

1 **Can early exposure to stress enhance resilience to ocean warming**
2 **in two oyster species?**

3 **Roberta R. C. Pereira¹, Elliot Scanes¹, Mitchell Gibbs¹, Maria Byrne^{1,2}, Pauline M. Ross^{1,*}**

4

¹School of Life and Environmental Science, The University of Sydney, Camperdown, NSW,
Australia,

²School of Medical Sciences, The University of Sydney, Camperdown, NSW, Australia

***Corresponding author:** pauline.ross@sydney.edu.au

Running page head: Early exposure to stress and resilience of oysters

5

6 ABSTRACT

7 Securing economically and ecologically significant molluscs, as our oceans warm and acidify
8 due to climate change, is a global priority. South eastern Australia receives warm water in a
9 strengthening East Australia Current and so resident species are vulnerable to elevated
10 temperature and marine heat waves. This study tested whether oysters pre exposed to
11 elevated temperature or heat stress enhances resilience to ocean warming later in life. Two
12 Australian species, the flat oyster, *Ostrea angasi*, and the Sydney rock oyster, *Saccostrea*
13 *glomerata*, were given a mild dose of warm water or “heat shock” stress in the laboratory
14 and then transferred to elevated temperature conditions where we used the thermal outfall
15 from power generation as a proxy to investigate the impacts of ocean warming. Shell
16 growth, condition index, lipid content and profile and survival of oysters was impacted by
17 elevated temperature in the field, with flat oysters being more impacted than Sydney rock
18 oysters. Flat oysters grew faster than Sydney rock oysters at ambient temperature, but were
19 more sensitive to elevated temperature. Early exposure to heat stress did little to
20 ameliorate the negative effects of increased temperature, although the survival of heat
21 shocked flat oysters was greater than non-heat shocked oysters. Further investigations are
22 required to determine if early exposure to heat stress can act to inoculate oysters to future
23 stress and overall enhance resilience of oysters to ocean warming.

24 **1. Introduction**

25 Climate change, the result of anthropogenic activities such as the burning of fossil fuels and
26 deforestation, has exponentially increased the concentration of carbon dioxide (CO₂) and
27 other greenhouse gasses in the atmosphere [1]. Since the onset of the industrial revolution,
28 atmospheric partial pressure of CO₂ (pCO₂) has increased from 280 ppm to 410 ppm causing

29 global warming with direct impacts on the oceans [1,2]. As a result, the world's oceans have
30 warmed by 0.68°C and for the East Australian coast are predicted to increase by up to 4°C by
31 2050 and 6°C before 2100 [3,4]. Ocean warming and the increased incidence of heatwaves
32 (abnormal high temperatures over multiple days [5]) negatively impacts diverse species [6].
33 Between 1925 and 2016 there has been a 54% annual increase in the duration of marine
34 heatwaves worldwide [7]. Climate change is also impacting ocean stratification, currents,
35 salinity, pH, sea level and increasing the frequency of extreme events [1,7,8].

36 Increasing frequency of thermal stress events will have consequences for fitness and
37 survival of marine species and there is concern for habitat engineers such as bivalves and
38 oysters [6,9]. If oysters and other molluscs are to persist during this century along the
39 southeast coast of Australia and in similar “hot spots” around the globe, they will need to be
40 resilient to marine heat waves and habitat warming. It has been suggested that organisms
41 can build resilience to environmental stress, through exposure to stress in early life. Studies
42 have found that exposure to a mild stress early in life can result in later life stress resistance
43 [10,11]. Rather like a vaccination or inoculation, resistance to stress after exposure to mild
44 stress in early life has been observed in a diverse array of organisms such as bacteria, plants,
45 insects, mammals and fish [10, 11, 12, 13 ,14].

46 An increase in resistance to stress has been shown in the tidepool fish *Oligocottus*
47 *maculosus* which after exposure to a +12°C heat stress had greater survival rates when
48 exposed to subsequent stressful levels of high salinity and low oxygen concentration
49 compared with fish that did not experience the heat shock [10]. The magnitude of the shock
50 and recovery time played an important role in the stress response later in life [10]. Baltic
51 Sea mussels *Mytilus edulis*, exposed to heat shock (+16°C) and then exposed to cadmium (20

52 $\mu\text{g L}^{-1}$) produced heat shock proteins at a faster rate than mussels not exposed to heat stress
53 [15]. Stress resistance may be enabled by production of protective heat shock proteins (e.g.
54 HSP 70), although this is energetically costly. The mechanisms behind stress inoculation, are
55 complex and likely not limited to production of heat shock proteins. Other processes such as
56 alterations in metabolism and epigenetics are also thought to be involved [16, 17].

57 While mobile species can migrate changing their distribution as the ocean warms, sessile
58 species are vulnerable because they are unable to move and the dispersive larval stages are
59 often short-lived [18,19]. It is predicted that sessile organisms such as oysters, which form
60 the basis of aquaculture across the globe, will be impacted by elevated temperature,
61 because of the energetic cost to physiological performance from climate change stress [20,
62 21]. Already, significant mortality has been reported for the north American oyster
63 *Crassostrea virginica* exposed to elevated temperature, due to impacts on energetic
64 reserves [22]. Reduced gametogenesis in *M. galloprovincialis* has been directly connected to
65 warming [23]. Parental exposure to stress (in this case ocean acidification has, however,
66 been shown to increase resilience of larval oysters, and this trait was carried over to
67 adulthood [24,25].

68 The flat oyster, *Ostrea angasi* and the Sydney rock oyster *Saccostrea glomerata* are native
69 to south eastern Australia [26, 27], where they historically formed extensive reefs and are
70 the basis of a USD \$30 million aquaculture industry [28,29]. *Saccostrea glomerata* is an
71 intertidal species that occurs along the east and west coast of Australia with a current upper
72 sea surface temperature (SST) range of 24-26 °C [28]. *Ostrea angasi* is distributed in shallow
73 subtidal sheltered waterways along a similar range with a current upper SST temperature
74 range of 22-24°C [27], however, this northern (warm) range is likely curtailed by historic

75 overharvesting and introduced parasites in New South Wales (*Polydora* spp.)(27). *O. angasi*
76 are mostly found subtidally in comparatively stable thermal conditions [30]. These species
77 are both currently the focus of reef restoration efforts along the south eastern coastline of
78 Australia [31,32] and are known to be vulnerable to acidification [33,34,35] and warming
79 [33,36]. South-eastern Australia receives warmer waters from the Coral Sea via the East
80 Australian Current (EAC), which is strengthening [37,38]. This region is considered a “hot
81 spot”, as the rise in mean temperatures will be 3-4 times higher than the average for the
82 world’s oceans and is also prone to marine heat waves [7,38,39].

83 The purpose of this study was to test the hypothesis that early exposure to heat stress or
84 heat shock can be used as a mechanism to build resilience of *O. angasi* and *S. glomerata* to
85 subsequent long-term exposure to warmed seawater. We used the thermal outfall from a
86 power generating station as a proxy for ocean warming conditions as in previous studies
87 [40]. Due to their different thermal ranges, distributions and habitats we predicted that *S.*
88 *glomerata* will be more resilient than *O. angasi* to elevated temperature. As momentum
89 gains to restore oyster reefs [31], knowledge of oyster responses and how to build resilience
90 is needed to ensure sustainability of restoration efforts and the aquaculture industry.

91 **Methods**

92 *Ostrea angasi* and *Saccostrea glomerata* were obtained from an oyster farm at Merimbula
93 Lake (Merimbula Gourmet Oysters; 36°89' 85"S, 149°88' 46"E) and approximately 200
94 oysters per species were transported to Port Stephens Fisheries Institute (PSFI; 32°44'47"S,
95 152°03'30"E), following the protocol for oyster movement in New South Wales, Australia,
96 during the Austral autumn 2018. The initial mean shell height was $69.68 \pm \text{S.E. } 0.34$ mm for
97 *O. angasi* and $69.86 \pm \text{S.E. } 0.33$ mm for *S. glomerata*. After arrival at PSFI the oysters were

98 placed in 40L tubs with seawater supplied from a 750L tank at 20°C. This temperature was
99 the same as in Merimbula Lake when oysters were collected. Oysters were fed a mixture of
100 microalgae cultured on-site containing 50 % *Chaetoceros muelleri* and 50 % *Tisochrysis lutea*
101 at a concentration equivalent to 2×10^9 cells oyster⁻¹ d⁻¹ [41]. The initial mean (\pm S.E.)
102 condition index for *O. angasi* and *S. glomerata* were 4.12 ± 0.42 g and 4.30 ± 0.39 g (n=6),
103 respectively (see below for methods).

104 **2.1 Heat shock**

105 To determine if exposure to heat shock would confer subsequent resilience to long term
106 exposure to elevated temperature, the following heat shock protocol was used. The oysters
107 were divided into two sub-groups; one “control” and a “heat shocked” group per species
108 into 750L tanks. Heat shock was administered by exposure to an elevated temperature of 26
109 °C for 18 hours and then 28°C for 6 more hours by slowly ramping up the temperature using
110 aquarium heaters (Titan G2 1500 W). This was an initial +6°C (from 20° to 26°C) and a
111 further increase of +2°C (from 26°C to 28°C). There was no mortality following heat shock
112 treatment. Following the 24 hours at elevated temperature, the water was left to slowly
113 cool to ambient (20°C). Oysters were submerged in ambient water overnight in the
114 laboratory. On the following day, they were placed in baskets and left submerged at
115 ambient conditions, in the adjacent estuary of PSFI (Tilligerry creek, Port Stephens) which
116 remained at 20 °C for one week. After this period, they were removed and shell height was
117 measured with a digital calliper.

118 A total of 40 oysters were randomly placed in baskets (600 x 250 x100 mm) divided into four
119 compartments with 10 “control” *O. angasi* and 10 “control” *S. glomerata* which were
120 exposed to ~20°C at all times, 10 “heat shocked” *O. angasi* and 10 “heat shocked” *S.*

121 *glomerata*, which were exposed to elevated temperature for 24 hours. The baskets were
122 transported and deployed into Lake Macquarie (33°.07'94", 151°.54'85", Figure 1).

123 **Figure 1.** (A) Map of Australia with the study area in red (New South Wales, NSW). (B) Map
124 of Lake Macquarie, NSW showing the field locations where the baskets were deployed for
125 approximately seven months. Yellow squares represent the warm seawater outfall of two
126 power stations (Eraring and Vales power stations). Black triangles are the ambient (control)
127 locations and the red triangles are the elevated locations (total 5 baskets).

