1 Short Title: Shifts in temperature and *Batrachochytrium* 2 Shifts in temperature influence how Batrachochytrium dendrobatidis infects 3 amphibian larvae Paul W. Bradley¹, Michael D. Brawner², Thomas R. Raffel³, Jason R. Rohr⁴, Deanna H. 4 Olson⁵, and Andrew R. Blaustein² 5 ¹ Department of Biology, University of San Diego, 5998 Alcala Park, San Diego CA, 6 7 93110, USA. ² Department of Integrative Biology, 3029 Cordley Hall, Oregon State University, 8 9 Corvallis, OR, 97331, USA. 10 ³ Department of Biology, 375 Dodge Hall, Oakland University, Rochester, MI, 48309, 11 USA. ⁴ Department of Integrative Biology, University of South Florida, 4202 East Fowler 12 13 Avenue, Tampa, FL, 33620, USA. ⁵ USDA Forest Service, Pacific Northwest Research Station, 3200 SW Jefferson Way, 14 15 Corvallis, OR, 97331, USA. 16 17 Corresponding author: paulwilliambradley@gmail.com 18 19 **Author contribution statement** 20 21 ARB, JRR, and TRR originally formulated the idea, PWB designed the experiment and developed the methodology. PWB performed the experiment. PWB and MDB performed the molecular analysis. PWB 22 23 and TRR performed the statistical analyses. ARB, JRR, TRR, and DHO obtained funding. PWB wrote the manuscript and other authors provided editorial advice. 24

Abstract:

25

26

27

28

29

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

47

Many climate change models predict increases in mean temperature, and increases in frequency and magnitude of temperature fluctuations. These potential shifts may impact ectotherms in several ways, including how they are affected by disease. Shifts in temperature may especially affect amphibians, a group with populations that have been challenged by several pathogens. Because amphibian hosts invest more in immunity at warmer than cooler temperatures and parasites may acclimate to temperature shifts faster than hosts (creating lags in optimal host immunity), researchers have hypothesized that a temperature shift from cold-to-warm might result in increased amphibian sensitivity to pathogens, whereas a shift from warm-to-cold might result in decreased sensitivity. Support for components of this climate-variability based hypothesis have been provided by prior studies of the fungus *Batrachochytrium dendrobatidis* (Bd) that causes the disease chytridiomycosis in amphibians. We experimentally tested whether temperature shifts before Bd exposure alter susceptibility to Bd in the larval stage of two amphibian species – western toads (*Anaxyrus boreas*) and northern red legged frogs (*Rana aurora*). Both host species harbored elevated Bd infection intensities under constant cold (15° C) temperature in comparison to constant warm (20° C) temperature. Additionally, both species experienced an increase in Bd infection abundance when shifted to 20° C from 15° C, compared to a constant 20° C but they experienced a decrease in Bd when shifted to 15° C from 20° C, compared to a constant 15° C. These results are in contrast to prior studies of adult amphibians that found increased susceptibility to Bd infection after a temperature shift in either direction, highlighting the potential for species and stage differences in the temperature-dependence of chytridiomycosis.

Keywords: amphibian declines, *Batrachochytrium dendrobatidis*, chytridiomycosis,

climate variability hypothesis, infectious disease, temperature, Rana aurora, Anaxyrus

boreas

48

49

50

51

52

53

54

55

56

57

58

59

60

61

62

63

64

65

66

67

68

69

Introduction

Climate change represents one of the greatest challenges to biodiversity and conservation because it might compromise ecosystem functions worldwide. Changes in climate have affected plant-animal interactions, predator-prey interactions and disease dynamics (Lafferty 2009, Rohr et al. 2011, Sheldon, Yang & Tewksbury 2011, Garcia et al. 2014). Changes to annual or seasonal mean temperatures often are used to predict climate-change-induced effects on disease risk (Paaijmans, Read & Thomas 2009, Paaijmans et al. 2010). However, many climate change models also predict increases in the frequency and magnitude of extreme weather events and increases in temperature variability at monthly to weekly timescales (Easterling et al. 2000, Meehl & Tebaldi 2004, Schar et al. 2004, Paaijmans et al. 2010, Rummukainen 2012). Yet few studies have investigated how increases in temperature variability affect disease dynamics despite the likelihood that such variability might differentially affect hosts and pathogens (Paaijmans et al. 2010, Ben-Horin, Lenihan & Lafferty 2012, Raffel et al. 2013, Bannerman & Roitberg 2014, Luis et al. 2014, Raffel et al. 2015). Ectotherms, such as amphibians, are particularly sensitive to climate change (Blaustein et al. 2010, Lawler et al. 2010, Shoo et al. 2011, Li, Cohen & Rohr 2013) and are experiencing diseaseassociated population declines and extinctions worldwide (Stuart et al. 2004, McCallum

2007, Rohr *et al.* 2008, Wake 2012), making them an ideal group to investigate the relationship between temperature shifts and disease risk.

70

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

Chytridiomycosis is an emerging infectious disease of amphibians caused by the aquatic chytrid fungal pathogens Batrachochytrium dendrobatidis (Bd) and B. salamandrivorans (Longcore, Pessier & Nichols 1999, Martel et al. 2013). Bd is widespread globally (Liu, Rohr & Li 2013, Olson et al. 2013) and is associated with worldwide amphibian population declines (Stuart et al. 2004, Skerratt et al. 2007). Moreover, models based on IPCC climate futures predict that Bd will shift to higher latitudes and altitudes due to increased environmental suitability in those regions under climate change, thus potentially affecting additional amphibian populations (Xie, Olson & Blaustein 2016). The negative effects of Bd infection are more pronounced in post-metamorphic stages, often leading to death (Blaustein et al. 2005, Garner et al. 2009, Gervasi et al. 2013, Gervasi et al. 2017). In larvae, Bd infection can cause host mortality in some species (Blaustein et al. 2005, Garner et al. 2009). However the infection is localized to keratinized larval mouthparts, (Marantelli et al. 2004, McMahon & Rohr 2015) often resulting in sublethal effects including inhibited foraging capacity, reduced growth and development, altered predator avoidance, or changes to other behaviors (Han, Bradley & Blaustein 2008, Venesky, Parris & Storfer 2010, Buck et al. 2012, Gervasi et al. 2013). Additionally, larvae of many species are important members of aquatic communities and alterations to larval feeding have the potential to cascade through the aquatic ecosystem (Alford 1989, Brönmark, Rundle & Erlandsson 1991, Lamberti et al. 1992, Kupferberg 1997).

