

# Regime shifts and hysteresis in the pitcher-plant microecosystem

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## Abstract

Changes in environmental conditions can lead to rapid shifts in the state of an ecosystem ("regime shifts"), which, even after the environment has returned to previous conditions, subsequently recovers slowly to the previous state ("hysteresis"). Large spatial and temporal scales of dynamics, and the lack of frameworks linking observations to models, are challenges to understanding and predicting ecosystem responses to perturbations. The naturally-occurring microecosystem inside leaves of the northern pitcher plant (*Sarracenia purpurea*) exhibits oligotrophic and eutrophic states that can be induced by adding insect prey. Here, we further develop a model for simulating these dynamics, parameterize it using data from a prey addition experiment and conduct a sensitivity analysis to identify critical zones within the param-

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eter space. Simulations illustrate that the microecosystem model displays regime shifts and hysteresis. Parallel results were observed in the plant itself after experimental enrichment with prey. Decomposition rate of prey was the main driver of system dynamics, including the time the system remains in an anoxic state and the rate of return to an oxygenated state. Biological oxygen demand influenced the shape of the system's return trajectory. The combination of simulated results, sensitivity analysis and use of empirical results to parameterize the model more precisely demonstrates that the *Sarracenia* microecosystem model displays behaviors qualitatively similar to models of larger ecological systems.

*Keywords:*

ecosystem dynamics, non-linear systems, food webs, ecological networks, nutrients, dissolved oxygen, hysteresis, regime shifts

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1 **1. Introduction**

2 Regime shifts in ecological systems are defined as rapid changes in the  
3 spatial or temporal dynamics of an otherwise resilient system. Ecological  
4 regime shifts are caused by slow, directional changes in one or more underly-  
5 ing state variables, such as species abundance, dissolved oxygen content, or  
6 nutrients [1, 2]. Regime shifts are of particular concern when the return rate  
7 to a previous (and perhaps more desirable) state is slow or requires a larger  
8 input of energy or resources relative to what initiated the state change (i.e.,  
9 hysteresis). In the last several years, many researchers have suggested that a

10 wide range of ecological systems are poised to “tip” into new regimes [2, 3],  
11 or even that we are approaching a planetary tipping point [4]; but see [5].  
12 Because identifying changes in the underlying state variables of most ecosys-  
13 tems require high frequency, long-term measurements [6], our understanding  
14 of the causes and consequences of ecological regime shifts has progressed  
15 relatively slowly. More rapid progress could be achieved by working with  
16 well-understood systems that can be described mathematically and manipu-  
17 lated experimentally over shorter time scales.

18 It is rare to find an ecological system in which the occurrence of a regime  
19 shift, and its cause-and-effect relationship with one or more underlying envi-  
20 ronmental drivers, is unambiguous [7]. This is primarily because long time  
21 series of observations collected at meaningfully large spatial scales are re-  
22 quired to identify the environmental driver(s), its relationship to the re-  
23 sponse variable of interest, the stability of each state, the breakpoint be-  
24 tween them, and hysteresis of the return time to the original state [3, 7].  
25 Detailed modeling, and decades of observations, and experiments have led to  
26 a thorough understanding of one canonical example of an ecological regime  
27 shift: the rapid shift from oligotrophic (low nutrient) to eutrophic (high nu-  
28 trient) states in lakes (e.g., [8, 9]). The primary difficulties with using lakes as  
29 models for studying alternative states and ecological regime shifts are their  
30 large size (which precludes extensive replication: [10]) and the long time scales  
31 (decades) required to observe a regime shift, subsequent ecosystem hystere-  
32 sis, and eventual recovery [11, 12]. Models of lake ecosystems and their food

33 webs, and associated empirical data have revealed that recovery of lakes from  
34 a eutrophic to an oligotrophic state can be very slow—on the order of decades  
35 to centuries [12]—and depends not only on slowing or reversing directional  
36 changes in underlying state variables but also on the internal feedback dy-  
37 namics of the system. Other aquatic systems, including fisheries [13], rocky  
38 intertidal communities, and coral reefs [3] have provided additional empirical  
39 support for these model results in terms of both dynamics and duration [14].