128 **2.2 Field location**

129 To determine the response of oysters in the real world of elevated temperature, we used
130 warmed water released into a saline coastal lake by two power stations at Lake Macquarie,
131 NSW. Lake Macquarie is a large coastal body of water in the centre of East Australian
132 warming "hot spot". Lake Macquarie is connected to the ocean and has daily tidal exchange.
133 There is little freshwater input from the surrounding catchment [42]. Two coal fired power
134 stations are located 23 kilometres apart on the shore of Lake Macquarie. Eraring power
135 station is located in Myuna Bay (33° 4'2.92"S, 151°33'19.13"E) and Vales Point power station
136 is located in Wyee Bay (33° 9'30.65"S, 151°31'48.37"E). Both stations use seawater from
137 Lake Macquarie for cooling. The seawater is circulated for cooling and then released back
138 into the estuary with no other treatment at a maximum of 37.5 °C as per licence
139 requirements (NSW Environment Protection Licences 761; 1429). One location was selected
140 near each power station outfall of the Eraring and Vales Point power stations in Lake
141 Macquarie during May 2018 (autumn). A control location was also selected that represented
142 the ambient mean temperature within Lake Macquarie which was not warmed by a power
143 station. At each location, two baskets were deployed within 20m of each other. Each

144 individual basket was attached to a 10 Kg concrete brick and contained a total of 40 oysters
145 from both species and treatments (control/non heat shock and heat shock) and were
146 deployed at a depth of 1.10 m by boat.

147 Temperature data were collected every 30 minutes by waterproof Hobo loggers (HOBO MX
148 Pendant Temperature, Onset) attached to the baskets. Study locations were visited five
149 times over seven months (late autumn to early summer) to download temperature data and
150 renew the loggers. Oysters were deployed in Lake Macquarie for approximately seven
151 months. At the end of the deployment (7 months) five baskets were retrieved; two from the
152 ambient (control) location and three from elevated locations; two from Wye Bay near
153 Vales power station and one from Rocky Point near Earing power station (120 oysters from
154 elevated temperature and 80 oysters from ambient temperature). Once retrieved, shell
155 growth, condition index, standard metabolic rate and survival of oysters was measured.
156 Total lipid and profile were measured in the laboratory.

157 **2.3 Shell growth and condition index**

158 To determine if exposure to heat shock confers subsequent resilience to long term exposure
159 to elevated temperature on growth and condition index. measurements of oysters were
160 done at the end of seven months of exposure in the field experiment.

161 There was no difference between shell height of oysters randomly allocated into heat shock
162 and control (non- heat shock) treatments (One-way ANOVA comparing heat shock vs non-
163 heat shock for each species, $n=60$; *O. angasi* = $p > 0.05$; *S. glomerata* = $p > 0.05$) at day zero.
164 Final shell growth was then calculated as the difference between the final size of each

165 individual oyster at seven months from an overall initial mean size of oysters per basket
166 (n=10). The difference in shell growth was calculated by the formula:

$$167 \quad SG = \frac{SH_1 - MSH_0}{t}$$

168 Where shell growth (SG) is the difference between final individual shell height (SH₁) in
169 millimetres and the mean initial shell height (MSH₀) divided by time (t) in days.

170

171 The condition index of oysters was measured at the end of the experiment. Oysters were
172 shucked, and body tissue and shell of individuals were dried in oven at 60°C for two days, to
173 determine the dry weight (grams). The condition index (Ci) of oysters was then calculated by
174 the formula [43,44]:

$$175 \quad Ci = \frac{\text{Dry body weight (g)}}{\text{Dry shell weight (g)}} \times 100$$

176 **2.4 Standard Metabolic Rate (SMR)**

177 To determine if exposure to heat shock would confer subsequent resilience to long term
178 exposure to elevated temperature on standard metabolic rate (SMR), the SMR of 9-11
179 oysters of each species, treatment and basket (total 51 oysters; heat shock and control/non-
180 heat shock) were measured at the end of the experiment using the methods of Parker et al.
181 [33]. Measurements were done adjacent to the locations of collection to minimise stress of
182 transport and to use seawater from Lake Macquarie.

183 To calculate SMR, oxygen consumption was measured by a closed respirometry system
184 (OXY-10 PreSens, AS1 Ltd, Regensburg, Germany). Seawater was collected from Lake

185 Macquarie and filtered through 0.47 μm glass filter paper before being used to fill
186 respirometry chambers. Respirometers were built to accommodate the maximum oyster
187 size (745ml and 830 ml). Each respirometer was connected to a fibre optic probe for
188 measurement of dissolved oxygen in seawater. The probe was previously calibrated using
189 two O_2 concentration points (0% and 100% oxygen saturation of seawater) following the
190 methods of Parker et al. [33]. Oysters were gently cleaned of any fouling organisms before
191 placed in filtered seawater (adjusted to the corresponding treatment levels). The time that
192 individuals took to lower the oxygen concentration in 20 % ($\sim 1.2 \text{ O}_2 \text{ mg L}^{-1}$) was recorded.
193 Following the procedure of Parker et al. [33], only the time that the oyster is open and
194 actively respiring (determined by observed decreasing oxygen) is used to calculate SMR. This
195 is done to guard against the oyster remaining closed from handling stress. After each trial,
196 each container was rinsed clean with filtered seawater (0.47 μm) and wiped clean with
197 paper towel. After measurement the oysters were removed from the container and shucked
198 to separate body tissues and shell. The tissue was then dried in an oven at 60°C for three
199 days to measure their constant dry body tissue and shell weight in grams ($\pm 0.0001\text{g}$,
200 Analytical Balance Sartorius Research). Standard metabolic rates (SMR) were calculated by
201 the formula:

$$202 \quad SMR = \frac{V_r(L) \times \Delta C_w \text{O}_2 (\text{mgO}_2\text{L}^{-1})}{\Delta t (\text{h}) \times \text{bw}(\text{g})}$$

203
204 where SMR is the oxygen consumption normalized to 1 g of dry tissue mass ($\text{mg O}_2 \text{ g}^{-1}$ dry
205 tissue mass h^{-1} , V_r is the volume of the respirometry chamber minus the volume of the

206 oyster (L), $\Delta C_w O_2$ is the change in water oxygen concentration measured ($\text{mg O}_2\text{L}^{-1}$), Δt is
207 measuring time (h) and b_w is the dry tissue mass (g) of the oyster.

208 **2.5 Total lipid and lipid profile**

209 To determine if exposure to heat shock and elevated temperature influences energy
210 allocation, total lipid and lipid profiles were analysed. Body tissues of the oysters were
211 placed in centrifuge tubes and frozen for analysis of total lipid content and lipid classes. The
212 tissues were kept at -22°C for transport and then stored at -80°C until analysis. The tissues
213 were then freeze dried (Alpha 1-4 LSCbasic, Martin Christ, Germany) and weighed in a
214 microbalance ($\pm 0.0001\text{g}$; Sartorius CPA225D). Lipids were extracted overnight using a
215 modified Bligh & Dyer [45] one-phase methanol-chloroform-water extraction (2:1:0.8
216 v/v/v). The phases were separated by the addition of chloroform-water (final solvent ratio,
217 1:1:0.9 v/v/v methanol-chloroform-water). The total solvent extract (TSE) was concentrated
218 using rotary evaporation at 40°C .

219 An aliquot of the TSE was analysed using an Iatroscan MK VI TH10 thin- layer
220 chromatography-flame ionization detector (TLC-FID) analyser (Tokyo, Japan) to quantify
221 individual lipid classes [46,47]. Samples were applied in duplicate to silica gel SIII
222 chromarods ($5\mu\text{m}$ particle size) using $1\mu\text{l}$ micropipettes. Chromorods were developed in a
223 glass tank lined with pre-extracted filter paper. The primary solvent system used for the lipid
224 separation was hexane-diethyl ether-formic acid (60:15:1.5), a mobile phase resolving non-
225 polar compounds such as steryl ester (SE), triacylglycerol (TAG), free fatty acids (FFA),
226 monoacylglycerol (MAG), Diacylglycerol (DAG). After development, the chromorods were
227 oven dried and analysed immediately to minimize absorption of atmospheric contaminants.
228 The FID was calibrated for each compound class (phosphatidylcholine (PL), cholesterol

229 (Chol), cholesteryl palmitate (SE), palmitic acid (FFA), monopalmitin (MAG), dipalmitin
230 (DAG), tripalmitin (TAG)). Peaks were quantified on an IBM compatible computer using
231 DAPA Scientific software (Kalamunda, Western Australia, Australia). TLC-FID results are
232 generally reproducible with a coefficient of variance of up to 3.46% of individual class
233 abundances [48].

234 **2.6 Survival**

235 Oyster survival was determined after seven months deployment by emptying baskets one
236 section at a time (to avoid mixing) and counting the total number of live oysters.

237 **2.7 Data analysis**

238 Statistical analyses were done using PRIMER v6+ software using either a three or two factor
239 nested PERMANOVA (PRIMER v6+). This analysis was selected because it is robust to
240 unbalanced designs [49].

241 For shell growth, condition index, SMR, and total lipids, data were analysed using a three
242 factor PERMANOVA with “heat shock” as fixed factor with two levels (heat shock or control),
243 “temperature” as fixed factor with two levels (ambient and elevated), and “basket” as
244 random factor with two levels (basket 1 and basket 2) nested in temperature and heat
245 shock. The analysis used 9999 permutations and only results with significance lower than
246 0.05 were considered as statistically different. The percentage survival at seven months was
247 analysed using a two factor PERMANOVA with heat shock as fixed factor with two levels
248 (heat shock or control) and temperature as fixed factor with two levels (ambient and
249 elevated).

250 The composition of lipid profiles were fourth root transformed to limit the influence of large
251 numbers [49] and analysed using a four factor multivariate PERMANOVA using the same
252 model as above; with heat shock as fixed factor with two levels (Heat shock or control),
253 temperature as fixed factor with two levels (control and elevated), and basket as random
254 factor with two levels (basket 1 and basket 2) nested in temperature and heat shock. The
255 analysis used 9999 permutations and only results with significance lower than 0.05 were
256 considered as statistically different.

257 **3. Results**

258 **3.1 Temperature**

259 The average temperature over seven months at the ambient location was $20.06^{\circ}\text{C} \pm 3.85$
260 (mean \pm S.D.) and the average temperature at the elevated locations was $24.56^{\circ}\text{C} \pm 4.59$
261 (Figure 2). The highest daily temperature experienced by oysters deployed at elevated
262 temperature locations was $32.81 \pm 0.39^{\circ}\text{C}$ (mean \pm S.D) in summer (December) and lowest
263 daily average for the ambient location during the experiment was $14.92^{\circ}\text{C} \pm 0.65$ (mean \pm
264 S.D) in winter (August).

265 **Figure 2.** Mean monthly temperatures \pm S.D. at control (ambient) and elevated temperature
266 locations in Lake Macquarie, NSW from May to December 2018 (approximately seven
267 months). Temperature data was measured every 30 minutes at 1.10m depth by water proof
268 loggers.

