

1 Revisiting reptile home ranges: moving beyond traditional 2 estimators with dynamic Brownian Bridge Movement Models

3 Inês Silva^{1*}, Matt Crane¹, Benjamin Michael Marshall², Colin Thomas Strine^{2**}

4 ¹ Conservation Ecology Program, School of Bioresources and Technology, King Mongkut's University of
5 Technology Thonburi, Bangkokunthien, Bangkok, Thailand.

6 ² School of Biology, Institute of Science, Suranaree University of Technology, Nakhon Ratchasima,
7 Thailand

8 * Email: imss.silva@gmail.com

9 ** Email: strine.conservation@gmail.com, Colin_Strine@sut.ac.th

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11 Abstract

12 Animal movement, expressed through home ranges, can offer insights into spatial and habitat requirements.
13 However, home range estimation methods vary, directly impacting conclusions. Recent technological
14 advances in animal tracking (GPS and satellite tags), have enabled new methods for home range estimation,
15 but so far have primarily targeted mammal and avian movement patterns. Most reptile home range studies
16 only make use of two older estimation methods: Minimum Convex Polygons (MCP) and Kernel Density
17 Estimators (KDE), particularly with the Least Squares Cross Validation (LSCV) and reference (h_{ref})
18 bandwidth selection algorithms. The unique characteristics of reptile movement patterns (*e.g.* low
19 movement frequency, long stop-over periods), prompt an investigation into whether newer movement-
20 based methods –such as dynamic Brownian Bridge Movement Models (dBBMMs)– are applicable to Very
21 High Frequency (VHF) radio-telemetry tracking data. To assess home range estimation methods for reptile
22 telemetry data, we simulated animal movement data for three archetypical reptile species: a highly mobile
23 active hunter, an ambush predator with long-distance moves and long-term sheltering periods, and an
24 ambush predator with short-distance moves and short-term sheltering periods. We compared traditionally
25 used home range estimators, MCP and KDE, with dBBMMs, across eight feasible VHF field sampling
26 regimes for reptiles, varying from one data point every four daylight hours, to once per month. Although
27 originally designed for GPS tracking studies, we found that dBBMMs outperformed MCPs and KDE h_{ref}
28 across all tracking regimes, with only KDE LSCV performing comparably at some higher-frequency
29 sampling regimes. The performance of the LSCV algorithm significantly declined with lower-tracking-
30 frequency regimes, whereas dBBMMs error rates remained more stable. We recommend dBBMMs as a
31 viable alternative to MCP and KDE methods for reptile VHF telemetry data: it works under contemporary
32 tracking protocols and provides more stable estimates, improving comparisons across regimes, individuals
33 and species.

34

35 **Keywords:**

36 Reptile, home range, simulation, spatial ecology, minimum convex polygon, kernel density, dynamic

37 Brownian Bridge Movement Models, snake, lizard, squamate, tortoise

38

39 1. Introduction

40 Animal movement is an underlying process in many ecological systems, and there is a growing
41 understanding of how individuals behave through space and time (Nathan *et al.*, 2008; Gurarie *et al.*, 2016).
42 Movement is often conceptualized then presented as a home range, defined as the area animals move
43 through during “normal” activities, including resource acquisition and reproduction (Burt, 1943; Powell
44 2012). While the utility of the home range concept has been questioned in recent years (Kie *et al.*, 2010;
45 Powell & Mitchell, 2012), home range estimates continue to have a range of applications: identifying
46 behavioural adaptations to predictable environmental features (Riotte-Lambert & Matthiopoulos, 2019) or
47 inferring habitat use (Fisher, 2000; Dickson & Beier, 2002; Tikkanen *et al.*, 2018; Marshall *et al.*, 2019).
48 Applying a home range approach to ecological research questions requires careful consideration (Péron,
49 2019), as any conclusions drawn can be profoundly impacted by the natural history of the target species.

50 Terrestrial reptiles —broadly lizards, snakes, and tortoises— have *distinct* natural histories from mammals
51 (*e.g.* as ectotherms), resulting in *distinct* movement patterns. Many reptiles move less frequently than
52 comparatively sized mammals (Hailey, 1989), but more importantly, many terrestrial reptiles spend
53 prolonged periods stationary under shelter (one day to several weeks; Guarino, 2002; Bruton, McAlpine,
54 Smith, & Franklin, 2014; Mata-Silva, DeSantis, Wagler, & Johnson, 2018). These inconsistent movement
55 patterns severely impact inferences drawn from home range analyses.

56 To properly inform desperately needed conservation actions (Gibbons *et al.*, 2000; Roll *et al.*, 2017), we
57 must tailor our methodologies to the peculiarities of reptile movement (Péron, 2019) –otherwise we risk
58 designing suboptimal solutions. We must assess the utility of newer methods designed for mammals, before
59 applying them to reptiles (Silva, Crane, Suwanwaree, Strine, Goode, 2018).

60 With the rise of Global Positioning System (GPS) animal tracking, researchers have developed new
61 statistical approaches for calculating home ranges that take advantage of the high number of location fixes.
62 However, GPS tracking currently has limited application in reptiles (see Schofield *et al.*, 2007; Campbell
63 *et al.*, 2013; Rosenblatt *et al.*, 2013; Smith, Hart, Mazzotti, Basille, & Romagosa, 2018) as their natural
64 history poses several problems (Hebblewhite & Haydon, 2010; Wolfe, Fleming, & Bateman, 2018); *e.g.*
65 weakened signal due to the surgical implantation or attachment of the tag, limited number of species which
66 can be ethically attached due to body size (Smith *et al.*, 2018), reduced fix rate and precision due to
67 sheltering underground (Bruton *et al.*, 2014, Wolfe *et al.*, 2018).

68 Given that traditional home range estimators –Minimum Convex Polygons (MCP) and Kernel Density
69 Estimators (KDE)– present major limitations for telemetry-based reptile studies (see Row & Blouin-
70 Demers, 2006), it is important to investigate whether newer methods developed for GPS tracking data can
71 be applied to reptile-targeted Very High Frequency (VHF) radio-telemetry studies. Dynamic Brownian
72 Bridge Movement Models (dBBMMs) are a technique intended for GPS telemetry, allowing for efficient
73 and repeatable analysis of high-resolution data –particularly useful for animals with behaviourally distinct

74 movement patterns. The method creates a one-dimensional fix-frequency independent behavioural measure
75 (Brownian motion variance; Kranstauber, Kays, LaPoint, Wikelski, & Safi, 2012) that have been employed
76 to elucidate avian and mammal home range and movement patterns (e.g. Palm *et al.*, 2015; Byrne, McCoy,
77 Hinton, Chamberlain, & Collier, 2014; Lai, Bêty, & Berteaux, 2015; Buechley, McGrady, Çoban, &
78 Şekercioglu, 2018).

79 Leveraging dBBMMs may benefit VHF studies (Silva *et al.*, 2018; Walter, Onorato, & Fischer, 2015); and
80 while multiple simulations studies have investigated how different methods interact with animal movement
81 and home range delineation (e.g. Katajisto & Moilanen, 2006; Row & Blouin-Demers, 2006; Knight *et al.*,
82 2009; Cohen, Prebyl, Collier, & Chamberlain, 2018), none have targeted reptile-specific movement
83 patterns.

84 We assess home range estimates resulting from variable study designs common in the reptile spatial ecology
85 literature: namely temporally low-resolution tracking regimes. We simulate movement data of three
86 archetypal reptile species, thoroughly examining the most common home range estimators —Minimum
87 Convex Polygons (MCPs) and Kernel Density Estimators (KDEs). We next compare traditional estimators
88 to a newer method: dynamic Brownian Bridge Movement Models (dBBMMs). Finally, we discuss the
89 implications of home range estimator choice, and present guiding principles for reptile spatial ecology
90 sampling designs.