94

95

96

97

98

99

100

101

102

103

104

105

106

107

108

109

110

111

112

113

114

115

Temperature is considered one of the most important environmental factors driving chytridiomycosis (Drew, Allen & Allen 2006, Bosch et al. 2007, Daskin, Alford & Puschendorf 2011, Forrest & Schlaepfer 2011, Voyles et al. 2017). Bd is non-linearly sensitive to temperature with an optimal growth range in culture between 17° and 25° (Piotrowski, Annis & Longcore 2004, Rohr & Raffel 2010, Raffel et al. 2013) and a temperature-dependent generation time of 4 to 10 days (Woodhams et al. 2008). The upper thermal limit for Bd growth in culture is between 25°C and 28°C, with Bd mortality occurring above 30°C (Longcore, Pessier & Nichols 1999, Piotrowski, Annis & Longcore 2004). Bd has been shown to be reliably cleared from multiple amphibian species by extended exposure to 30°C (McMahon et al. 2014). Its lower thermal limit is below 4°C (Piotrowski, Annis & Longcore 2004). Additionally, life history strategies of the pathogen can be altered by environmental temperature, where colder temperatures can cause Bd zoosporangia to develop and mature more slowly (Voyles et al. 2012), but produce more and longer-lived zoospores overall (Hyatt et al. 2007, Woodhams et al. 2008). Because physiologies of both the host and pathogen are strongly influenced by environmental temperature, climate change has been used to explain several major Bd outbreaks and amphibian population declines, (reviewed in Li, Cohen & Rohr 2013, Rohr et al. 2013). Yet, the host and pathogen are not expected to share a uniform response to a given temperature (Brown et al. 2004, Paull, LaFonte & Johnson 2012, Rohr et al. 2013), and thermal responses measured in constant-temperature artificial environments might not reflect organism responses in more realistic variable-temperature environments. Providing evidence of the lack of a uniform response between Bd and amphibians to

117

118

119

120

121

122

123

124

125

126

127

128

129

130

131

132

133

134

135

136

137

temperature shifts, Rohr and Raffel (2010) found a strong correlation between elevated month-to-month temperature variability and Bd-associated amphibian population declines of *Atelopus* spp. across Central and South America. Further support of the relationship between chytridiomycosis and temperature variation has been provided by laboratory studies. In one study, Cuban treefrogs (Osteopilus septentrionalis) displayed reduced resistance to Bd infection when exposed to random daily temperature fluctuations or when exposed to a temperature decrease after acclimation to a warmer temperature (Raffel et al. 2013). Similar results were obtained in newts (Notophthalmus viridescens) exposed to Bd, except both decreases and increases in temperature were associated with elevated Bd abundance relative to abundances at constant temperatures (Raffel et al. 2015). The potential for temperature variability to increase disease severity in amphibians was first postulated by Raffel et al. (2006) and has subsequently been referred to as the "climate variability hypothesis" (Rohr & Raffel 2010). This hypothesis posits that parasites acclimate to the new temperature more rapidly than their hosts, leading to lags in host acclimation following a temperature shift that could make hosts more susceptible to infection (Raffel et al. 2013). This hypothesis assumes that: 1) pathogens acclimate to the new temperature faster than the host because of their relatively smaller size and higher metabolic rate (Gillooly et al. 2001, Raffel et al. 2013); and 2) both host and parasite acclimation responses lead to increased performance at the new temperature, in accordance with the "beneficial acclimation hypothesis" of thermal biology (Angilletta 2009). However, Raffel et al. (2006) also pointed out potential

complexities in acclimation of the ectotherm immune system that might lead to alternative predictions.

138

139

140

141

142

143

144

145

146

147

148

149

150

151

152

153

154

155

156

157

158

159

160

According to the "lag effect" hypothesis (Raffel et al. 2006), changes in levels of temperature-dependent immune parameters might simply lag behind environmental temperature shifts (Fig. 1) because it takes time to produce necessary, or remove unnecessary, immune cells from the host. For example, amphibians are expected to require more immune cells at warmer temperatures to fight off faster-growing pathogens (Maniero & Carey 1997), and lags in production of new immune cells could lead to suboptimal immunity following a temperature increase (Raffel et al. 2006). Conversely, the amphibian immune system is expected to be down-regulated following a temperature decrease (Macela & Romanovsky 1970), with the removal of mature white blood cells determined by the rate of their respective half-lives (DeSantis & Strauss 1997, Janeway 2008). A lag in this process might lead to a brief period of elevated immune responsiveness relative to an already cold-acclimated host. Thus, the "lag effect" hypothesis predicts the opposite effect from the "climate variability hypothesis" following a temperature decrease, at least on a short timescale. These mechanistic hypotheses are not mutually exclusive, and it is unclear which effects might be more important for a given host-parasite combination. We tested the general prediction that an amphibian shifted to a new temperature

We tested the general prediction that an amphibian shifted to a new temperature before Bd exposure would respond to infection differently than a host already acclimated to the exposure temperature. We postulated that the direction of the effect would depend upon the direction of the temperature shift, in accordance with the "lag effect" hypothesis of Raffel *et al.* (2006). Given the differences in size between the host and the pathogen,

and associated physiological process rate differences, we assumed Bd would physiologically respond to the temperature shift faster than the host, such that an idealized host-immune response to Bd exposure would temporarily lag behind the temperature shift. Thus, we predicted that a temperature shift from cold-to-warm would result in an *increase* in susceptibility to Bd exposure, whereas a temperature shift from warm-to-cold would result in a *decrease* in susceptibility to Bd exposure. To test these predictions, we quantified susceptibility to Bd by measuring infection abundance after exposure to the pathogen.

Materials and Methods

To examine the how temperature shifts may alter larval amphibian infection dynamics, we selected two species of amphibian hosts, the northern red legged frog (*Rana aurora*) and the western toad (*Anaxyrus boreas*). Both species have been observed in the field with Bd infections (Pearl *et al.* 2007, Muths, Pilliod & Livo 2008, Piovia-Scott *et al.* 2011) and both species are susceptible to chytridiomycosis (Han, Bradley & Blaustein 2008, Gervasi *et al.* 2013). To ensure that the animals used in our experiment were not previously infected with Bd, amphibians were collected as eggs from natural oviposition sites. Red legged frog eggs were collected from a permanent pond located near Florence, Oregon, USA (Lincoln County, elevation 12 m; latitude/longitude: 44.088/-124.123) in the Oregon Coast Range on 11-Feb-2012. Western toad eggs were collected from Little Three Creeks Lake (Deschutes County, elevation 2,000 m; latitude/longitude: 44.009/-121.643) in the Cascade Range on 9-Jul-2011. Immediately after collection, eggs were transported to a laboratory at Oregon State University where

they were maintained at 14° C, under a 12-12 photoperiod in 40-liter aquaria filled with dechlorinated water. Upon hatching, larvae were maintained at a density of approximately 200 individuals per aquarium and fed *ad libitum* a mixture of Tetramin fish food and ground alfalfa pellets (1:3 ratio by volume). Water was changed every seven days. The 40-day trials for each species were not run concurrently, but identical protocols were used for both species and both trials consisted of individuals of identical larval stage (Gosner stage 26).

