

1 Estimating densities of larval Salmonflies (*Pteronarcys californica*) through multiple pass
2 removal of post-emergent exuvia in Colorado rivers.

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12 **Abstract**

13 Traditional methods of collecting and sorting benthic macroinvertebrate samples are useful
14 for stream biomonitoring and ecological studies; however, these methods are time consuming,
15 expensive, and require taxonomic expertise. Estimating larval densities through collection of
16 post-emergent exuvia can be a practical and time efficient alternative. We evaluated the use of
17 multiple pass depletion techniques of the post-emergent exuvia of *Pteronarcys californica* to
18 estimate larval densities at ten sites in three Colorado rivers. Exuvia density was highly
19 correlated with both final-instar larval density ($R^2 = 0.90$) and total larval density ($R^2 = 0.88$) and
20 the multiple pass removal technique performed well. Exuvia surveys found *P. californica* at
21 three low density sites where benthic sampling failed to detect it. At moderate and high density
22 sites the exuvia surveys always produced lower density estimates than benthic surveys. Multiple
23 pass depletion estimates of exuvia proved to be an accurate and efficient technique at estimating

24 larval densities and provided an effective alternative for traditional benthic sampling when
25 objectives are monitoring *P. californica* and detecting populations, especially at low density
26 sites.

27 **Introduction**

28 Evaluating the condition of freshwater ecosystems through benthic macroinvertebrate
29 communities is a common approach for stream health assessment and biomonitoring [1-3]. These
30 methods characterize and compare aquatic invertebrate communities among sites using
31 regionally developed standards. Benthic studies, while useful, are labor and time intensive,
32 expensive, sensitive to sampling techniques, and require taxonomic expertise. The costs can be
33 justified by the valuable data used by government agencies, researchers, and water managers to
34 maintain and monitor water quality and understand function of river ecosystems. But, if
35 sampling objectives are more specific and budgets are limited, whole community benthic
36 sampling may not be necessary or the most appropriate technique.

37 One ecologically important aquatic invertebrate commonly used as a bioindicator is the
38 Giant Salmonfly (*Pteronarcys californica* Newport). It is useful for biomonitoring because of its
39 sensitivity to habitat alteration, widespread distribution in western North America [4, 5], multi-
40 year larval life stage, large body size, easy identification, low larval dispersal, and well defined
41 larval habitat preferences [6-9]. *Pteronarcys californica* is among the largest and longest lived
42 stonefly in western North America [10-12]. In Colorado, larvae typically inhabit unpolluted,
43 medium to large, permanent streams with unconsolidated cobble and large gravel substrates
44 between 1,500 and 2,500 m in elevation [13, 14]. Adults emerge from late May to early July and
45 recruitment begins in April after a 9-10 month egg diapause [15] followed by a three to four year
46 aquatic larval stage [16, 17]. Mature larvae (larvae expected to hatch that year) migrate toward

47 the stream bank to stage a highly synchronous adult emergence. Salmonflies typically emerge at
48 night crawling out of the water onto riparian substrates to become winged terrestrial adults where
49 they leave post-emergent exuvia (hereafter, exuvia).

50 *Pteronarcys californica* plays an important ecological role, both in biomass and
51 abundance, in stream and riparian food webs. As shredders, larvae process coarse organic matter
52 like vascular plants and algae [9, 18] making the nutrients available to other feeding groups as
53 detritus or body biomass [19]. Salmonflies can comprise a large portion of the benthic biomass
54 because of their large body size and high densities in suitable habitat [20, 21], making them an
55 important component of stream food webs for crayfish, other invertebrates, and trout [22, 21].
56 Terrestrial adults are part of a critical link for aquatic-riparian nutrient and energy exchange [23]
57 as prey for frogs, birds, bats, and spiders [15, 24]. Despite its ecological importance, range-wide
58 declines of *P. californica* have been documented in the Logan and Provo Rivers in Utah [25, 26],
59 several rivers in Montana [27], and in the Gunnison and Colorado Rivers in Colorado [4, 28]
60 mostly due to effects of dams like decreased water quantity and quality, siltation, and pollution.