40 In a previous study, we experimentally demonstrated that organic-matter  
41 loading (i.e., the addition of excess insect prey to pitchers) can cause a shift  
42 from oligotrophic to eutrophic conditions in a naturally-occurring microe-  
43 cosystem: the water-filled leaves of the northern (or purple) pitcher plant,  
44 *Sarracenia purpurea* L. [15]. We use the term “microecosystem” here be-  
45 cause the pitcher plant and its inquiline food web is a naturally occurring,  
46 co-evolved community of organisms, which is not necessarily the case for  
47 microcosms [16]. In the typically five-trophic level *Sarracenia* microecosys-  
48 tem, bacteria reproduce rapidly and drive the nutrient-cycling dynamics [17].  
49 Prey additions cause shifts from oligotrophic to eutrophic states in hours or  
50 days rather than years or decades. Further, the comparatively small volume  
51 of individual pitchers, the ease of growing them in greenhouses and the oc-  
52 currence of large, experimentally manipulable populations in the field [18]  
53 have allowed for replicated studies of trophic dynamics and regime shifts in  
54 a whole ecosystem.

55 Here, we build on the original mathematical model of the *Sarracenia*

56 microecosystem [15], estimating parameter values using new empirical data  
57 and introducing more realism into the underlying environmental drivers of  
58 the model. We then use sensitivity analysis to identify the model parameters  
59 that most strongly control the dynamics of the system. We illustrate that  
60 once organic-matter input is stopped, the *Sarracenia* microecosystem—like  
61 large lakes—can eventually overcome the hysteresis in the system and return  
62 to an oligotrophic state. We conclude that the mathematical model illus-  
63 trates dynamic behaviors that are qualitatively similar to models of regime  
64 shifts in lakes and other ecosystems, and we suggest that the *Sarracenia*  
65 microecosystem is useful model for studying ecological regime shifts in real  
66 time.

## 67 2. Methods

### 68 2.1. *The pitcher-plant microecosystem*

69 The eastern North American pitcher plants (*Sarracenia* spp.) are peren-  
70 nial carnivorous plants that grow in bogs, low nutrient (“poor”) fens, seep-  
71 age swamps, and sandy out-wash plains [19]. Their leaves are modified into  
72 “pitchers” [20], tubular structures that attract and capture arthropods, and  
73 occasionally small vertebrate prey (e.g., [21, 22]). In the pitchers, prey are  
74 shredded by obligate pitcher-inhabiting arthropods, including histiostomatid  
75 *Sarraceniopus* mites, and larvae of sarcophagid (*Fletcherimyia fletcheri*) and  
76 chironomid flies (*Metrocnemius knabi*) [23, 24, 25]. The shredded organic  
77 matter is further decomposed and mineralized by a diverse assemblage of

78 microbes, including protozoa [26], yeasts [27], and bacteria [28].

79 Unlike other species of *Sarracenia* that also secrete and use digestive en-  
80 zymes to extract nutrients from their captured prey, *S. purpurea* pitchers  
81 secrete digestive enzymes for only a fraction of their lifespan [29]. Instead,  
82 *S. purpurea* relies on its aquatic food web to decompose the prey and min-  
83 eralize their nutrients [30]. As a result, the rainwater-filled pitchers of *S.*  
84 *purpurea* are best considered a detrital-based, “brown” ecosystems in which  
85 bacterially-mediated nutrient cycling determines whether it is in an oligo-  
86 trophic or eutrophic state [15, 17, 31].

## 87 2.2. Oxygen dynamics in lakes and pitchers

88 Oxygen dynamics, in both lakes and *Sarracenia* pitchers, can be described  
89 using a simple model that yields alternative oligotrophic and eutrophic states  
90 and hysteresis in the shift between them [1]:

$$\frac{dx}{dt} = a - bx + rf(x) \quad (1)$$

91 In this model, the observed variable  $x$  (e.g., oxygen concentration) is  
92 positively correlated with state variable  $a$  (e.g., rate of nutrient input or  
93 photosynthesis), and negatively correlated with state variable  $b$  (e.g., rate  
94 of nutrient removal or respiration). The function  $rf(x)$  defines a positive  
95 feedback that increases  $x$  (e.g., the rate of nutrient recycling between the  
96 sediment in lakes or mineralization-immobilization by bacteria of shredded  
97 prey in a water-filled *Sarracenia* pitcher). If  $r > 0$  and the maximum of

98  $\{rf(x)\} > b$ , there will be more than one equilibrium point (i.e., stable  
99 state) [1]; the function  $f(x)$  determines the shape of the switch between the  
100 states and the degree of hysteresis.