91 2. Materials and Methods

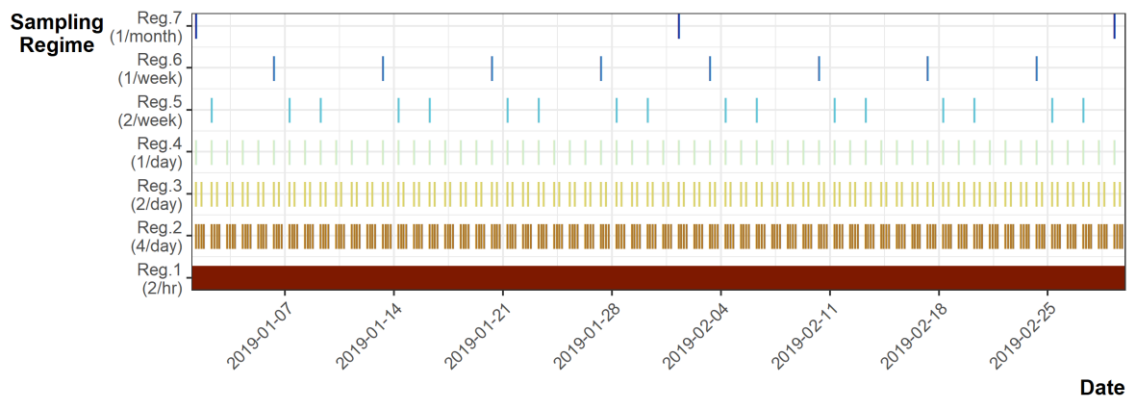
92 2.1. SIMULATED ANIMAL MOVEMENT AND TRACKING DATA

93 We used the *SimData* function in the *momentuHMM* package (McClintock & Michelot, 2018) to simulate
94 movement data from a Hidden Markov Model (HMM). HMMs are time-series models where the movement
95 pattern of an animal is assumed to depend on the underlying behavioural state of the animal (Langrock *et*
96 *al.*, 2014). We simulated data for 32 individuals from three archetypal reptile species, to represent three
97 main groups within reptile movement ecology: **Species 1** corresponds to highly mobile (active hunters)
98 with long-term shelter sites (e.g. monitor lizards, some skinks, and elapids like mambas and king cobras);
99 **Species 2** represents less mobile reptiles, capable of moving long distances but are ambush foragers, and
100 will still shelter for long periods (e.g. pythons); finally, **Species 3** represents smaller ambush predators,
101 infrequently moving and sheltering for shorter periods (e.g. viperid snakes, some smaller lizard species).

102 Each archetype had a unique set of state-dependent parameters and transition probabilities with the same
103 three behaviour states: “sheltering” (state 1), “moving” (state 2), “resting” (state 3). The state-dependent
104 data streams included step length (l_t) and turning angle (θ_t), which we generated from Gamma and von
105 Mises distributions, respectively. The simulations included a spatially correlated covariate for state 2, to
106 reflect habitat preferences, while states 1 and 3 followed a *cosinor* function, to reflect cyclical patterns of
107 long-term sheltering (state 1) and circadian rhythms (state 3). To simulate individual variation and
108 movement in a heterogeneous landscape we generated a random neutral landscape with fractal Brownian

109 movement, using the *NLMR* package (Sciaini, Fritsch, Scherer, & Simpkins, 2018). For further details on
110 these simulated species, as well as their specific step lengths, turning angles and transitional probabilities,
111 see Appendix S1, Supporting Information.

112 After creating the full simulated data set (regime 1), we generated six subsets of the data to represent various
113 field sampling regimes (regime 2-7): four locations per day, two locations per day, one location per day,
114 two locations per week, one location per week, and one location per month (Figure 1). For each subset, we
115 assumed a consistent regularly scheduled sampling protocol limited to the species' activity periods.



116

117 **Figure 1.** Example two-month period showing how data is thinned to represent different tracking regimes.

118 The autocorrelated nature of tracking data poses difficulties for home range estimators that assume
119 independence between points, namely KDEs. Attempting to remove autocorrelation to fit these assumptions
120 can reduce the biological relevance of the home range (De Solla *et al.*, 1999), but advocated in reptile home
121 range studies (Swihart & Slade, 1985; Worton, 1987).

122 We investigated the temporal autocorrelation present in our simulated dataset to determine whether our
123 coarser sampling regimes compiled with KDE independence assumptions. Other than less frequent
124 tracking, autocorrelation may be reduced by removing repeated locations. This method is particularly
125 relevant for reptiles that exhibit long term sheltering. We considered this special case –sampling regime 8–
126 where only animal relocations are included in the home range estimation. For regime 8 we used the four
127 location per day sampling regime, and then removed data points where the animal was stationary.

128 We described the autocorrelation in the simulated data using the *ctmm* package's variogram functionality
129 (Calabrese, Fleming, & Gurarie, 2016; Fleming *et al.*, 2017), and plotted the minimum number of days
130 until the autocorrelation became insignificant with raincloud plot code from Allen, Poggiali, Whitaker,
131 Marshall, & Kievit (2019).

132 2.2. HOME RANGE ESTIMATORS

133 2.2.1. *Minimum convex polygon*

134 We calculated the Minimum Convex Polygon (MCP) for each simulated individual that created the smallest
135 area convex polygon containing all animal locations. We used the 95% MCP, which removes outlying

136 points on the assumption that these represent exploratory movements and thus not part of the home range,
137 as originally defined by Burt (1943). The MCP method has long been lauded as a way of maintaining
138 comparability and historical consistency with previous studies (Jennrich & Turner, 1969), yet has well
139 documented issues: extreme sensitivity to sampling size and tracking duration (Anderson, 1982), and
140 overestimated boundary delineation (Robertson, Aebischer, Kenwards, Hanski, & Williams, 1998), with
141 the inclusion of areas that the animals never use (Börger *et al.*, 2006; Laver & Kelly, 2008). However, Row
142 & Blouin-Demers (2006) argued that MCPs are preferable to kernel density estimators specifically for
143 herpetofauna, and MCPs' use persists for "comparisons" in reptile telemetry studies (Petersen, Goetz,
144 Dreslik, Kleopfer, & Savitzky, 2019). An additional and considerable limitation of MCPs is that they do
145 not create a probabilistic utilization distribution.

146 *2.2.2. Fixed kernel home range*

147 Fixed KDE home ranges rely on a smoothing parameter (bandwidth) to generate a utilization distribution.
148 Bandwidth selection for KDE can dramatically influence home range estimation (Seaman *et al.*, 1999), and
149 thus we included two bandwidth selection algorithms, reference bandwidth (h_{ref}) and Least-Squares Cross-
150 Validation (LSCV), for our comparisons. Both bandwidth selection methods are frequently used in reptile
151 VHF studies, but potentially flawed for herpetofauna (Row & Blouin-Demer, 2006). The h_{ref} method tends
152 to overestimate home ranges while LSCV tends to underestimate (Hemson *et al.*, 2005). In general, fixed
153 KDE home ranges are not accurate when using autocorrelated data regardless of bandwidth selection
154 function (Noonan *et al.*, 2018).

155 *2.2.3. Dynamic Brownian Bridge Movement Model*

156 Dynamic Brownian Bridge Movement Models (dBBMMs) provide utilization distributions based on animal
157 movement paths. The method accounts for temporal autocorrelation, so it requires all locations to be time
158 stamped. In addition, dBBMMs incorporate error associated with each triangulated location, which we kept
159 consistent across species and regimes (at 5 metres) for the following reasons: (1) neither MPCs nor KDEs
160 account for location error, so the evaluation of the impact of this metric would be solely on one method and
161 not effective for comparison purposes; (2) location error associated with VHF telemetry is extremely
162 variable, dependent on macro and micro-habitat characteristics as well as tracking protocols (which we are
163 not assessing); and (3) we wanted to account for cases where GPS error can be greater than step length (e.g.
164 vipers, small lizards). The dBBMM method also allows calculation of Brownian motion variance (σ^2m),
165 which can help researchers determine how movement trajectories can occur due to a species' behaviour
166 and activity (Kranstauber *et al.*, 2012). Motion variance can help detect breeding and foraging behaviour
167 in reptiles, even with VHF telemetry data (Silva *et al.*, 2018).

168 2.3 METHOD COMPARISON

169 To compare the error generated from each home range estimator, we calculated the overlap with the
170 theoretical "true home range" for each individual. We generated an individual's "true home range" by
171 creating a buffer around all the simulated movement points with a width of two-times the step length

172 intersect from each simulated species' movement state (40-m for Species 1, 20-m for Species 2, 10-m for
173 Species 3). This provided a conservative home range estimate (excluding the impact of habitat), but more
174 generous and biologically sensible than only using simulated movement pathways. For each home range,
175 we calculated the omission (Type I, false positive) and commission (Type II error, false negative), using
176 the 95% contours for MCP, KDE and dBBMMs. We used the 95% contours, as this is the standard level
177 used in most home range estimates. We then calculated the F-measure [$2/(\text{recall}^{-1} + \text{precision}^{-1})$], which
178 provides a balanced metric of Type I and Type II errors and is insensitive to true negative rates (Sofaer,
179 Hoeting, & Jarnevich, 2019).

180 We explored the relationship between methods, regimes, and F-measures using a Bayesian generalized
181 linear mixed model with the *brms* package (Bürkner, 2017). We specified a model set for each species, with
182 F-measure as our response variable following a beta distribution (as it is bound between 0 and 1), with
183 individual as a random effect to account for individual variation and a varying slope for the effect of method.
184 We excluded regime 8 (four locations a day, relocations only) as this sampling regime was not systematic.
185 We ran models with six Markov Chain Monte Carlo (MCMC) chains, each with 6,000 iterations (1,000
186 burn-in iterations, thin = 1), and we set Δ to 0.99. We fitted each model with half-Cauchy weakly
187 informative priors (Lemoine, 2019). We checked model convergence by inspecting trace plots and \hat{R} values
188 (Bürkner, 2017), assessed model fit visually via posterior predictive diagnostic plots, and evaluated model
189 performance using leave-one-out cross-validation (Vehtari *et al.*, 2017) and Bayesian R^2 . For further details
190 on model selection and validation, see Appendix S2, Supporting Information.