61 Density of benthic macroinvertebrates is traditionally estimated by systematically
62 collecting samples from a fixed area of the stream bed. Alternative methods have been recently
63 developed to indirectly survey communities by identifying and enumerating exuvia. These
64 methods can reduce time and labor of traditional techniques while providing reliable population
65 density estimates, community structure, and life history information. Ruse [29] deduced
66 chironomid communities from larval and pupal exuvia and Foster and Soluk [30] estimated
67 densities of the endangered Hine's emerald dragonfly (*Somatochlora hineana*) more accurately
68 by sampling larval exuvia than by collecting adults. Raebel et al. [31] stated the importance of
69 exuvia collections to avoid bias in adult Odonata surveys. DuBois [32] enhanced these studies by

70 using a depletion population estimator to approximate exuvia densities and detection
71 probabilities of Anisoptera. Richards et al. [33] correlated *P. californica* exuvia densities and live
72 (wet) larval body weights with substrate embeddedness to demonstrate differences in life history,
73 distribution, and abundance above and below a main stem impoundment. Their work provided a
74 foundation in the development of our novel technique to estimate larval densities through
75 multiple pass removal sampling of exuvia.

76 Multiple pass removal sampling is a commonly used technique in wildlife and fisheries to
77 estimate population size of closed populations. Assumptions of this models used to analyze these
78 data are closure (no deaths, births, emigration, or immigration) and constant capture probability
79 [34, 35] that must be met to avoid bias [36, 37]. If more than two depletion events are completed
80 then assumptions about capture probabilities can be relaxed and capture rates for different passes
81 can be estimated. If populations can be considered geographically and demographically closed
82 (due to isolation or short sampling time period) then population estimation can be accomplished
83 rather simply if good unbiased estimates of detection probability are possible.

84 The objective of this study was to couple traditional benthic invertebrate sampling with
85 multiple pass removal techniques to evaluate if closed population estimation models can be used
86 to estimate the density *P. californica* larvae. We tested this by correlating densities of
87 systematically collected exuvia from the riparian area with densities of larvae from benthic
88 samples. Another goal was to provide a safer and more efficient method for estimating single
89 species densities.

90 **Methods**

91 **Study area**

92 Benthic and exuvia sampling was conducted at ten sites on three rivers in Colorado. Eight
93 sites were sampled on the Colorado River and one on the Fraser River both in north central
94 Colorado, and one site on the Gunnison River in southwest Colorado (Fig 1). Distance between
95 the lowest Colorado River site and the Fraser River sites is 74 km.

Fig 1. *Pteronarcys californica* benthic and exuvial collection sites in 2010 from the Colorado and Fraser Rivers. Gunnison River site shown only on inset map.

96 Benthic sampling

97 Three benthic subsamples were taken at each site between 15 -18 April 2010 from the
98 Colorado and Fraser Rivers and 10 May 2010 from the Gunnison River, approximately 1 month
99 prior to the typical adult emergence times of *P. californica*. All sites were located in riffle areas
100 dominated by cobble substrates interspersed with gravel except for sites 7 and 8 which were
101 dominated by sand and gravel. A modified Surber sampler with a 0.25 m² sampling frame (55.0
102 cm x 45.5 cm) and 150 µm mesh net was used. Cobbles larger than 10 cm in diameter were
103 individually scrubbed with a brush, invertebrates washed into the net, and then the cobbles
104 removed from the sampling frame. Remaining substrate within the frame was disturbed to a
105 depth of 10 cm to dislodge invertebrates into the net. Contents were preserved with 80% ethanol
106 in 2 L plastic jars.

107 In the lab, all *P. californica* larvae were sorted, sexed [21, 38], and measured for total
108 length (TL) from the anterior tip of head to the posterior tip of the epiproct to the nearest
109 millimeter under a dissecting microscope with a calibrated ocular micrometer. Length frequency
110 histograms for male and female larvae were constructed based on TL to separate annual year
111 classes. Densities of mature larvae and densities of all larvae were calculated and used in
112 separate analyses for correlation with exuvia densities. Mature larval cut off lengths were distinct

113 from the younger year class providing reliable data for analysis despite problems with TL
114 measurements. Separating cohorts and year classes of merovoltine species has proven difficult
115 because of varying growth rates [16] and contraction or expansion of abdomens in preserved
116 insect specimens can further confound this task. Our colleagues [39] used head capsule width
117 and combined head and thorax lengths to produce “body size” or “body area” to assign cohorts
118 within a stream.

