

1 **Integrating telemetry data into spatial capture-recapture modifies inferences on**
2 **multi-scale resource selection**

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14 **Abstract**

15 Estimating population size and resource selection functions (RSFs) are common approaches in
16 applied ecology for addressing wildlife conservation and management objectives. Traditionally
17 such approaches have been undertaken separately with different sources of data. Spatial capture-
18 recapture (SCR) provides a framework for jointly estimating density and multi-scale resource
19 selection, and data integration techniques provide opportunities for improving inferences from
20 SCR models. Here we illustrate an application of integrated SCR-RSF modeling to a population
21 of American marten (*Martes americana*) in alpine forests of northern New England. Spatial
22 encounter data from camera traps were combined with telemetry locations from radio-collared
23 individuals to examine how density and space use varied with spatial environmental features.
24 We compared multi-model inferences between the integrated SCR-RSF model with telemetry
25 and a standard SCR model with no telemetry. The integrated SCR-RSF model supported more
26 complex relationships with spatial variation in third-order resource selection (i.e., individual
27 space use), including selection for areas with shorter distances to mixed coniferous forest and
28 rugged terrain. Both models indicated increased second-order selection (i.e., density) for areas
29 close to mixed coniferous forest, while the integrated SCR-RSF model had a lower effect size
30 due to modulation from spatial variability in space use. Our application of the integrated SCR-
31 RSF model illustrates the improved inferences from spatial encounter data that can be achieved
32 from integrating auxiliary telemetry data. Integrated modeling allows ecologists to join
33 empirical data to ecological theory using a robust quantitative framework to better address
34 conservation and management objectives.

35 **Key words:** American marten; density estimation; integrated modeling; noninvasive sampling;
36 *Martes americana*; radio-telemetry; spatial capture-recapture

37 **Introduction**

38 Understanding the environmental features that influence variation in species abundance
39 or density is a common objective in wildlife conservation and management. Approaches to
40 estimating population size and habitat selection have traditionally required distinct forms of data
41 collection and separate modeling methods; spatial capture-recapture (SCR) allows for joint
42 estimation of both ecological processes using a single data source or through integration of
43 multiple data sources in a single analytical framework (Royle, Fuller & Sutherland 2017). The
44 development of SCR was initially motivated by the need to address the spatial dynamics of field
45 sampling and animal movement to improve density estimation from capture-recapture methods
46 (Efford 2004; Royle & Young 2008). By using spatial information on the location of
47 observations, SCR combines a point process model for the distribution of individuals in a
48 population with a probability model for the encounters or captures of those individuals. As with
49 other recent statistical advances in ecology (Gimenez *et al.* 2014), the hierarchical model
50 formulation of SCR has enabled custom data integration techniques and, therefore, expanded the
51 scope of inferences possible from trapping and other spatial encounter data (Sollmann *et al.*
52 2013a; Chandler & Clark 2014). While these developments are promising, as the complexity of
53 modeling methods increases so do the data demands for parameter estimation (Auger-Methe *et*
54 *al.* 2016) and the need for practical implementation options (Bolker *et al.* 2013). Producing valid
55 and useful inferences for helping achieve objectives in applied ecology requires finding a
56 balance between comprehensive models and logistically feasible data.

57 Royle *et al.* (2013) developed a spatial capture-recapture model that integrates telemetry
58 information and resource selection functions (RSFs) to provide improved accuracy and precision
59 for density estimation. The improved accuracy is particularly relevant when resource selection at

60 one or more spatial scales affects the distribution of individuals (i.e., second-order selection;
61 Johnson 1980) and/or individual space use (i.e., third-order selection). While the approach
62 seems promising and uses commonly collected wildlife data, it has been rarely applied in the
63 literature. Proffitt *et al.* (2015) included a resource selection function in an SCR model of a large
64 carnivore but estimated the function with telemetry data first before using the predictions as a
65 covariate in their SCR model fitting. They implied that the Royle *et al.* (2013) model
66 represented a “methodologically intensive joint estimation framework” which might preclude
67 interested users from easily applying it. The multi-step approach to RSF integration in SCR is a
68 potentially interesting and practical compromise, whereby a complex function is reduced to a
69 univariate prediction to serve as a single spatial covariate in the model (Efford 2015). Proffitt *et*
70 *al.* (2015) used this method to address spatial variation in density – they did not incorporate it in
71 their encounter probability model, despite having data that may have allowed for modeling
72 variation in individual space use. In general, SCR models are more sensitive to the structure of
73 the encounter model than the density model, as the latter has been found to be highly flexible to
74 misspecification (Efford & Fewster 2013; Royle *et al.* 2014) while the former has long been a
75 focus of refinements to capture-recapture methods (Dorazio & Royle 2003). As referenced
76 earlier, individual heterogeneity to capture was a primary motivation for SCR (Efford 2004).
77 Thus, in SCR applications where individual space use is hypothesized to be highly variable,
78 accounting for third-order selection in the encounter model may be important regardless of
79 whether ancillary telemetry data are available.

80 The task of estimating variation in space use with a spatial capture-recapture design alone
81 is made difficult by the required sample sizes in the various data dimensions. Importantly, the
82 number and configuration of trapping devices needs to be such that adequate coverage across

83 some gradient of habitat resources is achieved. This requires some traps to be placed within or
84 adjacent to relatively poor habitat, even though doing so may result in little information (i.e.,
85 zero encounters) being collected for the logistical effort expended. Since the placement of traps
86 will interact with both the distribution of individuals on the landscape and individual space use,
87 the number of encounters at a trap will be a function of the realized dynamics of both processes
88 (Royle *et al.* 2014). Thus, an optimal trap layout for estimating multi-scale resource selection
89 will purposefully expend trapping effort in locations where few individuals exist and where
90 individuals, even when present nearby, may be unlikely to visit. Such a design creates tension
91 with the general goal of obtaining as many encounters of as many individuals as possible to
92 enable model fitting and accurate parameter estimation (Royle *et al.* 2014).

93 Here, we illustrate an application of the integrated SCR-RSF model developed by Royle
94 *et al.* (2013) using a multi-year study on American marten (*Martes americana*) in New
95 Hampshire, USA (Sirén *et al.* 2016a; Sirén *et al.* 2016b). Marten are a forest-sensitive meso-
96 carnivore often used as an indicator species for forest conservation and management given their
97 vulnerability to anthropogenic disturbance and climate change (Carroll 2007). We used remote
98 camera stations to photograph and identify individual marten according to their unique pelage
99 patterns (Sirén *et al.* 2016a) and combined these data with telemetry locations collected on a
100 subset of radio-collared individuals (Sirén *et al.* 2016b). The sampling occurred across a
101 heterogeneous alpine forest landscape that was recently modified by a wind farm installment and
102 within which marten were shown to be differentially selecting resources at multiple scales.
103 Using the integrated SCR-RSF likelihood (Royle *et al.* 2013), we estimated marten density and
104 parameters associated with multi-scale resource selection and compared the resulting multi-
105 model inferences both with and without the telemetry data integration. Notably, we modified the

106 likelihood to accommodate a lack of independence between the data sources, given that all radio-
107 collared individuals were also photo-captured by camera traps. This integrated SCR-RSF model
108 was added to the R package `oSCR` (Sutherland, Royle & Linden 2016) to facilitate use by
109 interested researchers. Our example here provides further evidence that improved inferences on
110 variation in population density are possible when additional information on movement and
111 resource selection from telemetry data are integrated with spatial capture-recapture models.