191 We compared the special case of regime 8 (similar to regime 2 but only relocation points) to the original
192 regime 2 in its own Bayesian model set; this allowed us to evaluate the impact of removing stationary
193 locations as a method of reducing data autocorrelation. Additionally, for this special case we only compared
194 the best performing KDE bandwidth (LSCV) and dBBMMs.

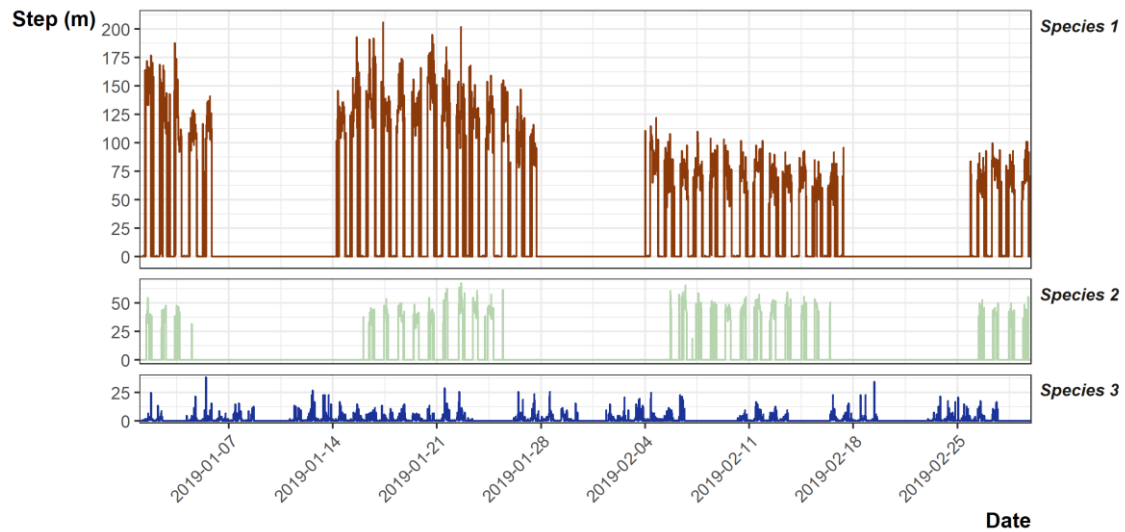
195 All datasets and R code to reproduce analyses is available at Zenodo repository platform
196 (DOI:10.5281/zenodo.3660796). We wrote code for R (v.3.5.2, R Core Team), using R studio (v.1.2.1335,
197 R Studio Team).

198 3. Results

199 3.1. SIMULATED ANIMAL MOVEMENT AND TRACKING DATA

200 The complete dataset for each simulated individual consisted of $n = 17,521$ data points for a full year, with
201 30-minute time steps (regime 1). Each regime progressively lowered the available data ($n^{\text{reg } 2} = 1,460$ data
202 points, $n^{\text{reg } 3} = 730$, $n^{\text{reg } 4} = 365$, $n^{\text{reg } 5} = 104$, $n^{\text{reg } 6} = 52$, $n^{\text{reg } 7} = 12$), while regime 8 varied for each species
203 and individual due to the variability in sheltering and resting behaviour ($n^{\text{species } 1} = 5,189 \pm 204$ data points
204 (mean \pm SD); $n^{\text{species } 2} = 3,501 \pm 1,099$; $n^{\text{species } 3} = 3,873 \pm 573$). Visual validation of movement patterns
205 matched with reported patterns in the literature (*e.g.* Parent & Weatherhead, 2000; Reed & Douglas, 2002;

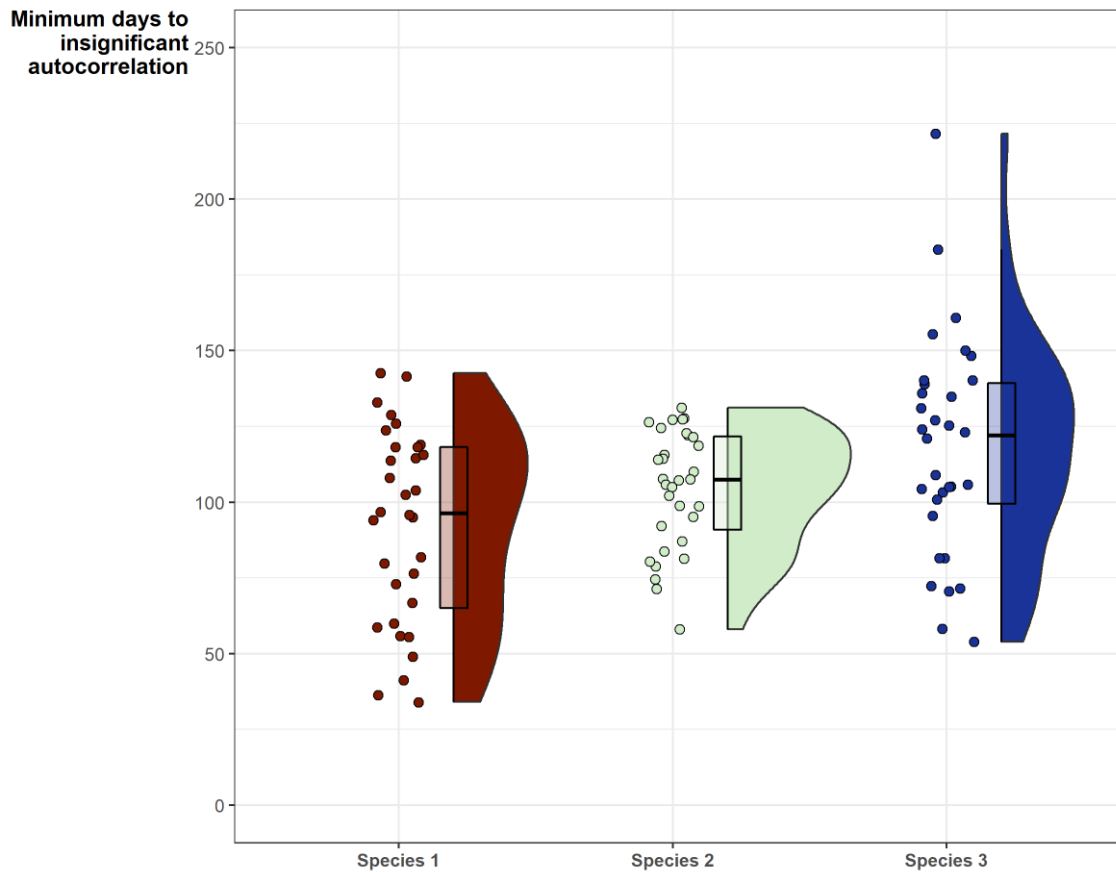
206 Wasko & Sasa, 2009; Hart *et al.*, 2015; Smith *et al.*, 2018; Silva *et al.*, 2018; Marshall *et al.*, 2019), and the
207 predicted patterns of the three archetypes (Figure 2).



208

209 **Figure 2.** Example two-month period illustrating how step distance (m) and its frequency differs between
210 our three species archetypes.

211 As expected, all simulated species and individual datasets showed strong autocorrelated structure. Time
212 until insignificant autocorrelation far exceeded even the coarsest tracking regime tested (regime 7, *i.e.*
213 1/month), indicating that all tracking regimes breach the assumption of independence required for KDE
214 methods (Figure 3).