119 **Exuvia sampling**

120 Sampling began with the onset of *P. californica* adult emergence on the Colorado River
121 at site nine on 2 June 2010 and proceeded upstream to end at site one on the Fraser River on 21
122 June 2010; sampling at site 10 on the Gunnison River lasted from 16-23 June 2010. Each site
123 was sampled beginning on the day when the first exuvia was found or winged adults were
124 observed and continued daily until exuvia were no longer found. Data collection was performed
125 by searching for exuvia within 10 m of the bank along two 30.5 m transects on one side of the
126 river directly adjacent to benthic sampling sites. Collections at a site were accomplished by 2-4
127 people in a matter of 2-6 hours completing 2-4 passes with identical effort and personnel.
128 Specimens were taken only when attached to dry riparian and emergent substrates; none were
129 taken from the water to avoid counting ones that possibly drifted into the site. Exuvia were
130 enumerated on hand held counters, stored in sealable bags, and removed from the search area.
131 Amount of time searching varied by site depending on the number of exuvia and complexity of
132 riparian search area.

133 **Data analysis**

134 Area of benthic habitat was estimated by multiplying the sampling section length (always
135 30.5 m) by the average wetted channel width derived from 10 evenly spaced cross-channel

136 transects. To evaluate the assumptions of the removal model and appropriateness of this
137 sampling technique, three and four pass removal data were compared to two pass data for twenty
138 of the sampling events. Three and four pass data were analyzed with the Huggins Closed Capture
139 model in Program Mark [40, 41] and two pass data were analyzed with the simpler two pass
140 removal model [34]. In Mark, models were built that varied capture probability by pass, allowing
141 a different capture probability for the first pass and the second pass and the third pass or third
142 and fourth passes. Declining capture probability with subsequent passes is a common source of
143 bias of removal models in fisheries data [36, 37] and comparing the population estimates and
144 capture probabilities allowed us to evaluate the assumption of constant capture probability of the
145 simpler two pass model. The assumptions of demographic and geographic closure were expected
146 to be met due to immobility of exuvia and the emergence occurring at night. To evaluate if
147 exuvia densities accurately estimated larval densities, we used simple linear regression in R [42].
148 Exuvia densities were the dependent variable and densities of mature larvae and all age class
149 larvae were both used in separate analyses as the independent variable.

150 **Results**

151 Adult emergence of *P. californica* lasted between 2-8 days at each site and proceeded
152 upstream approximately 4 km per day. Early in the emergence, male exuvia were dominant and
153 sex ratios were more even toward the end of the emergence. Approximately 97% of exuvia
154 (n=21,526) were collected within 2 m of the bank. A total of 592 larvae were collected. Larvae
155 from the Colorado and Fraser Rivers separated into four year classes; mature female larvae were
156 ≥ 39 mm TL (mean 46.5, SE 0.51) and males ≥ 35 mm TL (mean 39.2, SE 0.34) (Table 1).
157 Larvae from the Gunnison River separated into three year classes; mature female larvae ≥ 41 mm

158 TL (mean 49.1, SE 0.51) and mature males ≥ 37 mm TL (mean 41.9, SE 0.46) (Table 2). Mature
 159 females were significantly larger than mature males within each river ($p=0.0000$ for each).

Table 1. Year class lengths and frequency distribution of *Pteronarcys californica* larvae collected 30 April- 1 May 2010 from the Colorado and Fraser Rivers.

Lengths in mm from anterior tip of head to posterior tip of epiproct.

Year Class	Male larvae (n=149)	Female larvae (n=123)
1	≤ 15 mm (33)	≤ 17 mm (41)
2	16-25 (34)	18-25 (25)
3	26-34 (31)	26-38 (29)
4	≥ 35 (51)	≥ 39 (28)

Table 2. Year class lengths and frequency distribution of *Pteronarcys californica* larvae collected 14 April 2010 from the Gunnison River.

Lengths in mm from anterior tip of head to posterior tip of epiproct.

