  
290 the heatmap shows that kick-net samples are largely representative of the diversity of  
291 families detected within each conventional periphyton sampling method (Fig.6.). In  
292 some cases, kick-net samples failed to detect diatom families which were present in  
293 conventional periphyton samples (e.g. Sellaphoraceae and Diadesmidaceae in Clair15)  
294 and conversely, kick-net samples also detected families which were not detected in  
295 conventional periphyton samples (e.g. Eunotiaceae and Neidiaceae in Clair12; Fig. 6).  
296 Similar assemblages of diatoms communities were detected across both fair and good

297 quality sites, with the main difference observed between fair and good sites being the  
298 number of reads produced for families such as Thalassiosiraceae which was detected  
299 with a high number of reads (1000+) in fair sites and a lower number of reads (10-100)  
300 in good sites (Fig. 6).

301

## 302 **Discussion**

303 The demand for high-quality, reproducible ecological data is increasing in  
304 conjunction with the degradation of ecosystems globally [57]. There is a need to further  
305 streamline existing biomonitoring methodologies without sacrificing the quality of data  
306 produced [4,7,54,58]. With diatom assemblages providing a unique insight into the  
307 water quality status of lentic and lotic systems, fast-tracking diatom data collection for  
308 ecological assessments is a priority [39]. We have demonstrated that kick-net  
309 methodology with DNA metabarcoding provides sufficient taxonomic coverage to  
310 potentially be utilised as a for assessing diatom biodiversity in freshwater systems.

311

312 Kick-net sampling technique, whereby a zig-zag path is taken across the reach,  
313 provided sufficient representation of existing diatom community assemblages within  
314 site-specific microhabitats. Samples derived from the kick-net technique were highly  
315 comparable with conventional samples in terms of diatom taxa detected, despite the  
316 kick-net approach being more passive compared to direct periphyton scraping. Specific  
317 diatom taxa are known to have ecological preferences for different freshwater  
318 microhabitats [59,60]. For watershed-level health estimates, it is beneficial to be able to  
319 efficiently detect the diversity of diatom taxa present without directly sampling each

320 microhabitat within a reach. We have demonstrated that kick-net methodology can  
321 sufficiently capture the existing diatom biodiversity, ground truthed by comparing  
322 assemblages detected with periphyton scrapings.

323  
324 Ultimately, the detection of bioindicator species is a key variable to consider when  
325 comparing biomonitoring methods, as these taxa are pivotal for detecting subtle  
326 differences in freshwater health [3,5,14]. Naviculaceae contains diatom species  
327 sensitive to herbicide exposure, which is a family we observed in all sites and with all  
328 collection methods [61]. Additionally, the bioindicator family Stephanodiscaceae, (a  
329 known tolerant taxon) [62], has a higher read abundance in 'Fair' sites compared to  
330 'Good' in both conventional and kick-net sample types. Despite the direct sampling  
331 approach of periphyton rock scraping, this methodology failed to detect this family at  
332 one of the sites where kick-net samples were successful at detecting this benthic family.  
333 Rock scrapings are commonly used as the sole collection method for diatoms  
334 [14,39,63,64], which suggests that the kick-net approach facilitates the detection of taxa  
335 which otherwise may be missed from conventional sampling.

336

337

## 338 **Conclusion**

339 Overall, this study found that benthic kick-net methodology enables a robust and  
340 detailed assessment of freshwater diatom communities. This methodology is a scalable  
341 option for generating a holistic insight into the health of freshwater systems. The high  
342 similarity of diatom taxa detected between methods and significant differences between

343 diatom communities detected in sites of differing habitat quality, demonstrates that this  
344 rapid method can provide accurate, fine-resolution taxonomic results. Future research  
345 should examine the duo-analyses approach of macroinvertebrate and diatom  
346 communities from a single kick-net sample, to determine reproducibility of multi-taxa  
347 targeting with this method. Additionally, future studies should consider exploring the use  
348 of multiple markers (i.e. rbcL cpDNA versus 18S rRNA gene), to address level of  
349 taxonomic resolution that can be obtained with these markers commonly used for  
350 diatom DNA barcoding.

351

352

## 353 **Supporting Information**

354 **S1 Table. Information on study sites, including GPS coordinates and site status.**

355 **S2 Table. Outline of collections methods used in this study.** Samples for periphyton  
356 scraping were taken from a depth no greater than 1m (King et al. 2006).