Acclimation Period

184

185

186

187

188

189

190

191

192

193

194

195

196

197

198

199

200

201

202

203

204

205

206

Independent trials for each host species began with a 20-day acclimation period with 80 (Gosner stage 26) larvae randomly selected, and individually placed into 80 plastic 500-mL containers where they were housed for the duration of the acclimation period and experiment. Each container was filled with 14° C dechlorinated water and covered with a lid to help maintain water temperature and limit evaporation. Each container had 2-mm diameter holes drilled between the water line and the lid to allow air circulation into the container. Pairs of containers were then placed within 40 individual temperature-controlled chambers (to ensure independent replication of the temperature treatments) that were set at 15° C to avoid cold-shocking the larvae. Each temperaturecontrolled chamber was independently controlled via its own thermostat and the interior measured approximately 37 cm deep x 21 cm wide x 13 cm in height. Half of the 40 temperature-controlled chambers were then randomly selected to begin the acclimation period at 20° C (warm treatment) and the other half were kept at 15° C (cold treatment). The placement of temperature chambers within the laboratory was randomized, as was the placement of 500-mL containers within each temperature chamber.

Temperature Shifts

207

208

209

210

211

212

213

214

215

216

217

218

219

220

221

222

223

224

225

226

227

228

229

On day 20 of the experiment, half of the temperature chambers in each of the two acclimation temperatures (15° C and 20° C) were randomly selected to undergo a temperature shift, either from 20° to 15° C or from 15° C to 20° C. The other half of the temperature chambers underwent no shift in temperature. Thus, each of the temperature chambers was subjected to one of four temperature treatments: a constant 15° C (cold) throughout the experiment; a constant 20° C (warm) throughout the experiment; a temperature shift from 15° C to 20° C (cold-to-warm); or a temperature shift from 20° C to 15°C (warm-to-cold). Bd exposure On day 24, one of the two 500-mL containers within each temperature-controlled

chamber was randomly selected to undergo a Bd-exposure treatment and the other was selected as a control. Larvae in the Bd-exposure treatment were exposed to a single inoculate of Bd strain JEL 274, which was grown in pure culture on 1% tryptone agar in 10-cm diameter Petri dishes. This Bd strain was selected as it is one of the more virulent strains associated with major amphibian populations declines (Rosenblum et al. 2013). The Petri dishes were inoculated with liquid culture 10 days before the start of the experiment and incubated at 15° C. To harvest the zoospores, 10 plates were flushed with 15 mL of 15° C dechlorinated water and remained undisturbed for 10 minutes. The plates were scraped with a rubber spatula to release the zoospores and sporangia adhering to the agar. The inoculum from each plate was then pooled in a beaker and the number of moving zoospores was determined using a hemocytometer. After quantifying the zoospore concentration, the inoculum was diluted to 10,000 zoospores/mL. Individuals in

231

232

233

234

235

236

237

238

239

240

241

242

243

244

245

246

247

248

249

250

251

252

the Bd-exposed treatments were exposed to 10 mL of inoculum transferred into the 500mL container housing an individual larva. Control individuals were exposed to 10 mL of sham inoculum lacking the Bd culture (made from 1% tryptone sterile agar plates following the same methods), similarly transferred into the 500-mL container housing each larva. Thus, the individual larva underwent their exposure treatment on day 24, four days after the water temperature shift for chambers in the two temperature shift treatments. During the 40-d trial, survival and metamorphic status were checked daily. Water for each 500-mL container within the temperature chambers was changed every 12 days and consisted of dechlorinated water of the same temperature (15° C and 20° C). Individuals that survived until the end of the trial (i.e., day 40) were euthanized in a 2% solution of MS-222, and then preserved in 95% ethanol. Individuals that reached metamorphosis (Gosner stage 42: emergence of forelimbs) were euthanized, measured, and preserved as previously described. Determining infection status We used quantitative polymerase chain reaction (qPCR) to determine infection status and quantify Bd-infection intensity of all individuals in the Bd-exposure treatments. Additionally, we investigated Bd-infection status in eight randomly selected control individuals per species. To sample the individuals for Bd, we extracted whole mouthparts of the larvae using sterile dissection scissors. We conducted qPCR using an ABI PRISM 7500 sequencer (Applied Biosystems) according to the methods of Boyle et al. (2004) except that we used 60 μL of Prepman Ultra (Applied Biosystems, Carlsbad, California, USA), instead of the 40 µL in the DNA extraction. All samples were run in

triplicate and averaged.

253

254

255

256

257

258

259

260

261

262

263

264

265

266

267

268

269

270

271

272

273

274

275

Statistical Analyses Each temperature-controlled chamber was an experimental unit (whole plot) and the pairs of containers within each chamber acted as subplots. The whole plots were subjected to one of four temperature regimes consisting of a Bd-exposure temperature combined with a temperature shift status (constant cold, constant warm, shifted to cold, and shifted to warm). Further, subplots were subjected to one of two exposure treatments (Bd exposed and Bd unexposed). Survival was compared between temperature treatments for western toad larvae with a Cox proportional hazards model (Cox 1972) using TIBCO Spotfire S+ version 8.1. The model consisted of the main effects of the temperature treatment, temperature shift status (constant versus shifted), and an interaction between the two variables. Due to losses of western toad larvae prior to the application of the exposure treatment, we lacked the power to statistically compare survival in western toad larvae between the Bd exposure treatments Bd infection abundance (Bd genomic equivalents) among temperature treatments

and between host species was analyzed using R version 3.11. We used a zero-inflated negative-binomial generalized linear model (function 'zeroinf' in package 'pscl) as described by Raffel et al. (2010), which includes a zero-inflation component that models infection status as a binomial process (binomial distribution with a logit link) and a count component that models infection intensity as a negative binomial process (negative binomial distribution with a log link). Our full model investigated the effects of all of the explanatory variables including host species, exposure temperature, temperature shift

277

278

279

280

281

282

283

284

285

286

287

288

289

290

291

292

293

294

295

296

297

298

status, and all two- and three-way interactions on Bd (Batrachochytrium dendrobatidis) abundance. Interpretation of this analysis required further reduced models to investigate the effect of exposure temperature and temperature shift for each species (species model) and the effect of temperature shift for each Bd-exposure temperature and host species combination (Bd-exposure temperature model). Results Survival differences were not detected between exposure temperatures (Cox, Z = -1.099, p = 0.27) or temperature shift status (Cox, Z = -0.277, p = 0.78) in Bdexposed western toad larvae. We were unable to detect survival differences in red legged frog larvae, as only one individual larva experienced mortality after application of the exposure treatment (Table S1). Infection Abundance We detected a host species by temperature shift interaction ($\chi^2_1 = 3.83$, p = 0.050; Table S2) and a Bd-exposure temperature by temperature shift interaction ($\chi^2 = 7.50$, p =0.006; Table S2). We investigated these interactions with reduced models to investigate effects on Bd abundance at the levels of species and exposure temperature. Red legged frog larvae had higher Bd abundance when they were exposed to infection at 15° C when compared to 20° C ($\chi^2_1 = 3.88$, p = 0.049; Fig. 2). The main

effect of temperature shift was marginally significant in the reduced species model

analysis ($\chi^2_1 = 3.50$, p = 0.061), but there was a significant effect of temperature shift for

individuals exposed at 20° C in the reduced Bd-exposure model ($\chi^2_1 = 5.7$, p = 0.017),