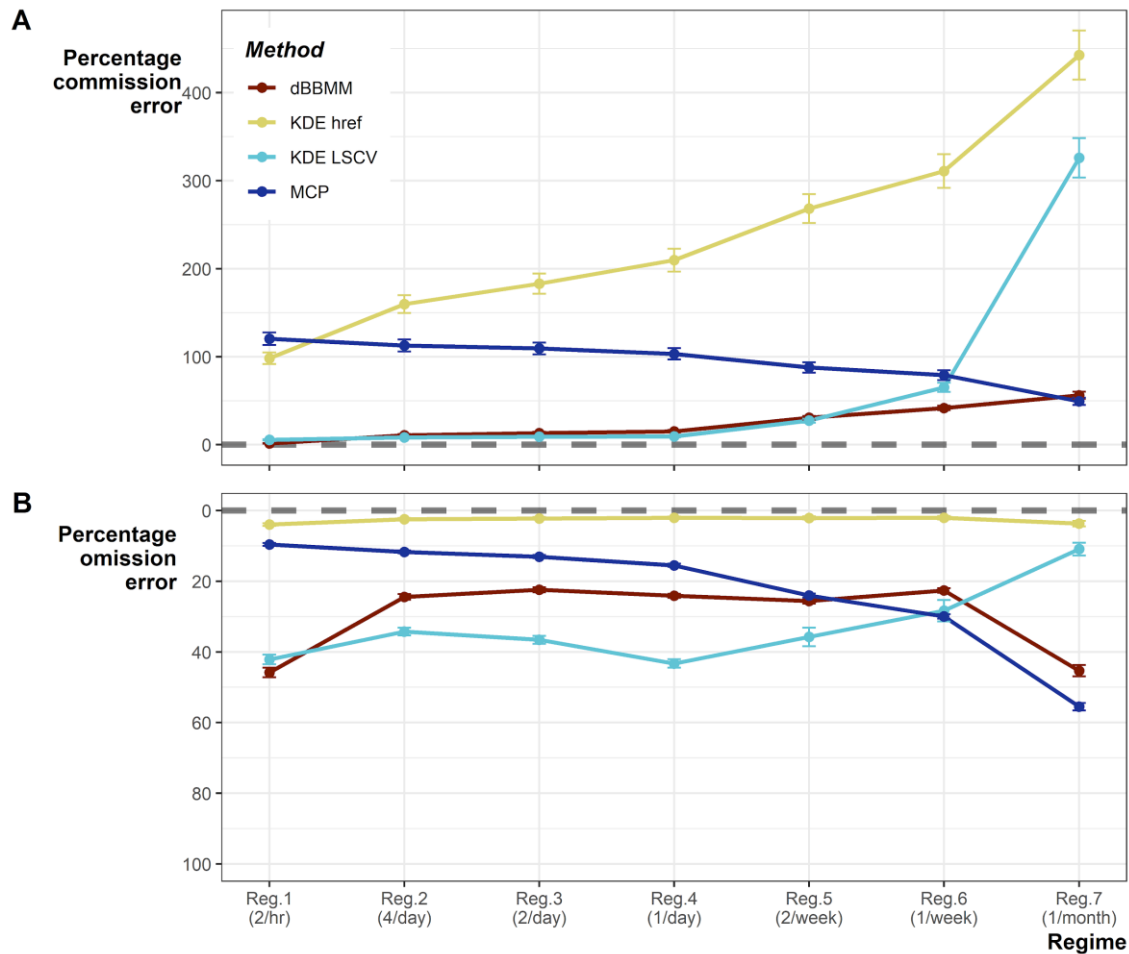


215

216 **Figure 3.** Minimum number of sampling days until the autocorrelation becomes insignificant and data
217 points can be considered independent.

218 3.2. METHOD COMPARISON: OMISSION VS. COMMISSION

219 Overall, coarser tracking regimes lead to greater % error when compared to true home ranges. However,
220 the balance between omission and commission is inconsistent and varies between home range estimation
221 methods (Figure 4). There is also a general trend towards commission error when estimating home ranges
222 because omission error is bounded between 0 and 100%.



223

224 **Figure 4.** Percentage error from the true home range using 95% contours. **A)** Commission error represented
 225 by positive values **B)** omission error represented by negative values. Error bars represent standard error of
 226 means across species (3) and individuals (96). Note the different scales for error, as omission error cannot
 227 exceed 100% of the true home range area.

228 3.2.1. Minimum convex polygon

229 Minimum convex polygons were the only method that showed a constant offset between omission and
 230 commission, as one increases the other decreases nearly 1:1. In addition, MCPs were the only method that
 231 decreased their commission error as tracking regime became temporally coarser. At frequent tracking
 232 regimes, MCPs only introduced minimal omission error, but their starkest failure is in their simple shape
 233 leading to the greatest commission error at highest resolution tracking regime (Figure 4, 5).

234 3.2.2. Fixed kernel home range

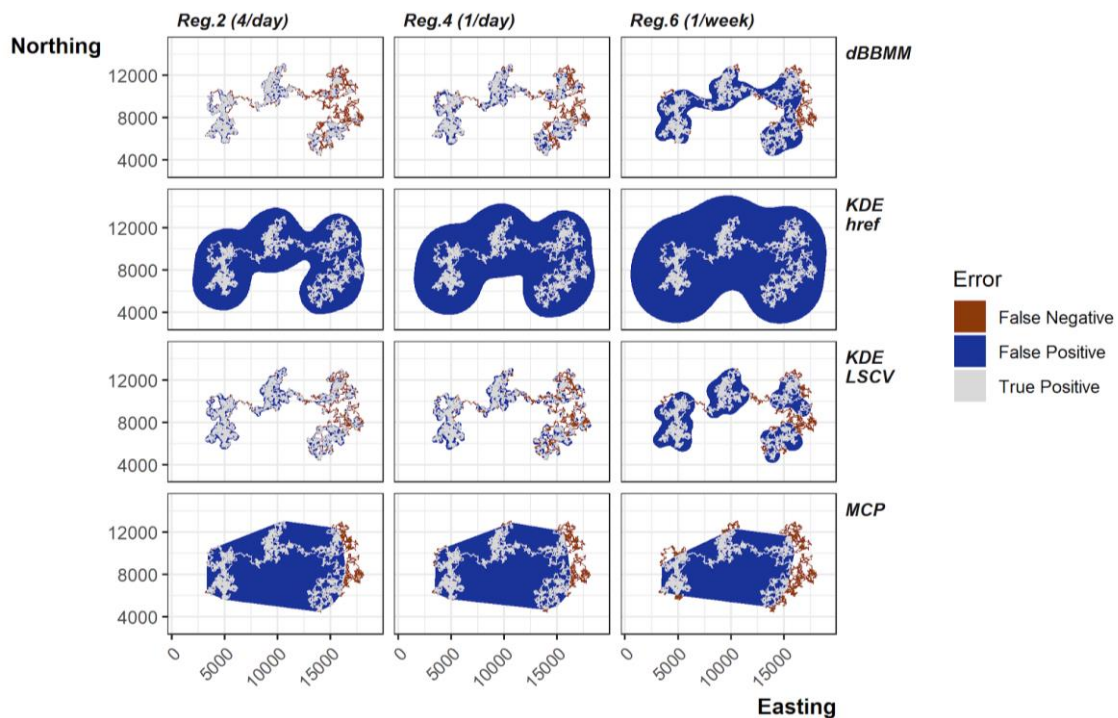
235 The fixed kernel home range using h_{ref} smoothing factor was by far the worst estimator for commission
 236 error. At low resolution tracking regimes, the >400% overestimation leads to near complete loss of home
 237 range edge fidelity (Figure 4). Due to this heavy emphasis on generous home range estimation KDE h_{ref}
 238 produced negligible omission error (Figure 5).

239 By comparison KDE LSCV produced consistently lower commission error at higher resolution tracking
240 regimes, but once the regime was once a week or coarser LSCV commission error spikes (up to 300%
241 overestimation). LSCV consistently performed worse in terms of omission error when applied to tracking
242 regimes with multiple tracks per day. Additionally, the LSCV algorithm frequently failed to converge
243 (68.5% of all LSCV home ranges failed). Only regime 7 converged consistently; the inclusion of more data
244 exacerbated convergence failure (regime 1-4, 100%; regime 5, 43.8%; regime 6, 33.3%). Using only
245 relocations reduced convergence failures (regime 8, 2.08%) compared to its closest parallel regime (regime
246 2,100%).

247 For both KDE methods, omission and commission error variability (displayed as SE on Figure 4) increased
248 as tracking regime became coarser.

249 3.2.3. *Dynamic Brownian bridge movement model*

250 Overall dBBMMs performed best. The method produced low commission error levels, matching KDE
251 LSCV performance (Figure 4). Unlike LSCV, dBBMMs commission error remained more stable and lower
252 when applied to coarser tracking regimes. Only MCPs produced a comparative level of commission error
253 at the coarsest tracking regimes, but dBBMMs kept some semblance of shape fidelity and connectivity
254 (Figure 5). Unlike other methods, dBBMM error remained low and balanced between omission and
255 commission, never exceeding 75%.