112

113 **Materials and methods**

114 **STUDY AREA AND SAMPLING**

115 We sampled marten in a ~62 km² area of New England-Acadian forest in northern New
116 Hampshire, USA during 2010–2012 (Figure S1). Extensive descriptions of the study area and
117 sampling details have been previously outlined (Sirén *et al.* 2016a; Sirén *et al.* 2016b). The area
118 was centered on some mountains containing a new 33-turbine wind farm with elevation ranging
119 624–1045 m. Forests were primarily mature conifers at high elevations (>823 m) and a mix of
120 mature and commercially harvested hardwoods at lower elevations. Topography was variable
121 with rugged terrain dispersed throughout and winters in the region (1948–2013) were cold with
122 high annual snowfall (average = 288 cm; range = 79–881 cm; National Climate Data Center:
123 <http://www.ncdc.noaa.gov>).

124 Marten were photo-captured by remote cameras during 2 winters (Sirén *et al.* 2016a) and
125 live-captured year-round to maintain 6–10 individuals with active VHF radio collars during the
126 study (Sirén *et al.* 2016b). Telemetry fixes from collared individuals were obtained weekly
127 using standard methods of triangulation; here, we restricted the data to fixes collected during the
128 2 leaf-off (i.e., winter) time periods (14 Nov 2010 to 15 May 2011 and 16 Oct 2011 to 15 May

129 2012) which overlapped with the winter sampling by remote cameras (14 Feb 2011 to 2 Apr
130 2011 and 3 Jan 2012 to 19 Jan 2012). Location error was estimated to be 2 ha (Sirén *et al.*
131 2016b). We used 30 total camera trapping stations during each sampling period that were baited
132 with sardines and a commercial skunk (*Mephitis mephitis*) lure. Stations were operational for 12
133 d in 2011 and 8 d in 2012, with bait replaced halfway through, and were located in a nonrandom
134 pattern to achieve a spacing of 500–950 m (Figure 1). Marten photo-captures were identifiable
135 to individual when ventral patches were clearly visible and multiple observers could agree on
136 assignment; sex was assigned to un-collared individuals based on morphological comparisons
137 with collared individuals (Sirén *et al.* 2016a).

138

139 INTEGRATED SCR-RSF MODEL

140 We estimated multi-scale habitat selection of marten using the spatial encounters of individuals
141 at camera traps and the telemetry locations of collared individuals with the integrated SCR-RSF
142 model (Royle *et al.* 2013). A discrete-space representation of the study area was required to
143 allow spatial environmental covariates to influence the space use of individuals (third-order
144 selection) and the distribution of individuals (second-order selection) on the landscape. The
145 discretization of the landscape accommodates a straightforward formulation of the space usage
146 model that links the camera trap and telemetry data under the integrated SCR-RSF model (Royle
147 *et al.* 2013). Using the notation of Royle *et al.* (2013), we can define the set of coordinates \mathbf{x}_1 ,
148 \dots , \mathbf{x}_{nG} for each pixel g on the landscape of nG pixels and estimate the likelihood of observing
149 the data for individuals in pixels conditional on \mathbf{s} , the latent centroids of activity (i.e., home range
150 centers) for individuals, and $z(\mathbf{x})$, the covariate value(s) for all pixels. Importantly, individual
151 use of a pixel is considered a Poisson random variable with an average rate $\lambda(\mathbf{x}|\mathbf{s})$ for both data

152 likelihoods. While the total probability of observing an individual in a given pixel differs
153 between the camera trap and telemetry data (as a function of the sampling rate and other features
154 of each device), the mechanisms underlying spatial variation are assumed to be the same,
155 allowing parameters to be shared in the joint likelihood (Royle *et al.* 2013).

156 Our discrete landscape for marten was defined by a 3.75 km buffer of the trapping array
157 composed of $200 \text{ m} \times 200 \text{ m}$ pixels for a total of $nG = 2,709$ non-water pixels. The buffer
158 accommodated individual movement in the sampled population and the pixel resolution was
159 small enough to distinguish differences in space usage within home ranges. We summarized the
160 telemetry data as pixel-specific counts, m_{ig} , for each of $i = 1, 2, \dots, N_{\text{tel}}$ collared individuals at each
161 pixel g . Spatial encounters at camera traps, y_{ijk} , were defined for each of $i = 1, 2, \dots, n$ photo-
162 captured individuals at camera trap j on survey k as binary variables, condensing any clusters of
163 encounters that occurred in a given day to a single “detection” (Siren *et al.* 2016a). We extracted
164 the spatial covariate values for trap j from the pixel within which the trap was located, such that
165 $\mathbf{z}_j \equiv \mathbf{z}(\mathbf{x}_j)$. The spatial covariates we considered for influencing habitat selection included the
166 average distance to mixed-coniferous forest and the vector ruggedness measure (VRM;
167 Sappington, Longshore & Thompson 2007) for each pixel; see Sirén *et al.* (2016b) for more
168 information on the remote sensing data and additional justification for the habitat covariates.

169 Following Royle *et al.* (2013), we modeled the spatial encounters, y_{ijk} , as Bernoulli
170 random variables such that $\Pr(y_{ijk} > 0 \mid \mathbf{x}_j, \mathbf{s}_i) = p_{ijk}$, and under a Gaussian hazard model, $p_{ijk} = 1 -$
171 $\exp(-\lambda_{ijk})$. In this way, λ_{ijk} represents the average encounter rate for the trap in pixel \mathbf{x}_j , assuming
172 the latent “use frequency” (i.e., 3rd order selection) is a Poisson random variable. This use
173 frequency is unobservable for the camera trap data due to difficulty in distinguishing
174 independent encounters within a survey but links directly to the use frequency indicated by the

175 telemetry data (Royle *et al.* 2013). Here, we modeled the encounter rate of marten as a function
176 of several components: 1) availability for an individual, given the location of its latent activity
177 center; 2) resource selection of spatial covariate values for the camera trap; 3) year during which
178 the individual was photo-captured; 4) sex of the individual; and 5) a trap-specific behavioral
179 response. The first two components address the crux of the methodological development
180 presented by Royle *et al.* (2013), while the remaining components are specific to the marten
181 study. Our log-linear model was therefore:

$$\begin{aligned} 182 \quad \log(\lambda_{ijk}) = & \alpha_0 - \alpha_{1i}d(\mathbf{x}_j, \mathbf{s}_i)^2 + \alpha_{2,\text{mixed}}\text{mixed}_j + \alpha_{2,\text{vrm}}\text{vrm}_j + \alpha_{2,\text{vrm}2}\text{vrm}_j^2 \\ 183 \quad & + \alpha_{2012}\text{yr}_i + \alpha_{\text{male}}\text{sex}_i + \alpha_{2012,\text{male}}\text{yr}_i \times \text{sex}_i + \alpha_{\text{behav}}C_{ijk} \end{aligned}$$