300

301

302

303

304

305

306

307

308

309

310

311

312

313

314

315

316

317

318

319

320

321

with individuals shifted from 15° C to 20° C having higher Bd abundance than red legged frog larvae experiencing constant 20° C (Fig. 2). In contrast, there was no evidence that a temperature shift influenced Bd infection when red legged frog larvae were exposed to Bd at 15° C ($\chi^2_1 = 0.6$, p = 0.4; Fig. 2). There was no statistically significant interaction between exposure temperature and temperature shift for red legged frog larvae ($\chi^2_1 = 2.4$, p = 0.13). We detected an interactive effect of exposure temperature and temperature shift on Bd abundance in western toad larvae ($\chi^2_1 = 5.2$, p = 0.023). This was driven by elevated Bd abundance in individuals under the constant 15° C temperature when compared to individuals that experienced a temperature shift from 20° to 15° C, but no evidence of an effect of shifting temperature from 15° C to 20° C (Fig. 2). There were no main effects of exposure temperature ($\chi^2_1 = 0.50$, p = 0.5) or temperature shift ($\chi^2_1 < 0.01$, p = 0.9) on Bd abundance in western toad larvae. Further, when investigating the exposure temperatures individually in the reduced Bd-exposure model, there was no evidence that a temperature shift influenced Bd infection in western toad larvae after exposure to Bd at 15° C (χ^2 ₁ = 3.4, p = 0.066) or 20° C (χ^2 ₁ = 2.5, p = 0.11). We failed to find evidence that the two host species differed in response to exposure to the pathogen, leading us to conclude that general patterns for both species were similar (Fig. 2). Both species experienced an increase in Bd abundance when shifted to 20° C compared to a constant 20° C, and both generally experienced a decrease in Bd abundance when shifted to 15° C compared to a constant 15° C. Additionally, both host species experienced elevated Bd abundance in the constant 15° C treatment when compared to the constant 20° treatment.

All red legged frog individuals survived until the end of the experiment but a number of western toad individuals died or metamorphosed earlier (Table S2). We therefore assessed the possibility that the timing of Bd sampling or the proximity of a larva to metamorphosis might drive observed patterns of Bd abundance in western toads. The model for Bd abundance on western toads was not significantly improved by adding either a variable coding whether individuals were near metamorphosis when sampled (χ^2_1 = 4.00, p = 0.150) or a covariate indicating the sampling date (χ^2_1 = 3.33, p = 0.068). Furthermore, neither variable qualitatively changed the contribution of exposure temperature or temperature shift status to the model. Therefore, we omitted both covariates from the final model for western toads.

Discussion

Our results suggest that Bd infection dynamics in larval amphibians can be affected by a shift in water temperature before host exposure to the pathogen, and that the direction of temperature shift determines the outcome of Bd exposure. Similar patterns were observed for the two host species when comparing individuals exposed to constant versus shifted temperatures. A shift from the warm temperature to the colder temperature was associated with a significant decrease in Bd abundance in western toad larvae and no significant decrease in red legged frog larvae. Likewise, a shift from the cold temperature to the warmer temperature significantly increased Bd abundance in red legged frog larvae and had no significant effect in western toad larvae. Importantly, we detected the effects of temperature shifts despite the host having a four-day head start on acclimating to the Bd exposure temperature relative to the pathogen. This suggests that we are likely

underestimating the strength of these effects and that their magnitudes might have been larger if the host and pathogen experienced the shifts concurrently.

345

346

347

348

349

350

351

352

353

354

355

356

357

358

359

360

361

362

363

364

365

366

367

Amphibian species do not all respond similarly to a given Bd exposure. Specieslevel differences in host tolerance to Bd infections have been well documented under controlled laboratory conditions (Searle et al. 2011, Gervasi et al. 2017). Under natural conditions, pathogen tolerance within a species may be affected by biotic factors such as inter- and intra-specific interactions, proximity to metamorphosis, or life stage (Parris & Cornelius 2004, Rachowicz & Vredenburg 2004, Blaustein et al. 2005, McMahon & Rohr 2015) or abiotic factors such as temperature, season, or resource availability (Berger et al. 2004, Raffel et al. 2010). For some susceptible host species, temperature-shift induced changes in Bd abundance might alter the outcome of infection by either pushing Bd abundance over or under a tolerance threshold. Such changes in relation to pathogen abundance and pathogen tolerance may result in altering the strength of negative effects of Bd infection. For example, temperature shifts in synergy with Bd infection may result in either positive or negative effects on growth and development rates, foraging efficiency, or predator avoidance (Parris & Cornelius 2004, Parris, Reese & Storfer 2006, Venesky, Parris & Storfer 2010, Venesky, Wassersug & Parris 2010). We hypothesized that hosts exposed to a shifted temperature would respond to infection differently than hosts exposed to a constant temperature, and under the

We hypothesized that hosts exposed to a shifted temperature would respond to infection differently than hosts exposed to a constant temperature, and under the framework of the "lag effect" hypothesis (Raffel *et al.* 2006, Rohr & Raffel 2010), the direction of the temperature shift would differentially affect infection severity. We predicted that a temperature shift from cold-to-warm would leave hosts in a temporarily immune-compromised state and result in an elevated Bd abundance after exposure when

369

370

371

372

373

374

375

376

377

378

379

380

381

382

383

384

385

386

387

388

389

compared to hosts exposed to a constant warm temperature. Conversely, we predicted that a temperature shift from warm-to-cold would provide hosts with a temporarily elevated-immune responsiveness and result in a decrease in Bd abundance after exposure when compared to hosts exposed to a constant cold temperature. Our results were consistent with predictions of the "lag effect" hypothesis, and were generally consistent with previous studies showing that a shift in temperature influences Bd infection in postmetamorphic amphibians (Raffel et al. 2013, Raffel et al. 2015). In particular, our finding of decreased resistance to infection following a temperature increase (relative to warm-acclimated individuals) mirrored a laboratory study of post-metamorphic redspotted newts (Notophthalmus viridescens), where juvenile newts exhibited decreased Bd resistance following a shift from 15° C to 25° C (Raffel et al. 2015). These findings of fluctuating temperature effects on Bd infection across four anuran taxonomic groups and life-stages suggest that effects of temperature shifts and Bd-related chytridiomycosis susceptibility might be widespread within amphibians. However, our finding of increased resistance to Bd infection following a temperature decrease (relative to cold-acclimated individuals) was opposite the pattern observed in red-spotted newts and Cuban treefrogs (Raffel et al. 2013, Raffel et al. 2015) These contrasting results suggests that there are important among-taxa or among-stage differences in the underlying mechanisms driving the effects of temperature fluctuation on Bd infection; whereas our results in premetamorphic life-stage of western toads and red legged frogs are consistent with the "lag effect" hypothesis, results of similar studies investigating post-metamorphic red-spotted newts and Cuban treefrogs support the "climate variability hypothesis."