101 Following [1], we used a Hill function for  $f(x)$ :

$$f(x) = \frac{x^p}{x^p + h^p} \quad (2)$$

102 The Hill function provides a simple model that can produce threshold be-  
103 haviors. The dynamics of the state variable  $x$  is determined by parameters  
104  $p$  and  $h$ , which determine the rate of change and the inflection point of the  
105 curve, respectively (Fig. 1A). If  $p$  is set such that more than one possible state  
106 exists for the system,  $h$  determines the threshold for the transition between  
107 these states. When viewed in phase-space (Fig. 1B), the transition between  
108 states can be seen as a path traversed by the system between distinct regions  
109 (i.e., phases). In part because of this threshold property, the Hill function  
110 has been applied to systems ranging from biochemistry and microbiology to  
111 ecology, whose dynamics depend on a limiting resource (e.g.,[32]).

112 We modeled the dynamics of the trophic state of the *Sarracenia* microe-  
113 cosystem using an equation of the same underlying form as Eq. 1:

$$x_{t+1} = \underbrace{a_t A_t}_{\text{Photosynthesis}} - \underbrace{\left\{ m + a_t \left[ \frac{w_{t-1}}{w_{t-1} + K_w} \right] \right\}}_{\text{BOD}} + \underbrace{D_t(x_t)}_{\text{Diffusion}} \quad (3)$$

114 Each model term is described below and summarized in Table 1.

115 The model (Eq. 3) of the *Sarracenia* microecosystem (Fig. 2A) is made  
116 up of the two main terms: production of oxygen by photosynthesis, and use  
117 of oxygen (respiration) during decomposition (BOD: biological oxygen de-  
118 mand). The pitcher fluid is oxygenated ( $x$ ) at each discrete time step ( $t$ ,  
119 in minutes) as the plant photosynthesizes ( $A_t$ ). The value of  $A_t$  is deter-  
120 mined by sunlight, which we modeled as a truncated sine function producing  
121 photosynthetically active radiation (PAR) (Fig. 2B), and by the maximum  
122 photosynthetic rate ( $A_{max}$ ) (Fig. 2C), which leads to the production of dis-  
123 solved oxygen in the pitcher fluid (Fig. 2D). Photoautotrophic bacteria, but  
124 not algae, have been detected in a recent molecular study of the proteome  
125 of *S. purpurea* inquiline communities; however, the total number of peptides  
126 mapping to these bacterial species was low relative to other species in the  
127 community [33]. Thus, although our model does not explicitly account for  
128 other photosynthetic organisms, these previous findings suggest that their  
129 contributions to the oxygen dynamics of the pitcher-plant fluid are likely to  
130 be small relative to that of the pitcher plant itself.

131 Decomposition of shredded prey by bacteria requires oxygen. The oxy-  
132 gen demand from respiration is modeled by the BOD term in Eq. 3. The  
133 parameter  $m$  is the basal metabolic respiration of the food web with no prey  
134 to decompose in the system. Adding prey ( $w$ ) induces decomposition, which  
135 we model as a negative exponential function with rate parameter  $\beta$ , and a  
136 constant  $W$  (maximum prey mass decomposed over 48 hours) using Eq. 4,  
137 and illustrated in Figure 2E.



$$w_{t+1} = w_t \cdot e^{-\beta \cdot W} \quad (4)$$

138 Bacterial populations increase at a rate determined by a half-saturation  
139 function with parameter  $K_w$  (Eq. 3), which increases BOD, and the deple-  
140 tion of oxygen from the pitcher fluid (Fig. 2F). Demand of the food web for  
141 oxygen (i.e., BOD) depends on the decomposition rate ( $\beta$ ) and the shape  
142 parameter ( $K_w$ ), but only when prey is present in the system ( $w_{t-1} > 0$   
143 in Eq. 3). When prey is absent (i.e.,  $w_{t-1} = 0$ ), BOD terms simplify by  
144 multiplication to the basal metabolic rate ( $m$ ).

145 Photosynthesis may be limited by available nutrients, primarily nitrogen  
146 and phosphorus [34, 35], that are mineralized by bacteria from the prey  
147 [17]. Photosynthesis is augmented ( $a_t$ ) by nutrient mineralization rate ( $s$ ).  
148 We model  $a_t$  as a saturating function with bounds determined by the range  
149 terms ( $a_{min}$ , and  $a_{max}$ ),  $s$ , and the point of saturation ( $d$ ):

$$a_{t+1} = a_t \times \left\{ \frac{a'_{max} - a'_{min}}{1 + e^{-s \cdot n_t - d}} + a'_{min} \right\} \quad (5)$$

150 The other impact of prey addition, and subsequent decomposition by the  
151 food web, is the release of nutrients into the pitcher fluid. The mineralization  
152 variable  $n_t$  (Eq. 5), which is modeled as proportional to the product of the  
153 amount of oxygen and prey in the system (i.e.,  $n_{t+1} = c \cdot (w_t \cdot x_t)$  where  $c$   
154 is a constant of proportionality), creates a feedback from decomposition to  
155 oxygen production by the plant (i.e., the path in Fig. 2A from the food web

156 to nutrients to pitcher to oxygen and back to the food web).

