

1 **Niche filtering, not interspecific resource competition, explains the geographical**
2 **co-occurrence of butterfly species**

3

4

5 Ryosuke Nakadai^{12*†}, Koya Hashimoto^{1*}, Takaya Iwasaki³, Yasuhiro Sato¹⁴

6 ¹Center for Ecological Research, Kyoto University, Hirano 2-509-3, Otsu, Shiga,
7 520-2113 Japan

8 ²Current address: Faculty of Science, University of the Ryukyus, Senbaru 1, Nishihara,
9 Okinawa, 903-0213 Japan.

10 ³Department of Biological Sciences, Faculty of Science, Kanagawa University,
11 Tsuchiya 2946, Hiratsuka, Kanagawa 259-1293 Japan

12 ⁴Current address: Department of Plant Life Sciences, Faculty of Agriculture, Ryukoku
13 University, Yokotani 1-5, Otsu, Shiga 520-2194 Japan.

14

15 *Equal contribution

16

17 †*Author Correspondence:* R. Nakadai

18 *Faculty of Science, University of the Ryukyus, Senbaru 1, Nishihara, Okinawa 903-0213,*
19 *Japan.*

20 *E-mail:* r.nakadai66@gmail.co.jp

21

22 Running head: Co-occurrence of butterfly species

23

24 Keywords: butterfly, co-occurrence, dispersal ability, climatic niche, herbivorous insect,
25 host use, taxonomic relatedness

26

27

28

Abstract

29 The relevance of interspecific resource competition in the context of community
30 assembly by herbivorous insects is a well-known topic in ecology. Most previous
31 studies focused on local species assemblies, that shared host plants. Few studies
32 evaluated species pairs within a single taxon when investigating the effects of host plant
33 sharing at the regional scale. Herein, we explore the effect of plant sharing on the
34 geographical co-occurrence patterns of 229 butterflies distributed across the Japanese
35 archipelago; we use two spatial scales (10×10 km and 1×1 km grids) to this end. We
36 considered that we might encounter one of two predictable patterns in terms of the
37 relationship between co-occurrence and host-sharing among butterflies. On the one
38 hand, host-sharing might promote distributional exclusivity attributable to interspecific
39 resource competition. On the other hand, sharing of host plants might promote
40 co-occurrence attributable to filtering by resource niche. At both grid scales, we found
41 significant negative correlations between host use similarity and distributional
42 exclusivity. Our results support the thesis that the butterfly co-occurrence pattern across
43 the Japanese archipelago is better explained by filtering via resource niche rather than
44 interspecific resource competition.

45

46

Introduction

47 Efforts to understand community assembly processes are of major importance in
48 ecological research (Diamond 1975, Cavender-Bares et al. 2009). All of dispersal
49 limitations, environmental filtering via both abiotic and biotic niches, and interspecific
50 interactions are thought sequentially to determine local community structures (reviewed
51 by Cavender-Bares et al. 2009). Of the various relevant factors, the significance of
52 interspecific interaction has often been assessed by examining how different species
53 co-occur spatially even when they share similar niches (Diamond 1975, Gotelli and
54 McCabe 2002). As different species with similar niches likely prefer similar

55 environmental habitats, but may compete strongly, recent studies have often compared
56 the significance of interspecific interactions (in terms of community assembly patterns)
57 from the viewpoint of niche filtering (Webb et al. 2002, Mayfield and Levine 2010).

58 In terrestrial ecosystems, many plant species are used by various herbivorous
59 insects as both food resources and habitats, suggesting that host use plays key roles in
60 both competition and filtering via resource niches. Earlier studies described niche
61 partitioning between co-occurring herbivores and considered that partitioning was
62 attributable to interspecific competition (Ueckert and Hansen 1971, Benson 1978,
63 Waloff 1979, Augustyn et al. 2016). Conversely, other studies described the frequent
64 co-occurrence of multiple herbivorous insect species on shared host plants despite the
65 fact that extensive niche overlap was in play (Ross 1957, Rathcke 1976, Bultman and
66 Faeth 1985, Hashimoto and Ohgushi 2017); the cited authors argued that the effects of
67 interspecific competition were relatively weak in terms of community organization
68 (Lawton and Strong 1981, Schoener 1983, Strong et al. 1984, Tack et al. 2009). Whether
69 the importance of competition was or was not supported, most previous studies focused
70 on how host plants shape the local assemblages of herbivorous insects. Therefore, the
71 question of whether interspecific resource competition is critical in terms of shaping
72 herbivore communities remains highly controversial (reviewed by Kaplan and Denno
73 2007). Very little is known about the extent to which the results of the cited studies can
74 be extrapolated to describe patterns, at the regional scale, of the distribution of
75 herbivorous insects within a single taxon.