269 **3.2 Shell growth and condition index**

270 Shell growth of *O. angasi* was almost ten-fold greater at ambient temperature compared to
271 elevated temperature treatment (Figure 3a). Mean shell growth (mm day^{-1}) was 0.10 ± 0.01

272 mm day⁻¹ (mean ± S.E) at ambient temperature compared to 0.02 ± 0.01 mm day⁻¹ and 0.01
273 ± 0.008 mm day⁻¹ at elevated temperatures (Figure 3a). At ambient temperature, *O. angasi*
274 which were heat shocked had lower growth than non-heat shocked oysters (Table 1a) and
275 there was a trend for heat shocked *O. angasi* to have greater growth than non-heat shocked
276 oysters at elevated temperature, but this was not significant. Shell growth of *S. glomerata*
277 was not affected by temperature or heat shock (Figure 3b, Table 1b). *O. angasi* grew an
278 order of magnitude greater than *S. glomerata* under ambient conditions, however, under
279 elevated temperature there was little growth of either species (Figure 3 a, b).

280 **Figure 3.** Mean difference in shell growth (± S.E.) for **a.** flat oysters, *Ostrea angasi* (FO
281 control and FO heat shocked) and **b.** Sydney rock oysters, *Saccostrea glomerata* (SRO
282 control and SRO heat shocked), exposed for seven months at ambient and elevated
283 temperature locations at Lake Macquarie.

284

285 **Table 1a.** Shell growth, condition index and SMR of *Ostrea angasi* exposed for seven months in Lake Macquarie. P values were created using Monte Carlo
 286 tests.

	Shell Growth				Condition index				SMR			
	df	MS	Pseudo-F	P(MC)	df	MS	Pseudo-F	P(MC)	df	MS	Pseudo-F	P(MC)
Heat Shock	1	9.97	1.55	0.27	1	166.93	0.42	0.65	1	241.95	0.19	0.78
Temperature	1	3960	616.92	<0.001	1	3201.8	7.87	0.04	1	1748.70	1.36	0.3
Heat Shock x Temperature	1	165.88	25.84	<0.001	1	626.55	1.55	0.29	1	701.91	0.54	0.54
Basket (Heat Shock x Temperature)	4	5.91	0.2	0.94	3	402.87	1.56	0.20	4.00	1337.00	2.46	0.06
Residuals	69	29.23			28.00	258.09			20.00	543.10		
Total	76				34.00				27.00			

287

288 **Table 1b.** Shell growth, condition index and SMR of *Saccostrea glomerata* exposed for seven months in Lake Macquarie. P values were created using Monte
 289 Carlo tests.

	Shell Growth				Condition index				SMR			
	df	MS	Pseudo-F	P(MC)	df	MS	Pseudo-F	P(MC)	df	MS	Pseudo-F	P(MC)
Heat Shock	1	2.60	0.04	0.85	1	132.82	0.42	0.65	1	102.31	0.12	0.90
Temperature	1	186.07	2.84	0.17	1	2651.80	8.29	0.02	1	1181.50	1.39	0.29
Heat Shock x Temperature	1	0.00	0.00	0.99	1	777.29	2.43	0.16	1	1571.90	1.85	0.21
Basket (Heat Shock x Temperature)	4	66.49	1.49	0.22	4	319.59	0.96	0.46	4	880.30	1.38	0.24
Residuals	79	44.58			19	331.50			18	635.75		
Total	86				26				25			

291 The condition index of both *O. angasi* and *S. glomerata* was significantly lower at elevated
292 temperature (Figure 4a,b, Table 1a,b) with no effect of heat shock treatment, although
293 there was a slight trend for heat shocked *O. angasi* oysters at elevated temperature to have
294 better condition.

295 **Figure 4.** Mean condition index (\pm S.E.) of **a.** flat oysters, *Ostrea angasi* (FO control and FO
296 heat shocked) and **b.** Sydney rock oysters, *Saccostrea glomerata* (SRO control and SRO heat
297 shocked) exposed for seven months at ambient and elevated locations at Lake Macquarie.

298 **3.3 Standard Metabolic Rate (SMR)**

299 Standard Metabolic Rate of control, non-heat shocked *O. angasi* was lower at elevated
300 temperature, but this was not significant (Figure 5a; Table 1a). SMR of control, non-heat
301 shocked *S. glomerata* was greater at elevated temperature, but this was not significant
302 (Figure 5b, Table 1b).

303 **Figure 5.** Mean standard metabolic rate (SMR) (\pm S.E.) of **a** flat oysters, *Ostrea angasi* (FO
304 control and FO heat shocked) and **b.** Sydney rock oysters, *Saccostrea glomerata* (SRO
305 control and SRO heat shocked) exposed for seven months at ambient and elevated
306 temperature locations at Lake Macquarie.

307 **3.4 Total lipids and lipid profiles**

308

309 Mean total lipid content of *O. angasi* was greater at in those from the elevated temperature
310 treatment compared to those held ambient (Figure 6a, Table 2a). There were no effects of
311 heat shock or temperature on total lipid content in *S. glomerata* (Figure 6b, Table 2b). Lipid
312 profile of *O. angasi*, was mostly driven by a greater amount of phospholipids in the oysters

313 in the elevated temperature treatment (Figure 7a, Table 2a) and significantly lower amounts
314 of TAGs (Figure 7a). Lipid profile of *S. glomerata* was similar across ambient and elevated
315 temperatures, but there were significantly greater phospholipids at elevated temperature
316 (Figure 7b, Table 2b).

317 **Figure 6.** Mean total lipids (\pm S.E.) for **a.** flat oysters, *Ostrea angasi* (FO control and FO heat
318 shocked) and **b.** Sydney rock oysters, *Saccostrea glomerata* (SRO control and SRO heat
319 shocked) exposed for seven months at ambient and elevated temperature locations (n=5;
320 except for HS oysters from Rocky Point – FO HS [n=3], SRO HS [n=2]) at Lake Macquarie.

321 **Figure 7.** Lipid profile of **a.** flat oysters, *Ostrea angasi* (FO control FO heat shocked) after
322 seven months exposure at ambient and elevated temperature locations and **b.** Lipid profile
323 of Sydney rock oysters *Saccostrea glomerata* (SRO control and SRO heat shocked) exposed
324 for seven months at ambient and elevated temperature. Lipid classes abbreviations are: SE –
325 steryl ester; TAG -Triacylglyceride; FFA – Free Fatty Acids; Chol – Cholesterol; DAG –
326 Diacylglyceride; MAG – Monoglyceride and PL – Polar Lipids.

327 **Table 2a.** Total lipids, amount of total lipids (mg/g), amount of Triacylglycerides (TAGs; mg/g) and Phospholipids (PLs; mg/g) of *Ostrea angasi* exposed for
 328 seven months in Lake Macquarie. P values were created using Monte Carlo tests.

	Total lipids				TAGs				PLs			
	df	MS	Pseudo-F	P(MC)	df	MS	Pseudo-F	P(MC)	df	MS	Pseudo-F	P(MC)
Heat Shock	1	204.36	0.58	0.55	1	3.58	0.02	0.96	1	780.33	1.20	0.36
Temperature	1	2652.8	7.55	0.03	1	798.48	3.65	0.13	1	8619.40	13.21	<0.001
Heat Shock x Temperature	1	270.63	0.77	0.48	1	72.98	0.33	0.62	1	347.69	0.53	0.64
Basket (Heat Shock x Temperature)	4	351.92	1.47	0.24	4	219.52	1.81	0.17	4	653.35	1.2	0.32
Residuals	14	239.88			15	121.39			15	545.05		
Total	21				2				22			

329

330 **Table 2b.** Total lipids (mg/g), amount of Triacylglycerides (TAGs; mg/g) and Phospholipids (PLs; mg/g), of *Saccostrea glomerata* exposed for seven months in
 331 Lake Macquarie. P values were created using Monte Carlo tests.

	Total lipids				TAGs				PLs			
	df	MS	Pseudo-F	P(MC)	df	MS	Pseudo-F	P(MC)	df	MS	Pseudo-F	P(MC)
Heat Shock	1	630.71	0.62	0.51	1	355.76	0.63	0.48	1	380.53	2.15	0.20
Temperature	1	585.15	0.57	0.54	1	40.98	0.07	0.86	1	1291.40	7.31	0.04
Heat Shock x Temperature	1	824.63	0.81	0.44	1	951.68	1.70	0.26	1	957.97	5.42	0.06
Basket (Heat Shock x Temperature)	4	1023.70	1.50	0.22	4	564.66	1.68	0.20	4	172.28	0.37	0.84
Residuals	17	682.38			17	337.06			17	461.41		
Total	24				24				24			

332

333 **3.5 Survival**

334 Survival of heat-shocked *O. angasi* was significantly greater than non-heat shocked oysters
335 at ambient and elevated temperature (Figure 8a; Table 3a). At the ambient and elevated
336 temperature locations, survival of *O. angasi* was greatest for the heat shocked oysters
337 (Ambient, control oysters = 90% and heat shocked oysters= 100%; Elevated, control = 53%
338 and heat shocked = 80%). Survival of *S. glomerata* was significantly lower for control oysters
339 at ambient temperature compared to heat-shocked oysters (Figure 8b, Table 3b).

340 **Figure 8.** Mean survival (\pm S.D.) of **a** flat oysters, *Ostrea angasi* (FO control and FO heat
341 shocked) and **b.** Sydney rock oysters, *Saccostrea glomerata*, (SRO control and SRO heat
342 shocked), exposed for seven months at ambient and elevated temperature locations at Lake
343 Macquarie.

344 **Table 3a.** Percentage survival of *Ostrea angasi* exposed for seven months in Lake Macquarie. P
345 values were created using Monte Carlo tests.

	df	MS	Pseudo- F	P(MC)
Heat Shock	1	359.59	16.16	0.007
Temperature	1	802.23	36.05	0.001
Heat Shock x Temperature	1	142.45	6.40	0.04
Residuals	6	22.25		
Total	9			

346

347 **Table 3b.** Percentage survival of *Saccostrea glomerate* exposed for seven months in Lake Macquarie.
348 P values were created using Monte Carlo tests.