157 Thus, the mineralization term couples respiration (oxygen depletion) to  
158 photosynthesis (oxygen production) when prey is introduced to the system,  
159 and the food web begins to decompose the prey and release nutrients into the  
160 pitcher fluid. Finally, a small amount of oxygen diffuses into the pitcher fluid  
161 directly from the atmosphere ( $D(x)$ ); however, the diffusion rate is generally  
162 so low that it is negligible relative to changes resulting from photosynthesis  
163 and BOD.

### 164 *2.3. Estimating decomposition rate*

165 We used data from a greenhouse pitcher-feeding experiment to estimate  
166 the decomposition parameter. The experiment was conducted over 35 days,  
167 starting on July 6, 2015, in a temperature-controlled greenhouse at the Uni-  
168 versity of Vermont's Biological Research Complex (Burlington, Vermont,  
169 USA). Eighteen newly-formed pitchers  $> 8$  ml in volume were rinsed with  
170 deionized water, and randomly assigned to one of three organic matter addi-  
171 tion treatments: 0.0, 0.5, or 5.0 mg mL<sup>-1</sup> of pitcher fluid. Pitcher fluid was  
172 collected from randomly-selected pitchers at Molly Bog (Morristown, Ver-  
173 mont, USA: 44.500 N,-72.642 W) on the morning of 6 July 2015. The fluid  
174 was transported to the greenhouse, filtered through the 30-micron frit bed  
175 of a chromatography column (BioRad, Hercules, California, USA) to remove  
176 larger organisms, homogenized and added to experimental pitchers in the  
177 greenhouse.

178 Pitchers were loaded with a single pulse of organic matter every morning  
179 for the first four days. In this experiment (and following [36]), the organic  
180 matter was bovine serum albumin (BSA), DNA from salmon testes at a  
181 concentration of  $1.5 \mu\text{g L}^{-1}$  of BSA, and trace elements, potassium, calcium,  
182 sodium, magnesium and manganese in a ratio of 1:0.115:0.044:0.026:0.0001.  
183 In a pilot study, BSA yielded similar changes in dissolved oxygen as we  
184 had obtained previously using ground wasps as organic matter [15], and its  
185 use enabled us to measure changes in organic matter content via a simple  
186 Bradford assay, and in a pilot study, yielded similar changes in dissolved  
187 oxygen as we had obtained previously using ground wasps as organic matter  
188 (Sirota et al. 2013).

189 Three  $100 \mu\text{L}$  aliquots of pitcher fluid were sampled from single pitchers  
190 of each of 14 individual plants. Sampling was conducted twice a day from  
191 day 0 to day 20 at 8:30am and 5:00pm ( $\pm 2$  hrs), once per day from day 20 to  
192 day 28 at 8:30am ( $\pm 2$  hrs), and once each on days 30, 31, 33, and 35 (8:30am  
193  $\pm 2$  hrs). Prior to sampling, the pitcher fluid was stirred with the pipette  
194 submerged to get fluid representative of an average of the entire pitcher.  
195 During sampling care was taken to minimize the introduction of additional  
196 oxygen from the sampling process itself. Pitcher fluid was topped off with  
197 deionized water after sampling to keep initial pitcher volumes consistent over  
198 the course of the 35 days. The initial volumes varied from 8 mL to 28.6 mL  
199 with a mean of 16.7 mL and standard deviation of 4.82 mL. Although we  
200 did control for the water taken with each sample, we did not adjust for

201 the removal of nutrients and microscopic organisms that occurred at each  
202 sampling.

203 Samples were centrifuged at 13000*g*, after which the supernatant contain-  
204 ing soluble BSA was removed, placed in a sterile tube and frozen at -80 °C  
205 until analyzed. A simple Bradford assay (Bradford 1976) was used to de-  
206 termine the concentration of BSA in each of the pitcher fluid samples. The  
207 assay was done using Bradford reagent (VWR), and the absorbance of each  
208 sample was measured on a Biophotometer Plus spectrophotometer (Eppen-  
209 dorf) at an optical density of 600 nm. Samples were read randomly on the  
210 spectrophotometer to avoid reaction time as a confounding variable. Sample  
211 concentrations were determined by comparison to standard curves generated  
212 using R [37].