76 The effects of interspecific competition and filtering via resource niches on
77 the co-occurrence patterns of herbivorous insects must be considered when attempting
78 to explain differences in species traits, with the exception of host use. For example,
79 climatic niches (often associated with differences in potential geographical distributions
80 in the absence of interspecific interactions; Warren et al. 2008, Takami and Osawa
81 2016) may drive niche filtering, which may in turn mediate the impact of host use on

82 assembly patterns. In addition, taxonomic relatedness should reflect niche similarity; the
83 niche of any organism should be partly determined by its phylogenetic history (Webb et
84 al. 2002). Thus, the outcomes of competition within shared niches should reflect both
85 resource and climatic factors (Mayfield and Levine 2010). As such factors may
86 correlate with host use by individual species, any focus on host use alone may yield
87 misleading results. Furthermore, dispersal ability may complicate assembly patterns;
88 these are often associated with both the extent of the geographical range and other
89 factors (Kneitel and Chase 2004). A high dispersal ability may enhance the extent of
90 co-occurrence. Thus, when attempting to explore the effects of competition and filtering
91 via resource niches on distribution patterns, it is important to consider all of taxonomic
92 relatedness, climatic niche preferences, and dispersal ability.

93 In the present study, we explored whether interspecific competition or niche
94 filtering better explained the geographical co-occurrence of a group of herbivorous
95 insects. We expected to discern one of two patterns when evaluating the significance of
96 interspecific competition in terms of the geographical patterns of species co-occurrence.
97 One possibility was that sharing of host plants would be associated with exclusive
98 distributions attributable to resource competition. The alternative was that sharing of
99 host plants would promote species co-occurrence attributable to filtering via resource
100 niches. The former pattern predicts that a positive correlation would be evident between
101 host use similarity (i.e., the extent of sharing of host species) and the exclusiveness of
102 geographic distribution. The latter pattern predicts that the correlation would be
103 negative. We focused on herbivorous butterflies of the Japanese archipelago, for the
104 following reasons. First, butterfly host specificity is relatively high; the insect larvae are
105 mostly leaf chewers (Novotny et al. 2010). Second, the Biodiversity Center of Japan has
106 extensive records of butterfly distributions. Third, forewing length (easily measured on
107 photographs) is a useful index of butterfly dispersal ability (Chai and Srygley 1990;
108 Shirôzu 2006). Finally, and most importantly, a great deal is known about the hosts

109 used by butterfly species (Saito et al. 2016). Thus, data on Japanese butterflies can be
110 used to explore the effects of host plant sharing and other factors on the co-occurrence
111 patterns at regional scales. Specifically, we determined the effect of host use similarity
112 on the exclusivity of geographical distribution between pairs of entirely herbivorous
113 Japanese butterflies; we considered taxonomic relatedness, climatic niche similarities,
114 and overall dispersal abilities in the course of our work.

115

116 **Methods**

117 *Study area*

118 The Japanese archipelago, including the Ryukyu Islands, forms a long chain of
119 continental islands lying off the eastern coast of Asia. The latitudinal range of the
120 archipelago (22°N to 45°N) embraces hemi-boreal, temperate, and subtropical zones.
121 The mean temperatures in the coldest and warmest months are -19.0°C and 31.5°C,
122 respectively; the annual precipitation ranges from 867 to 3,908 mm (Kubota et al.
123 2014).

124 *Study organisms*

125 The Japanese archipelago hosts over 280 species of butterflies of five families (the
126 Papilionidae, Pieridae, Lycaenidae, Nymphanidae, and Hesperidae) (Shirôzu 2006).
127 Over 95% of the larvae of Japanese butterflies feed on plants (Honda 2005). The host
128 plants are diverse and include both dicots and monocots (Saito et al. 2016). Lycaenid
129 butterflies include two non-herbivorous species, but all species of all other families are
130 exclusively herbivorous.

131 *Metadata compilation*

132 Butterfly census data are available on the website of the Biodiversity Center of Japan,
133 Ministry of the Environment (http://www.biodic.go.jp/index_e.html). We used the
134 results of the fourth and fifth censuses (1988 to 1991 and 1997 to 1998, respectively) of
135 the National Survey of the Natural Environment in Japan

136 (http://www.biodic.go.jp/kiso/15/do_kiso4.html). This database includes records of 273
137 species/subspecies of butterflies from the entire Japanese archipelago, in grid cells of
138 latitude 5 min and longitude 7.5 min (the Japanese Standard Second Mesh). These grid
139 dimensions are about 10 km × 10 km, and this grid is described below as the “10-km
140 grid.” Furthermore, the Biodiversity Center also contains records from grid cells of
141 latitude 30 s and longitude 45 s (the Japanese Standard Third Mesh). These grid
142 dimensions are about 1 km × 1 km, and this grid is described below as the “1-km grid.”
143 As processes driving community assemblies may vary between spatial scales
144 (Cavender-Bares et al. 2009), we evaluated data from both grids. We summarized data
145 at the species level, and converted all records into the presence or absence (1/0) of a
146 species in each grid. We used the taxonomy of Shirôzu (2006).

147 Data on host plants and forewing length were evaluated as possible variables
148 explaining, respectively, host use and dispersal ability. The host plants of 278 butterfly
149 species/subspecies were obtained from the data of Saito et al. (2016). Dispersal ability
150 was evaluated by reference to adult wing length. We compiled wing data on 284 species
151 using published illustrations (Shirôzu 2006). We used Image J software (Abramoff et al.
152 2004) to extract forewing lengths (in cm) from plates that included centimeter scale bars.
153 Multiple forewing lengths were extracted when individual, sexual, and/or geographical
154 variations were evident.