349

	df	MS	Pseudo- F	P(MC)
Heat Shock	1	39.189	2.019	0.2037
Temperature	1	57.668	2.971	0.1309
Heat Shock x Temperature	1	143.35	7.3852	0.0353
Residuals	6	19.41		
Total	9			

350

351 **4. Discussion**

352 Exposure to long term warming in the field had negative impacts on shell growth, condition
353 index, and survival of *O. angasi* and *S. glomerata*. Shell growth, condition index, lipid
354 content and profile and survival, but not SMR of oysters was impacted by elevated
355 temperature, with flat oysters more impacted than Sydney rock oysters. Flat oysters grew
356 faster than Sydney rock oysters at ambient temperature, but were more sensitive to
357 elevated temperature. Exposure early in life to heat shock did little to ameliorate the
358 negative effects of elevated temperature, although there was a trend for shell growth and
359 condition index and a significant effect of survival of heat shocked flat oysters to be greater
360 than control oysters at elevated temperature. SMR was not significantly impacted by
361 elevated temperature, although once again there was a trend for SMR of flat oysters to
362 decrease with increased temperature and for SMR of Sydney rock oysters to increase with
363 increased temperature. The lipid profile of *O. angasi* was also reduced by elevated
364 temperature, while the lipid profile for *S. glomerata* was not affected.

365 As oysters are ectothermic organisms, changes in external temperature away from their
366 optimum causes physiological processes to become less efficient and homeostasis begins to
367 require more energy [50]. The effects of elevated temperature on *O. angasi* and *S.*
368 *glomerata* are similar to those observed for other bivalve species. For example, Hiebenthal
369 et al. [51] found lower growth and condition for *Arctica islandica* at elevated temperature
370 (16°C) compared to the control (7.5°C) and an intermediate treatment (10°C). Condition
371 index and survivorship of *M. edulis* was reduced under elevated temperature (25°C)
372 compared with control [51]. Effects on these physiological processes, were attributed to
373 thermal sensitivity of *A. islandica* to temperatures outside its distribution and to

374 accumulation of lipofuscin, a disease related pigment [51]. For *Mercenaria mercenaria* and
375 *Argopecten irradians*, elevated temperature (28°C) impacted shell growth of juveniles [52].

376 Elevations in temperature increase the SMR of marine ectotherms until a point known as
377 the “Arrhenius Breakpoint Temperature” (ABT). When ABT is reached, SMR rapidly declines
378 indicating that the organism can no longer meet their energetic requirements at that
379 temperature [33]. The reduced growth of *O. angasi* at elevated compared to ambient
380 temperature was correlated with a trend for lower SMR, indicating that *O. angasi* may have
381 experienced temperatures beyond their ABT. Temperature had no effect on the SMR of *S.*
382 *glomerata*. Parker et al., [33] found that increases in seawater temperature can increase the
383 SMR of *S. glomerata*. SMR of *S. glomerata* increased with increased temperature up to 33 °C
384 (the upper temperature treatment in that study) indicating an ABT for *S. glomerata* of above
385 33 °C. Increased SMR can impact energy budget and may indicate a thermal response with
386 extra costs needed to cover basal metabolism [33,53]. For oysters, thermal stress can also
387 alter cardiac function, protein synthesis [53] and gametogenesis [23].

388 Oysters have the capacity to store surplus energy ingested from food in the form of lipids
389 which can assist in the persistence during stressful conditions. While the lipid profile of *S.*
390 *glomerata* was not impacted by elevated temperature, there were significant impacts of
391 elevated temperature on total lipids and lipid profile, especially Triacylglycerides (TAGs) of
392 *O. angasi*. TAGs are the primary source of stored lipid energy for bivalves [54], indicating
393 that *O. angasi* had begun to use stored lipid reserves. Studies have found that under
394 stressful conditions bivalves have lower lipid reserves. For example, exposure to elevated
395 $p\text{CO}_2$ decreased the lipid index of larvae of *A. irradians*, *M. mercenaria* and *C. virginica*
396 which further declined when combined with warming [52, 55]. Lipid levels in the eggs of *S.*

397 *glomerata* also decreased when exposed to the dual stress of elevated $p\text{CO}_2$ and copper
398 [56]. Further research on how lipids are used by oysters in response to stress could provide
399 insights into the ramifications of living in warmer oceans.

400 **Stress inoculation, and resilience**

401 This study tested the hypothesis that pre exposure of oysters to heat shock stress will build
402 resilience to later exposure to elevated temperature. Stress inoculation leading to stress
403 resilience has been observed in diverse phyla from bacteria to mammals e.g. [10,11]. Heat
404 shock may help to build resilience, but at the same time have costs. For example, heat
405 shocked *O. angasi* had significantly greater rates of survival at elevated temperatures, but
406 heat shocked oysters had less growth at ambient temperature. Perhaps energy was used to
407 produce heat shock proteins or other protective measures, thereby reducing energetic
408 reserves for growth and other important physiological processes.

409 Heat shocked *O. angasi* had greater rates of survival at elevated temperature, and a similar
410 trend was observed for *S. glomerata* which had greater than 90% survival at elevated
411 temperature. When organisms experience stressful temperatures, they undergo a thermal
412 response, which is energy dependent [57]. This thermal response includes producing
413 chaperones, such as energetically expensive heat shock proteins (HSPs) [57,58]. Species with
414 lower thermal tolerance might be induced to produce HSPs in response to elevated
415 temperatures before more tolerant species, which can endure longer periods under
416 warming stress (e.g. *M. trossulus* and *M. galloprovincialis*; [58]). Production of these
417 molecular chaperones (HSPs) is a common response to elevated temperature [57]. Heat
418 shock proteins have important functions when an organism is exposed to elevated
419 temperature, including degradation of denatured proteins and prevention of misfolding,

420 having a key function on cellular protection [59]. These responses (e.g. expression of heat
421 shock proteins, antioxidants, increased respiration rates) all incur an energetic cost which
422 can cause an imbalance in the energetic partitioning of individuals [33,57,60,61].

423 Overall, *S. glomerata* was found to be generally more tolerant of habitat warming than *O.*
424 *angasi*. *S. glomerata* had no change in shell growth although they were in poorer condition
425 at elevated temperature of 28-30 °C. As an intertidal species that experiences a highly
426 dynamic thermal environment, *S. glomerata* could be expected to be more thermally
427 tolerant as has been shown for other intertidal organisms [62,63]. These findings are
428 supported by previous work by Parker et al., [33] that showed 33 °C was not beyond their
429 ABT, and the distribution of *S. glomerata* which extends further north along the east coast
430 of Australia than *O. angasi*. Additionally, *S. glomerata* can experience air temperatures in
431 excess of 40 °C during emersion at low tide [64]. The lack of effect of elevated temperature
432 on *S. glomerata* indicates that they were not placed beyond their thermal limits in the
433 deployment used here, in contrast to *O. angasi*, which did not cope as well.

434 *O. angasi* had the greatest growth rate at ambient conditions. The shell growth of *O. angasi*
435 was over ten-fold greater than Sydney rock oysters after seven months, as expected from
436 growth in aquaculture [65]. While *O. angasi* grew well at ambient conditions, growth and
437 survival were impacted by warming. As this species lives in a relatively stable, sub-tidal
438 habitat we expected this species to be more sensitive to warming compared to *S.*
439 *glomerata*.

440 Globally and across Australia efforts are being made to restore oyster reefs [31,66,67].
441 Climate change will impact on oyster reef restoration [36]. Projected ocean warming for the
442 region (4°C) as well as contemporary marine heat waves, as seen in the region recently [39]

443 are an important consideration for reef restoration efforts along the south-eastern coastline
444 of Australia. Our study has shown that using thermal outfall as a proxy for ocean warming
445 can be useful for predicting future warming. This approach is similar to natural laboratories
446 using underwater CO₂ vents which have successfully tested the responses of marine
447 organisms to ocean acidification [68,69]. Our results indicate that habitat warming will be a
448 greater threat to *O. angasi* compared to *S. glomerata*. As ocean warming will not act alone,
449 oyster reef restoration is at risk from multiple stressors including ocean acidification,
450 salinity, and other environmental pollutants which will act simultaneously [36]. These co-
451 occurring stressors further threaten native species of oysters, other molluscs and marine
452 organisms and so mitigation strategies to build oyster resilience will be critical. Our results
453 indicate that early exposure to stress inoculation does not enhance resilience and may not
454 be useful strategy, especially for restoration ventures involving *O. angasi*.

455 **Acknowledgements.** We acknowledge and thank CNPq – Conselho Nacional de
456 Desenvolvimento Científico e Tecnológico – Brazil (PhD CNPq scholarship) for financially
457 supporting this work. We also would like to thank Port Stephens Fisheries Institute, Dr
458 Wayne O’Connor and the Office of Environment and Heritage for great support during
459 development of this study. We thank Kyle Johnston and Richard Grainger for assistance in
460 the field.

461

462

463 **References**

- 464 1. IPCC. Summary for policymakers. In: Climate Change 2014: Impacts, Adaptation, and
465 Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the
466 Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Field, C.B., V.R.
467 Barros, D.J. Dokken, K.J. Mach, M.D. Mastrandrea, T.E. Bilir, M. Chatterjee, K.L. Ebi, Y.O.
468 Estrada, R.C. Genova, B. Girma, E.S. Kissel, A.N. Levy, S. MacCracken, P.R. Mastrandrea, and
469 L.L.White (eds.)]. Cambridge University Press, Cambridge, United Kingdom and New York,
470 NY, USA. 2014;pp. 1-32.
- 471 2. Cheng L, Abraham J, Hausfather Z, Trenberth KE. How fast are the oceans warming?
472 Science. 2019;363(6423):128-9.
- 473 3. Hobday AJ, Okey TA, Poloczanska ES, Kunz TJ, Richardson AJ. Impacts of climate
474 change on Australian marine life. Report to the Australian Greenhouse Office, Canberra,
475 Australia. 2006.
- 476 4. Lenton A, McInnes KL, O'Grady JG. Marine projections of warming and ocean
477 acidification in the Australasian region. Australian Meteorological and Oceanographic
478 Journal. 2015;65(1):1-28.
- 479 5. Hobday AJ, Alexander LV, Perkins SE, Smale DA, Straub SC, Oliver EC, et al. A
480 hierarchical approach to defining marine heatwaves. Progress in Oceanography.
481 2016;141:227-38.
- 482 6. Ridgway K, Hill K. The East Australian Current. A marine climate change impacts and
483 adaptation report card for Australia. 2009;5(09).

