Running Head: DNA metabarcoding of degraded wolf scats

# Comparison of mechanical sorting and DNA metabarcoding for diet analysis with degraded wolf scats

Aimee Massey<sup>1</sup>, Gretchen Roffler<sup>2</sup>, Tessa Vermeul<sup>1</sup>, Jennifer Allen<sup>1</sup>, Taal Levi<sup>1</sup>

<sup>1</sup>Department of Fisheries and Wildlife, Oregon State University, Corvallis, OR 97331, USA <sup>2</sup>Alaska Department of Fish and Game, Division of Wildlife Conservation, P.O. Box 110024, Juneau, Alaska 99811 USA

Corresponding author email: <a href="mailto:aimeelmassey@gmail.com">aimeelmassey@gmail.com</a>, Tel: 207-314-7151

#### 1 Abstract

2 DNA metabarcoding has become a powerful technique for identifying species and 3 profiling biodiversity with the potential to improve efficiency, reveal rare prey species, and 4 correct mistaken identification error in diet studies. However, the extent to which molecular 5 approaches agree with traditional approaches is unknown for many species. Here, we compare diets from wolf scats profiled using both mechanical sorting and metabarcoding of amplified 6 7 vertebrate DNA sequences. Our objectives were: (1) compare findings from mechanical sorting 8 and metabarcoding as a method of diet profiling and (2) use results to better understand diets of 9 wolves on Prince of Wales Island, a population of conservation concern. We predicted metabarcoding would reveal both higher diversity of prey and identify rare species that are 10 11 overlooked with mechanical sorting. We also posited that the relative contribution of Sitka black-12 tailed deer (Odocoileus hemionus sitkensis) and beaver (Castor canadensis) would be 13 overestimated with mechanical sorting methods because of the failure to account for the full diet 14 diversity of these wolves. We found that there was substantial overlap in the diets revealed using 15 both methods, indicating that deer, beaver, and black bear (Ursus americanus) were the primary prey species. However, there was a large discrepancy in the occurrence of beaver in scats (54% 16 and 24% from mechanical sorting and metabarcoding, respectively) explained by the high rate of 17 18 false positives with mechanical sorting methods. Metabarcoding revealed more diet diversity 19 than mechanical sorting, thus supporting our initial predictions. Prince of Wales Island wolves appear to have a more diverse diet with greater occurrence of rare species than previously 20 21 described including 14 prey species that contribute to wolf diet. Metabarcoding is an effective 22 method for profiling carnivore diet and enhances our knowledge concerning the full diversity of 23 wolf diets, even in the extremely wet conditions of southeast Alaska, which can lead to DNA

24	degradation. Given the increasingly efficient and cost-effective nature of collecting eDNA, we
25	recommend incorporating these molecular methods into field-based projects to further examine
26	questions related to increased use of alternate prey coinciding with changes in abundance of
27	primary prey and habitat alteration.
28	
29	Keywords: Canis lupus, eDNA, noninvasive sampling, temperate rainforest, Prince of Wales
30	Island, Alaska, metabarcoding, diets, wolves, scats
31	
32	Introduction
33	Animal scats are a vital tool for answering scientific questions related to animal behavior,
34	diet, and species interactions. Traditionally, scat-based diet analysis has relied upon the
35	mechanical processing and sorting of scat remains. This typically includes processing a scat to
36	remove fecal material followed by meticulous sorting and identification of remaining hair and
37	hard parts. However, diet analysis with mechanical sorting of scats has well-known biases (Lake
38	et al. 2003); rare species or species that lack non-digestible hard parts are often overlooked or
39	misidentified. In addition, some species are challenging to distinguish based on bone fragments
40	or hair samples leading to additional misidentification. This is often the case for large mammals
41	that consume prey tissue rather than whole individuals such that diagnostic hard parts like teeth
42	and bones are frequently absent in scats. Metabarcoding of fecal DNA presents a new alternative
43	method for diet analysis (Shehzad et al. 2012, De Barba et al. 2014, Kartzinel et al. 2015,
44	McInnes et al. 2017, Eriksson et al. 2019). The DNA metabarcoding workflow includes
45	extracting DNA from environmental samples, DNA amplification using 'universal' primers
46	(Binladen et al. 2007), and mass-parallel sequencing of amplified product using next generation

47 sequencing technologies. This process allows DNA barcodes from multiple species in a bulk
48 sample to be sequenced simultaneously for an efficient and thorough profile of species present
49 within an environmental sample (Valentini et al. 2009).

50 The utility of metabarcoding for informing important management objectives, where 51 accurate taxonomic assignment and detection is paramount, is uncertain because unlike 52 mechanical sorting (1) it is unknown how quality of inference from metabarcoding depends on 53 acquiring relatively fresh scats with minimally degraded DNA, which can be challenging for rare 54 taxa, and (2) it is not yet clear if the relative read abundance from metabarcoding can yield 55 quantitative information that approximates the volume or biomass arising from each prev 56 species. The degree to which relative read abundance (RRA) from DNA metabarcoding is 57 correlated with the relative biomass of each prey species is a subject of substantial debate 58 (Deagle et al. 2018, Pinol et al. 2018). Limited empirical research validating RRA against 59 estimated biomass or volume from mechanical sorting informs this debate (Soininen et al. 2009, 60 Thomas et al. 2017), although no studies have done so with terrestrial carnivores. Pinol et al. 61 (2018) argued that metabarcoding results can only be interpreted quantitatively if amplification 62 of DNA through PCR with universal primers is avoided because different amplification 63 efficiencies among species can lead to poor representation of original biomass proportions. 64 While this is often true for invertebrates, for which primer mismatch is common (Krehenwinkel 65 et al. 2017), our 12S mtDNA primers rarely contains basepair mismatches for vertebrates and 66 contain no mismatches for the taxa considered here (Appendix S1: Fig. S1). In addition, recent 67 evidence suggests that as long as primer efficiency is high (no mismatches), the proportion of 68 sequences arising from each species in metabarcoding (RRA) can produce semi-quantitative 69 results (Kartzinel et al. 2015, Thomas et al. 2016, Krehenwinkel et al. 2017, Deagle et al. 2018).

This could allow metabarcoding to approximate relative biomass or volume information similar
to that produced by mechanical sorting of hard parts as well as frequency of occurrence
(proportion of scats that contain each species). Nevertheless, the degree to which degraded scats
yield suitable inference comparable to mechanical sorting is not currently well-understood
because of a lack of formal comparisons between metabarcoding and mechanical sorting (Deagle
et al. 2018, Pinol et al. 2018).

76 To provide this methods comparison, we focused on the Alexander Archipelago wolf 77 (*Canis lupus ligoni*) as a case-study. The Alexander Archipelago wolf has been repeatedly 78 petitioned for listing as threatened under the U.S. Endangered Species Act (ESA). These wolves 79 occur in relative geographic isolation in southeast Alaska, where continued pressure from habitat 80 loss, population decline of their primary prey, and wolf harvesting have raised concern about the 81 future of the population. Wolf population estimates at regional scales in southeast Alaska have 82 been based on expected Sitka black-tailed deer (Odocoileus hemionus sitkensis) abundance under 83 the assumption that wolves are closely tied to the abundance of their primary prey. This is 84 evident in the most recent ESA species status review where deer habitat quality metrics were 85 used to project wolf abundance (U.S. Fish and Wildlife Service 2015). 86 The wolves on Prince of Wales Island (POW) (Fig. 1) were of particular concern in the 87 most recent assessment (2015) because in addition to high levels of wolf harvest, POW has the 88 highest rate of old-growth logging in southeast Alaska (Albert and Schoen 2013, Person and 89 Brinkman 2017). Deer populations are predicted to decline as old-growth forests with palatable 90 understory forbs and shrubs are converted into dense, even-aged, closed canopy forests (Alaback 91 1982, Schoen et al. 1988, Person et al. 1996, Farmer and Kirchhoff 2007, Gilbert et al. 2016,