155 We assembled data on the distributions, host plants, climatic niches (described
156 below), and forewing lengths of 229 butterfly species. Twenty-four species were
157 excluded from the analysis, for the following reasons. First, the taxonomic status of
158 three species (*Papilio dehaanii*, *Pieris dulcinea*, and *Eurema hecabe*) changed in the
159 interval between the fourth and fifth biodiversity censuses (Inomata 1990; Shirôzu
160 2006). Thus, we excluded these species because identifications were unreliable. Second,
161 we excluded three species of non-herbivorous butterflies (*Taraka hamada*, *Spindasis*
162 *takanonis*, and *Niphanda fusca*). Finally, we excluded a further 18 species because the

163 models used to evaluate their ecological niches failed to satisfy the criteria that we
164 imposed (the details are in Appendix S2).

165 *Data analysis*

166 **Species distribution exclusiveness.** We used the checkerboard scores (C-scores; Stone
167 and Roberts 1990) to evaluate the exclusivity of distributions between each species pair.
168 We set r_i and r_j as the numbers of grids in which species i and j , respectively, were
169 present; the checker unit C_{ij} associated with the two species was defined as: $C_{ij} = (r_i -$
170 $S_{ij}) \times (r_j - S_{ij})$, where S_{ij} indicates the extent of co-presence (i.e., the number of grid
171 cells shared by the two species). Thus, the checker unit became larger as the two species
172 occurred more commonly in different grid cells. We simulated null models to allow the
173 observed checker units to be compared with stochastic distributions. We used the
174 method of Jonsson et al. (2001) to describe the frequencies of species occurrence and
175 randomized the presence/absence matrices for each pair of butterfly species. The null
176 models were run 999 times for each species pair. C_{obs} and C_{null} were the checker units of
177 the observed and null distributions, respectively; the checker unit was standardized as:
178 $C_{std} = (C_{obs} - C_{null})/SD_{null}$, where SD_{null} indicates the standard deviation of all checker
179 units of the null models. The checker unit of the null model, C_{null} , was the average
180 checker unit of all null models. Thus, positive and negative values of C_{std} indicate that
181 two species are allopatrically and sympatrically distributed, respectively, to extents
182 greater than indicated by the null models. All statistical analyses were performed with
183 the aid of R software version 3.2.0 (R Core Team 2015).

184 **Climatic niche similarities.** We used ecological niche modeling (ENM)
185 (Franklin 2010) to evaluate climatic niche similarities among butterfly species (Warren
186 et al. 2008). ENM associates distributional data with environmental characteristics, thus
187 estimating the response functions of, and the contributions made by, environmental
188 variables. Furthermore, potential distributional ranges may be obtained by projecting
189 model data onto geographical space. In the present study, potential distributional ranges

190 estimated by ENM should be influenced by abiotic environmental variables alone
191 (climate and altitude); we did not consider interspecific interactions among butterflies or
192 dispersal abilities in this context. Thus, comparisons of potential distribution patterns
193 estimated by ENM allow evaluation of climatic niche similarities among butterfly
194 species. The maximum entropy algorithm implemented in MaxEnt ver. 3.3.3e software
195 (Phillips et al. 2006) was employed in ENM analyses (Appendix S2 contains the details).
196 The logistic outputs of MaxEnt analyses can be regarded as presence probabilities.
197 Finally, we then used Schoener's (1968) D statistic to calculate climatic niche
198 similarities between pairs of butterfly species, based on the MaxEnt outputs. $P_{x,i}$ and $P_{y,i}$
199 were the probabilities (assigned by MaxEnt) that species x and y would mesh to the
200 extent of i on a geographic scale; the climatic niche similarity between the two species
201 was defined as: $D_{\text{env}} = 1 - 0.5 \times \sum |P_{x,i} - P_{y,i}|$, where D_{env} ranged from 0 (no niche
202 overlap) to 1 (completely identical niches). The probability assigned to the presence of
203 species x in grid i was $P_{x,i} = p_{x,i} / \sum p_{x,i}$, where $p_{x,i}$ was the logistic Maxent output for
204 species x in grid i .

205 **Explanatory variables.** We evaluated both host use similarity and other factors
206 that might explain exclusive species distributions (i.e., taxonomic relatedness). We
207 calculated the total dispersal abilities of species pairs and climatic niche similarities (as
208 explained above). Host use similarity was calculated as 1 minus Jaccard's dissimilarity
209 index (Koleff et al. 2003) when host plant species were shared by two butterflies. The
210 taxonomic relatedness of each species pair was classified as: 2: in the same genus; 1: in
211 the same family; and 0: in different families. The total dispersal ability was calculated
212 as the sum of the \ln (forewing lengths) of each species pair.

213 **Statistical tests.** We used the Mantel test, in which the response matrix
214 yielded pairwise C_{std} data, and which included explanatory matrices, to examine the
215 effects of host use similarity and other factors (i.e., taxonomic relatedness, climatic
216 niche similarity, and total dispersal ability) on the exclusivity of butterfly distribution.