Year Class	Male larvae (n=191)	Female larvae (n=129)
1	≤ 23 mm (162)	≤ 23 mm (95)
2	24-36 (14)	24-40 (12)
3	≥ 37 (15)	≥ 41 (22)

160

161 Exuvia densities were highly correlated with both mature larval densities ($R^2 = 0.90$) and
 162 total larval densities ($R^2 = 0.88$). Exuvia densities averaged 2.6/m² and ranged from 0.002/m² to
 163 11.443/m² (Table 3). Total larval density averaged 80.0/m² and ranged from 0 to 437.3/m².
 164 Density of mature larvae averaged 16.1/m² and varied from 0 to 101.3/m². The correlation
 165 between mature larvae and exuvia densities was high but the relationship was not 1:1. Larval
 166 estimates were generally higher than exuvia estimates except at sites 2, 3, and 7 where exuvia
 167 were found but no larvae. To predict the density of mature larvae, the linear equation was: larval
 168 density = 7.358*(exuvia density) - 2.854.

Table 3. Densities in m² of *Pteronarcys californica* pre-emergent larvae, all larvae, exuvia, and population estimates from exuvia collected from April-June 2010 from the Colorado, Fraser, and Gunnison Rivers.

Site	Densities m ²			
	All larvae	Pre-emergent larvae	Exuvia	Exuvia population estimate
1	8.00	1.33	0.854	0.872
2	0.00	0.00	0.064	0.065
3	0.00	0.00	0.077	0.079
4	4.00	4.00	2.391	2.477
5	4.00	1.33	0.686	0.697
6	44.00	4.00	0.306	0.315

7	0.00	0.00	0.002	0.002
8	5.33	0.00	0.007	0.007
9	297.33	101.33	11.001	11.443
10	437.33	49.33	9.392	9.849

169

170 Exuvia detected populations at all 10 sites whereas larvae were found at only seven of the
171 10 sites (Table 3). Capture probabilities of exuvia ranged from 0.45 to 0.89 (average 0.72).
172 Simple two pass population models were sufficient to produce unbiased population estimates.
173 Capture probabilities and population estimates were very similar for both the Huggins closed
174 capture model for three and four pass estimates and the Zippin two pass estimates (Fig 2). The
175 two pass depletion technique worked best at sites with moderate exuvia densities and there was
176 some variation in capture probability at very low densities (<80 exuvia per 30.5 m) and very high
177 densities (> 6,000 exuvia per 30.5 m) indicating that the assumption of equal capture
178 probabilities for all passes is violated with the simple two pass model. However, that bias was
179 small and population estimates of the two models were very close (Fig 2).

Fig 2. Population estimates (N) and capture probabilities (\hat{p}) of a three pass Huggins Closed Capture model in Program Mark (with time effects that varied capture probabilities) and a simple two pass removal model of Zippin 1956.

180 Discussion

181 Multiple pass removal estimates of *P. californica* exuvia effectively predicted densities of
182 mature larvae. Assumptions of the multiple pass depletion models appeared to be met and
183 capture probability varied minimally among passes. The two-pass depletion technique performed
184 well due to immobility of exuvia, high capture probability, and no size selective gear [36, 37,
185 43], suggesting two sampling passes can be adequate if three or four passes are cost prohibitive
186 as with Odonata exuvia [32].

187 Correlation between densities of exuvia and mature larvae were high but not 1:1. Exuvia
188 underestimated larvae at high density sites but at low density sites exuvia overestimated larvae.
189 The underestimation of larval densities is likely attributed to the behavior of mature larvae
190 congregating near the river bank prior to emergence in the shallow, wadeable water where
191 benthic sampling must occur, creating an artifact of unnaturally high densities. Other factors that
192 may contribute to an underestimate are imperfect detection probability of exuvia and dispersal of
193 larvae out of the sampling area or predation in the 1-2 month time period between benthic
194 sampling and emergence. Overestimates were likely due to the high capture probability of exuvia
195 in addition to the difficulty of collecting larvae that are rare at a site using Hess or Surber
196 samplers [44]. Therefore, exuvia sampling may more accurately estimate, not necessarily
197 overestimate, larval densities than benthic sampling at low density sites.