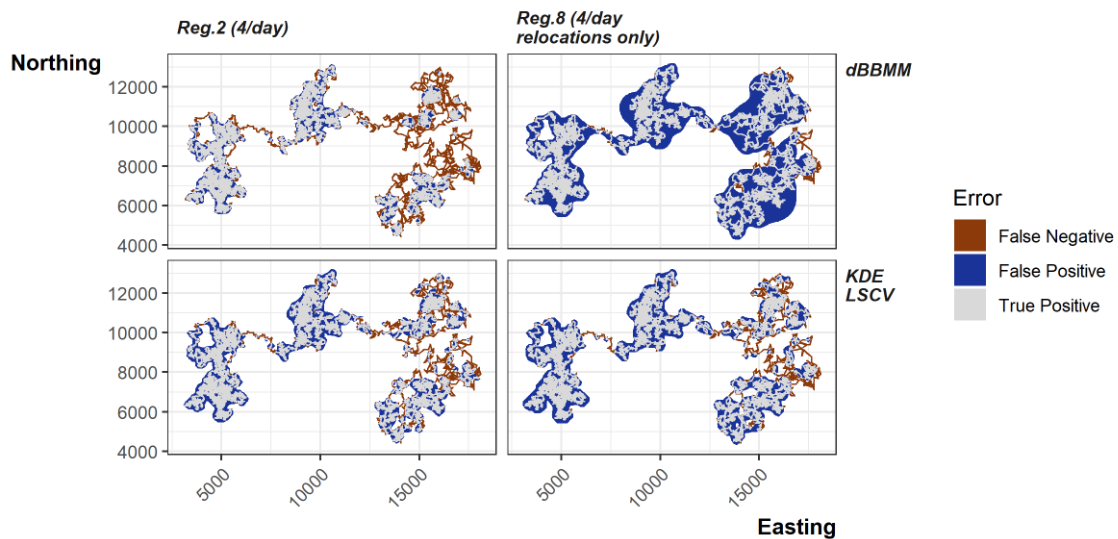


256

257 **Figure 5.** An example of how method and regime can interact to produce different levels of false negative,
258 false positive against the true home range. All contours shown are produced from the 95% utilization
259 distribution.

260 3.2.4. *Special case of regime 8*

261 Tracking regime 8 (four locations per day, relocations only) cannot be directly compared to the other
262 regimes as the structure of the tracking is different. A fairer comparison is between regime 8 and 2 (four
263 locations per day). Similar to all other regimes, regime 8 fails to remove autocorrelation to insignificance
264 (Figure 3); however, it did improve the performance of KDE LSCV estimation despite still breaching the
265 fundamental independence assumption (Figure 5, 6). The removal of repeated stationary points prevented
266 the LSCV smoothing from grouping too tightly to point concentrations (i.e. long-term shelter sites),
267 ultimately countering the tendency towards omission error for LSCV. However, on average, dBMMs
268 performed very similarly and balanced the omission and commission well (Figure 4). The dBMMs had
269 the added advantage of assuming serial dependence of points and, unlike LSCV, perform well when
270 provided high quantities of data.



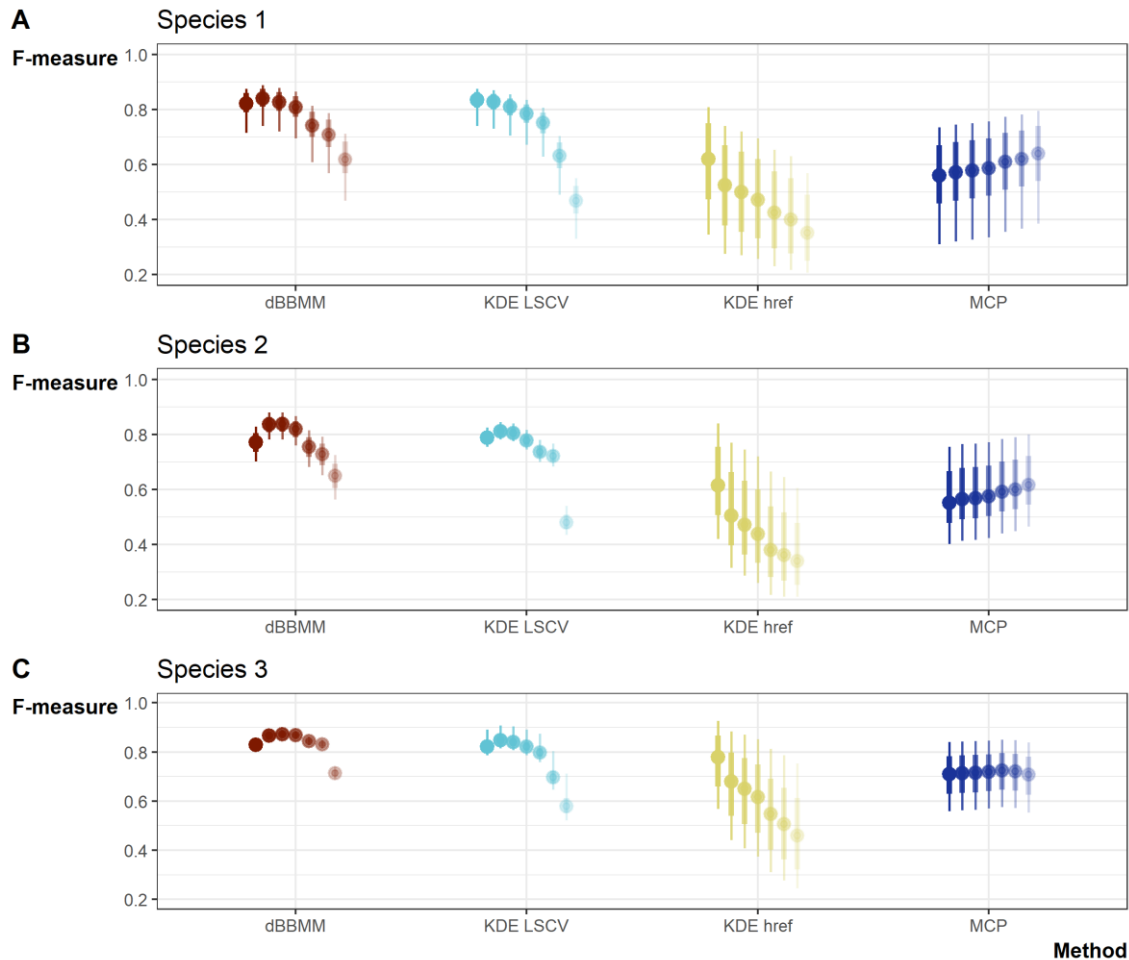
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272 **Figure 6.** Comparison between the error rates produced by the KDE LSCV and dBMM 95% contour
273 ranges when using data from sampling regime 2 (four locations per day) and regime 8 (four locations per
274 day, relocations only).

275 3.3. METHOD COMPARISON: F-MEASURES

276 The Bayesian models converged and performed well for all three species, with \hat{R} values ≈ 1.00 (Appendix
277 S2, Supporting Information), and R^2 values indicating considerable predictive power (Species 1: Bayesian
278 $R^2 = 0.960$, 95% CrI: 0.958–0.962; Species 2: Bayesian $R^2 = 0.946$, CrI: 0.755–0.786; Species 3: Bayesian
279 $R^2 = 0.905$, CrI: 0.897–0.911). Overall, our best models showed an interaction effect of methods and
280 regimes on F-measures; all species had a non-zero positive relationship between F-measures and regimes,

281 with higher estimates for dBBMM and KDE LSCV, while both MCP and KDE h_{ref} showed considerably
282 worse F-measures. However, Species 1 home range estimations were associated with lower F-measures,
283 suggesting that the home ranges of species with high movement and long periods of sheltering are harder
284 to model than those with more stable movement patterns.



285

286 **Figure 7.** Model results that aimed to predict F-measures using method, regime, and individual ID by
287 species. Tracking regime 1-7 are shown left to right with lowering levels of opacity. Fitted draws were
288 taken only from the first 5000 samples.

289 4. Discussion

290 Many published terrestrial reptile spatial ecology papers reuse the same two methods: Minimum Convex
291 Polygon (MCP), and Kernel Density Estimation (KDE), or variants. Both MCPs and KDEs produced high
292 error rates and failed to properly reflect simulated reptile home ranges. While originally intended for GPS
293 telemetry, we found that dBBMMs perform well across a range of lower fix rates sampling regimes, and
294 for our three archetypical reptile species.

295 4.1. CHOICE OF FIX FREQUENCY AND ESTIMATOR IMPACTS ESTIMATIONS

296 The data resampling throughout different tracking regimes led to a 91.7–99.9% data loss from our starting
297 point at 30-minute time steps: removing non-relocations (regime 8) still reduced data points by 70.4–80.0%.
298 Seamen *et al.*, (1999) suggested a minimum of 30-50 locations and both regimes 6 ($n = 52$) and 7 ($n = 12$)
299 failed to meet this criteria. A more stringent criteria (Girard *et al.*, 2002) recommending 300 locations also
300 excludes regime 5 ($n = 104$). Based on this fact alone, many reptile studies likely fail to meet KDE
301 requirements.

302 The use of MCP and KDE h_{ref} produced large false positive errors, which if carried forward are liable to
303 impact habitat and space-use inferences (Fieberg, 2007; Nilsen *et al.*, 2008). By comparison, both KDE
304 LSCV and dBBMM estimations fared better, although LSCV failed to produce F-measures comparable to
305 dBBMMs under low-resolution tracking regimes. Thus, dBBMMs can improve upon both traditional MCP
306 and fixed KDE methods. As a fix-frequency independent method (Kranstauber *et al.*, 2012), dBBMMs
307 performed most consistently across sampling regimes with the lowest error rates, even in low-resolution
308 datasets. To match dBBMM performance at the sparsest regimes ($n = 12$) KDEs required four times the
309 data. Maximizing performance under low-resolution regimes is critical for VHF studies because the data
310 are time, effort, and cost intensive (Recio *et al.*, 2011).