92 Person and Brinkman 2017, Porter 2018) that are strongly avoided by deer (Wallmo and Schoen

93 1980, Kohira and Rexstad 1997, Gilbert et al. 2017). Deer were the most frequently occurring 94 prey species for the Alexander Archipelago wolf based on previous research conducted on POW 95 (Kohira 1995, Person et al. 1996, Kohira and Rexstad 1997). However, mechanical sorting of 96 wolf scats has revealed other prey in significant quantities (Kohira 1995), and coastal wolves in 97 this region can consume substantial quantities of salmon seasonally and other marine resources 98 (Szepanski et al. 1999, Darimont et al. 2003, 2004, 2008a, Lafferty et al. 2014), suggesting that 99 wolf population abundance may also be dictated by the availability of prey other than deer. 100 Consequently, refining knowledge regarding the diet of wolves in the system has important 101 implications for wolf management, potential ESA considerations, and forest management in 102 southeast Alaska. 103 Here we provide the first formal comparison of carnivore diet analysis from mechanical 104 sorting and DNA metabarcoding using opportunistically collected scats across an assumed 105 degradation spectrum in a temperate rainforest which is hostile to DNA preservation. We 106 examined whether metabarcoding revealed a more diverse wolf diet than mechanical sorting, 107 achieved increased taxonomic precision, and identified infrequently consumed prey species. We 108 included both scats appearing highly degraded and those appearing fresh and assessed whether 109 age of scats or biases introduced during the molecular processing affected the diet profile shown 110 by metabarcoding. We additionally analyzed in detail Alexander Archipelago wolf diets with a 111 particular focus on Prince of Wales Island to determine the prey profile of wolves and their 112 dependence on deer.

113

114 Materials and Methods

#### 115 <u>Study area and field collection</u>

116	Southeast Alaska lies within the Alexander Archipelago composed of over 2,000 named							
117	islands (Fig. 1) (Cook et al. 2006). This region receives between 130 – 400 cm of precipitation							
118	annually (Shanley et al. 2015) thus making it particularly inhospitable to the preservation of							
119	DNA in exposed environmental samples. The mainland is buttressed by the rugged Coast							
120	Mountains and extensive temperate rainforests at lower elevations. As a result of natural							
121	fragmentation and isolation, the North Pacific coast region supports many endemic plant and							
122	animal lineages, particularly on Prince of Wales Island, the largest island in the archipelago							
123	(Cook et al. 2006, MacDonald and Cook 2007, Smith 2016). Most of the forested area is within							
124	the Tongass National Forest managed by the U.S. Forest Service. This ecosystem hosts a							
125	diversity of mammals including iconic species such as Sitka black-tailed deer (Odocoileus							
126	hemionus sitkensis), American black bear (Ursus americanus), North American beaver (Castor							
127	canadensis), American marten (Martes americanus), mountain goat (Oreamnos americanus),							
128	Steller sea lion (Eumetopias jubatus), harbor seal (Phoca vitulina), and moose (Alces alces).							
129	Species distribution and assemblages vary among island and mainland areas of this region.							
130	We collected wolf scats along wolf travel routes, near den sites, and on secondary roads							
131	during planned scat collection surveys during October 2014 - December 2015 (Fig. 1). We							
132	collected wolf scats primarily on Prince of Wales Island (55° 46'45.9480" N; 132° 49' 4.7748"							
133	W) ( $n = 145$ ), but also opportunistically collected samples in other mainland and island systems							
134	(n = 38). We estimated the age (fresh [<3 months] and old [>3 months]) of scat based on							
135	appearance, time since last site visit (Ciucci et al. 1996), and exposure time considering that scats							
136	decompose rapidly in rainforest environments (Wallmo et al. 1962, Ciucci et al. 1996, Darimont							
137	et al. 2008b) (Fig. 2). Collected wolf scats were stored in plastic bags, labeled with location,							

date, and perceived age of scat, and then frozen (-20°C). Frozen scats were shipped to Oregon
State University for sample preparation and analysis.

140

#### 141 <u>Mechanical sorting</u>

142 We stored a subsample of each scat for later molecular analysis (sterilized forceps and 143 razors were used to collect a sample from the middle section of each scat to minimize wolf DNA 144 (Stenglein et al. 2010)), and then placed each scat in a mesh bag (1/8") and soaked it in water for 145 48 hours in a mason jar. We power-washed the scat to remove as much remaining fecal matter as 146 possible. The remaining contents (i.e., hair, bones, other hard parts) were put in a labeled paper 147 bag and dried in an oven (at approximately 50°C) for at least five days. We weighed the 148 processed scat material (hair, bone, scale, feather, etc.) and mechanically homogenized and 149 sorted the remains by hand. On average, the fine-scale sorting took 3.6 hours per scat. We 150 examined hairs under a microscope and compared to hair samples from the Alaska Fur ID 151 project (Carrlee and Horelick 2011). We made slide mounts using clear nail polish to examine 152 scale pattern and medulla diameter in order to identify species. Following identification, the slide 153 was labeled with the species name and sample of origin. This exhaustive, fine-scale sorting (Fig. 154 3) ensured that even rare species could be identified. Along with species identification, we 155 estimated the volume of each prey species as a proportion of estimated hard parts for a species in 156 relation to all hard parts in an individual scat.

157

158 <u>Molecular analysis</u>

Using the stored subsamples from each scat, we extracted DNA from each sample using
the Qiagen DNeasy Blood and Tissue kit (Qiagen, Hilden, Germany) with slight modifications as

161	follows: 500 ul Buffer ATL, 50 ul Proteinase K, and 1.0 mm Zirconia/Silica beads (BioSpec
162	Products, Bartlesville, OK) were added to the 1.7 ml tube containing the scat. Samples were
163	vortexed for 10 minutes at maximum speed prior to incubation at 56°C for 4-6 hours. The DNA
164	was eluted in a total volume of 100 ul. A negative control was extracted with each round
165	(approximately 17 samples) of DNA extraction to identify possible cross contamination.
166	Following DNA extraction, each sample was amplified in three separate reactions using
167	the primer pair 12SV5F/12SV5R (Riaz et al. 2011). We used the forward primer
168	(TTAGATACCCCACTATGC) as Riaz et al. (2011) but modified the first base pair of the
169	reverse primer (YAGAACAGGCTCCTCTAG) to allow broader binding of vertebrate targets.
170	These primers target approximately 100 base pairs in the 12S region of the vertebrate
171	mitochondrial genome. The initial PCR was carried out using AmpliTaq Gold 360 Master Mix
172	(Life Technologies, Carlsbad, CA). To label samples for multiplexing, we used 384 unique 8 bp
173	dual matching indexes on the forward and reverse primers to eliminate contamination due to tag
174	jumping by filtering reads that did not have identical indexes, and we included 3 bp of random
175	nucleotides on the 5'end to increase sequence diversity and prevent degradation of indexes
176	during subsequent blunt-ending and ligation steps. PCR reactions were carried out in a volume of
177	20 ul with 10 ul AmpliTaq Gold 360 Master Mix for a final concentration of 1x, 5 ul of forward
178	and reverse primers for a final concentration of 0.25 uM, 3 ul of water, and 2 ul of DNA
179	template. PCR cycling included initial denaturing at 95°C for 10 minutes, followed by 40 cycles
180	of 95°C for 30 seconds, 58°C for 30 seconds, and 72°C for 30 seconds, with a final extension at
181	72°C for 7 minutes.
182	After the initial PCR all PCR amplicons were cleaned using PCRClean DX solid-phase

182 After the initial PCR, all PCR amplicons were cleaned using PCRClean DX solid-phase
183 reversible immobilization magnetic beads (Aline Biosciences, Woburn, MA). Each PCR reaction

184	was quantified using Accublue High Sensitivity dsDNA Quantitation kit (Biotium, Fremont, CA)						
185	and normalized to 6 ng/ul. Each group of 384 PCR products was then pooled into a single						
186	library, for a total of 3 libraries. Individual libraries were then tagged with an additional 6 base						
187	pair identifying index using the NEBnext Ultra II DNA Library Prep kit (New England Biolabs,						
188	Ipswich, MA). Pooled samples were analyzed on a Bioanalyzer to confirm fragment size. The						
189	libraries were then sequenced on one lane of Illumina HiSeq 3000 2 x 150 bp PE at the Center						
190	for Genome Research and Biocomputing at Oregon State University.						
191							
192	<u>Sequence analysis</u>						
193	Raw sequence reads were analyzed using a bioinformatics pipeline designed to trim and						

194 sort the sequence reads according to scat sample identification. An outline of the bioinformatic 195 process is as follows: (1) raw reads were paired using PEAR (Zhang et al. 2013); (2) followed by 196 demultiplexing using 8 basepair index sequences unique to each sample (mismatches discarded); 197 (3) lastly, sequences from each sample were clustered by 100% similarity and taxonomically 198 assigned using BLAST against 12S vertebrate sequences in GenBank and from a custom 12S 199 database.