198 Detection rates of populations through exuvia sampling were higher than for benthic
199 sampling. This is likely because the large amount of available benthic habitat was essentially
200 reduced to a much smaller, well defined and more easily accessible riparian sampling area.
201 Riparian sampling area among sites averaged 61 m² (30.5 m long x 2 m wide) compared to 742
202 m² (400-1500 m²) of unevenly distributed benthic habitat, much of which may not be accessible
203 by wading due to excessive depth or water velocities ≥ 2 m/s.

204 Presence of exuvia or adults is the only evidence of successful life cycle completion.
205 Varying densities can indicate habitat quality and help identify reference sites and priority areas
206 for river conservation, restoration, and monitoring of *P. californica*. In regions where *P.*
207 *californica* does not occur, this technique may be useful for other easily recognizable stoneflies
208 like *Pteronarcella badia*, *Claassenia sabulosa*, *Hesperoperla pacifica*, or mayflies like
209 *Timpanoga hecuba*. This technique eliminated the need for benthic sample collection,

210 preservation, and subsequent expense of processing in the laboratory. It also provided accurate
211 and less biased density estimates of *P. californica* larvae than those derived from benthic
212 samples.

213 Benthic sampling of aquatic invertebrates is a useful and productive biomonitoring
214 technique but the overall process to acquire data can be labor and cost intensive. In addition, it
215 can be difficult to find target species that are rare at a site with benthic sampling [44]. Using
216 multiple pass removal sampling of the recently shed exuvia can be an effective and efficient way
217 to estimate densities of *P. californica* and may be superior to traditional benthic sampling at
218 detecting the species at very low densities.

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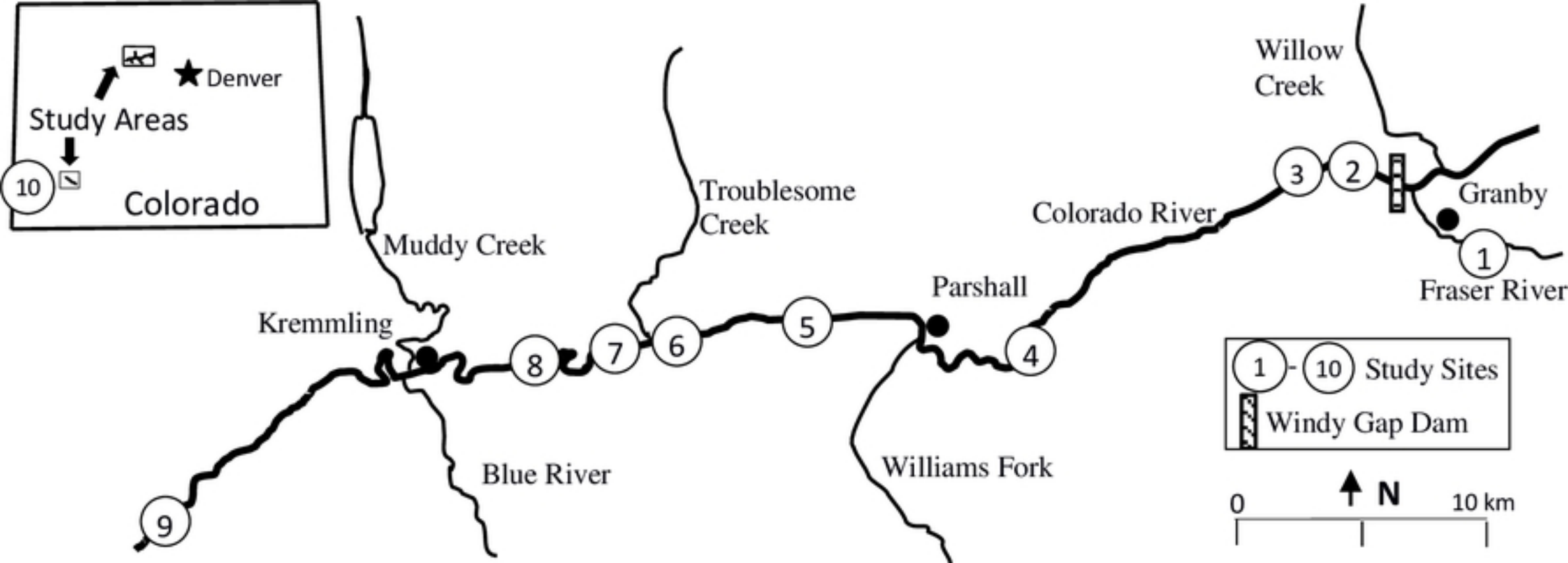
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Site	River	Location	N DecDeg, -W DecDeg
1	Fraser	Kaibab Park in Granby	40.0810, -105.9330
2	Colorado	1.2 km W of Windy Gap Dam	40.1087, -106.0020
3	Colorado	2.8 km W of Windy Gap Dam	40.1064, -106.0150
4	Colorado	below Byers Canyon	40.0520, -106.1320
5	Colorado	4.8 km W of Parshall	40.0654, -106.2330
6	Colorado	CR 39 bridge	40.0544, -106.2900
7	Colorado	150 m W of Troublesome Creek	40.0563, -106.3070
8	Colorado	2.4 km E of Kremmling	40.0494, -106.3450
9	Colorado	Pumphouse Recreation Area	39.9916, -106.5040
10	Gunnison	Ute Park in the Gunnison Gorge	38.6786, -107.8463