311 Furthermore, dBBMMs require no *a priori* knowledge of an animal's movements (necessary to identify the
312 correct smoothing bandwidth for KDEs), and can be put to use with current telemetry practices or to re-
313 analyse previously collected VHF data. The dBBMM method is easily compatible with low-resolution data
314 from herpetofauna spatial ecology studies still reliant on VHF. As gains from long-term high-resolution
315 tracking methods (GPS) still remain elusive for herpetofauna (Price-Rees, Brown, & Shine, 2013; Wolfe *et al.*
316 *et al.*, 2018), improving analytic methods represents a cheap, immediate alternative.

317 At high resolutions the KDE LSCV came closest to performing comparably with dBBMMs despite critical
318 flaws beyond failing the initial point independence assumption. Under higher resolution tracking regimes,
319 the LSCV algorithm fails to converge making the smoothing parameter estimate unusable (supporting
320 findings from Hemson *et al.*, 2015). High site fidelity in reptiles leads to unstable KDE LSCV because
321 non-convergence issues are compounded by large numbers of identical locations or very tight clusters (*i.e.*
322 site fidelity). We did not simulate any site fidelity which could inflate LSCV performance. Hemson *et al.*,
323 (2015) suggest ignoring site fidelity in simulation studies leads to inappropriate conclusions advocating
324 KDE LSCV (*e.g.* Worton, 1995; Seaman & Powell, 1996; Seaman *et al.*, 1999). Even with optimal
325 conditions for LSCV, dBBMMs performed similarly or better.

326 Removing non-relocations (regime 8) improved KDE LSCV while hindering dBBMMs. However, this fix
327 compromises the biological relevance of home range estimates (see De Solla, Bonduriansky, & Brooks,
328 1999) as the autocorrelated nature of animal movement is inherently biologically relevant (Cushman, Chase
329 & Griffin, 2005). The loss of stationary data points harms inferences drawn upon species that shelter for
330 long periods. Explorations using real GPS data show consistent problems with KDE LSCV omission error,
331 leading to severe undersmoothing, and frequent convergence failures (Hemson *et al.*, 2005). Jones, Marron,

332 & Sheather (1996) found that LSCV smoothed utilization distributions had unacceptable variability, that
333 can undermine comparisons between individuals, populations or studies.

334 Archetypal species movement characteristics influenced our range estimates (MCP, KDE and dBBMM).
335 The active hunter (Species 1), with its sporadic long-distance moves, had lower F-measures and higher
336 error rates than the ambush predators (Species 2 and 3). When comparisons between species are required,
337 researchers should explore how regime and estimation method effect comparisons. Ideally, researchers
338 should be able to access original data from previous studies. We encourage greater use of open data
339 repositories in reptile studies (*e.g.* Movebank). To date, reptile data on Movebank is lacking (11 species,
340 10 testudines and 1 serpentes). Without readily available data, researchers cannot confidently compare
341 between species.

342 4.2. CAVEATS

343 Herpetofauna and VHF tracking studies can be plagued with uncertainty due to inhospitable terrain and
344 associated costs. Failures to detect animals during tracking are inevitable, and we did not assess how the
345 frequency of missed or inconsistent tracks affects each method. Our results indicate that non-symmetrical
346 tracking regimes (*e.g.* tracks performed on Tuesdays and Thursdays) still appear to work well with
347 dBBMMs. Ultimately, accuracy of home range estimation will be dependent on resources, tracking
348 frequency and study duration (Mitchell, White, & Arnold, 2019). All directly impact the viability of
349 answering research questions. A clearly defined research question (Fieberg & Börger, 2012) enables
350 researchers to identify potential trade-offs in context.

351 While dBBMMs provide a more direct modelling approach for movements –a critical component of
352 assessing habitat use (Van Moorter, Rolandsen, Basille, & Gaillard, 2016)– there is scope for more
353 advanced methods when more is known about a species' movement patterns. dBBMMs provide an instant
354 option for estimating movement pathways of herpetofauna because they require no *a priori* knowledge. In
355 cases where more data are available, researchers can look at methods that integrate more about the
356 landscape, such as dBBMM with covariates (Kranstauber, 2019), or behaviour (Michelot & Blackwell,
357 2019). The more advanced methods may require data at higher resolution than is feasibly collectable by
358 VHF.

359 4.3. RECOMMENDATIONS AND CONCLUSIONS

360 Researchers must consider tracking regime during study design. There are practical considerations of cost,
361 time and ethics, but they must be paired with how the tracking regime will directly impact estimations and,
362 ultimately, the ability to answer research questions. There will always be spatial uncertainty. Tracking
363 regime should minimize spatial uncertainty with reference to the research question and targeted behaviours
364 (Fleming *et al.*, 2014; Schlägel & Lewis, 2016; Bastille-Rousseau *et al.*, 2017). Direct consideration of how
365 biology and movement impact home range will improve inferences drawn from telemetry studies.

366 The insights into reptile ecology can be invaluable despite data collection costs, and data utility should be
367 maximized. Better home range estimators are an inexpensive way of optimizing returns from tracking data
368 compared to technological advances or increasing field work. Reptile movement is peculiar: we revealed
369 the impact of long-term sheltering (essentially a zero-inflated movement dataset) on home range
370 estimations, which introduced error by under- and over-smoothing with traditional estimators. Inferences
371 based on traditional estimators have likely led to biases in reptile studies. Carrying these biases forward can
372 lead to misallocation of resources.

373 Our study concurs with previous studies *e.g.* Signer *et al.* (2015) stating problems with both MCP and
374 KDEs. Despite known problems researchers continue to justify use of MCPs and KDEs to maintain
375 comparability with previous studies. We find this deeply flawed in cases where tracking regime or estimator
376 differ which produce dramatically different error rates. However, we also demonstrate the stability of
377 dBBMMs and their suitability for comparisons. The information provided here can help optimise reptile
378 spatial ecology by yielding more accurate and reproducible home range estimations.

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383 References

- 384 Allen, M., Poggiali, D., Whitaker, K., Marshall, T. R., & Kievit, R. A. (2019). Raincloud plots: a multi-
385 platform tool for robust data visualization. *Wellcome Open Research*, 4, 63.
386 doi:10.12688/wellcomeopenres.15191.1
- 387 Anderson, D. J. (1982). The Home Range: A New Nonparametric Estimation Technique. *Ecology*, 63(1),
388 103–112. doi:10.2307/1937036
- 389 Bastille-Rousseau, G., Murray, D. L., Schaefer, J. A., Lewis, M. A., Mahoney, S. P., & Potts, J. R. (2017).
390 Spatial scales of habitat selection decisions: Implications for telemetry-based movement
391 modelling. *Ecography*, 40, 1–7. doi:10.1111/ecog.02655
- 392 Börger, L., Franconi, N., Michele, G. D., Gantz, A., Meschi, F., Manica, A., ... Coulson, T. (2006). Effects
393 of sampling regime on the mean and variance of home range size estimates. *Journal of Animal
394 Ecology*, 75(6), 1393–1405. doi:10.1111/j.1365-2656.2006.01164.x
- 395 Bruton, M. J., McAlpine, C. A., Smith, A. G., & Franklin, C. E. (2014). The importance of
396 underground shelter resources for reptiles in dryland landscapes: a woma python case study.

- 397 *Austral ecology*, 39(7), 819-829.
- 398 Buechley, E. R., McGrady, M. J., Çoban, E., & Şekercioğlu, Ç. H. (2018). Satellite tracking a
399 wide-ranging endangered vulture species to target conservation actions in the Middle East and
400 East Africa. *Biodiversity and Conservation*, 27(9), 2293-2310.
- 401 Bürkner, P. C. (2017). brms: An R package for Bayesian multilevel models using Stan. *Journal of Statistical*
402 *Software*, 80(1), 1-28.
- 403 Burt, W. H. (1943). Territoriality and home range concepts as applied to mammals. *Journal of*
404 *mammalogy*, 24(3), 346-352.
- 405 Byrne, M. E., McCoy, J.C., Hinton, J. W., Chamberlain, M. J., & Collier, B. A. (2014). Using
406 dynamic Brownian bridge movement modelling to measure temporal patterns of habitat selection.
407 *Journal of Animal Ecology*, 83(5), 1234-1243.
- 408 Calabrese, J. M., Fleming, C. H., & Gurarie, E. (2016). Ctm: an R Package for Analyzing Animal
409 Relocation Data As a Continuous-Time Stochastic Process. *Methods in Ecology and Evolution*,
410 7(9), 1124–1132. doi:10.1111/2041-210X.12559
- 411 Campbell, H. A., Dwyer, R. G., Irwin, T. R., & Franklin, C. E. (2013). Home range utilisation and long-
412 range movement of estuarine crocodiles during the breeding and nesting season. *PLoS One*, 8(5),
413 e62127.
- 414 Cohen, B. S., Prebyl, T. J., Collier, B. A., & Chamberlain, M. J. (2018). Home range estimator method and
415 GPS sampling schedule affect habitat selection inferences for wild turkeys. *Wildlife Society*
416 *Bulletin*, 42(1), 150-159.
- 417 Cushman, S.A., Chase, M. & Griffin, C. (2005). Elephants in space and time. *Oikos*, 109, 331–341.
- 418 De Solla, S. R., Bonduriansky, R., & Brooks, R. J. (1999). Eliminating autocorrelation reduces biological
419 relevance of home range estimates. *Journal of Animal Ecology*, 68(2), 221-234.
- 420 Dickson, B. G., & Beier, P. (2002). Home-range and habitat selection by adult cougars in southern
421 California. *The Journal of Wildlife Management*, 1235-1245.
- 422 Fieberg, J. (2007). Kernel density estimators of home range: Smoothing and the autocorrelation red herring.
423 *Ecology*, 88(4), 1059–1066. doi:10.1890/06-0930
- 424 Fieberg, J., & Börger, L. (2012). Could you please phrase “home range” as a question?. *Journal of*
425 *Mammalogy*, 93(4), 890-902.