Similar to the step-wise methods used by De Barba et al. (2014), a series of filtering and quality control measures were carried out on taxonomically assigned sequences. We initially removed sequences that were identified to be *Canis* spp. and contaminants based on read counts in no-template controls (which contained primarily human contamination). We then removed sample replicates that failed to amplify during PCR which included sample replicates with fewer than a total of 400 sequence reads. We compared taxonomic assignments with known fauna of southeast Alaska (MacDonald and Cook 2007) to replace non-regional species identified with

BLAST with closely-related regional taxa. We then excluded prey items occurring in fewer than 207 208 2 of 3 PCR replicates. Finally, we combined those sample replicates that amplified so that 209 sequence reads were totaled for each species within a sample and over the entire sample and 210 eliminated sequences that appeared in less than 1% of the total reads for an individual sample. 211 212 Age of scats 213 Prior to processing, we observed marked differences between the appearance and quality 214 of scats (Fig. 2). We performed t-tests to determine whether the perceived age of a scat made 215 during field collection correlated with either the average quantity of DNA (ng/ul) in a sample 216 (measured post normalization using Accublue High Sensitivity dsDNA Quantitation kit 217 (Biotium, Fremont, CA)), the total number of sequence reads in a sample including the wolf 218 defecator, or the total number of sequence reads excluding wolf. 219 220 Frequency of occurrence 221 We used both frequency of occurrence (FOO) and metrics of relative abundance (see 222 below) to describe the occurrence of prey in wolf diet. FOO was calculated to determine which 223 prey species were present and how often they were present based on the number of samples. For 224 mechanical sorting methods, a species was present if there was evidence (including trace 225 elements) of a prey species (e.g., hair, bone, scales, etc.) within a scat sample. FOO was then 226 calculated as the proportion of scats in which a prey species occurred. For metabarcoding, a 227 species occurrence was determined by whether sequence reads for a particular species were 228 found in an individual scat after quality control measures. We compared FOO from mechanical 229 sorting and metabarcoding using the subset of scats analyzed by both methods (n = 104), but we

additionally present diet analysis from all scats collected on Princes of Wales Island and close
surrounding islands (n = 118 metabarcoding; n = 98 mechanical sorting) to describe diet on
POW.

233 To analyze discrepancies between species present in samples with mechanical sorting and 234 not found with metabarcoding, we used generalized logistic regression with logit link to explore 235 whether false positives from mechanical sorting or false negatives generated from metabarcoding 236 best explained the absence of species. Statistical analyses were conducted in the R statistical 237 program using the 'stats' package (R Core Team 2018). We reasoned that false negatives could 238 arise if scats contained poor quality DNA or sequencing depth was insufficient. We therefore fit 239 three separate logistic regression models using average DNA quantity per sample (across the 240 three replicates PCRs), total number of sequence reads prior to quality control and including 241 wolf sequence reads, and total number of sequences reads post quality control and not including wolf sequences reads as univariate predictors in each model. In our analysis, zeroes were defined 242 243 as an absence in metabarcoding where mechanical sorting had indicated an occurrence of a 244 particular species in a sample; one indicated where metabarcoding was in agreement with 245 occurrence found in mechanical sorting. Therefore, positive coefficients imply an increasing rate 246 of proper assignment as DNA quality or sequencing depth increases. The absence of such an 247 effect would suggest that mismatch between metabarcoding and mechanical sorting is unlikely to 248 be due to false negatives by metabarcoding.

249

250 <u>Relative abundance</u>

To test whether metabarcoding and mechanical sorting yield similar metrics for relative abundance of a prey species within a scat, we compared percent estimated volume from

mechanical sorting with the relative read abundance (RRA) from metabarcoding. RRA for each
species *i* was calculated as

255

256 
$$RRA_{i} = \frac{1}{s} \sum_{k=1}^{S} \frac{n_{i,k}}{\sum_{i=1}^{T} n_{i,k}}$$
[1]

257

where  $n_{i,k}$  is the number of sequences of prey species *i* in sample *k*, *S* is the total number of samples, and *T* is the total number of species. We compared estimated volume of a prey species from mechanical sorting with RRA from metabarcoding using simple linear regression (R Core Team 2018).

For both the frequency of occurrence and relative abundance analyses we additionally revisited results from scats with mismatches from metabarcoding and mechanical sorting to assess whether metabarcoding found many sequence reads of an alternative species that was incorrectly assigned by mechanical sorting and was thus likely a false positive.

266

#### 267 **Results**

268 <u>Age of scats</u>

Purportedly fresh scats contained significantly more total sequence reads on average  $(\mu_{\text{fresh}} = 269,514 \pm 173,902)$  compared to the total number of reads from degraded wolf scats  $(\mu_{\text{degraded}} = 200,378 \pm 135,646)$  (t = 2.09, df = 85, p-value = 0.039). Likewise, fresh scats ( $\mu_{\text{fresh}} =$ 139,939 ± 135,858) had significantly more wolf sequence reads than degraded scats ( $\mu_{\text{degraded}} =$ 52,411 ± 75,531) (t = 3.80, df = 73.73, p-value < 0.001). However, we found no significant difference between degraded and fresh scats when considering only reads from prey items (excluding any wolf DNA reads), although degraded scats yielded a greater average number of

276 non-wolf reads per sample than fresh scats ( $\mu_{degraded} = 147,966 \pm 125,223$ ;  $\mu_{fresh} = 129,575 \pm$ 277 124,124; t = -0.69, df = 84.18, p-value = 0.49) (Fig. 4). Fresh scats had a higher average DNA 278 quantity post PCR (ng/ul;  $\mu_{degraded} = 4.12 \pm 1.97$ ;  $\mu_{fresh} = 4.55 \pm 2.20$ ) but the difference was not 279 statistically significant (t = 0.97, df = 85.93, p-value = 0.33). 280 *Comparing wolf diet by mechanical sorting and metabarcoding – frequency of occurrence* 281 282 We compared wolf diet from 104 scat samples that were analyzed with both mechanical 283 sorting and metabarcoding. Metabarcoding revealed a number of rare species that were not found 284 using mechanical sorting methods and thus revealed greater dietary diversity (Fig. 5). Species 285 that were found with metabarcoding methods but were absent when using mechanical sorting 286 methods include: duck (Anas spp.), dusky grouse (Bonasa umbellus), elk (Cervus elaphus), raven 287 (Corvus species), Northern collared lemming (Dicrostonyx groenlandicus), Steller sea lion 288 (Eumetopias jubatus), American marten (Martes americana), and American red squirrel 289 (Tamiasciurus hudsonicus). Mechanical methods identified moose (Alces alces) in a single scat 290 where metabarcoding did not, although moose was identified by metabarcoding in this particular 291 scat prior to quality filtering. 292 Frequency of occurrence (FOO) (Fig. 5) results were similar with both methods. 293 However, there was substantial discrepancy between the primary prey species (Sitka black-tailed 294 deer) and the secondary prey species (beaver) between metabarcoding and mechanical 295 occurrence results. The occurrence of deer was greater in the mechanical sorting results 296  $(FOO_{mech} = 0.962)$  compared to metabarcoding results  $(FOO_{MB} = 0.8)$  and the occurrence of beaver was twice as frequent in the mechanical sorting (FOO<sub>mech</sub> = 0.519) results compared to 297

298 metabarcoding (FOO<sub>MB</sub> = 0.236).