217 We calculated Spearman's correlations because one of the explanatory variables
218 (taxonomic relatedness) was a rank variable. *P*-values were determined by running
219 9,999 permutations. In addition, when analyzing correlations between host use
220 similarity and other explanatory matrices, we ran Mantel tests with 9,999 permutations,
221 and calculated Spearman's correlations. We used partial Mantel tests, in which the
222 response matrix yielded pairwise C_{std} data, and in which a matrix of host use similarity
223 including the other three explanatory matrices served as a co-variable, to evaluate the
224 effects of confounding factors associated with host use similarity on the exclusivity of
225 butterfly distribution. We ran 9,999 permutations and calculated Spearman's
226 correlations. We employed the `vegdist` function of the `vegan` package (Oksanen et al.
227 2015) and the `mantel` function of the `ecodist` package (Goslee and Urban 2007)
228 implemented in R. We performed all analyses using data from both the 10-km and the
229 1-km grids.

230 **Results**

231 The standardized checkerboard scores (C_{std} values) of most species pairs were negative,
232 indicating that, in general, Japanese butterflies were more likely to co-occur than
233 expected by chance (Fig. 1). The Mantel test showed that host use similarity was
234 significantly, and negatively, correlated with the C_{std} values at both geographical scales
235 (Fig. 1, Table 1a); all three explanatory variables exhibited significant negative
236 correlations with the C_{std} values at both scales (Table 1a). Host use similarity was
237 significantly and positively correlated with both taxonomic relatedness and climatic
238 niche similarity, but we found no significant correlation with total dispersal ability
239 (Table 1b). The partial Mantel tests revealed negative correlations between the C_{std}
240 values and host use similarity, attributable to both taxonomic relatedness and dispersal
241 ability, at both geographical scales (Table 1c). In contrast, we found a significant
242 positive correlation in terms of climatic niche similarity in the 10-km grid dataset, but
243 no significant correlation in the 1-km grid dataset (Table 1c).

244

245

Discussion

246 Significant negative correlations were clearly evident between the C_{std} scores and host
247 use similarities at both grid scales (Fig. 1, Table 1a), indicating that pairs of Japanese
248 butterflies were more likely to share host plants than not. Significant negative
249 correlations between C_{std} scores and host use similarities were evident after controlling
250 for other potentially confounding factors, with the exception of climatic niche similarity
251 (Table 1c). Interspecific resource competition (i.e., a positive correlation between the
252 C_{std} score and host use similarity) was detected only after controlling for the effects of
253 climatic niche similarity in the 10-km grid data, but the correlation coefficient was low
254 (Table 1c). Our results are consistent with the idea that interspecific resource
255 competition is too weak to organize communities of herbivorous insects effectively
256 (Lawton and Strong 1981, Strong et al. 1984). Rather, our results suggest that the
257 geographic pattern of species co-occurrence among Japanese butterflies is better
258 explained by niche filtering. However, the niche axes that drive the observed patterns
259 remain unclear because the factors tested exhibited mutual correlations (Table 1b).

260 The most likely explanation of our data is that the relative strength of
261 structuring via resource competition may be weaker than that associated with niche
262 filtering. As the geographical distributions of host plants would be expected to be
263 strongly associated with the local climatic environment, the impacts of resource and
264 climatic niche filtering may combine to ensure that butterfly species sharing host plants
265 assemble in the same places. In addition, the dispersal of adult butterflies from the
266 patches in which they were born may counteract the structuring force imposed by
267 interspecific competition. Indeed, co-occurrence was facilitated by the overall total
268 dispersal ability (Table 1a). However, negative correlations between the C_{std} scores and
269 host use similarity were evident even when we controlled for the effects of total
270 dispersal ability (Table 1c). This means that dispersal alone may not explain the weak

271 impact of resource competition on the co-occurrence patterns of Japanese butterflies.
272 Other potential factors reducing the effects of interspecific resource competition may
273 also be in play; we did not address these topics in the present study. For example, the
274 presence of natural enemies is known to reduce interspecific competition during
275 community assembly markedly (Strong 1982, Nakadai and Kawakita 2017). It is very
276 difficult to assess the effects of natural enemies at regional scales.

277 The negative correlation evident between taxonomic relatedness and the C_{std}
278 scores (Table 1a) suggests that niche filtering is in play among Japanese butterflies,
279 given that taxonomic relatedness serves as a proxy of niche similarity including host use.
280 Indeed, we found significant (positive) correlations between host use similarity and the
281 taxonomic relatedness of Japanese butterflies (Table 1b), as has often been shown for
282 other herbivorous insects (e.g., Nyman et al. 2010). Moreover, when host use similarity
283 was controlled using the partial Mantel test, taxonomic relatedness did not significantly
284 affect co-occurrence at the 10-km grid scale (Appendix S1: Table S4). These results
285 suggest that, at least at the 10-km grid scale, the effects of taxonomic relatedness largely
286 reflect host use similarity.