- 484 7. Oliver EC, Donat MG, Burrows MT, Moore PJ, Smale DA, Alexander LV, et al. Longer
485 and more frequent marine heatwaves over the past century. *Nat Commun.* 2018;9(1):1324.
- 486 8. Doney SC, Ruckelshaus M, Duffy JE, Barry JP, Chan F, English CA, et al. Climate
487 change impacts on marine ecosystems. *Annu Rev Mar Sci.* 2012;4:11–37.
- 488 9. Oliver EC, Benthuisen JA, Bindoff NL, Hobday AJ, Holbrook NJ, Mundy CN, et al. The
489 unprecedented 2015/16 Tasman Sea marine heatwave. *Nat Commun.* 2017;8:16101.
- 490 10. Todgham AE, Schulte PM, Iwama GK. Cross-tolerance in the tidepool sculpin: the role
491 of heat shock proteins. *Physiological and Biochemical Zoology.* 2005;78(2):133-44.
- 492 11. Parker KJ, Buckmaster CL, Sundlass K, Schatzberg AF, Lyons DM. Maternal mediation,
493 stress inoculation, and the development of neuroendocrine stress resistance in primates.
494 *Proceedings of the National Academy of Sciences.* 2006;103(8):3000-5.
- 495 12. Laplace JM, Boutibonnes P, Auffray Y. Unusual resistance and acquired tolerance to
496 cadmium chloride in *Enterococcus faecalis*. *J Basic Microbiol.* 1996;36(5):311-7.
- 497 13. Krebs RA, Feder ME. Hsp70 and larval thermotolerance in *Drosophila melanogaster*:
498 how much is enough and when is more too much? *J Insect Physiol.* 1998;44(11):1091-101.
- 499 14. Munne-Bosch S, Alegre L. Cross-stress tolerance and stress"memory" in plants.
500 *Environ Exp Bot.* 2013;94:1-88.
- 501 15. Tedengren M, Olsson B, Reimer O, Brown DC, Bradley BP. Heat pretreatment
502 increases cadmium resistance and HSP 70 levels in Baltic Sea mussels. *Aquat Toxicol.*
503 2000;48(1):1-12.

- 504 16. Chapple JP, Smerdon GR, Berry R, Hawkins AJ. Seasonal changes in stress-70 protein
505 levels reflect thermal tolerance in the marine bivalve *Mytilus edulis* L. J Exp Mar Biol Ecol.
506 1998;229(1):53-68.
- 507 17. Huey RB, Bennett AF. Physiological adjustments to fluctuating thermal
508 environments: an ecological and evolutionary perspective. In: Morimoto RI, Tissieres A,
509 Georgopoulos C, editors. Stress proteins in biology and medicine. Cold Spring Harbor, NY:
510 Cold Spring Harbor Lab. Press; 1990;19:37-59.
- 511 18. Pechenik JA. On the advantages and disadvantages of larval stages in benthic marine
512 invertebrate life cycles. Mar Ecol Prog Ser. 1999;177:269-97.
- 513 19. Gattuso J-P, Magnan A, Billé R, Cheung WW, Howes EL, Joos F, et al. Contrasting
514 futures for ocean and society from different anthropogenic CO₂ emissions scenarios.
515 Science. 2015;349(6243):aac4722.
- 516 20. Pörtner H-O. Integrating climate-related stressor effects on marine organisms:
517 unifying principles linking molecule to ecosystem-level changes. Mar Ecol Prog Ser.
518 2012;470:273-90.
- 519 21. Sokolova IM. Energy-limited tolerance to stress as a conceptual framework to
520 integrate the effects of multiple stressors. Integ and Comp Biol. 2013;53(4):597-608.
- 521 22. Ivanina AV, Dickinson GH, Matoo OB, Bagwe R, Dickinson A, Beniash E, et al.
522 Interactive effects of elevated temperature and CO₂ levels on energy metabolism and
523 biomineralization of marine bivalves *Crassostrea virginica* and *Mercenaria mercenaria*.
524 Comp Biochem Physiol A Mol Integr Physiol. 2013;166(1):101-11. doi:
525 <https://doi.org/10.1016/j.cbpa.2013.05.016>.

- 526 23. Fearman J, Moltschaniwskyj N. Warmer temperatures reduce rates of gametogenesis
527 in temperate mussels, *Mytilus galloprovincialis*. *Aquaculture*. 2010;305(1-4):20-5.
- 528 24. Parker LM, Ross PM, O'Connor WA, Borysko L, Raftos DA, Pörtner HO. Adult
529 exposure influences offspring response to ocean acidification in oysters. *Glob Change Biol*.
530 2012;18(1):82-92.
- 531 25. Parker LM, O'Connor WA, Raftos DA, Pörtner H-O, Ross PM. Persistence of positive
532 carryover effects in the oyster, *Saccostrea glomerata*, following transgenerational exposure
533 to ocean acidification. *PloS one*. 2015;10(7):e0132276.
- 534 26. Scanes E, Johnston EL, Cole VJ, O'Connor WA, Parker LM, Ross PM. Quantifying
535 abundance and distribution of native and invasive oysters in an urbanised estuary. *Aquatic*
536 *Invasions*. 2016;11(4):425-36.
- 537 27. Ogburn DM, White I, Mcphee DP. The disappearance of oyster reefs from eastern
538 Australian estuaries—impact of colonial settlement or mudworm invasion? *Coast Manage*.
539 2007;35(2-3):271-87.
- 540 28. Nell JA. The history of oyster farming in Australia. *Marine Fisheries Review*.
541 2001;63(3):14-25.
- 542 29. NSW Department of Primary Industries, *Aquaculture Production Report 2010-2011*.
543 Port Stephens, New South Wales: Department of Primary Industries, 2011. ISSN 1444-840.
- 544 30. Crawford, C, *National review of *Ostrea angasi* aquaculture: historical culture, current*
545 *methods and future priorities*, University of Tasmania Institute for Marine and Antarctic
546 *Studies*, Hobart, Tasmania 2016.

- 547 31. Gillies CL, Crawford C, Hancock B. Restoring Angasi oyster reefs: What is the
548 endpoint ecosystem we are aiming for and how do we get there? *Ecol Manage Restor.*
549 2017;18(3):214-22.
- 550 32. McLeod I, Boström-Einarsson L, Creighton C, D'Anastasi B, Diggles B, Dwyer P, et al.
551 Habitat value of Sydney rock oyster (*Saccostrea glomerata*) reefs on soft sediments. *Mar*
552 *and Freshw Res.*2019. <https://doi.org/10.1071/MF18197>
- 553 33. Parker LM, Scanes E, O'Connor WA, Coleman RA, Byrne M, Pörtner H-O, et al. Ocean
554 acidification narrows the acute thermal and salinity tolerance of the Sydney rock oyster
555 *Saccostrea glomerata*. *Mar Pollut Bull.* 2017;122(1-2):263-71.
- 556 34. Parker LM, Ross PM, O'Connor WA. Comparing the effect of elevated $p\text{CO}_2$ and
557 temperature on the fertilization and early development of two species of oysters. *Mar Biol.*
558 2010;157(11):2435-52.
- 559 35. Cole VJ, Parker LM, O'Connor SJ, O'Connor WA, Scanes E, Byrne M, et al. Effects of
560 multiple climate change stressors: ocean acidification interacts with warming, hyposalinity,
561 and low food supply on the larvae of the brooding flat oyster *Ostrea angasi*. *Mar Biol.*
562 2016;163(5):1-17.
- 563 36. Pereira RRC, Scanes E, Parker LM, Byrne M, Cole VJ, Ross PM. Restoring the flat
564 oyster *Ostrea angasi* in the face of a changing climate. *Mar Ecol Prog Ser.* 2019;625:27-39.
- 565 37. Ridgway K. Long-term trend and decadal variability of the southward penetration of
566 the East Australian Current. *Geophys Res Lett.* 2007;34(13).

- 567 38. Hobday AJ, Pecl GT. Identification of global marine hotspots: sentinels for change
568 and vanguards for adaptation action. *Rev Fish Biol Fish.* 2014;24(2):415-25.
- 569 39. Babcock, R.C., Bustamante, R.H., Fulton, E.A., Fulton, D.J., Haywood, M.D., Hobday,
570 A.J., Kenyon, R., Matear, R.J., Plagányi, E.E., Richardson, A.J. and Vanderklift, M.A. Severe
571 continental-scale impacts of climate change are happening now: Extreme climate events
572 impact marine habitat forming communities along 45% of Australia's coast. *Frontiers in*
573 *Marine Science*, 6, 2019;p.411.
- 574 40. Schiel, D.R., Steinbeck, J.R. and Foster, M.S.,. Ten years of induced ocean warming
575 causes comprehensive changes in marine benthic communities. *Ecology*, 85(7), 2004;
576 pp.1833-1839.
- 577 41. Nell JA, O'Connor WA. The evaluation of fresh algae and stored algal concentrates as
578 a food source for Sydney rock oyster, *Saccostrea commercialis* (Iredale & Roughley), larvae.
579 *Aquaculture*. 1991;99(3-4):277-84.
- 580 42. Roy P, Williams R, Jones A, Yassini I, Gibbs P, Coates B, et al. Structure and function
581 of south-east Australian estuaries. *Estuar Coast Shelf Sci.* 2001;53(3):351-84.
- 582 43. Lucas A, Beninger PG. The use of physiological condition indices in marine bivalve
583 aquaculture. *Aquaculture*. 1985;44(3):187-200.
- 584 44. Mann R. A comparison of methods for calculating condition index in eastern oysters
585 *Crassostrea virginica* (Gmelin, 1791). *J Shellfish Res.* 1992;11(1):55.
- 586 45. Bligh EG, Dyer WJ. A rapid method of total lipid extraction and purification. *Can J*
587 *Biochem Physiol.* 1959;37(8):911-7.

- 588 46. Volkman JK, Nichols PD. Applications of thin layer chromatography-flame ionization
589 detection to the analysis of lipids and pollutants in marine and environmental samples. J
590 Planar Chromatogr. 1991;4:19-26.
- 591 47. Ackman R. [11] Flame ionization detection applied to thin-layer chromatography on
592 coated quartz rods. Methods Enzymol. 72: Elsevier; 1981. p. 205-52.
- 593 48. Sinanoglou VJ, Strati IF, Bratakos SM, Proestos C, Zoumpoulakis P, Miniadis-
594 Meimaroglou S. On the combined application of Iatroscan TLC-FID and GC-FID to identify
595 total, neutral, and polar lipids and their fatty acids extracted from foods. ISRN
596 Chromatography. 2013;2013.
- 597 49. Anderson M, Gorley RN, Clarke RK. Permanova+ for primer: Guide to software and
598 statisticl methods: Primer-E Limited; 2008, Plymouth, UK.
- 599 50. Pörtner H-O, Reipschläger A, Heisler N. Acid-base regulation, metabolism and
600 energetics in *Sipunculus nudus* as a function of ambient carbon dioxide level. J Exp Biol.
601 1998;201(1):43-55.
- 602 51. Hiebenthal C, Philipp EE, Eisenhauer A, Wahl M. Effects of seawater $p\text{CO}_2$ and
603 temperature on shell growth, shell stability, condition and cellular stress of Western Baltic
604 Sea *Mytilus edulis* (L.) and *Arctica islandica* (L.). Mar Biol. 2013;160(8):2073-87.
- 605 52. Talmage SC, Gobler CJ. Effects of elevated temperature and carbon dioxide on the
606 growth and survival of larvae and juveniles of three species of northwest Atlantic bivalves.
607 PloS one. 2011;6(10):e26941.