- 426 Fisher, D. O. (2000). Effects of vegetation structure, food and shelter on the home range and habitat use of
427 an endangered wallaby. *Journal of Applied Ecology*, 37(4), 660-671.
- 428 Fleming, C. H., Calabrese, J. M., Dong, X., Winner, K., Péron, G., Kranstauber, B., ... Mueller, T. (2017).
429 Package 'ctmm'. doi:10.1086/675504>
- 430 Gibbons, J. W., Scott, D. E., Ryan, T. J., Buhlmann, K. A., Tuberville, T. D., Metts, B. S., ... Winne, C. T.
431 (2000). The Global Decline of Reptiles, Déjà Vu Amphibians. *BioScience*, 50(8), 653–666.
432 doi:10.1641/0006-3568(2000)050[0653:TGDORD]2.0.CO;2
- 433 Girard, I., Ouellet, J. P., Courtois, R., Dussault, C., & Breton, L. (2002). Effects of sampling effort based
434 on GPS telemetry on home-range size estimations. *The Journal of wildlife management*, 1290-
435 1300.
- 436 Guarino, F. (2002). Spatial ecology of a large carnivorous lizard, *Varanus varius* (Squamata: Varanidae).
437 *Journal of Zoology*, 258(4), 449–457. doi:10.1017/S0952836902001607
- 438 Gurarie, E., Bracis, C., Delgado, M., Meckley, T. D., Kojola, I., & Wagner, C. M. (2016). What is the
439 animal doing? Tools for exploring behavioural structure in animal movements. *Journal of Animal*
440 *Ecology*, 85(1), 69-84.
- 441 Hailey, A. (1989). How far do animals move? Routine movements in a tortoise. *Canadian Journal of*
442 *Zoology*, 67(1), 208-215.
- 443 Hart, K. M., Cherkiss, M. S., Smith, B. J., Mazzotti, F. J., Fujisaki, I., Snow, R. W., & Dorcas, M. E. (2015).
444 Home range, habitat use, and movement patterns of non-native Burmese pythons in Everglades
445 National Park, Florida, USA. *Animal Biotelemetry*, 3(8), 1–13. doi:10.1186/s40317-015-0022-2
- 446 Hebblewhite, M., & Haydon, D. T. (2010). Distinguishing technology from biology: a critical review of
447 the use of GPS telemetry data in ecology. *Philosophical Transactions of the Royal Society B:*
448 *Biological Sciences*, 365(1550), 2303-2312.
- 449 Hemson, G., Johnson, P., South, A., Kenward, R., Ripley, R., & McDonald, D. (2005). Are kernels the
450 mustard? Data from global positioning system (GPS) collars suggests problems for kernel home-
451 range analyses with least-squares cross-validation. *Journal of Animal Ecology*, 74(3), 455–463.
452 doi:10.1111/j.1365-2656.2005.00944.x
- 453 Jennrich, R. I., & Turner, F. B. (1969). Measurement of non-circular home range. *Journal of Theoretical*
454 *Biology*, 22(2), 227–237. doi:10.1016/0022-5193(69)90002-2

- 455 Jones, M. C., Marron, J. S., & Sheather, S. J. (1996). A Brief Survey of Bandwidth Selection for Density
456 Estimation. *Journal of the American Statistical Association*, 91(433), 401–407.
457 doi:10.1080/01621459.1996.10476701
- 458 Katajisto, J., & Moilanen, A. (2006). Kernel-based home range method for data with irregular sampling
459 intervals. *Ecological Modelling*, 194(4), 405-413.
- 460 Kie, J.G., Matthiopoulos, J., Fieberg, J., Powell, R.A., Cagnacci, F., Mitchell, M.S., Gaillard, J.M. and
461 Moorcroft, P.R. (2010). The home-range concept: are traditional estimators still relevant with
462 modern telemetry technology?. *Philosophical Transactions of the Royal Society B: Biological*
463 *Sciences*, 365(1550), 2221-2231.
- 464 Knight, C. M., Kenward, R. E., Gozlan, R. E., Hodder, K. H., Walls, S. S., & Lucas, M. C. (2009). Home-
465 range estimation within complex restricted environments: importance of method selection in
466 detecting seasonal change. *Wildlife Research*, 36(3), 213-224.
- 467 Kranstauber, B. (2019). Modelling animal movement as Brownian bridges with covariates. *Movement*
468 *Ecology*, 7(1), 22. doi:10.1186/s40462-019-0167-3
- 469 Kranstauber, B., Kays, R., Lapoint, S. D., Wikelski, M., & Safi, K. (2012). A dynamic Brownian bridge
470 movement model to estimate utilization distributions for heterogeneous animal movement. *Journal*
471 *of Animal Ecology*, 81(4), 738–746. doi:10.1111/j.1365-2656.2012.01955.x
- 472 Lai, S., Bêty, J., & Berteaux, D. (2015). Spatio–temporal hotspots of satellite–tracked arctic foxes
473 reveal a large detection range in a mammalian predator. *Movement ecology*, 3(1), 37.
- 474 Langrock, R., Hopcraft, J. G. C., Blackwell, P. G., Goodall, V., King, R., Niu, M., ... & Schick, R. S. (2014).
475 Modelling group dynamic animal movement. *Methods in Ecology and Evolution*, 5(2), 190-199.
- 476 Laver, P. N., & Kelly, M. J. (2008). A critical review of home range studies. *The Journal of Wildlife*
477 *Management*, 72(1), 290-298.
- 478 Lemoine, N. P. (2019). Moving beyond noninformative priors: why and how to choose weakly informative
479 priors in Bayesian analyses. *Oikos*, oik.05985. doi:10.1111/oik.05985
- 480 Marshall, B. M., Strine, C. T., Jones, M. D., Artchawakom, T., Silva, I., Suwanwaree, P., & Goode, M.
481 (2019). Space fit for a king: spatial ecology of king cobras (*Ophiophagus hannah*) in Sakaerat
482 Biosphere Reserve, Northeastern Thailand. *Amphibia-Reptilia*, 40(2), 163–178.
483 doi:10.1163/15685381-18000008

- 484 Mata-Silva, V., DeSantis, D. L., Wagler, A. E., & Johnson, J. D. (2018). Spatial Ecology of Rock
485 Rattlesnakes (*Crotalus lepidus*) in Far West Texas. *Herpetologica*, 74(3), 245-254.
- 486 McClintock, B. T., & Michelot, T. (2018). momentuHMM: R package for generalized hidden Markov
487 models of animal movement. *Methods in Ecology and Evolution*, 9(6), 1518-1530.
- 488 Michelot, T., & Blackwell, P. G. (2019). State-switching continuous-time correlated random walks.
489 *Methods in Ecology and Evolution*, 10(5), 637-649.
- 490 Mitchell, L.J., White, P.C. and Arnold, K.E., 2019. The trade-off between fix rate and tracking duration on
491 estimates of home range size and habitat selection for small vertebrates. *PloS one*, 14(7).
- 492 Nathan, R., Getz, W. M., Revilla, E., Holyoak, M., Kadmon, R., Saltz, D., & Smouse, P. E. (2008). A
493 movement ecology paradigm for unifying organismal movement research. *Proceedings of the*
494 *National Academy of Sciences*, 105(49), 19052-19059.
- 495 Nilsen, E. B., Pedersen, S., & Linnell, J. D. (2008). Can minimum convex polygon home ranges be used to
496 draw biologically meaningful conclusions?. *Ecological Research*, 23(3), 635-639.
- 497 Noonan, M. J., Tucker, M. A., Fleming, C. H., Akre, T., Alberts, S. C., Ali, A. H., ... Calabrese, J. M.
498 (2018). A comprehensive analysis of autocorrelation and bias in home range estimation.
499 *Ecological Monographs*. doi:10.1002/ecm.1344
- 500 Palm, E. C., Newman, S. H., Prosser, D. J., Xiao, X., Ze, L., Batbayar, N., ... & Takekawa, J. Y.
501 (2015). Mapping migratory flyways in Asia using dynamic Brownian bridge movement models.
502 *Movement ecology*, 3(1), 3.
- 503 Parent, C., & Weatherhead, P. J. (2000). Behavioral and life history responses of eastern massasauga
504 rattlesnakes (*Sistrurus catenatus catenatus*) to human disturbance. *Oecologia*, 125(2), 170–178.
505 <https://doi.org/10.1007/s004420000442>
- 506 Péron, G. (2019). The time frame of home-range studies: from function to utilization. *Biological*
507 *Reviews*, 94(6), 1974-1982.
- 508 Petersen, C. E., Goetz, S. M., Dreslik, M. J., Kleopfer, J. D., & Savitzky, A. H. (2019). Sex, Mass, and
509 Monitoring Effort: Keys to Understanding Spatial Ecology of Timber Rattlesnakes (*Crotalus*
510 *horridus*). *Herpetologica*, 75(2), 162–174. doi:10.1655/D-18-00035
- 511 Powell, R. A. (2012). Diverse perspectives on mammal home ranges or a home range is more than location
512 densities. *Journal of mammalogy*, 93(4), 887-889.