184 where $d(\mathbf{x}_j, \mathbf{s}_i)$ is the Euclidean distance between trap j and the latent activity center for individual
185 i , and $\alpha_{1i} = 1/(2\sigma_i^2)$, where σ_i represents the scale parameter of the half-normal distance function.
186 In this way, $\alpha_{1i}d(\mathbf{x}_j, \mathbf{s}_i)^2$ quantifies the “availability” of a trap pixel conditional on \mathbf{s}_i for the
187 individual. The spatial covariates were standardized with mean 0 and unit variances, after
188 distance to mixed conifers (mixed_j) was square-root transformed and terrain ruggedness (vrm_j)
189 was log-transformed. Year and sex were both binary variables indicating whether an individual
190 was encountered in 2012 ($\text{yr}_i = 1$) and its sex was male ($\text{sex}_i = 1$), including a potential
191 interaction. Finally, the behavioral response used $C_{ijk} = 1$ for all k after the initial encounter of
192 individual i at trap j , and 0 otherwise. We allowed σ_i to vary according to year and sex:

$$193 \quad \log(\sigma_i) = \delta_0 + \delta_{2012}\text{yr}_i + \delta_{\text{male}}\text{sex}_i + \delta_{2012,\text{male}}\text{yr}_i \times \text{sex}_i$$

194 This accounted for potential differences in the scale of movement between sexes and years. We
195 also treated sex as a random variable and estimated $\phi_{\text{male}} = \Pr(\text{sex}_i = 1)$ using the model extension
196 described in (Royle *et al.* 2015).

197 We used the pixel-specific frequencies from the telemetry data to estimate use
198 probabilities under a multinomial distribution such that $m_{ig} \sim \text{Multinomial}(R_i, \pi_{ig})$, where R_i is the
199 number of telemetry fixes for individual i and π_{ig} is the relative probability of use as defined by

$$200 \quad \pi_{ig} = \frac{\exp(-\alpha_{1i}d(\mathbf{x}_g, \mathbf{s}_i)^2 + \alpha_{2,\text{mixed}}\text{mixed}_g + \alpha_{2,\text{vrm}}\text{vrm}_g + \alpha_{2,\text{vrm}2}\text{vrm}_g^2)}{\sum_g \exp(-\alpha_{1i}d(\mathbf{x}_g, \mathbf{s}_i)^2 + \alpha_{2,\text{mixed}}\text{mixed}_g + \alpha_{2,\text{vrm}}\text{vrm}_g + \alpha_{2,\text{vrm}2}\text{vrm}_g^2)}$$

201 Here, the individual- and pixel-specific usage rate (i.e., the numerator) is formulated with a
202 similar log-linear model as the encounter rate for the spatial encounters at camera traps, with the
203 exception that the rate is not survey specific and is only a function of attributes that vary by
204 location. Another difference is that the usage rate is defined at all pixels (g) in the state space,
205 not only the pixels with camera traps. Since the usage rate is a function of availability, the
206 individual rates and relative probabilities are conditional on \mathbf{s}_i and variation in the movement
207 scale (σ_i) is defined by the same log-linear model.

208 We modeled variation in the distribution of activity centers, representing 2nd order
209 resource selection by marten, by specifying an inhomogeneous point process (Borchers & Efford
210 2008). Here, the expected density in a given pixel g was a linear function (on the log scale) of
211 the distance to mixed conifer forest:

$$212 \quad \log(\mathbb{E}(D_g)) = \beta_0 + \beta_{\text{mixed}}\text{mixed}_g$$

213 This density model determined the prior probability of an activity center being located in any
214 given pixel according to:

$$215 \quad \Pr(s_i = g | \boldsymbol{\beta}) = \frac{\mathbb{E}(D_g)}{\sum_g \mathbb{E}(D_g)}$$

216 where s_i now represents a pixel ID instead of two-dimensional coordinates. Since the likelihoods
217 of the data for both the spatial encounters and the telemetry fixes are conditional on s_i , the

218 marginal likelihood is calculated by integrating over all possible pixel values (i.e., the Poisson-
219 integrated likelihood approach; Borchers & Efford 2008). Note, when considering the photo-
220 captured individuals that were also radio-collared, the conditional likelihoods must be combined
221 before calculating a single marginal likelihood for each individual (Appendix S1). Royle *et al.*
222 (2013) used data from different sets of individuals and assumed the data likelihoods were
223 independent. The consequences of accommodating non-independent data include more precise
224 estimation of activity centers for the captured individuals with collars and reduction of the
225 effective sample size of individuals used to estimate the resource selection functions. Assuming
226 independence between the data when the overlap of individuals is high would result in
227 pseudoreplication problems (e.g., artificially reduced standard errors).

228

229 MODEL COMPARISONS WITH AND WITHOUT TELEMETRY

230 We compared inferences on multi-scale selection between the integrated SCR-RSF model and a
231 standard SCR model (without telemetry data) using an information theoretic approach followed
232 by parameter estimate comparisons for the relevant coefficients. We used model selection to
233 identify the covariate structures with the best predictive performance for the data at each level of
234 the hierarchical models, starting with movement scale (σ), then encounter rate (λ), and finally
235 density (D). A multi-staged approach was used to reduce the total set of candidate models: 1)
236 select among covariates for σ with full covariate structures for λ and D ; 2) select among
237 covariates for λ using the top covariates for σ and full structure for D ; and, 3) select among
238 covariates for D using the top covariates for σ and λ . Candidate models for the encounter rate
239 always included the behavioral effect (α_{behav}) given previous findings (Sirén *et al.* 2016a). This
240 resulted in 5 candidate models for movement scale, 30 candidate models for encounter rate, and

241 2 candidate models for density. Our main focus was on parameter comparisons for α_2 (third-
242 order selection) and β (second-order selection), both in terms of best supported model structures
243 and differences in effect size and precision of the estimates, where applicable.

244 All models were fit using maximum likelihood methods with the `oSCR` package
245 (Sutherland, Royle & Linden 2016) in R (R Core Team 2016). The `oSCR` package is an open-
246 source alternative to `secr` (Efford 2016) for fitting certain spatial capture-recapture models
247 using maximum likelihood. We integrated the R code from Royle *et al.* (2013) into `oSCR` and
248 included an option for adjusting the marginal likelihood calculations to accommodate non-
249 independent data. The candidate models were ranked using AIC and, to improve clarity for
250 encounter rate comparisons, we removed from our model selection tables those variables with
251 85% confidence intervals that included zero (Arnold 2010). Note, AIC cannot be directly
252 compared between the standard SCR and integrated SCR-RSF model types given differences in
253 the data likelihoods; we focus on the relative rankings of model structure. We used the top-
254 ranked models to map the predicted encounter probability and expected density across the study
255 area, illustrating variation in resource selection at each scale.