- 608 53. Bayne B, Bayne C, Carefoot T, Thompson R. The physiological ecology of *Mytilus*
609 *californianus* Conrad. 1. Metabolism and energy balance. *Oecologia*. 1976:211-28.
- 610 54. Abad, M., Ruiz, C., Martinez, D., Mosquera, G. and Sánchez, J. Seasonal variations of
611 lipid classes and fatty acids in flat oyster, *Ostrea edulis*, from San Cibrán (Galicia, Spain).
612 *Comparative Biochemistry and Physiology Part C: Pharmacology, Toxicology and*
613 *Endocrinology*, 1995;110(2). pp.109-118.
- 614 55. Fields PA, Zuzow MJ, Tomanek L. Proteomic responses of blue mussel (*Mytilus*)
615 congeners to temperature acclimation. *J Exp Biol*. 2012;215(7):1106-16.
- 616 56. Scanes E, Parker LM, O'Connor WA, Gibbs MC, Ross PM. Copper and ocean
617 acidification interact to lower maternal investment, but have little effect on adult physiology
618 of the Sydney rock oyster *Saccostrea glomerata*. *Aquat Toxicol*. 2018;203:51-60.
- 619 57. Somero GN. Thermal physiology and vertical zonation of intertidal animals: optima,
620 limits, and costs of living. *Integr Comp Biol*. 2002;42(4):780-9.
- 621 58. Anestis A, Lazou A, Pörtner HO, Michaelidis B. Behavioral, metabolic, and molecular
622 stress responses of marine bivalve *Mytilus galloprovincialis* during long-term acclimation at
623 increasing ambient temperature. *American Journal of Physiology-Regulatory, Integr Comp*
624 *Physiol*. 2007;293(2):R911-R21.
- 625 59. Sørensen JG, Kristensen TN, Loeschcke V. The evolutionary and ecological role of
626 heat shock proteins. *Ecol Lett*. 2003;6(11):1025-37.

- 627 60. Ivanina A, Taylor C, Sokolova I. Effects of elevated temperature and cadmium
628 exposure on stress protein response in eastern oysters *Crassostrea virginica* (Gmelin). *Aquat*
629 *Toxicol.* 2009;91(3):245-54.
- 630 61. Abele D, Heise K, Pörtner H-O, Puntarulo S. Temperature-dependence of
631 mitochondrial function and production of reactive oxygen species in the intertidal mud clam
632 *Mya arenaria*. *J Exp Biol.* 2002;205(13):1831-41.
- 633 62. Somero, G.N. The physiology of global change: linking patterns to mechanisms.
634 *Annual Review of Marine Science*, 2012;4, pp.39-61.
- 635 63. Rivest, E.B., Comeau, S. and Cornwall, C.E. The role of natural variability in shaping
636 the response of coral reef organisms to climate change. *Current Climate Change Reports*,
637 2017;3(4), pp.271-281.
- 638
- 639 64. McAfee, D., O'connor, W.A. and Bishop, M.J. Fast-growing oysters show reduced
640 capacity to provide a thermal refuge to intertidal biodiversity at high temperatures. *Journal*
641 *of Animal Ecology*, 2017;86(6), pp.1352-1362.
- 642 65. Mitchell, I.M., Crawford, C.M. and Rushton, M.J. Flat oyster (*Ostrea angasi*) growth
643 and survival rates at Georges Bay, Tasmania (Australia). *Aquaculture*, 2000;191(4), pp.309-
644 321.
- 645 66. Lipcius RN, Burke RP. Successful recruitment, survival and long-term persistence of
646 eastern oyster and hooked mussel on a subtidal, artificial restoration reef system in
647 Chesapeake Bay. *PloS one.* 2018;13(10):e0204329.

- 648 67. Laing I, Walker P, Areal F. Return of the native—is European oyster (*Ostrea edulis*)
649 stock restoration in the UK feasible? Aquatic Living Resources. 2006;19(3):283-7.
- 650 68. Rodolfo-Metalpa R, Lombardi C, Cocito S, Hall-Spencer JM, Gambi MC. Effects of
651 ocean acidification and high temperatures on the bryozoan *Myriapora truncata* at natural
652 CO₂ vents. Mar Ecol. 2010;31(3):447-56.
- 653 69. Calosi P, Rastrick S, Graziano M, Thomas S, Baggini C, Carter H, et al. Distribution of
654 sea urchins living near shallow water CO₂ vents is dependent upon species acid–base and
655 ion-regulatory abilities. Mar Pollut Bull. 2013;73(2):470-84.
- 656
- 657