Fig 1.tif

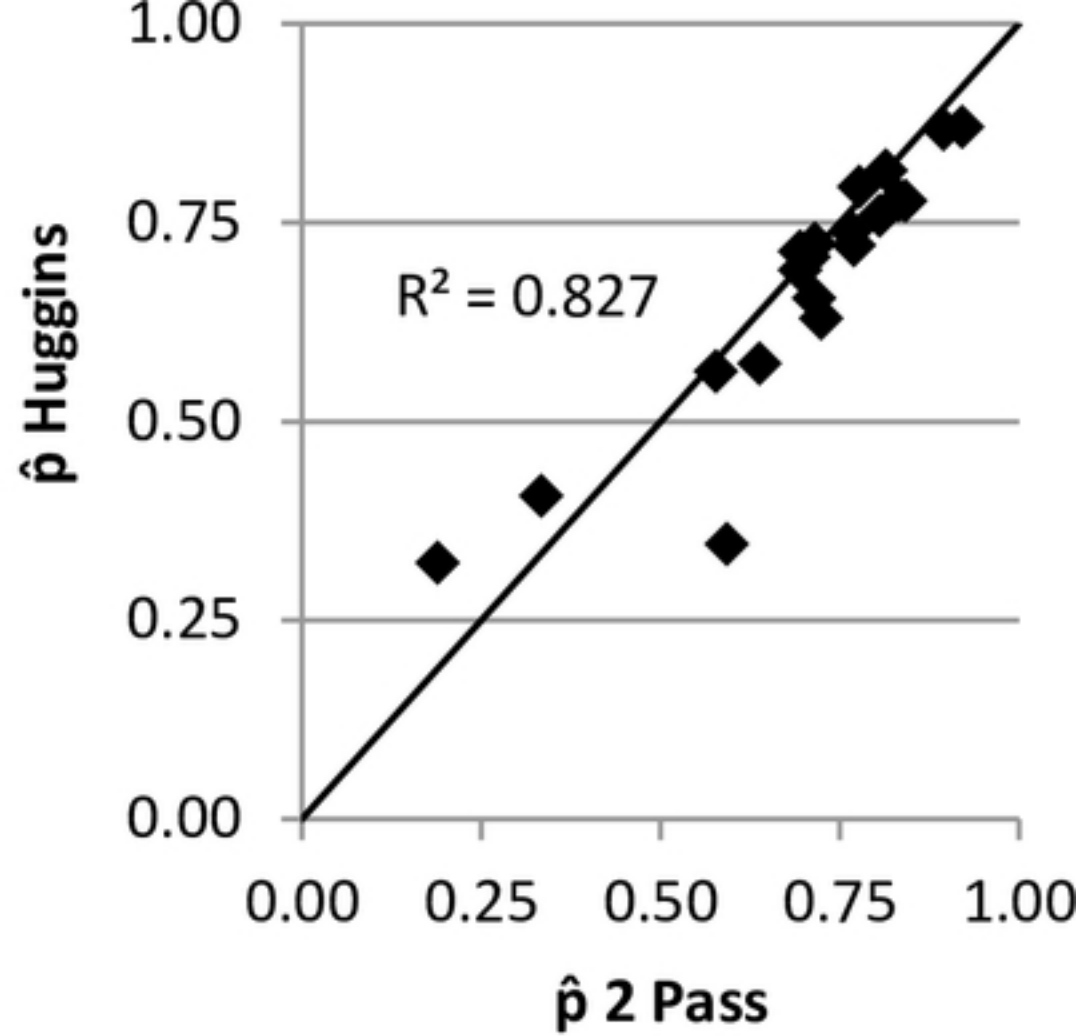