357 **S3 Table. Summary table of decontamination and sterilisation procedures**  
358 **undertaken for the equipment in this study.**

359 **S4 Table. Mean ESV values (replicates pooled) for each sample type across the**  
360 **four sites.** Based on normalised data.

361 **S5 Table. rbcL exact sequence variants (ESVs) are not significantly different**  
362 **between sampling methods (kick-net versus conventional periphyton sampling).**

363 No significant beta dispersion was detected within groups (method, site, status). Only  
364 significant difference detected was rbcL ESVs between sites and status. Summary of



365 PERMANOVA results based on a Sorensen dissimilarity matrix of rbcL ESVs.

366 Significant p-values are bolded.

367 **S1 Fig. Example of confirmation of diatom presence from preservative of kick-net**  
368 **sample.** Image: CBG Photography Group.

369 **S2 Fig. All samples show that ESV sampling reached saturation.** Samples were  
370 color-coded by site or method as shown in the legend. The vertical dashed line  
371 indicates the 15th percentile of sampling read depth, which is the number of reads that  
372 would be used in any future analysis based on normalized data.

373 **S3 Fig. Naviculales is the most abundant diatom order detected.** Results for the  
374 top 10 orders are shown with respect to proportion of ESVs and reads recovered.  
375 Based on raw unnormalized data.

376 **S4 Fig. Non-metric multi-dimensional scaling plots of microhabitat samples**  
377 **pooled show clustering by due to site and status.** Specifically, a) depicts overlap  
378 between the binary Bray Curtis (Sorensen) dissimilarities between different sampling  
379 approaches, b) sample site clustering c) clustering based on habitat quality status  
380 (stress = 0.012, R<sup>2</sup> = 0.98). Based on rarefied data.

381

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392

## 393 **Author Contributions and Competing Interests**

394 V.C.M. collected samples, conducted molecular and genomic analyses and wrote  
395 the manuscript with help from all authors, C.V.R and M.H. designed the study, C.V.R  
396 contributed to sampling, bioinformatic processing and statistical analyses, T.M.P trained  
397 the classifier, contributed to bioinformatic processing and advised on data analysis. All  
398 authors helped to write/edit the manuscript. The authors have declared that no  
399 competing interests exist.

400

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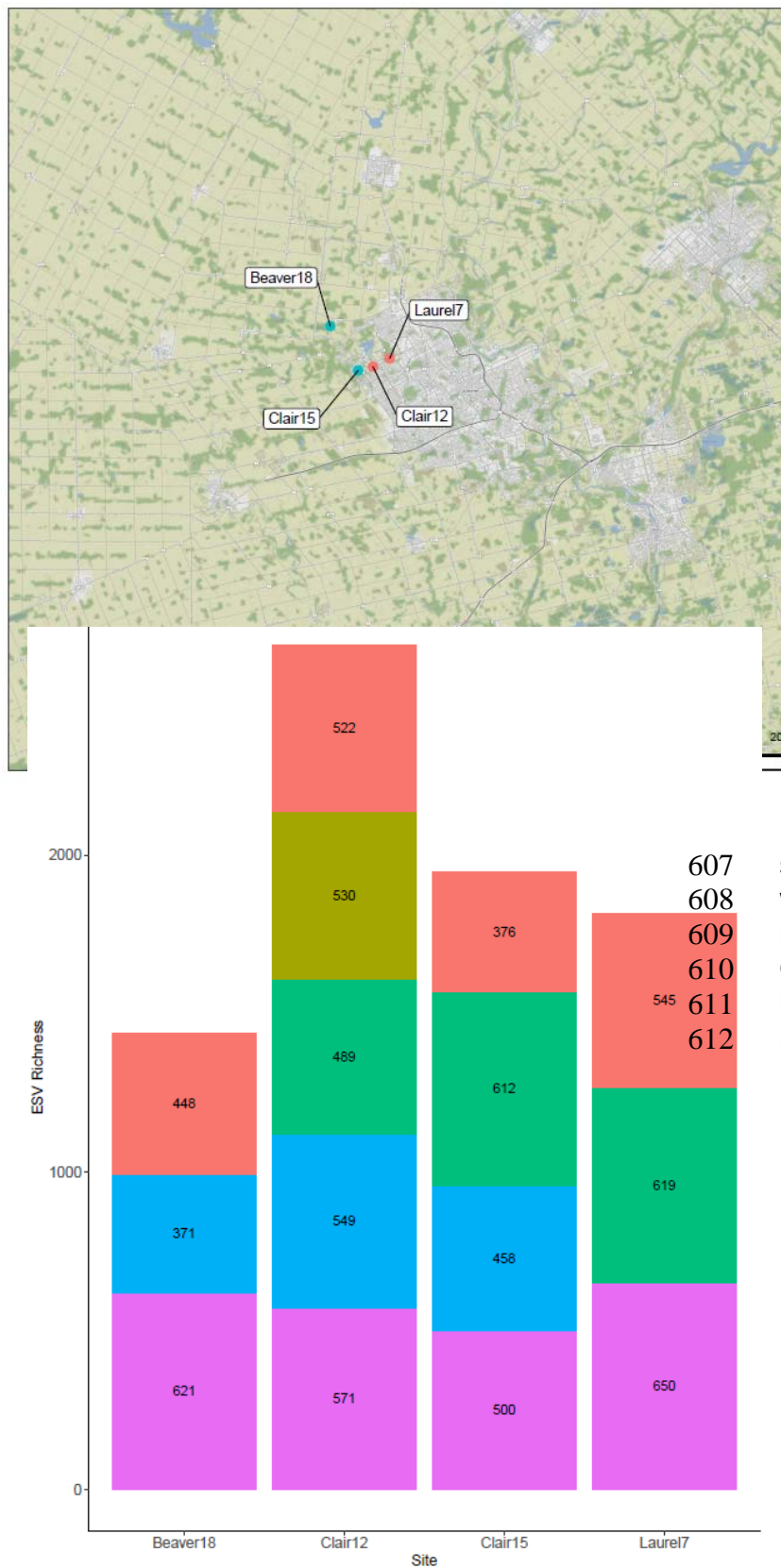
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607 **Fig. 1. Map of**  
 608 **located**  
 609 **Waterloo**  
 610 **(Ontario,**  
 611 **bar shown in**  
 612 **status**  
 613 **legend.**