287 In the present study, we used ENM to evaluate the effects of climatic niche
288 similarity on co-occurrence patterns. When we controlled for the effects of such niche
289 similarity, the negative correlations between the C_{std} scores and host plant similarities
290 disappeared at both spatial scales (Table 1c). This suggests that the explanatory power
291 of climatic niche filtering is stronger than that of resource niche filtering. It should be
292 noted that, although ENM has been widely used to quantify climatic niches (e.g., Kozak
293 et al. 2008, Warren et al. 2008, Takami and Osawa 2016), ENM data should be treated
294 with caution (Peterson et al. 2011, Warren 2012, Warren et al. 2014). For example,
295 Warren et al. (2014) noted that ENM always includes non-targeted factors that limit real
296 distributions if those distributions correlate spatially with environmental predictors.
297 Such confounding effects may cause overestimation of any positive correlation between

298 climatic niche similarity and butterfly co-occurrence, and may also cause the effects of
299 host use similarity to be underestimated when controlling for the effects of climatic
300 niche similarity. The use of ENM to study species co-occurrence patterns requires
301 careful consideration; more case studies evaluating the relative importance of climatic
302 niches are required.

303

304

Conclusions

305 The significance of interspecific resource competition in terms of the structuring of
306 herbivorous insect communities is a source of long-standing controversy. Many
307 researchers have sought to explain the general patterns of relationships between the
308 co-occurrence of, and the use of different niches by, herbivorous insects. However, the
309 data remain limited because most previous studies employed narrow taxonomic and
310 spatial scales. In this context, our study is the first to provide a comprehensive picture of
311 the co-occurrence patterns among a single taxonomic group over a large region.

312 Co-occurrence of Japanese butterflies is more likely to be driven by niche filtering than
313 interspecific resource competition. It is essential to employ broad taxonomic and spatial
314 scales when attempting to reveal general patterns of community assembly among
315 herbivorous insects. Future studies should explore the relative importance of each
316 assembly stage not only ecologically but also over evolutionary time (Rabosky 2009).

317 Such work would answer the important question: “Why have herbivorous insects
318 become one of the most diverse groups of the natural world?”

319

320

Acknowledgement

321 We thank K. Kadowaki for his comments and advice on our manuscript. We also thank
322 the Biodiversity Center of Japan, Ministry of the Environment, Japan, for allowing
323 access to butterfly data at the 1-km grid scale. This work was supported by a grant from
324 the Grant-in-Aid Program for JSPS Fellows (grant no. 15J00601).

325

326

Literature cited

327 Abramoff, M. D., P. J. Magalhães, and S. J. Ram. 2004. Image processing with ImageJ.

328 Biophotonics international 11:36–42.

329 Augustyn, W. J., B. Anderson, and A. G. Ellis. 2016. Experimental evidence for

330 fundamental, and not realised, niche partitioning in a plant-herbivore community

331 interaction network. *Journal of Animal Ecology* 85:994–1003.

332 Benson, W. W. 1978. Resource Partitioning in Passion Vine Butterflies. *Evolution*

333 32:493–518.

334 Bultman, T. L. and S. H. Faeth. 1985. Patterns of intra- and inter- specific association in

335 leaf-mining insects on three oak host species. *Ecological Entomology* 10:121–129.

336 Cantor, S. B., C. C. Sun, G. Tortolero-Luna, R. Richards-Kortum, and M. Follen. 1999.

337 A Comparison of C/B Ratios from studies using receiver operating characteristic

338 curve analysis. *Journal of Clinical Epidemiology* 52:885–892.

339 Cavender-Bares, J., K. H. Kozak, P. V. A. Fine, and S. W. Kembel. 2009. The merging

340 of community ecology and phylogenetic biology. *Ecology Letters* 12:693–715.

341 Chai, P., and R. B. Srygley. 1990. Predation and the flight, morphology, and temperature

342 of neotropical rainforest butterflies. *The American Naturalist* 135:748–765.

343 Diamond, J. 1975. Assembly of species communities. In: Cody ML, Diamond JM (eds)

344 Ecology and evolution of communities. Harvard University Press, Cambridge, pp

345 342–444

346 Franklin, J. 2010. Mapping species distributions: spatial inference and prediction.

347 Cambridge University Press.

348 Goslee, S. C. and D. L. Urban. 2007. The ecodist package for dissimilarity-based

349 analysis of ecological data. *Journal of Statistical Software* 22:1–19.