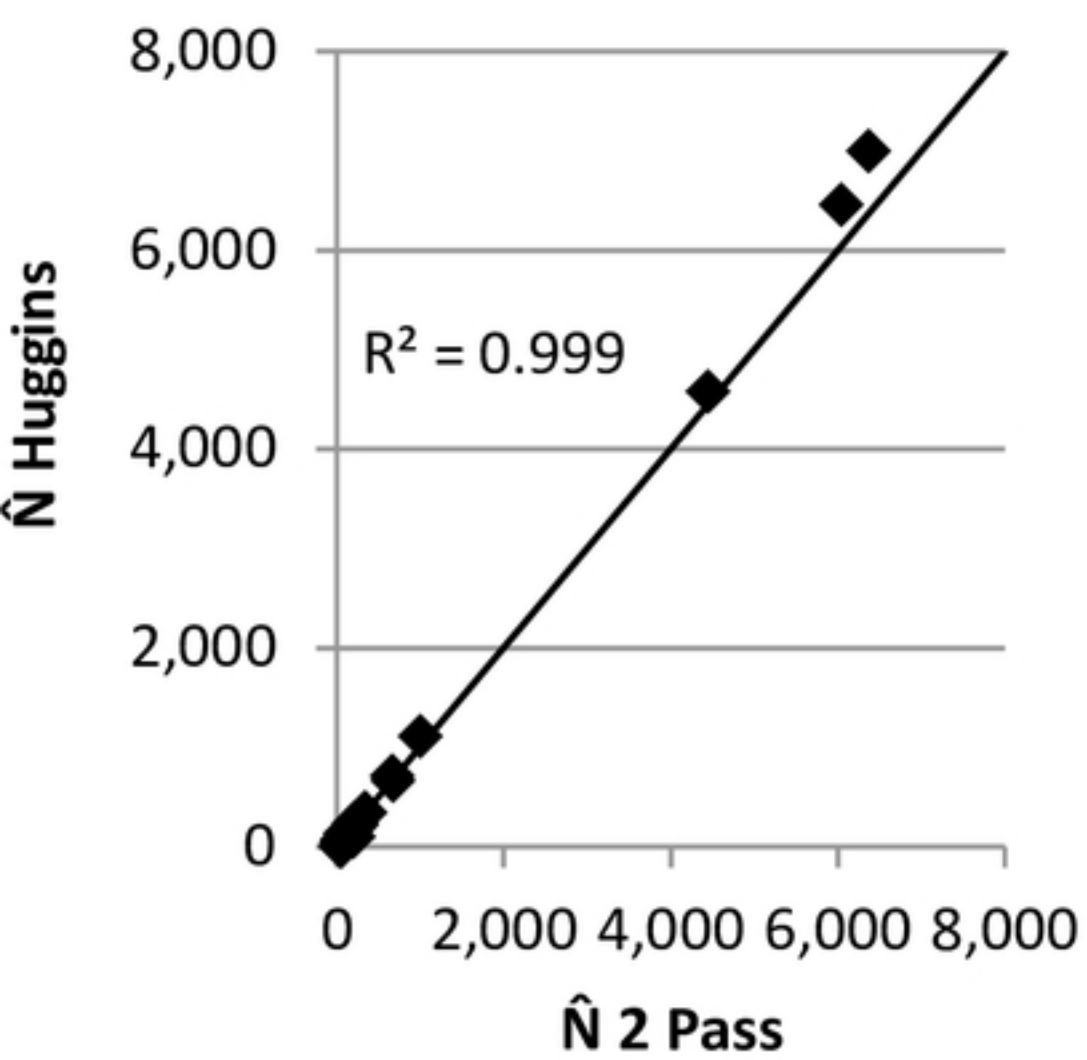


Fig 2.tif