299 Logistic regression to assess mismatch between metabarcoding and mechanical sorting 300 revealed that neither average DNA quantity, total sequence reads, nor total sequence reads of 301 prey (i.e. excluding wolf) were associated with failing to detect species that were identified by 302 metabarcoding (Table 1). However, contrary to predictions, increasing number of prey sequence 303 reads (i.e. excluding wolves) was associated with increasing mismatch with beaver occurrences 304 detected by mechanical sorting (p = 0.025), which suggests that the error was due to 305 misassignment by mechanical sorting rather than by metabarcoding. Thirty-two of the 59 beaver 306 occurrences had disagreement between mechanical sorting and metabarcoding results. Notes and 307 hair slides taken during mechanical sorting showed that 18 of the 32 mismatches could be 308 attributed to false positives generated from mechanical sorting. In addition, a substantial number 309 of definitive deer occurrences (i.e. high relative read abundance for deer) were mistakenly assigned to beaver by mechanical sorting (Fig. 6), further suggesting that mismatch between 310 311 methods was due to misassignment by mechanical sorting.

312

#### 313 <u>*Comparing wolf diet by mechanical sorting and metabarcoding – relative read abundance*</u>

There was minimal discrepancy between RRA of primary prey species (metabarcoding) and their estimated volume in scats (mechanical sorting); the difference between RRA and estimated volume for deer was 2% (RRA<sub>deer</sub> = 68.3%; estimated volume<sub>deer</sub> = 66.3%) and for beaver it was less than 7% (RRA<sub>beaver</sub> = 14.1%; estimated volume<sub>beaver</sub> = 20.5%). For the rarer species, we found a close association (within 2%) between the RRA and the estimated volume for that species.

320 The estimated volume from mechanical sorting was positively correlated with RRA of 321 deer ( $\beta = 0.53$ ; R<sup>2</sup> = 0.26; p < 0.01, n = 87), beaver ( $\beta = 0.57$ ; R<sup>2</sup> = 0.28; p < 0.01, n = 25), and

322	black bear ( $\beta = 0.80$ ; R <sup>2</sup> = 0.28; $p = 0.17$ , n = 6) (Fig. 6), supporting a positive but variable
323	relationship between the volume of parts of a particular species found in the physical scat and the
324	proportion of DNA sequence reads for that species. However, substantial variability is likely due
325	to species misidentification by mechanical sorting such as deer falsely identified as beaver (Fig.
326	6).
327	
328	<u>Prince of Wales</u>
329	Metabarcoding of scats found only within Prince of Wales Island (POW) (Fig. 7)
330	revealed 14 species (Supplementary table) including Sitka black-tailed deer (FOO <sub>MB_POW</sub> =
331	0.852), beaver (FOO <sub>MB_POW</sub> = 0.231), and black bear (FOO <sub>MB_POW</sub> = 0.157) were the most
332	common prey items (Fig. 7). Other common prey species were salmon (Oncorhynchus spp.)
333	(FOO <sub>MB_POW</sub> = 0.056), American marten ( <i>Martes americana</i> ) (FOO <sub>MB_POW</sub> = 0.046), North
334	American river otter ( <i>Lontra canadensis</i> ) (FOO <sub>MB_POW</sub> = 0.037), and bald eagle ( <i>Haliaeetus</i>
335	<i>leucocephalus</i> ) (FOO <sub>MB_POW</sub> = 0.019). Additional prey items in less than 1% of scats include
336	American red squirrel (Tamiasciurus hudsonicus), deermouse (Peromyscus spp.), vole (Myodes
337	and Microtus spp.), dusky grouse (Bonasa umbellus), duck (Anas spp.), and unidentified bird
338	species.
339	Mechanical sorting revealed a total of 10 prey species (Fig. 7), including harbor seal

which was not found with metabarcoding for the POW samples. However, it should be noted that for this sample, mechanical sorting estimated only 2% harbor seal and metabarcoding instead found otter, which could have been mistaken for harbor seal during sorting. Deer (FOO<sub>mech\_POW</sub> = 0.969) and beaver (FOO<sub>mech\_POW</sub> = 0.561) (the two primary prey species) showed greater FOO compared to metabarcoding, although mechanical sorting did not show any American marten

and had a lower FOO of salmon species compared to the metabarcoding results. There was also substantial occurrence of material from unknown species in the mechanical results (FOO<sub>mech\_POW</sub> = 0.163) that is not seen with metabarcoding.

348

#### 349 Discussion

350 DNA metabarcoding has emerged as a novel method for diet analysis because of the 351 ability to reveal rare or difficult to identify species (Shehzad et al. 2012, De Barba et al. 2014, 352 Berry et al. 2015, Srivathsan et al. 2015, Kartzinel et al. 2015, McInnes et al. 2017, Buglione et 353 al. 2018). However, substantial uncertainty remains as to whether inference from mechanical 354 sorting and DNA metabarcoding produce comparable results, particularly if scats are of 355 uncertain age and quality. Our results suggest that excluding purportedly degraded scats from 356 DNA metabarcoding analyses does not improve inference about diet. Perceived fresh scats 357 contained on average a greater number of reads per scat when including wolf sequence reads, but 358 there was no significant difference in the average number of reads between fresh and degraded 359 scats when only including reads from prey species (Fig. 4). The average quantity of DNA was 360 also not significantly different between fresh and degraded scats; this is likely because fresh scats 361 contained more fecal material relative to hair and bone, and total DNA quantity per sample is 362 normalized prior to sequencing such that abundant wolf DNA leads to dilution of prev DNA. 363 Many degraded scats were primarily clusters of hair and bone that were washed of fecal material. 364 Importantly, these results suggest that metabarcoding is sensitive enough to determine prey 365 assemblages in degraded scats and thus scat collection and processing should not be predicated 366 upon perceived scat quality.

367	FOO and RRA metrics were qualitatively similar among methods. RRA of each species
368	was significantly correlated with estimated volume determined with mechanical sorting ( $p_{all} <$
369	0.01, $p_{deer} < 0.01$ , $p_{beaver} < 0.01$ , $p_{bear} = 0.16$ ) suggesting that RRA can be a reasonable proxy for
370	volume of prey species obtained from mechanical sorting (which in turn could be used to
371	estimate relative biomass using biomass equations that correct for body size (Weaver 1993)).
372	Both mechanical sorting and metabarcoding agreed that Sitka black-tailed deer was the primary
373	prey item, followed by beaver, and then black bear as suggested by previous research in this
374	region (Kohira, 1995; Kohira & Rexstad, 1997) (Fig. 5). However, both deer and beaver
375	occurred substantially more frequently in mechanically sorted scats than in metabarcoded scats.
376	The divergence between the two methods examined in our study was more substantial for beaver
377	which were identified mechanically in 52% of scats while only detected by metabarcoding in
378	24%.
379	We closely examined scats that were mismatched (i.e. the prey species was found in a
380	scat during mechanical sorting but not found in the same scat with DNA metabarcoding) with a

380 scat during mechanical sorting but not found in the same scat with DNA metabarcoding) with a 381 focus on beaver to assess whether mismatches were due to false positives produced from 382 mechanical sorting or false negatives produced from metabarcoding. Eighteen of the 32 383 mismatched samples show evidence of false positive resulting from mechanical sorting. In these 384 scats, beaver was thought to be present, but notes during sorting specified uncertainty that these 385 small amounts of unknown hair samples could also be attributed to deer or black bear. In fact, we 386 found that in all mismatched beaver samples metabarcoding showed a high RRA of deer and 387 mechanical sorting found low volume of deer, strongly suggesting that mechanical sorting mis-388 assigned deer hair to beaver as the primary prey species in that scat (highlighted in Fig. 6). Our

logistic regression analysis additionally suggests that these errors resulted from mis-assignmentby mechanical sorting rather than metabarcoding (Table 1).

391 Why do we see these potential false positives generated from mechanical sorting? One 392 explanation is that relying on mechanical sorting of scats results in the overestimation of primary 393 prey species (i.e. deer and beaver) due to search image bias. Mechanical sorting can lead to 394 mislabeling difficult to identify parts as common species rather than an infrequently detected 395 species because the researcher is accustomed to encountering the common prey species. The 396 pronounced difference seen in beaver FOO could also be attributed to the difficulty in 397 distinguishing between beaver and guard hair from other species such as deer and black bear 398 (Fig. 8).