350 Gotelli, N. J., and D. J. McCabe. 2002. Species co-occurrence: A meta-analysis of J. M.

351 Diamond’s assembly rules model. *Ecology* 83:2091–2096.

- 352 Hashimoto, K., and T. Ohgushi. 2017. How do two specialist butterflies determine
353 growth and biomass of a shared host plant? *Population Ecology* 59:17–27.
- 354 Honda, K. 2005. Larval feeding habit and host selection. In: Honda K, Kato Y (eds)
355 *Biology of butterflies*. University of Tokyo Press, Tokyo, pp 255–301 (in
356 Japanese).
- 357 Inomata, T. 1990. Keys to the Japanese butterflies in natural color. Hokuryukan.
- 358 Jonsson, B. G. 2001. A null model for randomization tests of nestedness in species
359 assemblages. *Oecologia* 127:309–313.
- 360 Kaplan, I., and R. F. Denno. 2007. Interspecific interactions in phytophagous insects
361 revisited: a quantitative assessment of competition theory. *Ecology Letters*
362 10:977–994.
- 363 Kneitel, J. M., and J. M. Chase. 2004. Trade-offs in community ecology: Linking spatial
364 scales and species coexistence. *Ecology Letters* 7:69–80.
- 365 Koleff, P., K. J. Gaston, and J. J. Lennon. 2003. Measuring beta diversity for
366 presence-absence data. *Journal of Animal Ecology* 72:367–382.
- 367 Kozak, K. H., C. H. Graham, and J. J. Wiens. 2008. Integrating GIS-based
368 environmental data into evolutionary biology. *Trends in Ecology and Evolution*
369 23:141–148.
- 370 Kubota, Y., T. Hirao, S. Fujii, T. Shiono, and B. Kusumoto. 2014. Beta diversity of
371 woody plants in the Japanese archipelago: the roles of geohistorical and ecological
372 processes. *Journal of Biogeography* 41:1267–1276.
- 373 Lawton, J. H. and D. R. Strong. 1981. Community patterns and competition in
374 folivorous insects. *The American Naturalist* 118:317–338.
- 375 Liu, C., P. M. Berry, T. P. Dawson, and R. G. Pearson. 2005. Selecting thresholds of
376 occurrence in the prediction of species distributions. *Ecography* 28:385–393.
- 377 Mayfield, M. M., and J. M. Levine. 2010. Opposing effects of competitive exclusion on
378 the phylogenetic structure of communities. *Ecology letters* 13:1085–1093.

- 379 Nakadai, R., and A. Kawakita. 2017. Patterns of temporal and enemy niche use by a
380 community of leaf cone moths (*Caloptilia*) coexisting on maples (*Acer*) as revealed
381 by metabarcoding. *Molecular Ecology*
- 382 Novotny, V., S. E. Miller, L. Baje, S. Balagawi, Y. Basset, L. Cizek, K. J. Craft, F. Dem,
383 R. A. I. Drew, J. Hulcr, J. Leps, O. T. Lewis, R. Pokon, A. J. A. Stewart, G. Allan
384 Samuelson, and G. D. Weiblen. 2010. Guild-specific patterns of species richness
385 and host specialization in plant-herbivore food webs from a tropical forest. *Journal*
386 *of Animal Ecology* 79:1193–1203.
- 387 Nyman, T., V. Vikberg, D. R. Smith, and J.-L. Boevé. 2010. How common is ecological
388 speciation in plant-feeding insects? A “Higher” Nematinae perspective. *BMC*
389 *Evolutionary Biology* 10:266.
- 390 Oksanen, J., F. G. Blanchet, R. Kindt, P. Legendre, P. R. Minchin, R. B. O’Hara. 2015.
391 *Vegan: community ecology package*
392 [<https://cran.r-project.org/web/packages/vegan/index.html>].
- 393 Peterson, A. T., J. Soberon, R. G. Pearson, R. P. Anderson, E. Martinez-Meyer, M.
394 Nakamura, M. B. Araújo 2011. *Ecological niches and geographic distributions.*
395 *Monographs in Population Biology.* Princeton University Press. Princeton. New
396 Jersey, USA
- 397 Phillips, S. J., R. P. Anderson, and R. E. Schapire. 2006. Maximum entropy modeling of
398 species geographic distributions. *Ecological Modelling* 190:231–259.
- 399 Rabosky, D. L. 2009. Ecological limits and diversification rate: alternative paradigms to
400 explain the variation in species richness among clades and regions. *Ecology Letters*
401 12: 735– 743.
- 402 Rathcke, B. J. 1976. Competition and co-existence within a guild of herbivorous insects.
403 *Ecology* 57:76–87.
- 404 R Core Team 2015. *R: a language and environment for statistical computing.* R
405 Foundation for Statistical Computing, Vienna, Austria. <http://www.R-project.org/>

- 406 Ross, H. H. 1957. Principles of natural coexistence indicated by leafhopper populations.
407 Evolution 11:113–129.
- 408 Saito, M. U., U. Jinbo, M. Yago, O. Kurashima, and M. Ito. 2016. Larval host records of
409 butterflies in Japan. Ecological Research 31:491–491.
- 410 Schoener, T. W. 1968. The anolis lizards of bimini: Resource partitioning in a complex
411 fauna. Ecology 49:704–726.
- 412 Schoener, T. W. 1983. Field experiments on interspecific competition. The American
413 Naturalist 122:240–285.
- 414 Shirôzu, T. 2006. The butterflies of Japan in color. In Japanese. Gakken Holdings,
415 Tokyo, Japan.
- 416 Stone, L., and A. Roberts. 1990. The checkerboard score and species distributions.
417 Oecologia 85:74–79.
- 418 Strong, D. R. 1982. Harmonious coexistence of *Hispine* beetles on *Heliconia* In
419 experimental and natural communities. Ecology 63:1039–1049.
- 420 Strong, D. R., J. H. Lawton, and S. R. Southwood. 1984. Insects on plants. Community
421 patterns and mechanisms. Harvard University Press, Cambridge, MA.
- 422 Tack, A. J. M., O. Ovaskainen, P. J. Harrison, and T. Roslin. 2009. Competition as a
423 structuring force in leaf miner communities. Oikos 118:809–818.
- 424 Takami, Y., and T. Osawa. 2016. Ecological differentiation and habitat unsuitability
425 maintaining a ground beetle hybrid zone. Ecology and Evolution 6:113–124.
- 426 Ueckert, D. N., and R. M. Hansen. 1971. Dietary overlap of grasshoppers on sandhill
427 rangeland in northeastern Colorado. Oecologia, 8: 276-295.
- 428 Waloff, N. 1979. Partitioning of resources by grassland leafhoppers (Auchenorrhyncha,
429 Homoptera). Ecological Entomology 4:379-385.
- 430 Warren, D. L., M. Cardillo, D. F. Rosauer, and D. I. Bolnick. 2014. Mistaking
431 geography for biology: inferring processes from species distributions. Trends in
432 Ecology & Evolution 29:572–580.