607 **sample sites**  
 608 **within the**  
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 610 **(Canada)** Scale  
 611 km, site habitat  
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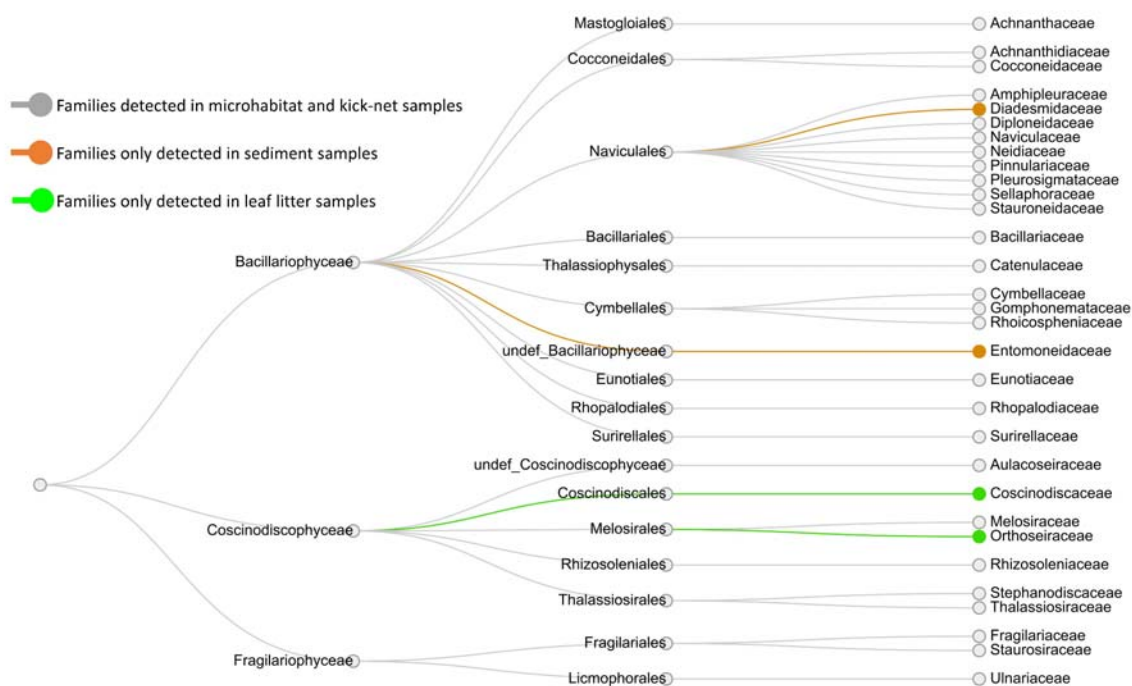
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625 **Fig. 2. ESV richness varies across different sample types.** Methods refer to the  
626 different sampling approaches analyzed (i.e. Kick-net, Macrophyte, Leaf Litter, Rock  
627 and Sediment). Replicates are pooled. Based on rarefied data.