256

257 **Results**

258 The winter camera trapping resulted in 13 individuals (6 F; 7 M) captured 121 times in 2011 and
259 15 individuals (6 F; 9 M) captured 86 times in 2012. Across years, the number of spatial
260 encounters at unique trap pixels per individual ranged 1–10, with 50% having ≤ 2 and 25%
261 having ≥ 5 spatial encounters. The number of radio-collared individuals was 6 (2 F; 4 M) in
262 2011 and 8 (2 F; 6 M) in 2012, with a total of 147 and 218 telemetry locations, respectively,
263 collected during each winter. Across years, the number of telemetry locations at unique pixels

264 per collared individual ranged 11–27 (median = 17.5). The proportion of encounters at camera
265 traps attributed to radio-collared individuals was 0.73 in 2011 and 0.71 in 2012.

266 Model selection results indicated that an increased complexity in model structure was
267 supported by the integrated SCR-RSF model with telemetry data compared to the standard SCR
268 model without telemetry data (Tables 1–3). The top model structure for movement scale (σ)
269 included no covariates (i.e., null; AICwt = 0.42) in the absence of telemetry and a sex \times year
270 interaction (AICwt = 0.99) with telemetry (Table 1). Sex was also included as a covariate in the
271 top model for encounter rate (λ) without telemetry (Table 2), though model selection uncertainty
272 suggested it was a marginal predictor at best (AICwt = 0.52); none of the spatial covariates were
273 important predictors for encounter rate in the standard SCR model. The integrated SCR-RSF
274 model supported variables corresponding to third-order resource selection (distance to mixed
275 conifer forest and terrain ruggedness) in the top model structures for encounter rate (Table 2).
276 Regardless of telemetry integration, variation in second-order resource selection was supported
277 as distance to mixed conifers was considered an important predictor for density in both the
278 standard SCR and integrated SCR-RSF models (Table 3).

279 Differences in parameter estimates for the top-ranked models further indicated how
280 telemetry integration modified inferences on multiscale resource selection (Table 4). Population
281 density decreased with increasing distance to mixed conifers for both top-ranked models and the
282 effect size was larger without telemetry ($\beta_{\text{mixed}} = -2.09$ [SE: 0.792]) than with telemetry ($\beta_{\text{mixed}} =$
283 -0.79 [0.397]). This reduction in effect size for density variation was due to the integrated SCR-
284 RSF model attributing additional variation in the observed encounters to differences in encounter
285 rate, including a decrease with increased distance to mixed conifers ($\alpha_{2,\text{mixed}} = -0.11$ [SE:
286 0.057]). Encounter rate also appeared to vary by terrain ruggedness, with a positive quadratic

287 relationship suggesting increased space use of terrain that was flat or extremely rugged, and
288 lower use at moderate ruggedness ($\alpha_{2, \text{vrm}} = -0.11$ [SE: 0.055]; $\alpha_{2, \text{vrm}^2} = 0.06$ [SE: 0.034]). While
289 maps of expected density illustrated similar spatial patterns of 2nd order resource selection
290 between the top-ranked models (Figure 2a,b), the standard SCR model with no telemetry data
291 exhibited greater variation consistent with the larger estimate for β_{mixed} . We also mapped the
292 predicted probability of encounter when $d(\mathbf{x}_j, \mathbf{s}_i)^2 = 0$ to illustrate the spatial variation in 3rd order
293 selection as indicated by the top-ranked integrated SCR-RSF model (Figure 2c). The overall
294 predicted marten density (#/km²) for 2010–2011 was slightly lower for the standard SCR model
295 (0.39/km² [95% CI: 0.29–0.56/ km²]) than the integrated SCR-RSF model (0.43/km² [95% CI:
296 0.32–0.61/ km²]).

297 The top-ranked models also differed with regards to movement scale (σ) variation and the
298 predicted probabilities of activity center locations for collared individuals. The standard SCR
299 model did not support any variation in movement scale, with an average $\sigma = \exp(\delta_0) = 0.79$ km
300 (95% CI: 0.70–0.90 km), while the integrated SCR-RSF model suggested an interaction between
301 sex and year such that female marten in 2012 had a lower movement scale than all other
302 individuals (Table 4). The precision of the activity center predictions was significantly increased
303 for collared individuals when telemetry data were integrated (Figure 3), underscoring the value
304 of using a non-independent likelihood.

305

306 **Discussion**

307 Integrated modeling methods allow ecologists to bring together multiple data sources to expand
308 the scope of potential inferences for and enhance understanding of complex ecological systems.
309 The recent growth of these techniques coincides with the development of hierarchical models

310 (Gimenez *et al.* 2014), which posit data generation as a combination of observation processes
311 and latent ecological processes (Royle & Dorazio 2008). When multiple sources of data provide
312 information on the same ecological processes of interest, joint data likelihoods can be defined to
313 explicitly link diverse observations and improve estimation of model parameters describing
314 relevant phenomena. The result is an increased ability for ecologists to join empirical data to
315 ecological theory using robust quantitative approaches.

316 The integrated SCR-RSF model developed by Royle *et al.* (2013) connects observed
317 spatial encounters of animals to established theory on multi-scale resource selection (Johnson
318 1980). By combining data from devices that differ in their ability to target variation at specific
319 scales (i.e., individual vs. population level), the model is better able to simultaneously describe
320 both scales. Our application of the integrated SCR-RSF model allowed for more complex
321 covariate relationships with parameters related to individual movement and multi-scale resource
322 selection of marten than possible under standard SCR model fitting. In the absence of telemetry
323 data, spatial variation in third-order selection (i.e., through encounter rate) could not be detected
324 with the camera trap observations and instead, the standard SCR model suggested more variation
325 in second-order selection. Both model types indicated that expected marten density decreased
326 with increasing distance to mixed conifers. So while the camera trap data alone were sufficient
327 for identifying an existing density relationship, the slope of this relationship was potentially
328 biased due to unmodeled heterogeneity in encounter rate. The integrated SCR-RSF model
329 indicated that areas of the landscape with relatively short distances to mixed conifer forest had
330 more marten, and those marten spent more time in these locations. Additional model complexity
331 supported by our telemetry integration included an interaction between sex and year on
332 movement scale (σ), with female marten in 2012 being encountered at shorter distances from

333 their activity centers than individuals of either sex in other years. The standard SCR model had
334 little power to detect variation in σ according to locations of spatial encounters at camera traps
335 alone, likely due to small sample sizes. Similarly, the small number and strategic placement
336 (i.e., to increase encounters) of camera traps made it difficult to detect variation in encounter rate
337 according to spatial environmental variables without the additional data provided by telemetry.