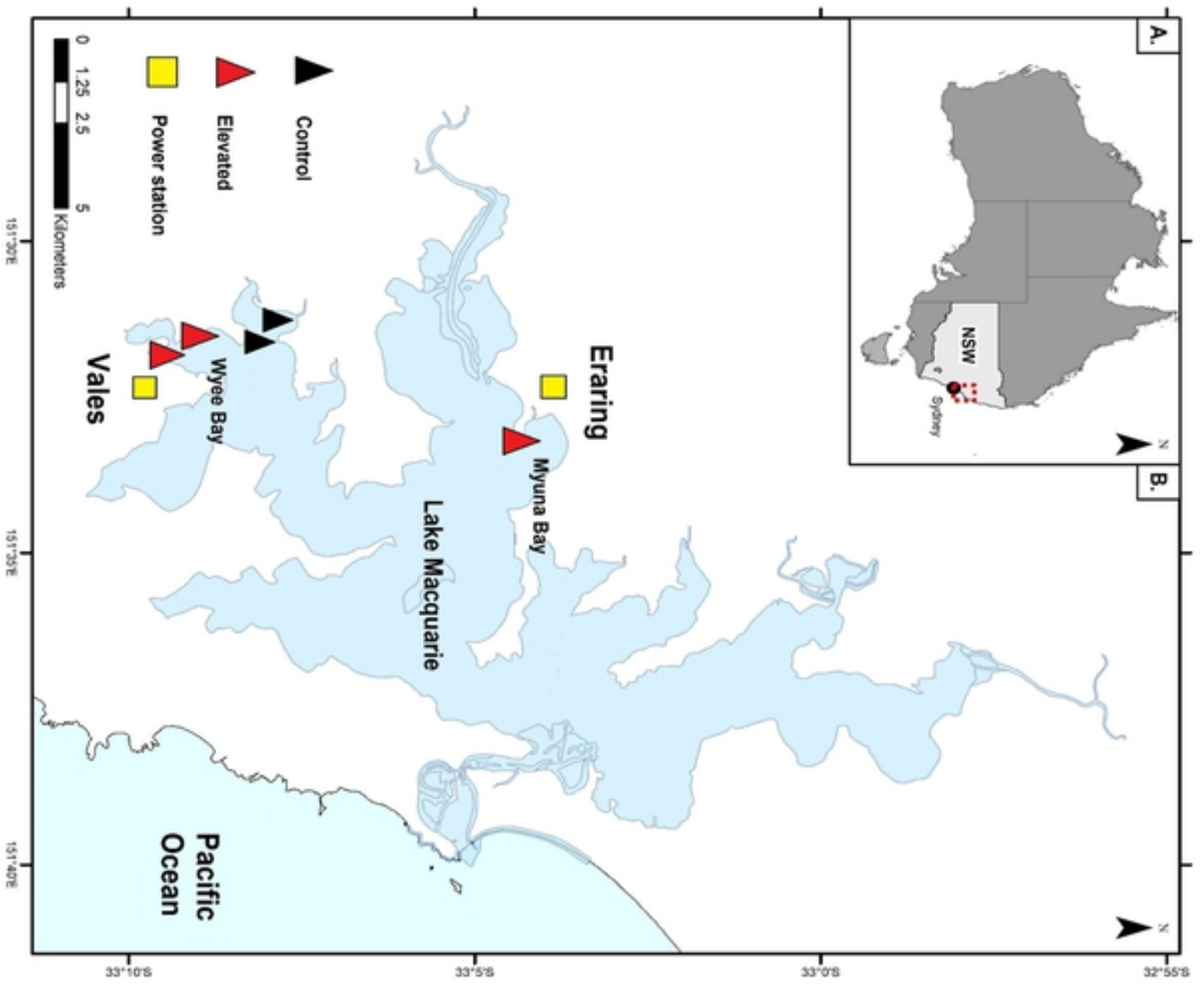


Figure 1

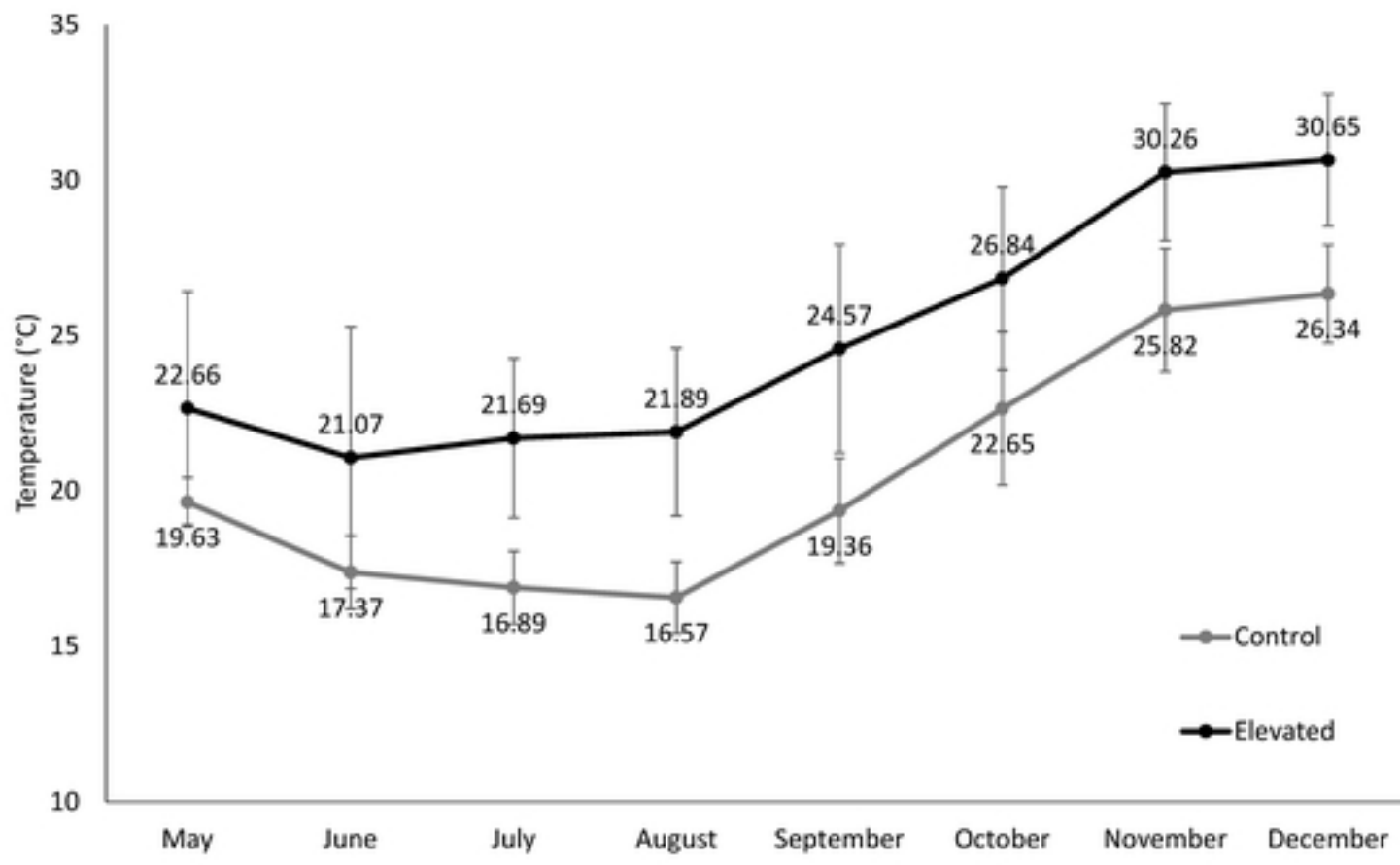


Figure 2

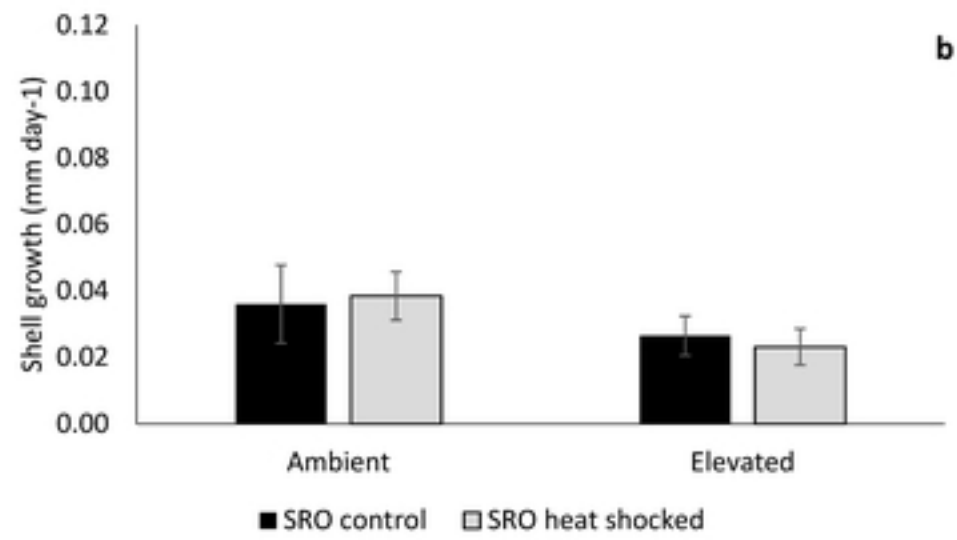
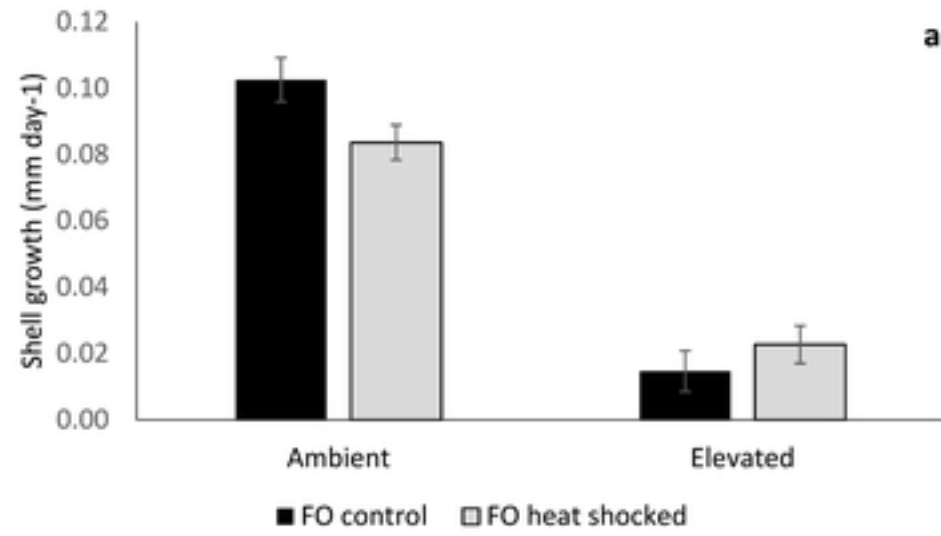