- 513 Powell, R. A., & Mitchell, M. S. (2012). What is a home range?. *Journal of mammalogy*, 93(4), 948-958.
- 514 Price-Rees, S. J., Brown, G. P., & Shine, R. (2013). Habitat selection by bluetongue lizards (Tiliqua,
515 Scincidae) in tropical Australia: a study using GPS telemetry. *Animal Biotelemetry*, 1(1), 7.
- 516 R Core Team. (2019). R: A language and environment for statistical computing. Vienna, Austria: R
517 Foundation for Statistical Computing. Retrieved from <https://www.r-project.org/>
- 518 R Studio Team. (2019). RStudio: Integrated Development Environment for R. Boston, MA: RStudio, Inc.
519 Retrieved from <http://www.rstudio.com/>
- 520 Recio, M. R., Mathieu, R., Maloney, R., & Seddon, P. J. (2011). Cost comparison between GPS-and VHF-
521 based telemetry: case study of feral cats *Felis catus* in New Zealand. *New Zealand Journal of*
522 *Ecology*, 35(1), 114.
- 523 Reed, R. N., & Douglas, M. E. (2002). Ecology of the Grand Canyon rattlesnake (*Crotalus viridis abyssus*)
524 in the Little Colorado River Canyon, Arizona. *Southwestern Naturalist*, 47(1), 30–39.
525 <https://doi.org/Doi.10.2307/3672799>
- 526 Riotte-Lambert, L., & Matthiopoulos, J. (2019). Environmental predictability as a cause and
527 consequence of animal movement. *Trends in Ecology & Evolution*, 35(2), 163-174.
- 528 Robertson, P. A., Aebischer, N. J., Kenwards, R. E., Hanski, I. K., & Williams, N. P. (1998). Simulation
529 and jack-knifing assessment of home-range indices based on underlying trajectories. *Journal of*
530 *Applied Ecology*, 35(6), 928–940. doi:10.1111/j.1365-2664.1998.tb00010.x
- 531 Roll, U., Feldman, A., Novosolov, M., Allison, A., Bauer, A. M., Bernard, R., ... Meiri, S. (2017). The
532 global distribution of tetrapods reveals a need for targeted reptile conservation. *Nature Ecology &*
533 *Evolution*, 1(11), 1677–1682. doi:10.1038/s41559-017-0332-2
- 534 Rosenblatt, A. E., Heithaus, M. R., Mazzotti, F. J., Cherkiss, M., & Jeffery, B. M. (2013). Intra-population
535 variation in activity ranges, diel patterns, movement rates, and habitat use of American alligators
536 in a subtropical estuary. *Estuarine, Coastal and Shelf Science*, 135, 182-190.
- 537 Row, J. R., & Blouin-Demers, G. (2006). Kernels are not accurate estimators of home-range size
538 for herpetofauna. *Copeia*, 2006(4), 797-802.
- 539 Schlägel, U. E., & Lewis, M. A. (2014). Detecting effects of spatial memory and dynamic information on
540 animal movement decisions. *Methods in Ecology and Evolution*, 5, 1236–1246.
541 doi:10.1111/2041-210X.12284

- 542 Schofield, G., Bishop, C. M., MacLean, G., Brown, P., Baker, M., Katselidis, K. A., ... & Hays, G. C.
543 (2007). Novel GPS tracking of sea turtles as a tool for conservation management. *Journal of*
544 *Experimental Marine Biology and Ecology*, 347(1-2), 58-68.
- 545 Sciaini, M., Fritsch, M., Scherer, C., & Simpkins, C. E. (2018). NLMR and landscapetools: An integrated
546 environment for simulating and modifying neutral landscape models in R. *Methods in Ecology*
547 *and Evolution*, 9(11), 2240-2248.
- 548 Seaman, D. E., & Powell, R. A. (1996). An Evaluation of the Accuracy of Kernel Density Estimators for
549 Home Range Analysis. *Ecology*, 77(7), 2075–2085.
- 550 Seaman, D. E., Millsbaugh, J. J., Kernohan, B. J., Brundige, G. C., Raedeke, K. J., & Gitzen, R. A. (1999).
551 Effects of sample size on kernel home range estimates. *The journal of wildlife management*, 739-
552 747.
- 553 Signer, J., Balkenhol, N., Ditmer, M., & Fieberg, J. (2015). Does estimator choice influence our ability to
554 detect changes in home-range size?. *Animal Biotelemetry*, 3(1), 16.
- 555 Silva, I., Crane, M., Suwanwaree, P., Strine, C., & Goode, M. (2018). Using dynamic Brownian Bridge
556 Movement Models to identify home range size and movement patterns in king cobras. *PLOS ONE*,
557 13(9), e0203449. doi:10.1371/journal.pone.0203449
- 558 Smith, B. J., Hart, K. M., Mazzotti, F. J., Basille, M., & Romagosa, C. M. (2018). Evaluating GPS
559 biologging technology for studying spatial ecology of large constricting snakes. *Animal*
560 *Biotelemetry*, 6(1), 1.
- 561 Sofaer, H. R., Hoeting, J. A., & Jarnevich, C. S. (2019). The area under the precision-recall curve as a
562 performance metric for rare binary events. *Methods in Ecology and Evolution*, 10(4), 565–577.
563 doi:10.1111/2041-210X.13140
- 564 Swihart, R. K., & Slade, N. A. (1985). Influence of sampling interval on estimates of home-range size. *The*
565 *Journal of Wildlife Management*, 1019-1025.
- 566 Tikkanen, H., Rytönen, S., Karlin, O. P., Ollila, T., Pakanen, V. M., Tuohimaa, H., & Orell, M. (2018).
567 Modelling golden eagle habitat selection and flight activity in their home ranges for safer wind
568 farm planning. *Environmental Impact Assessment Review*, 71, 120-131.
- 569 Van Moorter, B., Rolandsen, C. M., Basille, M., & Gaillard, J.-M. (2016). Movement is the glue connecting
570 home ranges and habitat selection. *Journal of Animal Ecology*, 85(1), 21–31. doi:10.1111/1365-

571 2656.12394

572 Vehtari, A., Gelman, A., & Gabry, J. (2017). Practical Bayesian model evaluation using leave-
573 one-out cross-validation and WAIC. *Statistics and computing*, 27(5), 1413-1432.

574 Walter, W. D., Onorato, D. P., & Fischer, J. W. (2015). Is there a single best estimator? Selection
575 of home range estimators using area-under-the-curve. *Movement ecology*, 3(1), 10.

576 Wasko, D. K., & Sasa, M. (2009). Activity patterns of a neotropical ambush predator: Spatial ecology of
577 the fer-de-lance (*Bothrops asper*, serpentes: Viperidae) in Costa Rica. *Biotropica*, 41(2), 241–249.
578 <https://doi.org/10.1111/j.1744-7429.2008.00464.x>

579 Wolfe, A. K., Fleming, P. A., & Bateman, P. W. (2018). Impacts of translocation on a large urban-adapted
580 venomous snake. *Wildlife Research*. <https://doi.org/10.1071/WR17166>

581 Worton, B. J. (1989). Kernel Methods for Estimating the Utilization in Home-Range Studies. *Ecology*,
582 70(1), 164–168.

583 Worton, Bruce J. (1995). Using Monte Carlo Simulation to Evaluate Kernel-Based Home Range
584 Estimators. *The Journal of Wildlife Management*, 59(4), 794–800. doi:10.2307/3801959

585