- 433 Warren, D. L., R. E. Glor, and M. Turelli. 2008. Environmental niche equivalency
434 versus conservatism: Quantitative approaches to niche evolution. *Evolution*
435 62:2868–2883.
- 436 Warren, D. L. 2012. In defense of ‘niche modeling’. *Trends in Ecology & Evolution*
437 27:497–500.
- 438 Webb, C. O., D. D. Ackerly, M. A. McPeck, and M. J. Donoghue. 2002. Phylogenies
439 and community ecology. *Annual Review of Ecology and Systematics* 33:475–505.

440

441

Supporting information

442 **Appendix S1: Table S1.** The Japanese butterfly species analyzed in the present study.

443 **Appendix S1: Table S2.** Summary of the MaxEnt data at the 10-km grid scale.

444 **Appendix S1: Table S3.** Summary of the MaxEnt data at the 1-km grid scale.

445 **Appendix S1: Table S4.** Correlations among host use similarity (Host), taxonomic
446 relatedness (Taxon), climate niche similarity (Climate), and total dispersal ability
447 (Dispersal), at two spatial scales. (a) Summary of Mantel test data on pairwise
448 correlations between the explanatory matrices. (b) Summary of partial Mantel test data
449 on standardized C_{std} scores between pairs of butterfly species. The “Taxon-Dispersal”
450 data (b) were obtained using the datasets at either grid mesh scale.

451 **Appendix S2:** Details of the methods used for ecological niche modeling.