338 Telemetry has been used in multiple ways as an auxiliary data source to improve
339 inferences from population size estimation, including both ad-hoc adjustments (Soisalo &
340 Cavalcanti 2006) and more recent joint modeling techniques (Ivan, White & Shenk 2013; Bird *et al.*
341 *et al.* 2014). Spatially-explicit applications have generally focused on the contribution of telemetry
342 data to informing the scale of individual movement in SCR models for the purposes of
343 improving the precision and/or identifiability of parameter estimates (Sollmann *et al.* 2013a;
344 Sollmann *et al.* 2013b). Even as few as $N_{tel} = 3$ collared individuals with an adequate collection
345 of telemetry locations can greatly increase the precision of the σ estimate (Royle *et al.* 2013),
346 though such small sample sizes will limit the exploration of potentially important relationships
347 with individual attributes (e.g., sex; Sollmann *et al.* 2013a). Here we had some flexibility with
348 $N_{tel} = 14$ collared martens and were able to model more variation in σ , though at the expense of a
349 significant improvement in precision (Table 4). Most importantly, we were better able to model
350 variation in encounter rate as it related to individual space use due to the relatively large
351 collection of telemetry locations.

352 Proffitt *et al.* (2015) is the only other application we could find where telemetry-informed
353 resource selection was integrated into an SCR framework in the vein of Royle *et al.* (2013).
354 Their analysis involved a two-stage approach where 30+ years of radiotelemetry data on
355 mountain lions (*Puma concolor*) were used to estimate and validate an RSF that corresponded to

356 second-order selection (Robinson *et al.* 2015); spatial predictions from the RSF were then
357 specified as a density covariate in an SCR model fitted to search encounter data collected during
358 a 4-month sampling period. The RSF covariate appeared to improve the accuracy and precision
359 of density estimates (Proffitt *et al.* 2015), yet, without integrating the telemetry data into the
360 likelihood directly, their analysis could not incorporate information on third-order selection or
361 benefit from increased precision in σ estimates. The latter may have provided a significant
362 improvement due to the small number of spatial encounters and could have alternatively been
363 achieved in their Bayesian modeling approach by using an informed prior for σ calculated from
364 the 18,000+ telemetry locations on 85 individuals (Proffitt *et al.* 2015). The individuals in the
365 telemetry data need to be representative of the focal population being exposed to capture and the
366 degree to which this is true will dictate how many model parameter estimates can reasonably be
367 shared between the data.

368 The marten study consisted of individuals that overlapped entirely between the telemetry
369 and camera-trap data, as all radio-collared individuals were also photo captured. While this
370 obviously addressed concerns about collared individuals being representative of the focal
371 population targeted by camera traps, it required a modification to the joint likelihood (Appendix
372 S1) originally constructed under an assumption of data independence (Royle *et al.* 2013). The
373 primary benefit to our joint likelihood came from the increased precision of predicted activity
374 centers for collared individuals (Figure 3), which theoretically should have improved estimation
375 of the inhomogeneous point process model for density and the conditional probabilities of
376 encounter. The resulting inferences were mostly similar to the separate analyses previously
377 described by Sirén *et al.* (2016a) and Sirén *et al.* (2016b) regarding average density and multi-
378 scale resource selection, respectively, of American marten in alpine forests of New England.

379 Sirén *et al.* (2016b) used a “design III” approach to fitting an RSF (Manly *et al.* 2002) and found
380 that regenerating forest was the most important factor influencing second-order resource
381 selection by marten, as individuals tended to avoid it on the landscape. Here, our
382 inhomogeneous point process was akin to a “design II” approach (Manly *et al.* 2002) and we
383 used proximity metrics (e.g., distance to feature) instead of compositional metrics for the spatial
384 habitat covariates; therefore, some differences were to be expected. In other landscapes of
385 eastern North America, marten have similarly exhibited second- and third-order selection for
386 older-aged mixedwood forest (Potvin, Bélanger & Lowell 2000; Cheveau *et al.* 2013). In the
387 present analysis, density estimation was simultaneously modulated by both second-order and
388 third-order resource selection, which were found to be important predictors of variation in the
389 observed spatial encounters for both camera traps and telemetry locations.

390 In summary, the integrated SCR-RSF model addresses concerns regarding heterogeneity
391 in capture due to individual space use which can otherwise generate bias in the estimation of
392 density using spatial capture-recapture models (Royle *et al.* 2013). The increased popularity in
393 using SCR to estimate density of rare, wide-ranging species (e.g., carnivores) will result in many
394 sparse datasets that are unlikely to support complex encounter models (Sollmann *et al.* 2013a).
395 Adding several individuals with VHF or GPS collars to provide an auxiliary source of movement
396 information can increase the accuracy and precision of inferences from spatial encounter designs,
397 particularly when species are selecting resources at multiple scales. Using the modified
398 likelihood as made available in \circ SCR (Sutherland, Royle & Linden 2016) will allow researchers
399 with data sets containing heavy overlap of individuals to fit integrated SCR-RSF models that can
400 accommodate the lack of independence and improve parameter estimation. Integrated modeling
401 approaches allow ecologists to make inferences with the best available information and improve

402 our understanding of ecological systems and our ability to develop effective conservation and
403 management strategies.

404

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414 **Data accessibility**

415 Data will be archived with Dryad Digital Repository.

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505

506 Table 1. Model selection results for covariates influencing marten movement scale (σ_i) using
 507 fully parameterized models for density (D_g) and encounter rate (λ_{ijk}) both without telemetry and
 508 with telemetry integration. Covariates specified here include year (yr) and sex; models with
 509 interactions or quadratic terms always included the linear terms.

Model	nPar	AIC	Δ AIC	AICwt	LogLik
No telemetry					
$D(*) \lambda(*) \sigma(\cdot)$	12	1496.10	0.00	0.42	736.05
$D(*) \lambda(*) \sigma(\text{yr})$	13	1497.16	1.06	0.25	735.58
$D(*) \lambda(*) \sigma(\text{sex})$	13	1498.05	1.95	0.16	736.02
$D(*) \lambda(*) \sigma(\text{yr} + \text{sex})$	14	1498.82	2.72	0.11	735.41
$D(*) \lambda(*) \sigma(\text{yr} \times \text{sex})$	15	1500.11	4.01	0.06	734.06
Telemetry					
$D(*) \lambda(*) \sigma(\text{yr} \times \text{sex})$	15	5327.34	0.00	0.99	2648.67
$D(*) \lambda(*) \sigma(\text{yr} + \text{sex})$	14	5336.79	9.45	0.01	2654.39
$D(*) \lambda(*) \sigma(\text{sex})$	13	5345.30	17.96	0.00	2659.65
$D(*) \lambda(*) \sigma(\text{yr})$	13	5346.79	19.45	0.00	2660.39
$D(*) \lambda(*) \sigma(\cdot)$	12	5350.74	23.39	0.00	2663.37

510

511 Table 2. Model selection results for covariates influencing marten encounter rate (λ_{ijk}) using the
 512 top model for movement scale (σ_i) and the fully parameterized model for density (D_g) both
 513 without telemetry and with telemetry integration. Covariates specified here include distance to
 514 mixed forest (mixed), terrain ruggedness (vrm, vrm^2), year (yr), sex, and a trap-specific
 515 behavioral response (b); models with interactions or quadratic terms always included the linear
 516 terms. Models having variables with 85% confidence intervals that included zero are not listed.