391

392

393

394

395

396

397

398

399

400

401

402

403

404

405

406

407

408

409

410

411

412

We observed differences in Bd abundance on our two amphibian species at the two constant temperature treatments. Higher Bd abundances were observed for both host species under the constant cold temperature treatment compared to the constant warm temperature treatment. These results are consistent with previous experiments that showed increased Bd abundance (Raffel et al. 2015) and Bd-induced mortality (Kilpatrick, Briggs & Daszak 2010, Murphy, St-Hilaire & Corn 2011, Raffel et al. 2015) associated with lower temperatures. This is despite Bd growing best in culture at about 23° C, which is much closer to the warm than cold temperature treatments in this experiment (Piotrowski, Annis & Longcore 2004, Woodhams et al. 2008). This might be because the larval immune response to Bd infection increases with increasing temperatures at a faster rate than the infectivity or growth rate of Bd (Raffel et al. 2013), or alternatively because of the differences between the growth rate of Bd in culture compared to the growth rate on host tissue (Venesky et al. 2013). Our results provide further evidence to suggest patterns of Bd growth in culture differ from patterns of Bd growth on a host and that it is important to assess the host-parasite interaction when predicting effects of climate and climate change on disease risk. Alternatively, differences in Bd abundance between the two constant temperature treatments might be due to temperature effects on the pathogen rather than the host (Woodhams et al. 2008, Voyles et al. 2012). The Bd was cultured at 15° C; it is possible that the temperature shift experienced by the pathogen in the warm exposure treatment caused the depressed Bd abundances observed in both host species compared to the elevated Bd abundance in the cold exposure temperature treatment. A decrease in temperature may cause an increase in the number of Bd zoospores released from

zoosporangia (Hyatt *et al.* 2007, Woodhams *et al.* 2008), however the effect of a similar increase in temperature on Bd physiology is unclear.

In conclusion, our results provide additional evidence for climate variability affecting Bd infection in amphibians but suggest important among-taxa differences in the directionality of these effects. Our finding of increased host resistance to infection following a temperature decrease is consistent with the "lag effect" hypothesis of Raffel et al. (2006) but contradicts components of the "climate variability hypothesis", which has been proposed as an explanation for patterns of Bd-associated amphibian population declines (Rohr & Raffel 2010, Raffel et al. 2013, Raffel et al. 2015). Our study highlights the complexity that temperature plays in determining the outcome of Bd-amphibian interactions and the role that a fluctuating temperature might play in altering these interactions. Furthermore, this study increases the diversity of amphibian species and stages that have been shown to exhibit thermal acclimation effects on disease, and the broad generality of this pattern across four disparate taxa suggests that fluctuatingtemperature effects on amphibian infection may be widespread. Accurately predicting the effects of global climate change on infectious diseases, such as chytridiomycosis will require further understanding of how infectious agents respond to heterogeneity in temperatures and temperature fluctuations.

Acknowledgments

413

414

415

416

417

418

419

420

421

422

423

424

425

426

427

428

429

430

431

432

433

434

435

All applicable institutional and national guidelines for the care and use of animals were followed; this research was conducted under Oregon State University IACUC animal care and use permit 3917. Collection of amphibian eggs was approved by the

437

438

439

440

441

442

443

444

445

446

447

448

449

450

451

452

453

material for this article.

Oregon Department of Fish and Wildlife (Oregon Scientific Taking Permit #006-12 issued to ARB). We thank S. Bauer, E. Davis, E. Hunt, A. Koosman, B. Meyers, M. Ouspenskaya, E. Peseke, V. Raffeale, and C. Rains for their help performing the experiment, K. Boersma for her help with the experimental design, and E. Boersley for her support and assistance. Additionally we thank J. Spatafora, V. Weis, and the Center for Genome Research and Biocomputing at Oregon State University for providing laboratory space for qPCR. This research was supported by grants from the National Science Foundation (EF-1241889), National Institutes of Health (R01GM109499, R01TW010286), U.S. Department of Agriculture (NRI 2006-01370, 2009-35102-0543), and U.S. Environmental Protection Agency (CAREER 83518801) to JRR and NSF grant IOS 1121529 to TRR. Support was provided by the U.S. Forest Service Pacific Northwest Research Station, Corvallis, Oregon to DHO. Conflict of Interest: The authors declare that they have no conflict of interest. **Supporting Information** Additional supporting information may be found in electronic supplementary

454 **Literature Cited** 455 Alford, R. A. (1989) Variation in predator phenology affects predator performance 456 and prey community composition. *Ecology*, **70**, 206-219. 457 Angilletta, M. J. (2009) Thermal adaptation a theoretical and empirical synthesis. 458 New York, New York. USA.: Oxford University Press. 459 Bannerman, J. A. & B. D. Roitberg (2014) Impact of extreme and fluctuating 460 temperatures on aphid-parasitoid dynamics. *Oikos*, **123**, 89-98. 461 Ben-Horin, T., H. S. Lenihan & K. D. Lafferty (2012) Variable intertidal temperature 462 explains why disease endangers black abalone. *Ecology*, **94**, 161-168. 463 Berger, L., R. Speare, H. B. Hines, G. Marantelli, A. D. Hyatt, K. R. McDonald, L. F. 464 Skerratt, V. Olsen, J. M. Clarke, G. Gillespie, M. Mahony, N. Sheppard, C. 465 Williams & M. J. Tyler (2004) Effect of season and temperature on mortality 466 in amphibians due to chytridiomycosis. Aust. Vet. J., 82, 434-439. 467 Blaustein, A. R., S. S. Gervasi, P. T. J. Johnson, J. T. Hoverman, L. K. Belden, P. W. 468 Bradley & G. Y. Xie (2012) Ecophysiology meets conservation: 469 understanding the role of disease in amphibian population declines. *Philos*. 470 Trans. R. Soc., B, 367, 1688-1707. 471 Blaustein, A. R., J. M. Romansic, E. A. Scheessele, B. A. Han, A. P. Pessier & J. E. 472 Longcore (2005) Interspecific variation in susceptibility of frog tadpoles to the 473 pathogenic fungus Batrachochytrium dendrobatidis. Conserv. Biol., 19, 1460-474 1468.