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630 **Fig. 3. A majority of diatom families were detected in both microhabitat and kick-**  
 631 **net samples.**

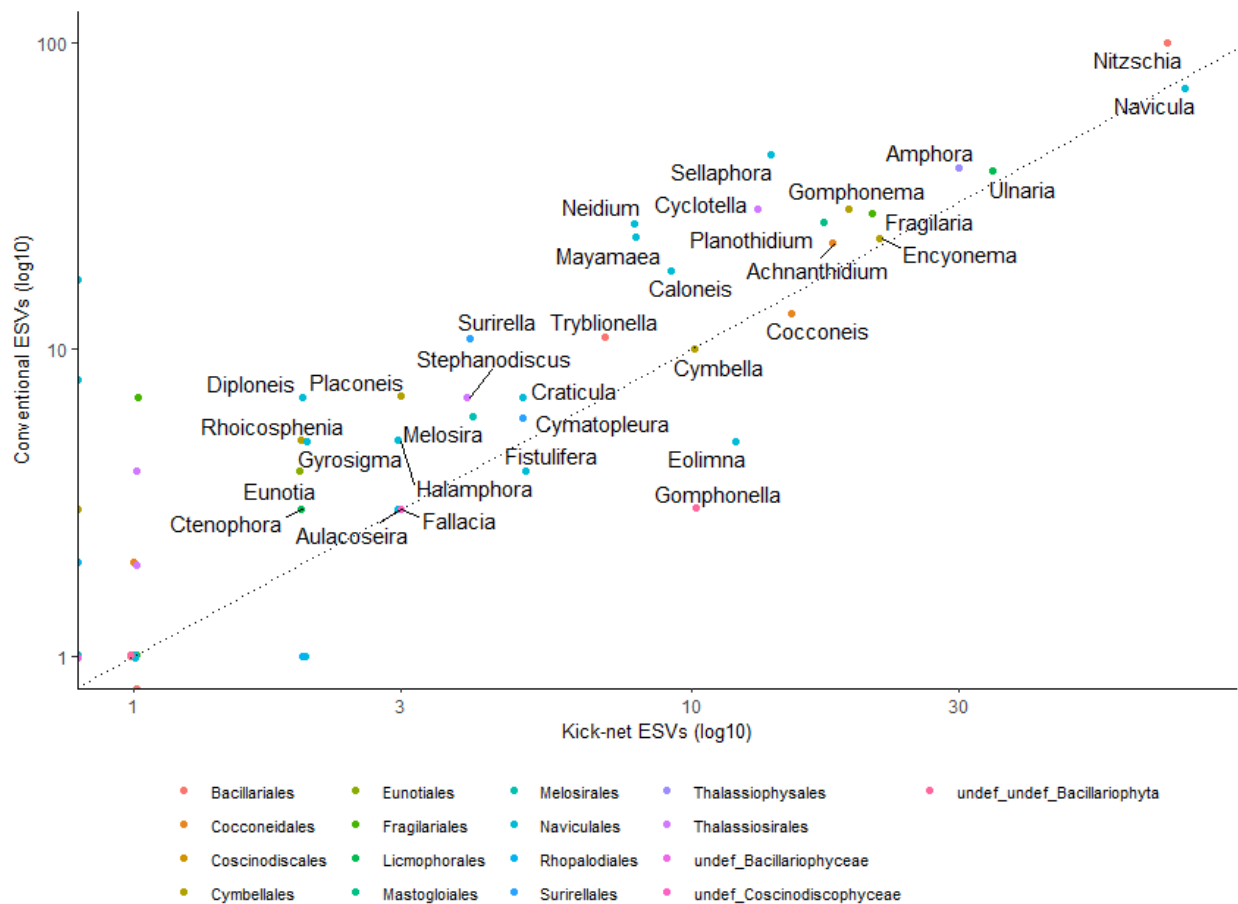
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638 **Fig. 4. Number of ESVs detected from genera detected from kick-net versus**

639 **conventionally sampled diatoms are similar.** The points are color-coded for the

640 orders detected in this study. A 1:1 correspondence line (dotted) is also shown. A log10

641 scale is shown on each axis to improve the spread of points with small values. Based

642 on rarefied data.