Model	nPar	AIC	Δ AIC	AICwt	LogLik
No telemetry					
$D(*) \lambda(\text{behav} + \text{sex}) \sigma(\cdot)$	7	1487.70	0.00	0.52	736.85
$D(*) \lambda(\text{behav}) \sigma(\cdot)$	6	1487.85	0.15	0.48	737.93
Telemetry					
$D(*) \lambda(\text{behav} + \text{mixed} + \text{vrm}^2) \sigma(\text{yr} \times \text{sex})$	12	5324.80	0.00	0.36	2650.4
$D(*) \lambda(\text{behav} + \text{mixed} + \text{vrm}) \sigma(\text{yr} \times \text{sex})$	11	5326.10	1.29	0.19	2652.05
$D(*) \lambda(\text{behav} + \text{vrm}^2) \sigma(\text{yr} \times \text{sex})$	11	5326.40	1.59	0.16	2652.2
$D(*) \lambda(\text{behav} + \text{mixed}) \sigma(\text{yr} \times \text{sex})$	10	5327.03	2.22	0.12	2653.51
$D(*) \lambda(\text{behav} + \text{mixed} + \text{vrm}^2) \sigma(\text{yr} \times \text{sex})$	9	5327.41	2.61	0.10	2654.71
$D(*) \lambda(\text{behav} + \text{vrm}) \sigma(\text{yr} \times \text{sex})$	10	5327.67	2.86	0.08	2653.83

517

518 Table 3. Model selection results for covariates influencing marten density (D_g) using the top
 519 models for movement scale (σ_i) and encounter rate (λ_{ijk}) both without telemetry and with
 520 telemetry integration. Covariates specified here include distance to mixed forest (mixed), terrain
 521 ruggedness (vrm, vrm^2), year (yr), sex, and a trap-specific behavioral response (b); models with
 522 interactions or quadratic terms always included the linear terms.

Model	nPar	AIC	Δ AIC	AICwt	LogLik
No telemetry					
$D(\text{mixed}) \lambda(\text{behav} + \text{sex}) \sigma(\cdot)$	7	1487.70	0.00	0.99	736.85
$D(\cdot) \lambda(\text{behav} + \text{sex}) \sigma(\cdot)$	6	1497.88	10.18	0.01	742.94
Telemetry					
$D(\text{mixed}) \lambda(\text{behav} + \text{mixed} + \text{vrm}^2) \sigma(\text{yr} \times \text{sex})$	12	5324.80	0.00	0.85	2650.40
$D(\cdot) \lambda(\text{behav} + \text{mixed} + \text{vrm}^2) \sigma(\text{yr} \times \text{sex})$	11	5328.32	3.52	0.15	2653.16

523

524

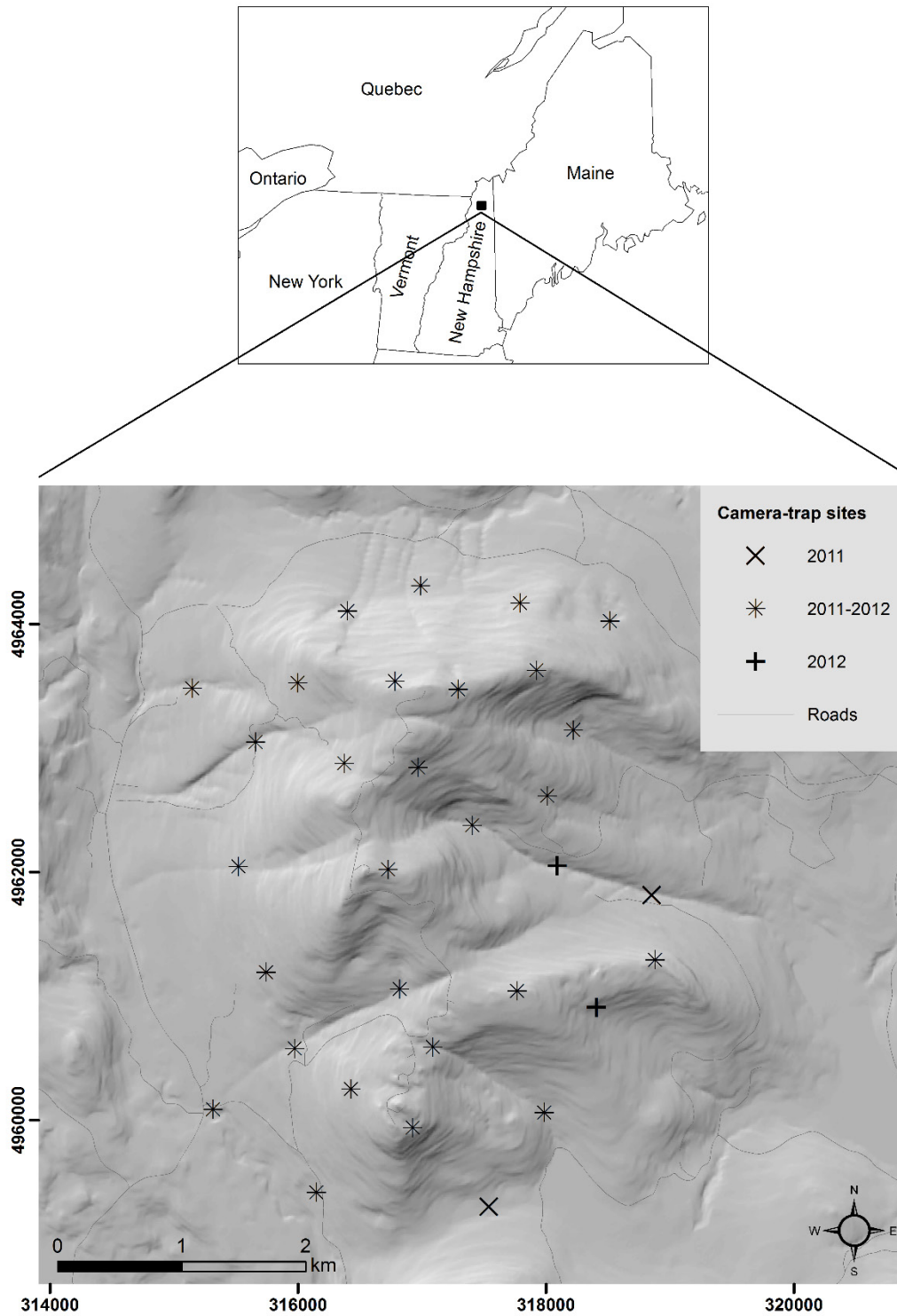
525 Table 4. Maximum likelihood estimates (with standard errors) from the top-ranked spatial
 526 capture-recapture models.