475 Blaustein, A. R., S. C. Walls, B. A. Bancroft, J. J. Lawler, C. L. Searle & S. S. 476 Gervasi (2010) Direct and indirect effects of climate change on amphibian 477 populations. *Diversity*, **2**, 281-313. 478 Bosch, J., L. Carrascal, L. Dur·n, S. Walker & M. Fisher (2007) Climate change and 479 outbreaks of amphibian chytridiomycosis in a montane area of central Spain; 480 is there a link? *Proceedings of the Royal Society B*, **274**, 253. 481 Boyle, D. G., D. B. Boyle, V. Olsen, J. A. T. Morgan & A. D. Hyatt (2004) Rapid 482 quantitative detection of chytridiomycosis (Batrachochytrium dendrobatidis) 483 in amphibian samples using real-time Taqman PCR assay. Dis. Aquat. Org., 484 **60,** 141-148. 485 Brönmark, C., S. D. Rundle & A. Erlandsson (1991) Interactions between freshwater 486 snails and tadpoles: competition and facilitation. *Oecologia*, **87**, 8-18. 487 Brown, J. H., J. F. Gillooly, A. P. Allen, V. M. Savage & G. B. West (2004) Toward 488 a metabolic theory of ecology. *Ecology*, **85**, 1771-1789. 489 Buck, J. C., E. A. Scheessele, R. A. Relyea & A. R. Blaustein (2012) The effects of 490 multiple stressors on wetland communities: pesticides, pathogens and 491 competing amphibians. Freshwat. Biol., **57**, 61-73. 492 Daskin, J. H., R. A. Alford & R. Puschendorf (2011) Short-term exposure to warm 493 microhabitats could explain amphibian persistence with Batrachochytrium 494 dendrobatidis. PLoS ONE, 6, e26215.

495 DeSantis, D. E. & R. G. Strauss, (1997) Cell biology and disorders of neutrophils. In: 496 *Clinical Hematology and Fundamentals of Hemostasis*: 265–282. D. M. 497 Harmening (Ed.). Davis, Philadelphia, Pennsylvania. USA. 498 Drew, A., E. J. Allen & L. J. S. Allen (2006) Analysis of climatic and geographic 499 factors affecting the presence of chytridiomycosis in Australia. Dis. Aquat. 500 Org., 68, 245-250. 501 Easterling, D. R., G. A. Meehl, C. Parmesan, S. A. Changnon, T. R. Karl & L. O. 502 Mearns (2000) Climate extremes: observations, modeling, and impacts. 503 Science, **289**, 2068-2074. 504 Forrest, M. J. & M. A. Schlaepfer (2011) Nothing a hot bath won't cure: infection 505 rates of amphibian chytrid fungus correlate negatively with water temperature 506 under natural field settings. *PLoS ONE*, **6**, e28444. 507 Garcia, R. A., M. Cabeza, C. Rahbek & M. B. Araujo (2014) Multiple dimensions of 508 climate change and their implications for biodiversity. *Science*, **344**, 1247579. 509 Garner, T. W. J., S. Walker, J. Bosch, S. Leech, J. M. Rowcliffe, A. A. Cunningham 510 & M. C. Fisher (2009) Life history tradeoffs influence mortality associated 511 with the amphibian pathogen Batrachochytrium dendrobatidis. Oikos, 118, 512 783-791. 513 Gervasi, S., C. Gondhalekar, D. H. Olson & A. R. Blaustein (2013) Host identity 514 matters in the amphibian-Batrachochytrium dendrobatidis system: fine-scale

515 patterns of variation in responses to a multi-host pathogen. PLoS ONE, 8, 516 e54490. 517 Gervasi, S. S., P. R. Stephens, J. Hua, C. L. Searle, G. Y. Xie, J. Urbina, D. H. Olson, 518 B. A. Bancroft, V. Weis, J. I. Hammond, R. A. Relyea & A. R. Blaustein 519 (2017) Linking Ecology and Epidemiology to Understand Predictors of Multi-520 Host Responses to an Emerging Pathogen, the Amphibian Chytrid Fungus. 521 *PLoS ONE*, **12**, e0167882. 522 Gillooly, J. F., J. H. Brown, G. B. West, V. M. Savage & E. L. Charnov (2001) 523 Effects of size and temperature on metabolic rate. *Science*, **293**, 2248-2251. 524 Han, B. A., P. W. Bradley & A. R. Blaustein (2008) Ancient behaviors of larval 525 amphibians in response to an emerging fungal pathogen, Batrachochytrium 526 dendrobatidis. Behav. Ecol. Sociobiol., 63, 241-250. 527 Hyatt, A. D., D. G. Boyle, V. Olsen, D. B. Boyle, L. Berger, D. Obendorf, A. Dalton, 528 K. Kriger, J. M. Hero, H. Hines, R. Phillott, R. Campbell, G. Marantelli, F. 529 Gleason & A. Colling (2007) Diagnostic assays and sampling protocols for 530 the detection of Batrachochytrium dendrobatidis. Dis. Aquat. Org., 73, 175-531 192. 532 Janeway, C. (2008) Janeway's Immunobiology, 7 edn. New York, New York. USA.: 533 Garland Science.

534 Kilpatrick, A. M., C. J. Briggs & P. Daszak (2010) The ecology and impact of 535 chytridiomycosis: an emerging disease of amphibians. Trends Ecol. Evol., 25, 536 109-118. 537 Kupferberg, S. J. (1997) The role of larval diet in anuran metamorphosis. Am. Zool., 538 **37,** 146-159. 539 Lafferty, K. D. (2009) The ecology of climate change and infectious diseases. 540 Ecology, **90**, 888-900. 541 Lamberti, G., S. Gregory, C. Hawkins, R. Wildman, L. Ashkenas & D. Denicola 542 (1992) Plant—herbivore interactions in streams near Mount St Helens. 543 Freshwat. Biol., 27, 237-247. 544 Lawler, J. J., S. L. Shafer, B. A. Bancroft & A. R. Blaustein (2010) Projected climate 545 impacts for the amphibians of the Western hemisphere. Conserv. Biol., 24, 38-546 50. 547 Li, Y., J. M. Cohen & J. R. Rohr (2013) Review and synthesis of the effects of 548 climate change on amphibians. Integr Zool, 8, 145-161. 549 Liu, X., J. R. Rohr & Y. Li (2013) Climate, vegetation, introduced hosts and trade 550 shape a global wildlife pandemic. *Proceedings of the Royal Society B*: 551 *Biological Sciences*, **280**, 20122506. 552 Longcore, J., A. Pessier & D. Nichols (1999) Batrachochytrium dendrobatidis gen. et 553 sp. nov., a chytrid pathogenic to amphibians. Mycologia, **91**, 219-227.

554 Luis, A. D., R. J. Douglass, J. N. Mills & O. N. Bjørnstad (2014) Environmental 555 fluctuations lead to predictability in Sin Nombre hantavirus outbreaks. 556 Ecology, **96**, 1691-1701. 557 Macela, A. & A. Romanovsky (1970) The role of temperature in separate stages of 558 the immune reaction in anurans. *Folia Biologica*, **15**, 157-160. 559 Maniero, G. D. & C. Carey (1997) Changes in selected aspects of immune function in 560 the leopard frog, Rana pipiens, associated with exposure to cold. J. Comp. 561 Physiol., B, 167, 256-263. 562 Marantelli, G., L. Berger, R. Speare & L. Keegan (2004) Distribution of the 563 amphibian chytrid Batrachochytrium dendrobatidis and keratin during tadpole 564 development. Pac. Conserv. Biol., 10, 173-179. 565 Martel, A., A. Spitzen-van der Sluijs, M. Blooi, W. Bert, R. Ducatelle, M. C. Fisher, 566 A. Woeltjes, W. Bosman, K. Chiers, F. Bossuyt & F. Pasmans (2013) 567 Batrachochytrium salamandrivorans sp. nov. causes lethal chytridiomycosis 568 in amphibians. Proc. Natl. Acad. Sci. U. S. A., 110, 15325-15329. 569 McCallum, M. L. (2007) Amphibian decline or extinction? Current declines dwarf 570 background extinction rate. J. Herpetol., 41, 483-491. 571 McMahon, T. A. & J. R. Rohr (2015) Transition of chytrid fungus infection from 572 mouthparts to hind limbs during amphibian metamorphosis. EcoHealth, 12, 573 188-193.