399 The remaining 14 of the 32 beaver mismatches were attributed to false negatives 400 generated by metabarcoding; we concluded this because beaver was verified to have occurred in 401 mechanical sorting but was absent from metabarcoding results. However, for 10 of these scats 402 beaver occurred in the metabarcoding results prior to quality filtering that removed prey that 403 occurred in fewer than 2 of 3 PCR replicates and with fewer that 1% of the total reads 404 (importantly, beaver was nearly absent from our negative controls), which had the effect of 405 underestimating prey items that occurred in only a small portion of a scat. It is important to note 406 that our conservative quality filtering thresholds following De Barba et al. (2014) led to some of 407 these false negatives at the expense of false positives. Thus, it is imperative to explicitly reason 408 through quality control protocols to balance false positives and false negatives when using 409 bioinformatically-generated metabarcoding data.

There was also divergence in the detection of rare species among methods. Although
metabarcoding revealed several clear false negatives, this was substantially more common with

412	mechanical sorting where 8 species in final metabarcoding results were not found by mechanical
413	sorting for the same subset of samples. In particular, American marten, Northern collared
414	lemming, and a number of bird species were missing from mechanical sorting but evident in the
415	metabarcoding results. This conclusion supports our initial prediction that metabarcoding would
416	be more advantageous in identifying rare species.
417	
418	<u>POW wolf diet – policy and management</u>
419	The issue of what wolves eat and how much is an important question in southeast Alaska
420	and in particular on Prince of Wales Island where there are concerns about the long-term
421	viability of wolves given the trophic linkage between wolves, Sitka black-tailed deer, and old-
422	growth forest. The population of Sitka black-tailed deer is expected to decline in this region with
423	continued logging of old-growth forests (Person and Brinkman 2017). Given this, wolf
424	populations are predicted to decline and these declines are most significant under scenarios
425	where wolves rely heavily on deer in the future (Gilbert et al. 2016).
426	Our study shows the promise of eDNA and metabarcoding methods to examine wolf diet
427	diversity and diet changes. Comparing our results with previous work indicated that the
428	occurrence of the primary prey species (Sitka black-tailed deer) is comparable on POW; Person
429	et al. (1996) reported a >90% occurrence while we report 85.2% occurrence using DNA
430	metabarcoding and 96.9% occurrence using mechanical sorting. However, the occurrence of
431	beaver is greater compared to previous work; the frequency of occurrence of beaver was 13.7%
432	(Person et al. 1996) and 31% (Kohira and Rexstad 1997), whereas we report 23.1% occurrence
433	using DNA metabarcoding and 56.1% using mechanical sorting. These previous studies found
434	that aside from Sitka black-tailed deer, beaver, and black bear, the only significant other prey

435 were small mustelid species, river otter, and fish. Our results show a diverse diet with 14 total 436 prey species identified from mechanical sorting that contribute to wolf diet on POW (Fig. 7), 437 which more closely resembles the diversity found by Darimont et al. (2004) in their study of 438 wolf diet using scats along the coastal region in British Columbia. Importantly, our findings 439 suggest that metabarcoding was able to reveal the breadth of Alexander Archipelago wolf diet 440 diversity more accurately than mechanical sorting. (24 vs. 14 refer to Appendix S1: Table S1). 441 Continued diet analysis using metabarcoding of wolf scats found on POW could reveal 442 whether this increase in diversity is due to the increased power in the method used 443 (metabarcoding vs. mechanical sorting), or if wolves are beginning to exhibit increased 444 opportunistic predation on species other than Sitka black-tailed deer. Given that we also found 445 greater dietary diversity using mechanical sorting compared to results using the same methods 446 from the mid-1990's points towards a potential dietary shift in wolves on POW (Kohira and 447 Rexstad 1997). The rate of clear-cut logging in this region peaked during the late 1980's and 448 1990's and while this rate has slowed in recent years, a total of nearly 30% of old-growth forests 449 have been logged on POW (U.S. Fish and Wildlife Service 2015). Because young-growth stands 450 older than 25 years are the least productive in terms of deer forage (U.S. Fish and Wildlife 451 Service 2015), the effects of potential deer abundance decline on wolf populations are only just 452 being realized. As well-known diet generalists, it remains to be seen whether wolves on POW 453 are resilient to landscape-level ecological changes expected from old-growth logging. 454 Metabarcoding has revealed a more diverse and precise diet for wolves on POW and in 455 southeast Alaska, potentially pointing towards these wolves making greater use of alternate prey. 456 In general, DNA metabarcoding can be used as a tool to reliably describe diet for other carnivore 457 species. Even in a hostile environment for the preservation of eDNA, we have shown that DNA

458	metabarcoding is an effective and powerful method for describing carnivore diet. Diet analysis						
459	remains one of the most important avenues of wildlife study as it is a necessary component of						
460	understanding species interactions, predator-prey dynamics, and the biodiversity of systems. This						
461	nuanced profiling of diet is especially important as vulnerable wildlife populations face						
462	continued habitat loss and degradation, and thus we can use changes in diet can as potential						
463	indicators of environmental health.						
464							
465	Acknowledgements						
466	This research was funded by the Alaska Department of Fish and Game and by Oregon State						
467	University. The authors would like to thank all those involved with collecting wolf scats						
468	including T. Bentz, S. Bethune, D. Gregovich, J. Jemmison, M. Kampnich, K. Larson, J. Manuel,						
469	G. Roffler, J. Reeves, S. Sell, Y. Shakeri, and R. Slayton as well as C. Cousins, D. Martinez, and						
470	A. Pepper for the assistance with the mechanical sorting of scats. The Center for Genome						
471	Research and Biocomputing (CGRB) at Oregon State University provided both access to the						
472	Illumina HiSeq 3000 for all metabarcoding and support for initial bioinformatic analysis.						
473							
474	Literature Cited						
475	Alaback, P. B. 1982. Dynamics of Understory Biomass in Sitka Spruce-Western Hemlock Forest						
476	of Southeast Alaska: Ecological Archives E063-004. Ecology 63:1932–1948.						
477	Albert, D. M., and J. W. Schoen. 2013. Use of historical logging patterns to identify						
478	disproportionately logged ecosystems within temperate rainforests of southeastern						
479	Alaska. Conservation Biology 27:774–784.						

- 480 Berry, O., C. Bulman, M. Bunce, M. Coghlan, D. C. Murray, and R. D. Ward. 2015. Comparison
- 481 of morphological and DNA metabarcoding analyses of diets in exploited marine fishes.
- 482 Marine Ecology Progress Series 540:167–181.
- 483 Binladen, J., M. Gilbert, J. Bollback, F. Panitz, C. Bendixen, R. Nielsen, and E. Willerslev. 2007.
- 484 The use of coded PCR primers enables high-throughput sequencing of multiple homolog
  485 amplification products by 454 parallel sequencing. PLoS ONE 2:e197.
- 486 Buglione, M., V. Maselli, D. Rippa, G. de Filippo, M. Trapanese, and D. Fulgione. 2018. A pilot
- 487 study on the application of DNA metabarcoding for non-invasive diet analysis in the
- 488 Italian hare. Mammalian Biology 88:31–42.
- 489 Carrlee, E., and L. Horelick. 2011. The Alaska Fur ID Project: A virtual resource for material
  490 identification. Objects Specialty Group Postprints 18:149–171.
- 491 Ciucci, P., L. Boitani, E. R. Pelliccioni, M. Rocco, and I. Guy. 1996. A comparison of scat-

492 analysis methods to assess the diet of the wolf Canis lupus. Wildlife biology 2:37–48.

493 Cook, J. A., N. Dawson, and S. MacDonald. 2006. Conservation of highly fragmented systems:

494 The north temperate Alexander Archipelago. Biological Conservation 133:1–15.

495 Darimont, C. T., P. C. Paquet, and T. E. Reimchen. 2008a. Spawning salmon disrupt tight

496 trophic coupling between wolves and ungulate prey in coastal British Columbia. BMC497 Ecology 8.