Parameter	No telemetry	Telemetry
Movement scale: $\log(\sigma_i)$		
δ_0	-0.23 (0.063)	-0.39 (0.055)
δ_{2012}	–	-0.37 (0.083)
δ_{male}	–	0.07 (0.066)
$\delta_{2012,\text{male}}$	–	0.32 (0.097)
Encounter rate: $\log(\lambda_{ijk})$		
α_0	-2.54 (0.228)	-2.11 (0.141)
$\alpha_{2,\text{mixed}}$	–	-0.11 (0.057)
$\alpha_{2,\text{vrm}}$	–	-0.11 (0.055)
$\alpha_{2,\text{vrm}2}$	–	0.06 (0.034)
α_{2012}	–	–
α_{male}	0.29 (0.201)	–
$\alpha_{2012,\text{male}}$	–	–
α_{behav}	1.33 (0.174)	1.19 (0.151)
Density: $\log(E(D_g))$		
β_0	-5.58 (0.869)	-4.32 (0.311)
β_{mixed}	-2.09 (0.792)	-0.79 (0.397)

527

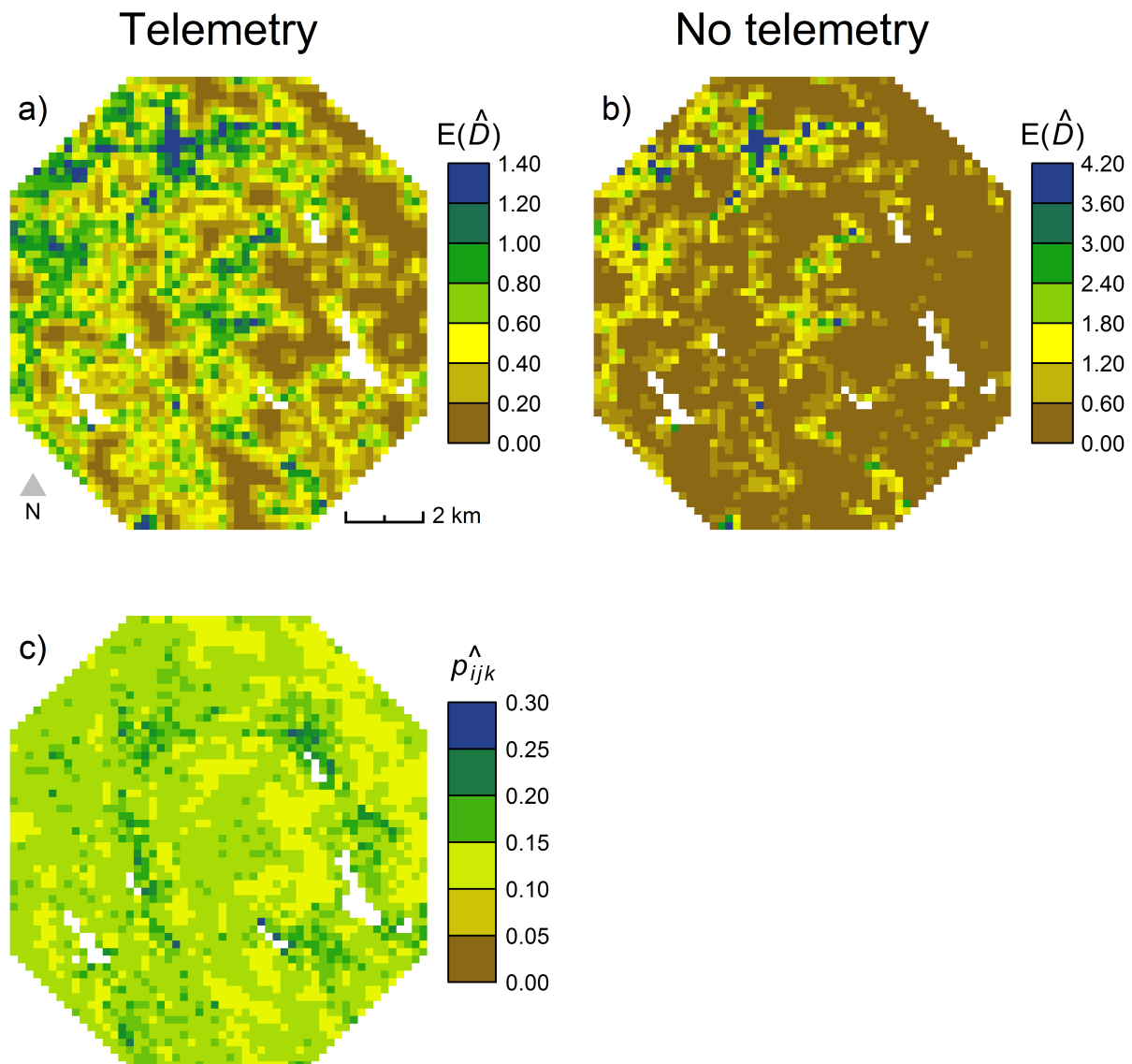
528

529 Figure 1. Map of marten study area in northern New Hampshire, USA, reproduced with some
530 modification from Sirén *et al.* (2016a).



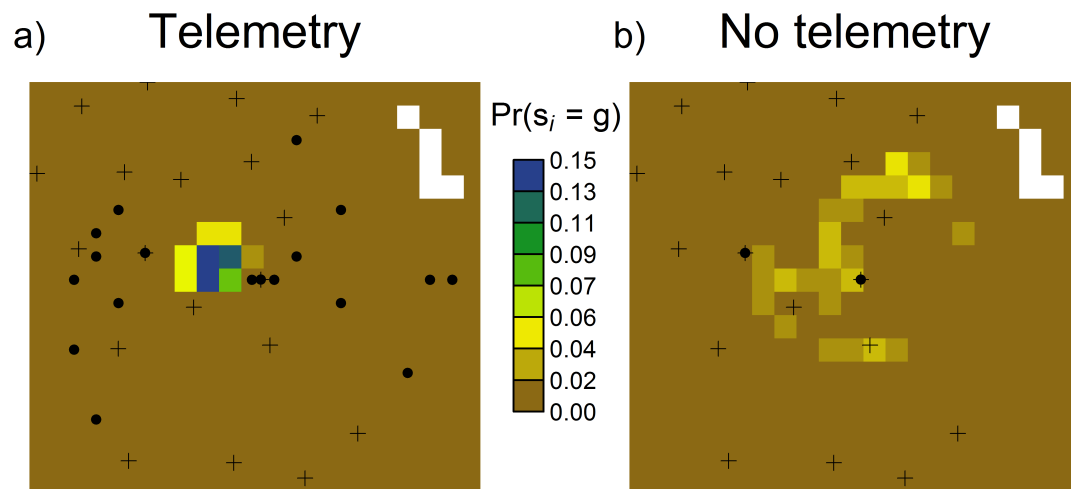
531

532 Figure 2. Predicted density of marten ($\#/km^2$) under the integrated SCR-RSF model with
533 telemetry (a) and a standard SCR model with no telemetry (b), illustrating 2nd order resource
534 selection. In addition, we map the predicted encounter probability for the integrated model (c),
535 illustrating 3rd order resource selection. Blank interior pixels represent water.
536