574 McMahon, T. A., B. F. Sears, M. D. Venesky, S. M. Bessler, J. M. Brown, K. 575 Deutsch, N. T. Halstead, G. Lentz, N. Tenouri, S. Young, D. J. Civitello, N. 576 Ortega, J. S. Fites, L. K. Reinert, L. A. Rollins-Smith, T. R. Raffel & J. R. 577 Rohr (2014) Amphibians acquire resistance to live and dead fungus 578 overcoming fungal immunosuppression. *Nature*, **511**, 224-227. 579 Meehl, G. A. & C. Tebaldi (2004) More intense, more frequent, and longer lasting 580 heat waves in the 21st century. Science, **305**, 994-997. 581 Murphy, P. J., S. St-Hilaire & P. S. Corn (2011) Temperature, hydric environment, 582 and prior pathogen exposure alter the experimental severity of 583 chytridiomycosis in boreal toads. Dis. Aquat. Org., 95, 31-42. 584 Muths, E., D. S. Pilliod & L. J. Livo (2008) Distribution and environmental 585 limitations of an amphibian pathogen in the Rocky Mountains, USA. *Biol.* 586 Conserv., 141, 1484-1492. 587 Olson, D. H., D. M. Aanensen, K. L. Ronnenberg, C. I. Powell, S. F. Walker, J. 588 Bielby, T. W. J. Garner, G. Weaver, M. C. Fisher & T. B. M. Group (2013) 589 Mapping the global emergence of *Batrachochytrium dendrobatidis*, the 590 amphibian chytrid fungus. *PLoS ONE*, **8,** e56802. 591 Paaijmans, K. P., S. Blanford, A. S. Bell, J. I. Blanford, A. F. Read & M. B. Thomas 592 (2010) Influence of climate on malaria transmission depends on daily 593 temperature variation. *Proc. Natl. Acad. Sci. U. S. A.*, **107**, 15135-15139.

594 Paaijmans, K. P., A. F. Read & M. B. Thomas (2009) Understanding the link between 595 malaria risk and climate. Proc. Natl. Acad. Sci. U. S. A., 106, 13844-13849. 596 Parris, M., E. Reese & A. Storfer (2006) Antipredator behavior of chytridiomycosis-597 infected northern leopard frog (Rana pipiens) tadpoles. Can J Zool, 84, 58-65. 598 Parris, M. J. & T. O. Cornelius (2004) Fungal pathogen causes competitive and 599 developmental stress in larval amphibian communities. *Ecology*, **85**, 3385-600 3395. 601 Paull, S. H., B. E. LaFonte & P. T. J. Johnson (2012) Temperature-driven shifts in a 602 host-parasite interaction drive nonlinear changes in disease risk. Global 603 Change Biol., 18, 3558-3567. 604 Pearl, C. A., E. L. Bull, D. E. Green, J. Bowerman, M. J. Adams, A. Hyatt & W. H. 605 Wente (2007) Occurrence of the amphibian pathogen *Batrachochytrium* 606 dendrobatidis in the Pacific Northwest. J. Herpetol., 41, 145-149. 607 Piotrowski, J. S., S. L. Annis & J. E. Longcore (2004) Physiology of 608 Batrachochytrium dendrobatidis, a chytrid pathogen of amphibians. 609 *Mycologia*, **96**, 9-15. 610 Piovia-Scott, J., K. L. Pope, S. P. Lawler, E. M. Cole & J. E. Foley (2011) Factors 611 related to the distribution and prevalence of the fungal pathogen 612 Batrachochytrium dendrobatidis in Rana cascadae and other amphibians in 613 the Klamath Mountains. Biol. Conserv., 144, 2913–2921.

614 Rachowicz, L. J. & V. T. Vredenburg (2004) Transmission of Batrachochytrium 615 dendrobatidis within and between amphibian life stages. Dis. Aquat. Org., 61, 616 75-83. 617 Raffel, T. R., N. T. Halstead, T. A. McMahon, A. K. Davis & J. R. Rohr (2015) 618 Temperature variability and moisture synergistically interact to exacerbate an 619 epizootic disease. Proceedings of the Royal Society B: Biological Sciences, 620 **282,** 20142039. 621 Raffel, T. R., P. J. Michel, E. W. Sites & J. R. Rohr (2010) What drives chytrid 622 infections in newt populations? Associations with substrate, temperature, and 623 shade. *EcoHealth*, **7**, 526-536. 624 Raffel, T. R., J. R. Rohr, J. M. Kiesecker & P. J. Hudson (2006) Negative effects of 625 changing temperature on amphibian immunity under field conditions. Funct. 626 *Ecol.*, **20**, 819-828. 627 Raffel, T. R., J. M. Romansic, N. T. Halstead, T. A. McMahon, M. D. Venesky & J. 628 R. Rohr (2013) Disease and thermal acclimation in a more variable and 629 unpredictable climate. Nat. Clim. Change, 3, 146–151. 630 Rohr, J. R., A. P. Dobson, P. T. Johnson, A. M. Kilpatrick, S. H. Paull, T. R. Raffel, 631 D. Ruiz-Moreno & M. B. Thomas (2011) Frontiers in climate change-disease 632 research. Trends Ecol. Evol., 26, 270-277.

633 Rohr, J. R. & T. R. Raffel (2010) Linking global climate and temperature variability 634 to widespread amphibian declines putatively caused by disease. *Proc. Natl.* 635 Acad. Sci. U. S. A., 107, 8269-8274. 636 Rohr, J. R., T. R. Raffel, A. R. Blaustein, P. T. J. Johnson, S. H. Paull & S. Young 637 (2013) Using physiology to understand climate-driven changes in disease and 638 their implications for conservation. *Conserv. Physiol.*, **1**, cot022. 639 Rohr, J. R., T. R. Raffel, J. M. Romansic, H. McCallum & P. J. Hudson (2008) 640 Evaluating the links between climate, disease spread, and amphibian declines. 641 Proc. Natl. Acad. Sci. U. S. A., 105, 17436. 642 Rosenblum, E. B., T. Y. James, K. R. Zamudio, T. J. Poorten, D. Ilut, D. Rodriguez, 643 J. M. Eastman, K. Richards-Hrdlicka, S. Joneson, T. S. Jenkinson, J. E. 644 Longcore, G. Parra Olea, L. F. Toledo, M. L. Arellano, E. M. Medina, S. 645 Restrepo, S. V. Flechas, L. Berger, C. J. Briggs & J. E. Stajich (2013) 646 Complex history of the amphibian-killing chytrid fungus revealed with 647 genome resequencing data. Proc. Natl. Acad. Sci. U. S. A., 110, 9385-9390. 648 Rummukainen, M. (2012) Changes in climate and weather extremes in the 21st 649 century. Wiley Interdiscip. Rev.: Clim. Change, 3, 115-129. 650 Schar, C., P. L. Vidale, D. Luthi, C. Frei, C. Haberli, M. A. Liniger & C. Appenzeller 651 (2004) The role of increasing temperature variability in European summer 652 heatwaves. *Nature*, **427**, 332-336.