- 498 Darimont, C. T., M. H. H. Price, N. N. Winchester, J. Gordon-Walker, and P. C. Paquet. 2004.
- 499 Predators in natural fragments: foraging ecology of wolves in British Columbia's central
  500 and north coast archipelago. Journal of Biogeography 31:1867–1877.

- 501 Darimont, C. T., T. E. Reimchen, H. M. Bryan, and P. C. Paquet. 2008b. Faecal-centric
- approaches to wildlife ecology and conservation: methods, data and ethics. Wildlife
  Biology in Practice 4:73–87.
- 504 Darimont, C. T., T. E. Reimchen, and P. C. Paquet. 2003. Foraging behavior by gray wolves on
- salmon streams in coastal British Columbia. Canadian Journal of Zoology 81:349–353.
- 506 De Barba, M., C. Miquel, F. Boyer, C. Mercier, D. Rioux, E. Coissac, and P. Taberlet. 2014.
- 507 DNA metabarcoding multiplexing and validation of data accuracy for diet assessment:
  508 application to omnivorous diet. Molecular Ecology 14:306–323.
- 509 Deagle, B. E., A. C. Thomas, J. C. McInnes, L. J. Clarke, E. J. Vesterinen, E. L. Clare, T. R.
- 510 Kartzinel, and J. P. Eveson. 2018. Counting with DNA in metabarcoding studies: How
  511 should we convert sequence reads to dietary data? Molecular Ecology:1–16.
- 512 Eriksson, C. E., K. M. Moriarty, M. A. Linnell, and T. Levi. 2019. Biotic factors influencing the
- 513 unexpected distribution of a Humboldt marten (Martes caurina humboldtensis)

514 population in a young coastal forest. PLoS ONE 14:e0214653.

- 515 Farmer, C. J., and M. Kirchhoff. 2007. Ecological classification of deer habitat in the Tongass
- 516 National Forest, Alaska. Northwestern Naturalist:73–84.
- 517 Gilbert, S. L., T. Haynes, M. S. Lindberg, D. Albert, M. Kissling, and D. K. Person. 2016. Future

population trends and drivers of change for Alexander Archipelago wolves on and near
Prince of Wales Island, Alaska.

- 520 Gilbert, S. L., K. J. Hundertmark, D. K. Person, M. S. Lindberg, and M. S. Boyce. 2017.
- 521 Behavioral plasticity in a variable environment: snow depth and habitat interactiosn drive
- deer movement in winter. Journal of Mammalogy 98:246–259.

- 523 Kartzinel, T. R., P. A. Chen, T. C. Coverdale, D. L. Erickson, W. J. Kress, M. L. Kuzmina, D. I.
- 524Rubenstein, W. Wang, and R. M. Pringle. 2015. DNA metabarcoding illuminates dietary
- 525 niche partitioning by African large herbivores. PNAS 112:8019–8024.
- 526 Kohira, M. 1995. Diets of wovles on Prince of Wales Island, southeast Alaska. University of
- 527 Alaska Fairbanks, Fairbanks, Alaska.
- Kohira, M., and E. A. Rexstad. 1997. Diets of Wolves, Canis lupus, in Logged and Unlogged
  Forest of Southeastern Alaska. Canadian Field Naturalist 111:429–435.
- 530 Krehenwinkel, H., M. Wolf, J. Y. Lim, A. J. Rominger, W. B. Simison, and R. G. Gillespie.
- 5312017. Estimating and mitigating amplification bias in qualitative and quantitative
- arthropod metabarcoding. Scientific Reports 7:17668.
- 533 Lafferty, D. J. R., J. L. Belant, K. S. White, W. N. Jamie, and A. T. Morzillo. 2014. Linking
- 534 Wolf Diet to Changes in Marine and Terrestrial Prey Abundance. Arctic 67:143–148.
- Lake, S., H. Burton, and J. van den Hoff. 2003. Regional, temporal and fine-scale spatial
- 536 variation in Weddell seal diet at four coastal locations in east Antarctica. Marine Ecology
- 537 Progress Series 254:293–305.
- MacDonald, S., and J. Cook. 2007. Mammals and Amphibians of Southeast Alaska. The
  Museum of Southwestern Biology, University of New Mexico.
- 540 McInnes, J. C., R. Alderman, M.-A. Lea, B. Raymond, B. E. Deagle, R. A. Phillips, A.
- 541 Stanworth, D. R. Thompson, P. Catry, H. Weimerskirch, C. G. Suazo, M. Gras, and S. N.
- 542 Jarman. 2017. High occurrence of jellyfish predation by black-browed and Campbell
- albatross identified by DNA metabarcoding. Molecular Ecology 26:4831–4845.
- Person, D. K., and T. J. Brinkman. 2017. Succession debt and roads. Pages 143–167 North
- 545 Pacific Temperate Rainforests: Ecology and Conservation.

- 546 Person, D. K., M. Kirchhoff, V. Van Ballenberghe, G. C. Iverson, and E. Grossman. 1996. The
  547 Alexander Archipelago wolf: a conservation assessment.
- 548 Pinol, J., M. A. Senar, and W. O. C. Symondson. 2018. The choice of universal primers and the
- 549 characteristics of the species mixture determin when DNA metabarcoding can be
- 550 quantitative. Molecular Ecology:1–13.
- 551 Porter, B. 2018. Wolf management report and plan, Game Management Unit 2: Report period 1
- 552 July 2010-30 June 2015, and plan period 1 July 2015-30 June 2020. Alaska Department 553 of Fish and Game.
- R Core Team. 2018. R: A language and environment for statistical computing. R Foundation for
  Statistical Computing, Vienna, Austria.
- Riaz, T., W. Shehzad, A. Viari, F. Pompanon, P. Taberlet, and E. Coissac. 2011. ecoPrimers:
  inference of new DNA barcode markers from whole genome sequence analysis. Nucleic
  Acids Research 39:e145.
- Schoen, J., M. Kirchhoff, and J. H. Hughes. 1988. Wildlife and old-growth forests in Southeast
  Alaska. Natural Areas Journal 8:138–145.
- 561 Shanley, C. S., S. Pyare, M. I. Goldstein, P. B. Alaback, D. M. Albert, C. M. Beier, T. J.
- 562 Brinkman, R. T. Edwards, E. Hood, A. MacKinnon, and M. V. McPhee. 2015. Climate
- 563 change implications in the northern coastal temperate rainforest of North America.
- 564 Climatic Change 130:155–170.
- 565 Shehzad, W., T. Riaz, M. A. Nawaz, C. Miquel, C. Poillot, S. A. Shah, F. Pompanon, E. Coissac,
- and P. Taberlet. 2012. Carnivore diet analysis based on next-generation sequencing:
- 567 application to the leopard cat (Prionailurus bengalensis) in Pakistan. Molecular ecology
- 568 21:1951–1965.

569	Smith, M. A.	2016. Ecological	Atlas of Southeast	Alaska. Audubon Alaska.

- 570 Soininen, E. M., A. Valentini, E. Coissac, C. Miquel, L. Gielly, C. Brochmann, A. K. Brysting, J.
- 571 H. Sonstebo, N. Yoccoz, and P. Taberlet. 2009. Analysing diet of small herbivores: The
- 572 efficiency of DNA barcoding coupled with high-throughput pyrosequencing for
- deciphering the composition of complex plant mixtures. Frontiers in Zoology 6:16.
- 574 Srivathsan, A., J. C. M. Sha, A. P. Vogler, and R. Meier. 2015. Comparing the effectiveness of
- 575 metagenomics and metabarcoding for diet analysis of a leaf-feeding monkey (Pygathrix
  576 nemaeus). Molecular Ecology Resources 15:250–261.
- 577 Stenglein, J. L., M. De Barba, D. E. Ausband, and L. P. Waits. 2010. Impacts of sampling
- 578 location within a faeces on DNA quality in two carnivore species. Molecular Ecology
  579 Resources 10:109–114.
- 580 Szepanski, M. M., M. Ben-David, and V. Van Ballenberghe. 1999. Assessment of anadromous
- salmon resources in the diet of the Alexander Archipelago wolf using stable isotopeanalysis. Oecologia 120:327–335.
- 583 Thomas, A. C., B. E. Deagle, J. P. Eveson, C. H. Harsch, and A. W. Trites. 2016. Quantitative
- 584 DNA metabarcoding: improved estimates of species proportional biomass usning
- 585 correction factors derived from control material. Molecular Ecology Resources 16:714–
  586 726.
- Thomas, A. C., B. W. Nelson, M. M. Lance, B. E. Deagle, and A. W. Trites. 2017. Harbour seals
  target juvenile salmon of conservation concern. Canadian Journal of Fisheries and
  Aquatic Sciences 74:907–921.
- 590 U.S. Fish and Wildlife Service. 2015. Species status assessment for the Alexander Archipelago
  591 wolf (Canis lupus ligoni). Version 1.0. Alaska Region, Anchorage, Alaska. 162 pp.