537

538 Figure 3. Probability densities of the activity center s_i in pixel g for one radio-collared male
539 marten in 2012 from the integrated SCR-RSF model with telemetry (a) and a standard SCR
540 model with no telemetry (b). Crosses represent camera traps and black dots represent spatial
541 encounters within a 200 m pixel recorded by cameras and/or telemetry fixes.
542



543

544 **Appendix S1. Likelihood for integrated SCR-RSF model with non-independence**

545

546 Royle et al. (2013; Supplement 1) described the full likelihood for the integrated SCR-RSF
547 model assuming that the capture data and telemetry data were independent. To accommodate the
548 marten study where all collared individuals were also photo-captured, the conditional likelihoods
549 need to be combined before calculating the marginal likelihood for each individual. We do not
550 describe all components of the likelihoods here as they are fully described in Royle et al. (2013;
551 Supplement 1); we instead emphasize the main differences between the independent and non-
552 independent formulations. In addition, our model description here is simplified to match that of
553 Royle et al. (2013), removing some details specific to the marten study.

554 When the datasets are independent, the total likelihood for the integrated SCR-RSF
555 model is the product of the likelihoods for the capture data (SCR) and the telemetry data (RSF):

$$556 L_{rsf+scr}(\boldsymbol{\alpha}, N; \mathbf{y}, \mathbf{m}) = L_{scr}(\alpha_0, \alpha_1, \alpha_2, N; \mathbf{y}) \times L_{rsf}(\alpha_1, \alpha_2; \mathbf{m})$$

557 Here the $\boldsymbol{\alpha}$ parameters correspond to variation in the encounter rate (for spatial encounters, \mathbf{y}) and
558 usage rate (for telemetry locations, \mathbf{m}), while population size (N) is only estimated from the SCR
559 model. Note that α_0 only appears in the SCR likelihood as it corresponds to the baseline
560 encounter rate and does not involve spatial variation. The other parameters represent availability
561 (α_1), conditional on the latent activity centers (\mathbf{s}), and resource selection (α_2).

562 The conditional-on- \mathbf{s} likelihoods, here represented as $f(\text{data} | \mathbf{s}, \text{parameters})$, differ
563 according to the observation models for the capture data and the telemetry data. The marginal
564 distributions, here represented as $f(\text{data} | \text{parameters})$, for each individual i are calculated by
565 integrating the conditional-on- \mathbf{s} likelihoods over the possible locations for the individual activity
566 centers, \mathbf{s}_i :

567
$$f(\mathbf{y}_i | \boldsymbol{\alpha}) = \int_S f(\mathbf{y}_i | \mathbf{s}_i, \boldsymbol{\alpha}) g(\mathbf{s}_i) d\mathbf{s}_i$$

568
$$f(\mathbf{m}_i | \boldsymbol{\alpha}) = \int_S f(\mathbf{m}_i | \mathbf{s}_i, \boldsymbol{\alpha}) g(\mathbf{s}_i) d\mathbf{s}_i$$

569 Here, S represents the continuous state space over which the integral is calculated and $g(\mathbf{s}_i) =$
 570 $1/||S||$ when density does not vary, indicating a homogeneous point process for activity centers
 571 (note: we used a discrete state space and an inhomogeneous point process (Borchers & Efford
 572 2008) for the marten study). The likelihoods for all observed individuals in the data (n for
 573 captures, N_{tel} for telemetry) are then the products of the individual likelihoods:

574
$$L_{scr}(\boldsymbol{\alpha} | \mathbf{y}_1, \mathbf{y}_2, \dots, \mathbf{y}_n) = \prod_{i=1}^n f(\mathbf{y}_i | \boldsymbol{\alpha})$$

575
$$L_{rsf}(\boldsymbol{\alpha} | \mathbf{m}_1, \mathbf{m}_2, \dots, \mathbf{m}_{N_{tel}}) = \prod_{i=1}^{N_{tel}} f(\mathbf{m}_i | \boldsymbol{\alpha})$$

576 As described earlier, the product of these two data likelihoods then provides the total likelihood.
 577 Royle et al. (2013; Supplement) provide a more complete description of the likelihoods,
 578 including the important contribution of unobserved individuals for estimating N from the capture
 579 data. We do not highlight the components of the likelihood that remain the same regardless of
 580 independence between datasets.

581 In applications where the data sources are not independent (including the marten study),
 582 the individual conditional-on-s likelihoods for each data source must be combined before
 583 computing the marginal likelihood for each individual. When all collared individuals are also
 584 captured, the marginal likelihood for collared individuals is as follows:

585
$$f(\mathbf{y}_i, \mathbf{m}_i | \boldsymbol{\alpha}) = \int_S f(\mathbf{y}_i | \mathbf{s}_i, \boldsymbol{\alpha}) f(\mathbf{m}_i | \mathbf{s}_i, \boldsymbol{\alpha}) g(\mathbf{s}_i) d\mathbf{s}_i$$

586 The marginal likelihood for the $n - N_{tel}$ individuals that were captured but not collared does not
587 change. The joint likelihood for the observations, assuming collared individuals are sorted first,
588 is then:

$$589 \quad L_{scr|rsf}(\boldsymbol{\alpha} | \mathbf{y}, \mathbf{m}) = \prod_{i=1}^{N_{tel}} f(\mathbf{y}_i, \mathbf{m}_i | \boldsymbol{\alpha}) \times \prod_{i=N_{tel}+1}^n f(\mathbf{y}_i | \boldsymbol{\alpha})$$

590 In this way, the likelihood for the telemetry data becomes embedded in the likelihood for
591 the capture data for those collared individuals that were also captured. If there were also some
592 collared individuals that were never captured, the joint likelihood would involve the product of
593 the two components above and a third representing the marginal distribution of the telemetry
594 data. Thus, depending on how individuals overlap between data sets, there are three potential
595 marginal distributions to calculate for the observed individuals: 1) captures only; 2) collars only;
596 and 3) captures and collars. Assuming independence, only #1 and #2 are used in the joint
597 likelihood. The marten study here consisted of #1 and #3. The integrated likelihood in $\circ SCR$
598 (Sutherland, Royle & Linden 2016) currently supports model fitting to data consisting of #1 and
599 combinations involving #1 and #2, and #1 and #3.

600

601

602

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608 Capture-Recapture Models. R package version 0.30.1. <https://github.com/jaroyle/oSCR>
609