452

453

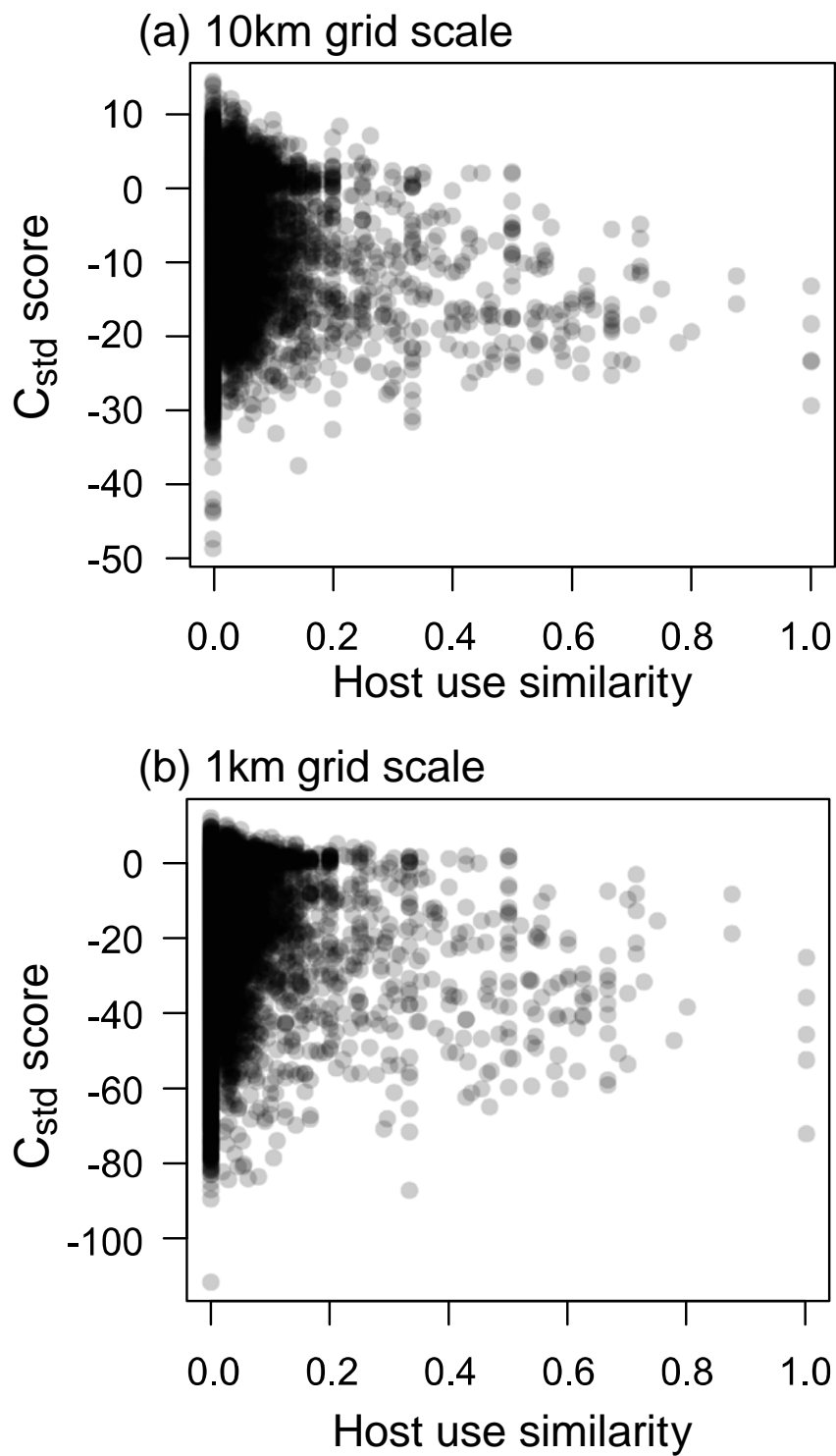
Legend

454 **Figure 1.** The relationships between C_{std} scores and host use similarities (10-km grid
455 scale: Spearman $\rho = -0.132$, $P = 0.0001$; 1-km grid scale: Spearman $\rho = -0.132$, $P =$
456 0.0001; Mantel test).

457

458 **Figure 1.**

459



460 **Table 1.** Correlations among host use similarity (Host), taxonomic relatedness (Taxon),
 461 climate niche similarity (Climate), and total dispersal ability (Dispersal), at two spatial
 462 scales. (a) Summary of Mantel test data on standardized C_{std} scores between pairs of
 463 butterfly species. (b) Summary of Mantel test data on pairwise correlations between
 464 host use similarity and the data of other explanatory matrices. (c) Summary of partial
 465 Mantel test data on standardized C_{std} scores between pairs of butterfly species. The
 466 “Host-Taxon” and “Host-Dispersal” data (b) were analyzed using the datasets of either
 467 grid mesh scale.

(a) Factor	10-km grid		1-km grid	
	Mantel ρ	<i>P</i> -values	Mantel ρ	<i>P</i> -values
Host	-0.126	0.0001	-0.126	0.0001
Taxon	-0.027	0.0063	-0.037	0.0008
Climate	-0.723	0.0001	-0.482	0.0001
Dispersal	-0.047	0.0161	-0.072	0.0008

468

(b) Factor	10-km grid		1-km grid	
	Mantel ρ	<i>P</i> -values	Mantel ρ	<i>P</i> -values
Host-Taxon	0.097	0.0001	0.097	0.0001
Host-Climate	0.209	0.0001	0.240	0.0001
Host-Dispersal	0.067	0.0659	0.067	0.0659

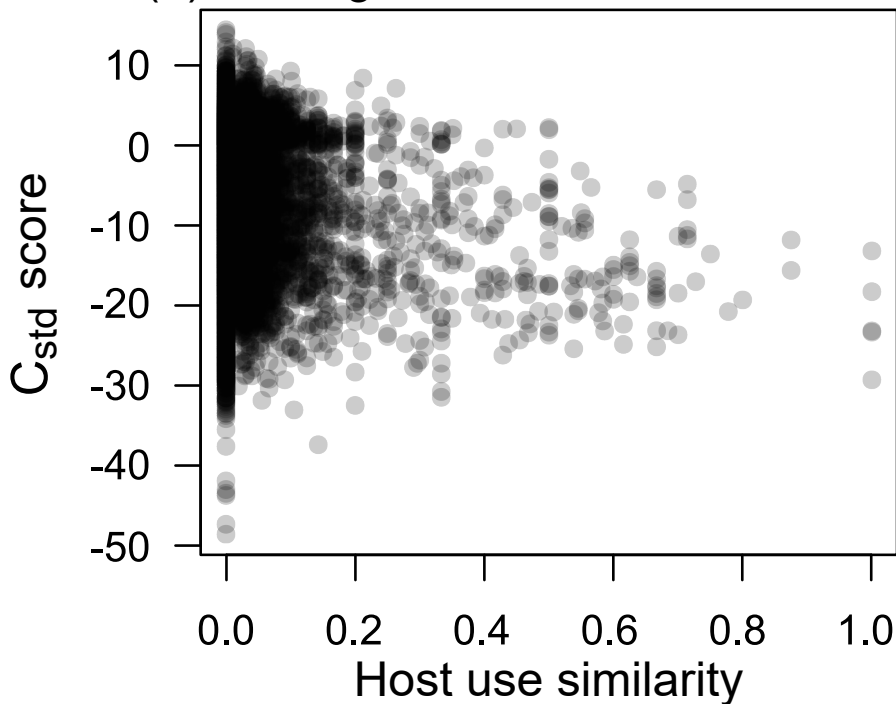
469

(c) Factor	Covariate	10-km grid		1-km grid	
		Mantel ρ	<i>P</i> -values	Mantel ρ	<i>P</i> -values
Host	Taxon	-0.124	0.0001	-0.123	0.0001
Host	Climate	0.036	0.0020	-0.012	0.3660
Host	Dispersal	-0.124	0.0001	-0.121	0.0001

470 Spearman's correlation coefficients (ρ values) are shown for all four factors.

471 Bold: $P < 0.05$; Underlined: $0.05 < P < 0.1$ after 9,999 permutations.

(a) 10km grid scale



(b) 1km grid scale