653 Searle, C. L., S. S. Gervasi, J. Hua, J. I. Hammond, R. A. Relyea, D. H. Olson & A. 654 R. Blaustein (2011) Differential host susceptibility to *Batrachochytrium* 655 dendrobatidis, an emerging amphibian pathogen. Conserv. Biol., 25, 965-974. 656 Sheldon, K. S., S. Yang & J. J. Tewksbury (2011) Climate change and community 657 disassembly: impacts of warming on tropical and temperate montane community structure. Ecol. Lett., 14, 1191-1200. 658 659 Shoo, L. P., D. H. Olson, S. K. McMenamin, K. A. Murray, M. Van Sluys, M. A. 660 Donnelly, D. Stratford, J. Terhivuo, A. Merino-Viteri, S. M. Herbert, P. J. 661 Bishop, P. S. Corn, L. Dovey, R. A. Griffiths, K. Lowe, M. Mahony, H. McCallum, J. D. Shuker, C. Simpkins, L. F. Skerratt, S. E. Williams & J.-M. 662 663 Hero (2011) Engineering a future for amphibians under climate change. J. 664 Appl. Ecol., 48, 487-492. 665 Skerratt, L., L. Berger, R. Speare, S. Cashins, K. McDonald, A. Phillott, H. Hines & 666 N. Kenyon (2007) Spread of chytridiomycosis has caused the rapid global 667 decline and extinction of frogs. *EcoHealth*, **4**, 125-134. 668 Stuart, S. N., J. S. Chanson, N. A. Cox, B. E. Young, A. S. L. Rodrigues, D. L. 669 Fischman & R. W. Waller (2004) Status and trends of amphibian declines and 670 extinctions worldwide. Science, 306, 1783-1786. 671 Venesky, M., M. Parris & A. Storfer (2010) Impacts of *Batrachochytrium* 672 dendrobatidis infection on tadpole foraging performance. EcoHealth, 6, 565-673 575.

674 Venesky, M. D., T. R. Raffel, T. A. McMahon & J. R. Rohr (2013) Confronting 675 inconsistencies in the amphibian-chytridiomycosis system: implications for 676 disease management. Biol. Rev. Camb. Philos. Soc., 89, 477-483. 677 Venesky, M. D., R. J. Wassersug & M. J. Parris (2010) Fungal pathogen changes the 678 feeding kinematics of larval anurans. J. Parasitol., 96, 552-557. 679 Voyles, J., L. R. Johnson, C. J. Briggs, S. D. Cashins, R. A. Alford, L. Berger, L. F. 680 Skerratt, R. Speare & E. B. Rosenblum (2012) Temperature alters 681 reproductive life history patterns in Batrachochytrium dendrobatidis, a lethal 682 pathogen associated with the global loss of amphibians. Ecology and 683 Evolution, 2, 2241-2249. 684 Voyles, J., L. R. Johnson, J. Rohr, R. Kelly, C. Barron, D. Miller, J. Minster & E. B. 685 Rosenblum (2017) Diversity in growth patterns among strains of the lethal 686 fungal pathogen Batrachochytrium dendrobatidis across extended thermal 687 optima. Oecologia, 1-11. 688 Wake, D. B. (2012) Facing extinction in real time. Science, 335, 1052-1053. 689 Woodhams, D. C., R. A. Alford, C. J. Briggs, M. Johnson & L. A. Rollins-Smith 690 (2008) Life-history trade-offs influence disease in changing climates: 691 strategies of an amphibian pathogen. *Ecology*, **89**, 1627-1639. 692 Xie, G. Y., D. H. Olson & A. R. Blaustein (2016) Projecting the Global Distribution 693 of the Emerging Amphibian Fungal Pathogen, *Batrachochytrium* 694 dendrobatidis, Based on IPCC Climate Futures. PLoS ONE, 11, e0160746.

698

699

700

701

702

703

704

705

706

707

708

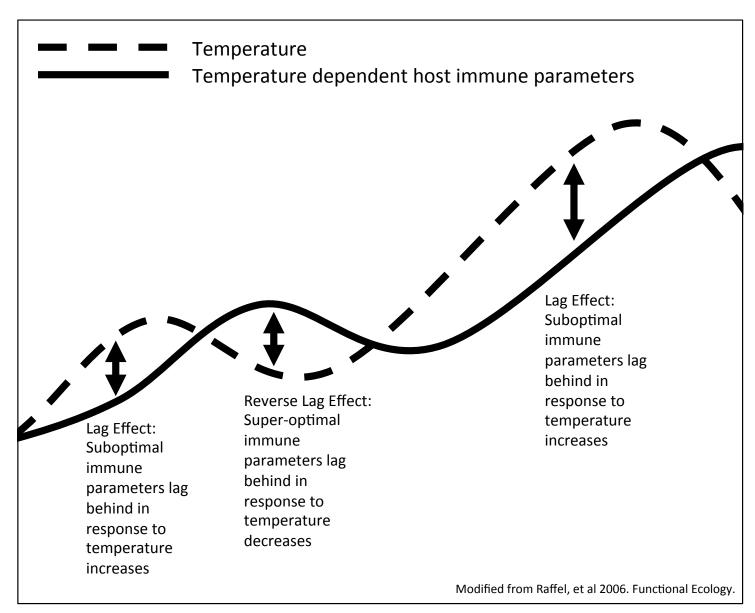
709

710

711

712

Fig 1. Hypothesized lag effect showing the relationship between fluctuating temperatures (over days to weeks) and the optimal levels of a hypothetical temperature-dependent host immune parameter. The immune parameter follows and lags behind temperature changes – resulting in periods of a compromised immune status after a temperature increase, and resulting in an over-active (or unnecessarily costly) immune status after a temperature decrease. Modified from Raffel et al. (2006).Fig 2. Mean Batrachochytrium dendrobatidis (Bd) infection abundance (± SE) measured at death, or at euthanasia 16-days after Bd exposure, in both western toad (Anaxyrus boreas) larvae and red legged frog (Rana aurora) larvae from Oregon, USA, and between the two temperatures at the time of Bd-exposure (cold [15° C] versus warm [20° C]) and between larvae having experienced either a constant or shifted temperature. Bd infection abundance is quantified as the log (1 + Bd genomic equivalents) per excised larval mouthparts of all individuals exposed to the pathogen.



Time in Days