- Valentini, A., F. Pompanon, and P. Taberlet. 2009. DNA barcoding for ecologists. Trends in
  Ecology and Evolution 24:110–117.
- Wallmo, O. C., A. W. Jackson, T. L. Hailey, and R. L. Carlisle. 1962. Influence of rain on the
- count of deer pellet groups. The Journal of Wildlife Management 26:50–55.
- 596 Wallmo, O. C., and J. Schoen. 1980. Response of deer to secondary forest succession in
- 597 southeast Alaska. Forest Sciences 26:448–462.
- 598 Weaver, J. L. 1993. Refining the equation for interpreting prey occurrence in gray wolf scats.
- 599 The Journal of Wildlife Management 57:534–538.
- 600 Zhang, J., K. Kobert, T. Floui, and A. Stamatakis. 2013. PEAR: a fast and accurate Illumina
- 601 Paired-End reAd mergeR. Bioinformatics 30:614–620.
- 602
- 603
- 604
- 605
- 606
- 607
- 608
- 609
- 610
- 611
- 612
- 613
- 614
- 615
- 616

617 **Table 1:** Summary statistics for all generalized logistic regression models. Predictor variables

618 include average DNA quantity per sample (avg DNA quant), total number of sequence reads

619 prior to quality control and including wolf sequence reads (reads with wolf), and total number of

620 sequences reads post quality control and not including wolf sequences reads (reads no wolf).

621 Models were tested against all mechanically sorted samples that had a positive occurrence for a

622 species and against all mechanically sorted samples that had a positive occurrence for beaver.

623

Model	Estimate	SE	z value	$\Pr(> z )$
spp.presence ~ avg DNA quantity	0.13	0.14	0.966	0.33
spp.presence ~ total reads with wolf	4.16e-07	1.5e-06	0.273	0.79
spp.presence ~ total reads no wolf	-1.5e-07	1.63e-06	-0.095	0.93
beaver.presence ~ avg DNA quantity	-0.034	0.139	-0.25	0.81
beaver.presence ~ total reads with wolf	4.4e-07	1.78e-06	0.25	0.81
beaver.presence ~ total reads no wolf	-5.85e-06	2.61e-06	-2.24	0.025*

624

625

626

627

628

629

630

### 632 Figure Legends

- **Figure 1**: Study area map showing Alexander Archipelago in Southeast Alaska. Red and yellow
- 634 points represent individual scat collection sites. Most scats collections were concentrated on
- 635 Prince of Wales Island (yellow points).
- **Figure 2**: Examples of wolf scats collected in southeast Alaska near Prince of Wales Island.
- 637 Left-sided panels (a, c, and e) are examples of fresh scats (< 3 months old) and the right-sided
- panels (b, d, f) are examples of old/degraded scats (> 3 months old). Age was determined by the
- 639 collector; scats were collected throughout 2014 2015.
- **Figure 3**: Photographs depicting examples of fine-scale mechanical sorting results of prey
- 641 species in wolf scats collected in Southeast Alaska, 2014-2015. Starting from the top left panel
- and moving clockwise, species shown are salmon, black bear, bald eagle, harbor seal, and
- 643 sculpin.

Figure 4: Boxplots depicting the total number of reads and the DNA quantity (measured postnormalization) for scat samples binned by the age of the scat.

Figure 5: Diet summary from analysis of wolf scats (Southeast Alaska, 2014-2015) using (a) metabarcoding methods and (b) mechanical sorting methods. For the diet trees, each branch and terminal node represent a prey species identified in the wolf scats with the size and color of the branch showing the number of occurrences of that prey species. Frequency of Occurrence and RRA and estimated volume are compared.

Figure 6: Correlation between relative read abundance data for metabarcoding methods and
estimated volume for mechanical sorting methods by scat sample for the three most prevalent
prey species from wolf scats, Southeast Alaska, 2014-2015. Estimated volume is measured as the

654	proportion of	a nrev species	consumed ner s	cat relative to	the whole scat	RRA is the relative
034	proportion or	a prev species	consumed per s		the whole scat.	

- read abundance. Data points highlighted in brown show samples where deer was thought to be
- 656 mistakenly identified as beaver in mechanical sorting.
- **Figure 7**: Diet diversity, frequency of occurrence (FOO), and RRA and estimated volume found
- with a) metabarcoding results and b) mechanical results for scats found on Prince of Wales

659 Island, Alaska.

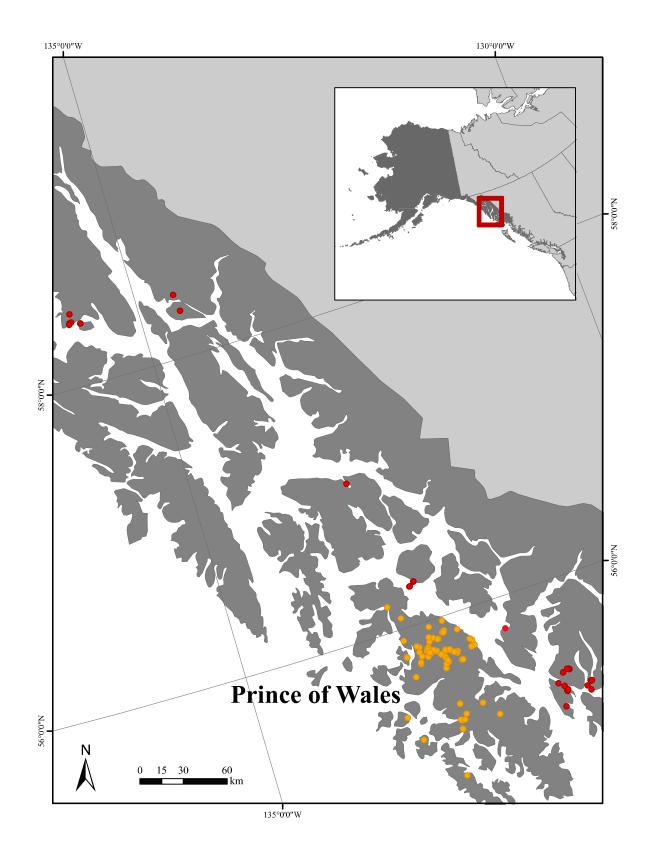
- **Figure 8**: Panel of hair samples. The top row shows examples of guard hairs from the Alaska
- 661 Fur ID project of Sitka black-tailed deer, beaver, and black bear. The bottom row shows
- examples of scale pattern from scale casts from the Alaska Fur ID project of Sitka black-tailed

deer, beaver, and black bear (left to right). The last panel in each row is an example of a difficult

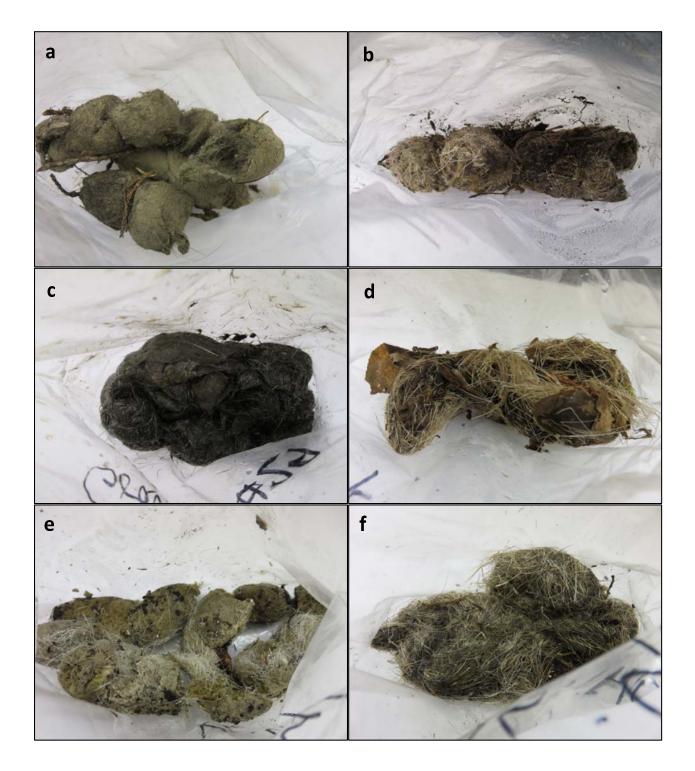
to identify hair and scale pattern from a wolf scat sample, Southeast Alaska 2014-2015.