Figure 3

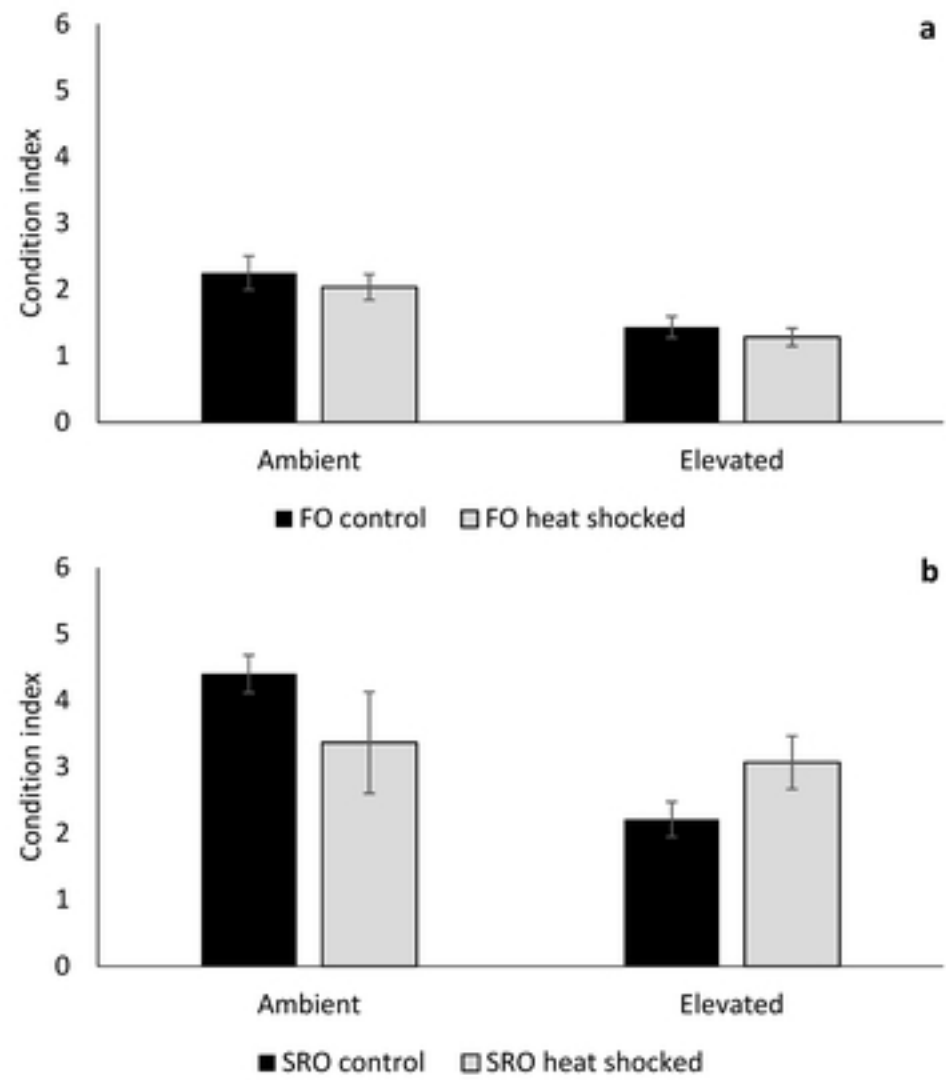


Figure 4

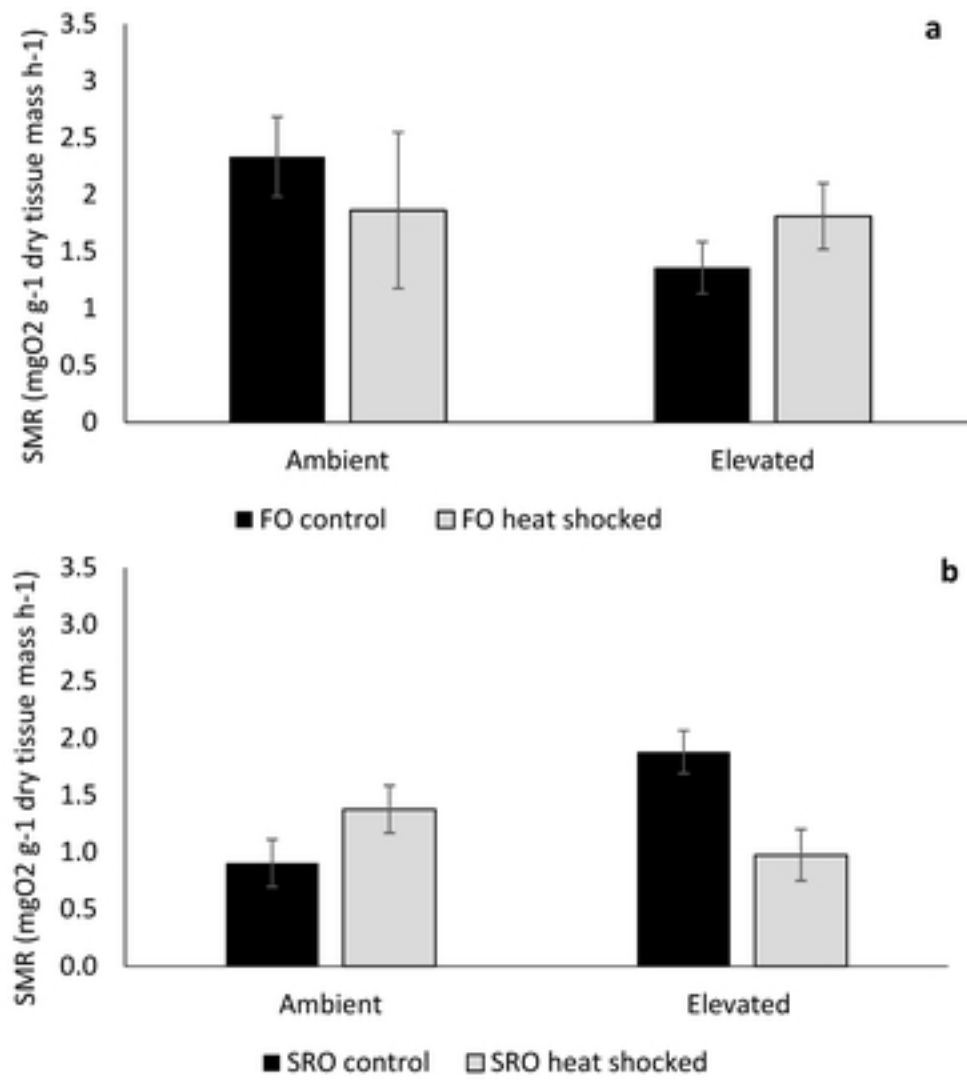


Figure 5

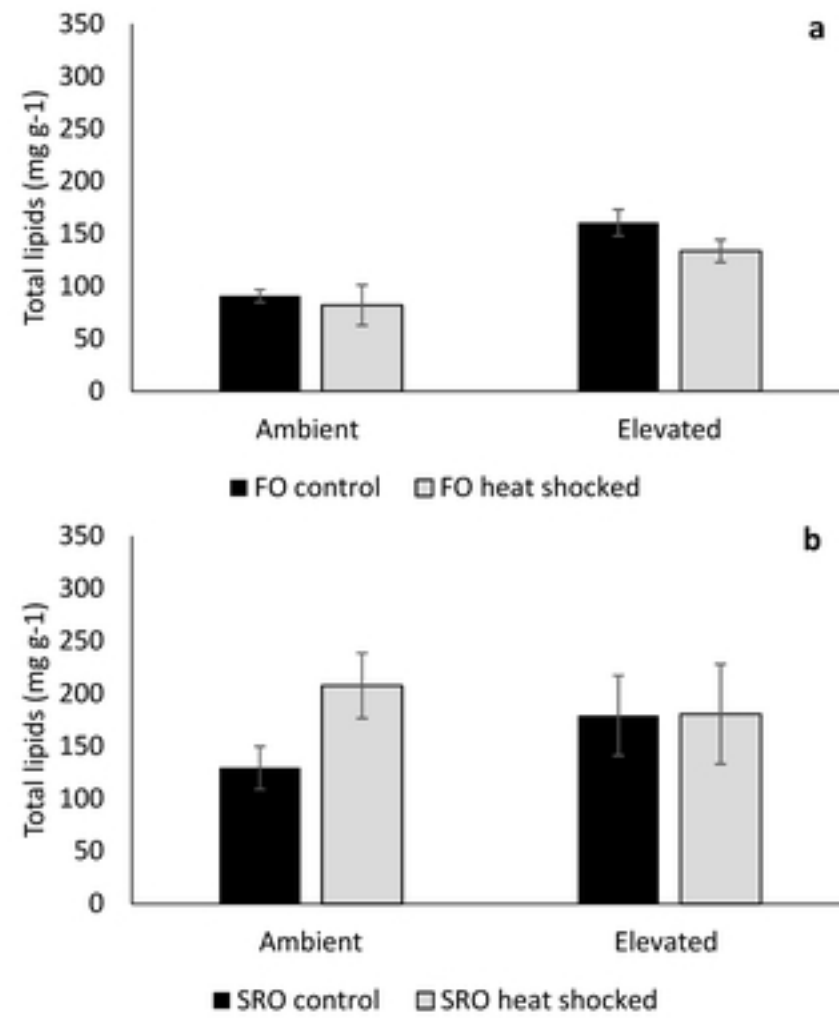


Figure 6

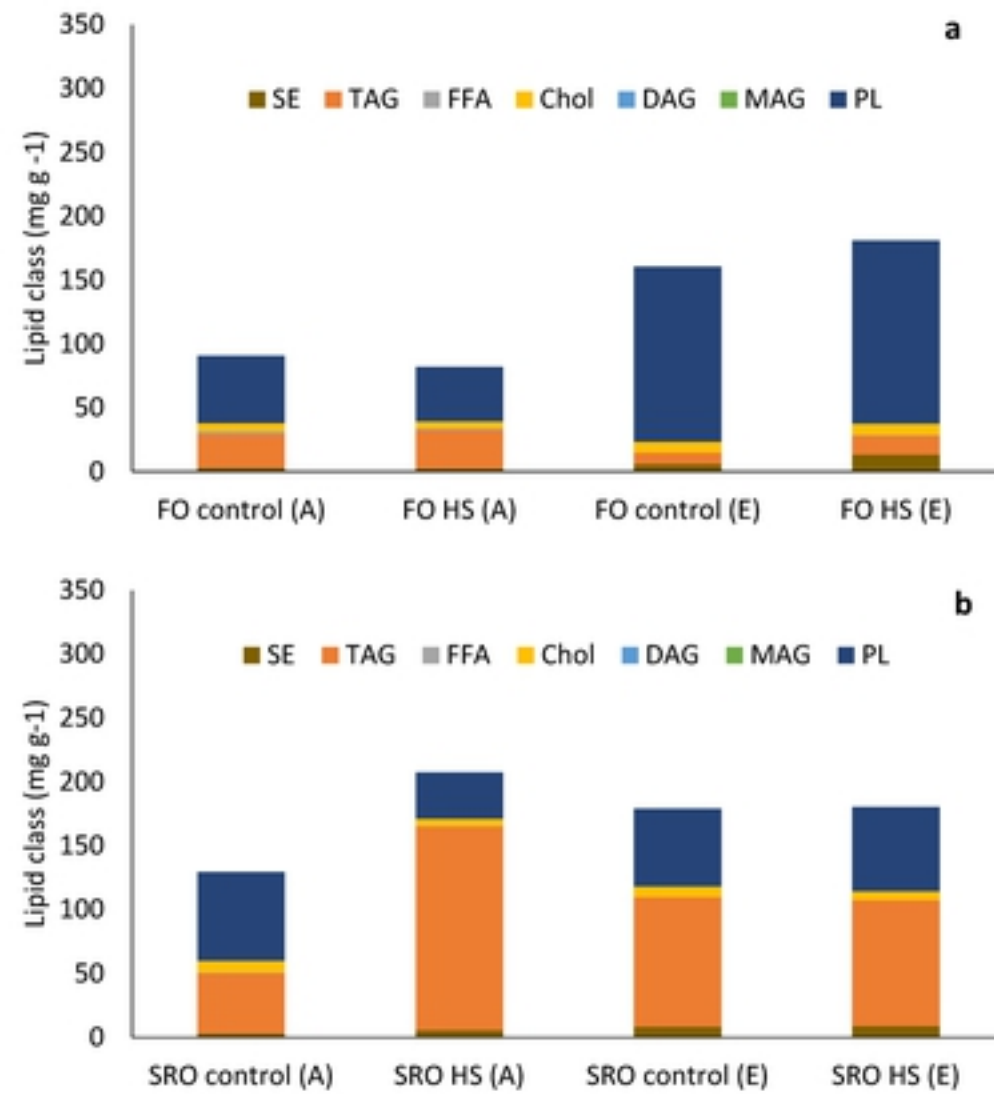


Figure 7

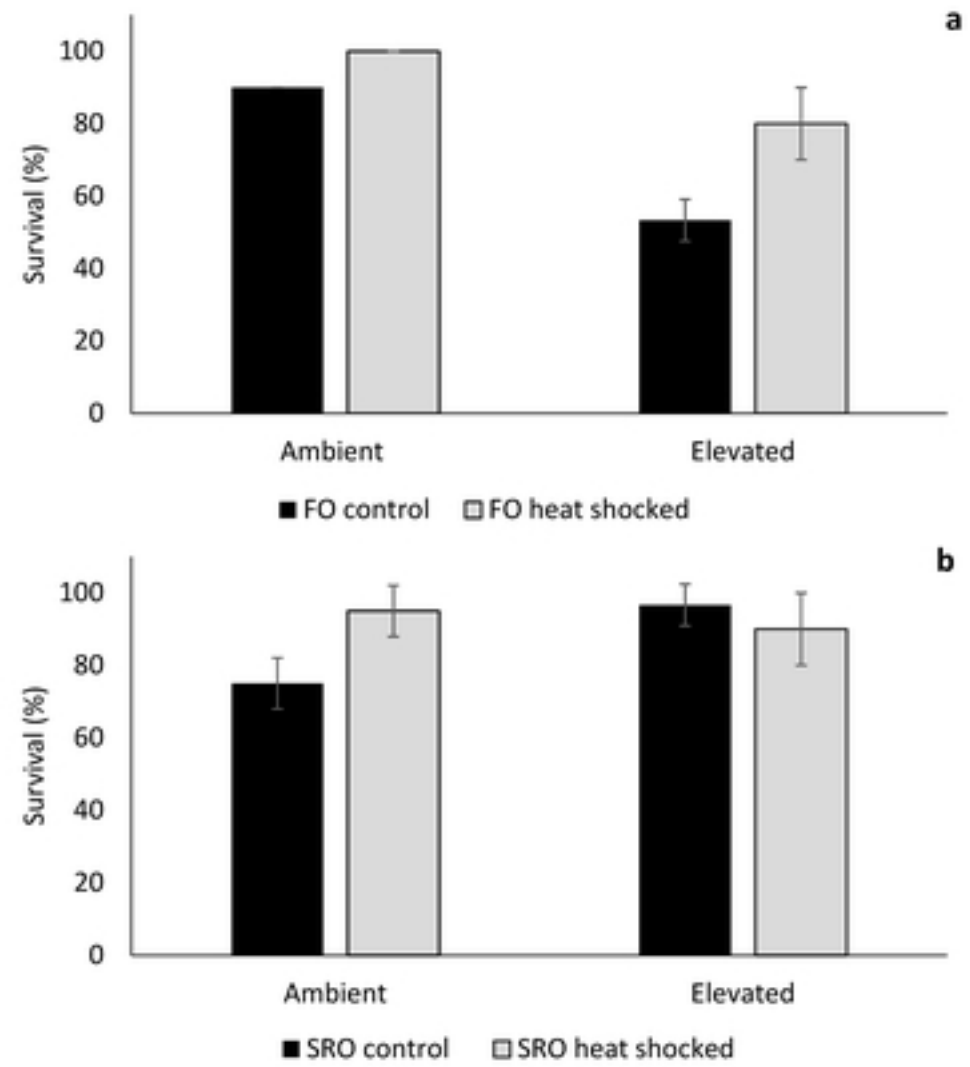


Figure 8