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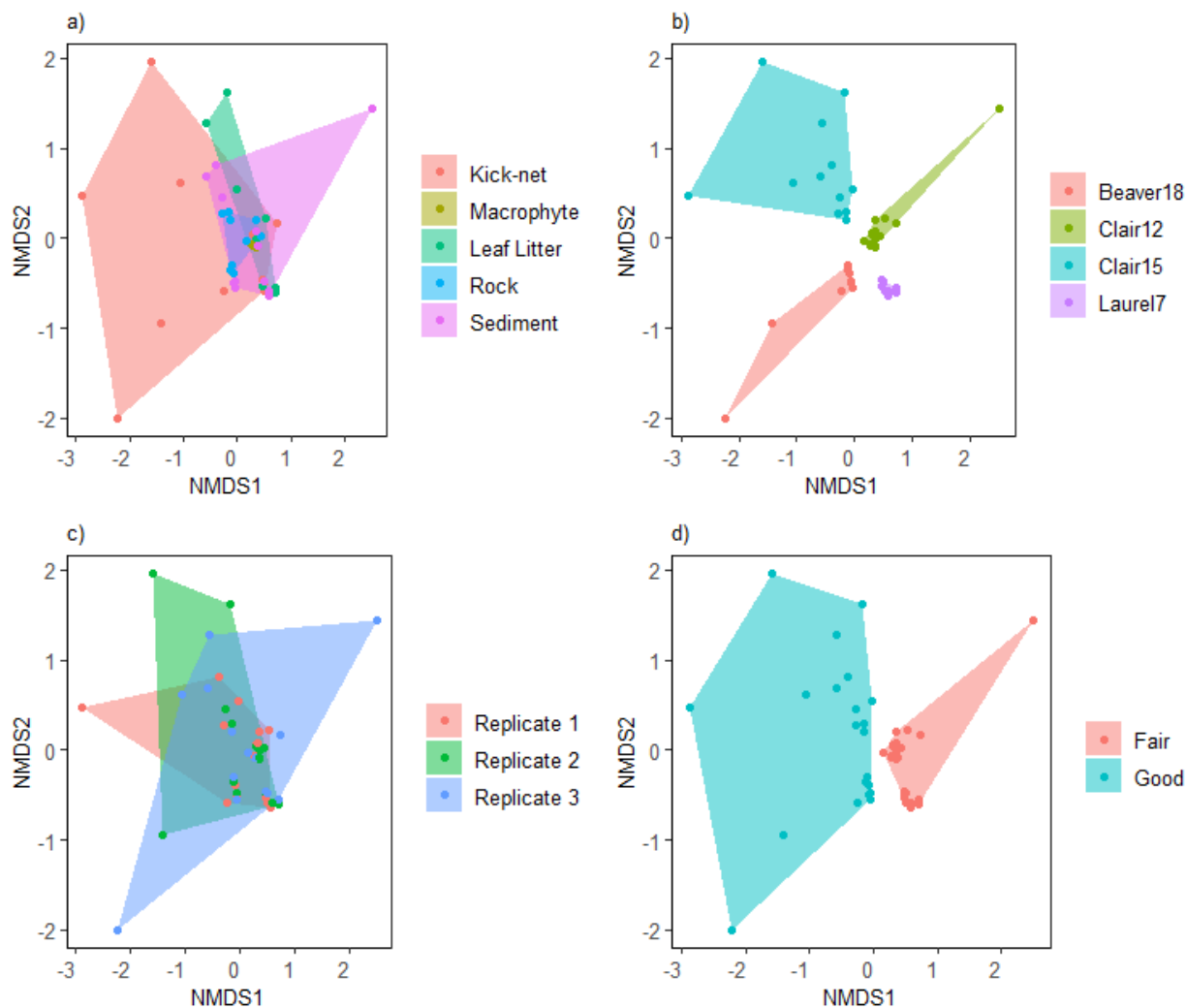
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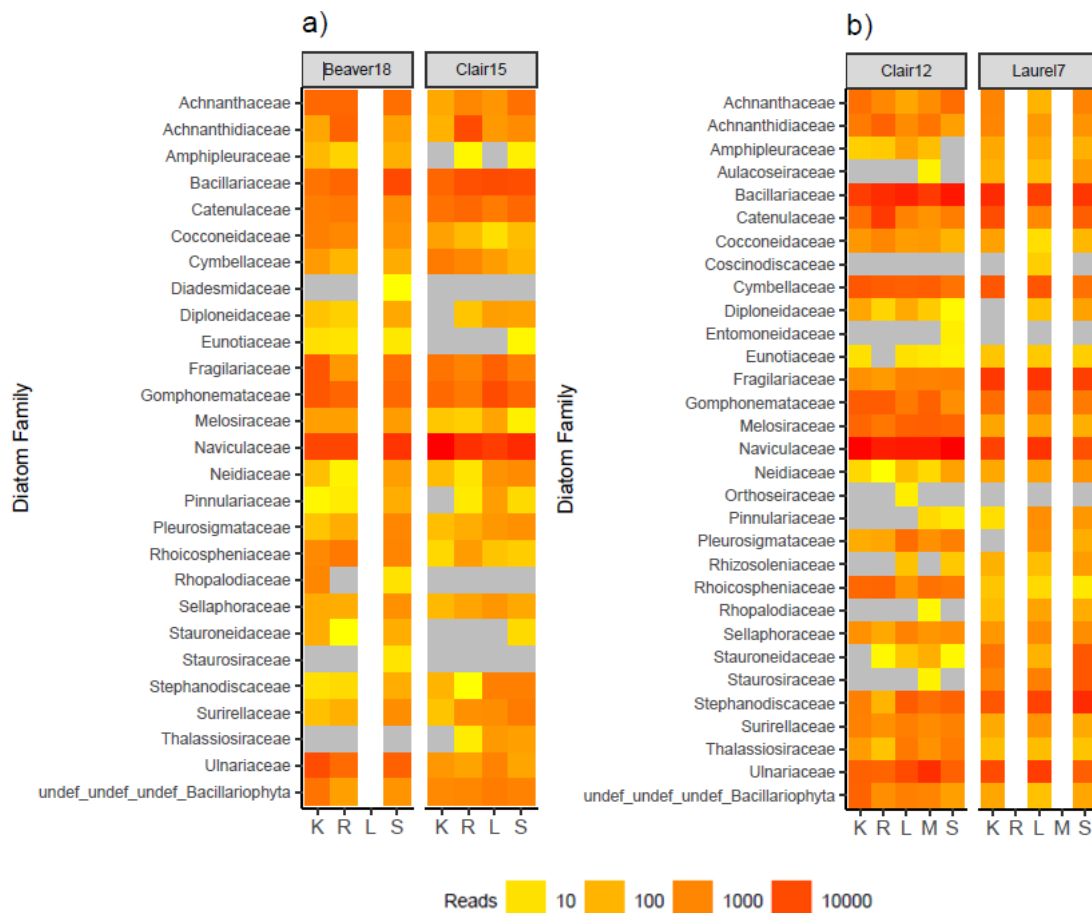
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**Fig. 5. Non-metric multi-dimensional scaling plots show clustering mainly due to site and status.** Specifically, a) binary Bray Curtis (Sorensen) dissimilarities overlapping across different sampling approaches, b) clustering by site, c) overlap between replicates, and d) clustering based on habitat quality status (stress = 0.111,  $R^2 = 0.98$ ). Based on rarefied data.



667

668 **Fig. 6. Samples detect similar diatom families across sampling methods and site**

669 **status.** Only ESVs taxonomically assigned to families with high confidence (bootstrap

670 support  $\geq 0.60$  for 95% accuracy) are included. Part a) shows sites with a ‘good’

671 quality status b) sites with a ‘fair’ quality status. Sampling methods: K = kick-net; R =

672 rock scraping; L = leaf litter; M = macrophyte; S = sediment. Empty lanes indicate the

673 corresponding microhabitat was not present at the site. For each site, three replicates

674 for each sampling method are pooled. Based on normalized data.

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## 677 **Data Availability**

678 Raw sequences will be available from NCBI SRA on acceptance. The SCVURL rbcL

679 metabarcode pipeline-1.0.2 is available from

680 [https://github.com/terrimporter/SCVURL\\_rbcL\\_metabarcode\\_pipeline](https://github.com/terrimporter/SCVURL_rbcL_metabarcode_pipeline) and the

681 rbcLdiatomClassifier v1 we used is available on GitHub at

682 <https://github.com/terrimporter/rbcLdiatomClassifier>.

683