- 665
- 666
- 667
- 668
- 669
- 670
- 671
- 672
- 673

- 675



### **678 Figure 2**



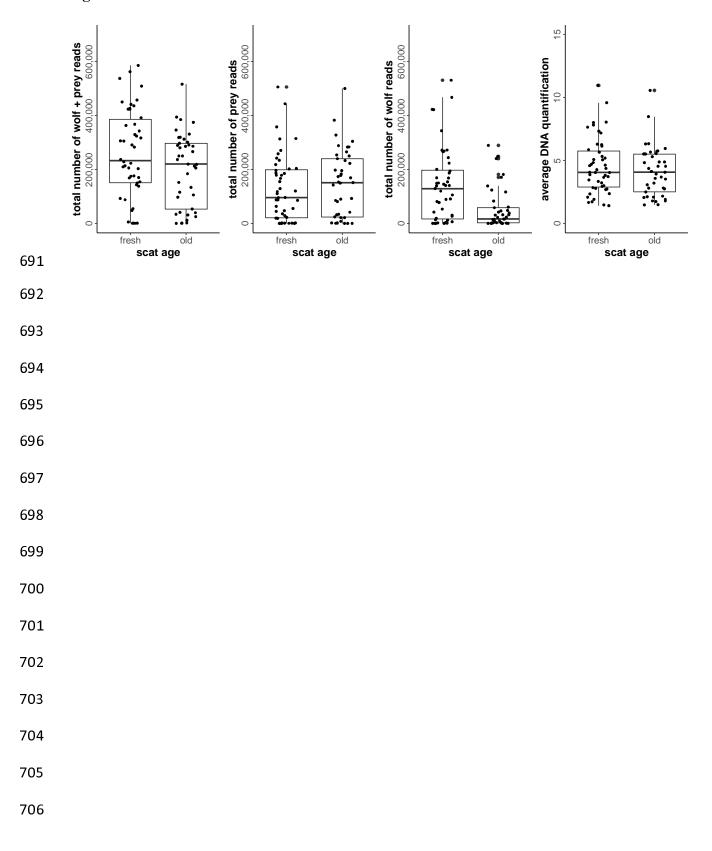


680

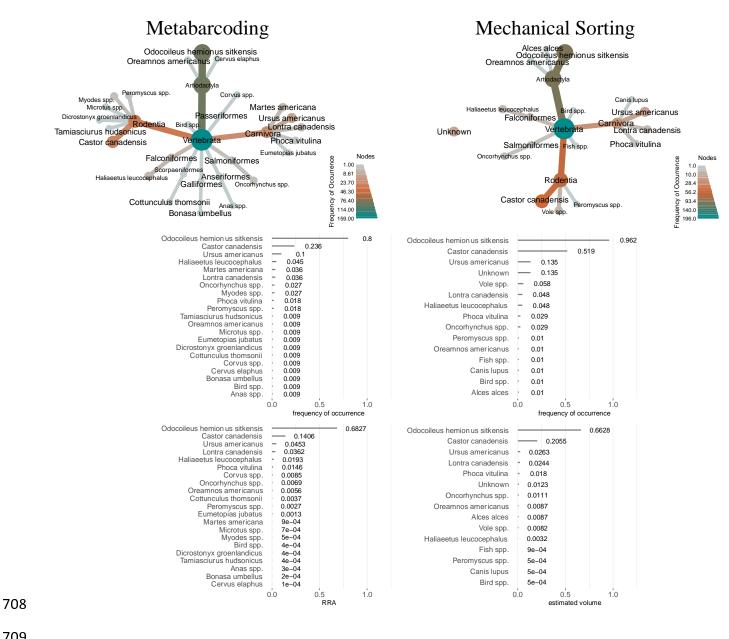
## 682 Figure 3



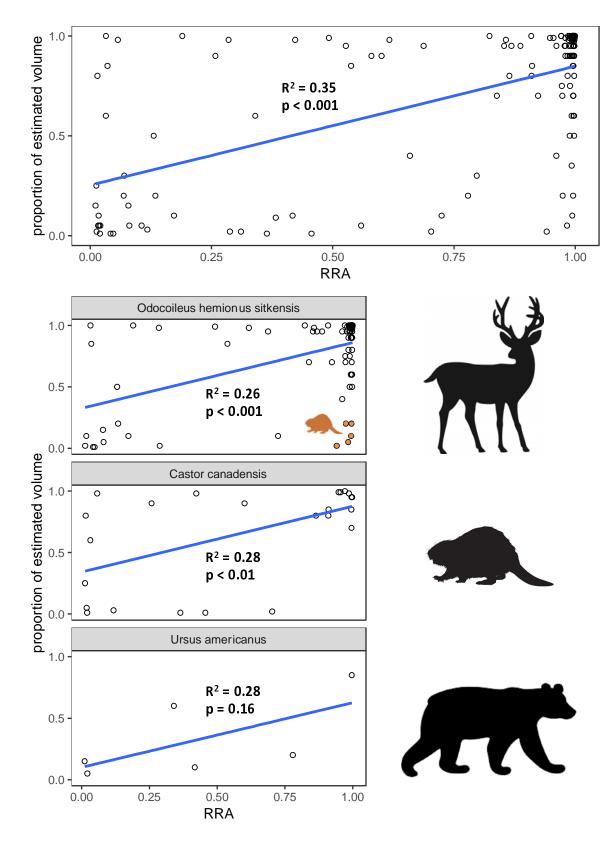
### 690 Figure 4



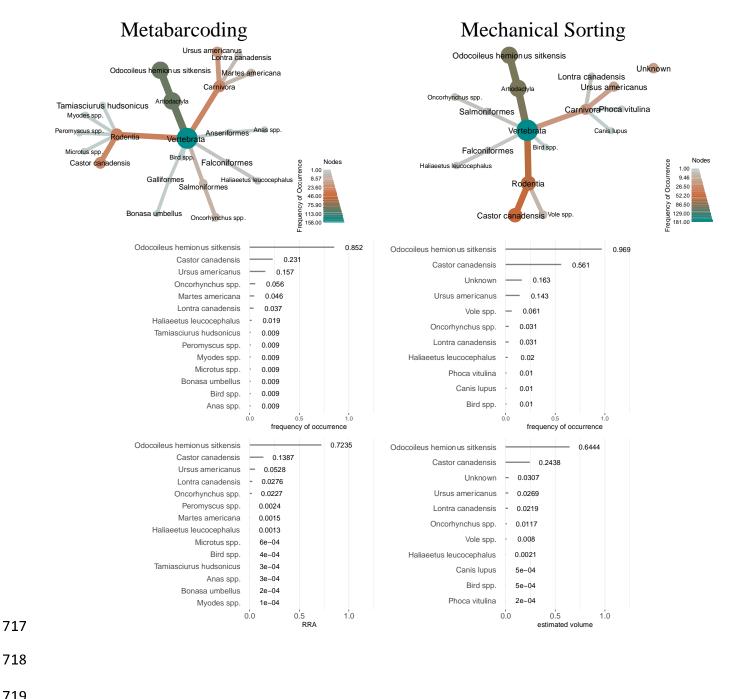
#### Figure 5



### 714 Figure 6



### 716 Figure 7



- 719
- 720
- 721
- 722

### **Figure 8**