213 We used an empirical least-squares estimator (LSE) approach to generate  
214 the best-fitting values for the decomposition parameter ( $\beta$ ) in Eq. 4, given  
215 the quantity of prey added in the experiment and the duration of the prey  
216 addition. As  $K_w$  is not a part of the decomposition function, we did not vary  
217 it during parameter estimation and model fitting. We then ran a series of 35-  
218 day simulations (equivalent to the run-time of the prey-addition experiment)  
219 in which  $\beta$  was sampled from a grid of values ranging from 1E-8 to 0.0007,  
220 and the amount of prey in the simulation was recorded at each simulated  
221 minute. For each run, the sum of squared errors (SSE) was recorded as  
222  $\sum (sim - obs)^2$ . The  $\beta$  that minimized the SSE in each simulation was  
223 considered to be the best-fit value for each replicate pitcher ( $n = 12$ ). All

224 model fitting was also done using R [37].

#### 225 *2.4. Sensitivity Analysis*

226 We used a sensitivity analysis in which we varied the prey addition rate  
227 ( $w$ ), decomposition rate ( $\beta$ ), and the half-saturation constant  $K_w$  to explore  
228 the behavior of the microecosystem model across a wide range of parameter  
229 space. Rather than set combinations of fixed values for the three parameters  
230 of interest, we sampled the model parameter space by drawing values inde-  
231 pendently from uniform distributions:  $K_w \sim U(0.0001, 3)$ ,  $\beta \sim U(1.0E-6,$   
232  $2.0E-3)$  and  $w \sim U(0, 100)$ . To characterize baseline (oligotrophic) oxygen  
233 concentrations, for each combination of  $\beta$  and  $K_w$  we ran one simulation  
234 in which no prey was added to the system ( $w = 0$ ). In all simulations ( $n$   
235  $= 15,000$ ) variables were initialized to 0 with the exception of oxygen ( $x$ ),  
236 which was initialized using the photosynthesis term ( $x_0 = 7.55$ ). Simulated  
237 prey additions occurred at mid-day on days 4–6 (i.e.,  $t = 6480$  to  $t = 9360$ ),  
238 each simulation ran for 30 simulated days (43,200 minutes), and output was  
239 saved for each simulated minute. The simulations were initialized using a  
240 random sample of parameter values, and run in parallel. Because the model  
241 is completely deterministic, the resulting runs can be reproduced by starting  
242 the simulations with the exact values used to initialize and parameterize the  
243 models.

244 To aid in the detection of the impact of the most important parameters  
245 and variables, in all simulations we set some parameter values to zero, which

246 altered the model in the following two ways. First, we ignored the  $D(x_t)$  term  
247 because we assumed that the amount of oxygen diffusing directly into the  
248 pitcher fluid from the atmosphere would be orders of magnitude lower than  
249 oxygen produced by pitcher photosynthesis [38]. Second, we noted that since  
250 the basal metabolic respiration of the food web parameter ( $m$ ) is an additive  
251 constant, any change in the value of the constant  $m$ , (basal respiration of the  
252 microbial community) would result only in a proportional change in value of  
253  $x$ , not in the shape (i.e.,  $dx/dt$ ), of the oxygen production over time. Thus,  
254 we could set  $m = 0$  without loss of generality.

255 By setting  $m = 0$ , we also observed that the photosynthetic augmentation  
256 term ( $a_t$ ) influenced photosynthesis ( $A_t$ ), and BOD ( $\frac{w_{t-1}}{w_{t-1}+K_w}$ ) identically.  
257 Therefore, the parameters  $s$  and  $d$  in Eq. 5 could be set as constants in the  
258 sensitivity analysis. By ignoring diffusion, setting  $m = 0$ , and fixing  $s$  and  $d$ ,  
259 we reduced the dimensionality of the sensitivity analysis to three ( $w$ ,  $\beta$ , and  
260  $K_w$ ), which eased the interpretation of the results.

261 We calculated two measures of the state of the system from the time  
262 series of oxygen concentration ( $x_t$ ): hypoxia, and return rate. We defined  
263 hypoxia in the model to be an oxygen concentration of  $\leq 1.6$  mg L<sup>-1</sup>, which  
264 is the median lethal O<sub>2</sub> concentration ( $[O_2]$ ) for aquatic animals [39]. We  
265 measured the return rate of the system as the linear trend in  $[O_2]$  (i.e., after  
266 removing the daily cycle in oxygen resulting from photosynthesis) across the  
267 entire simulation using Pearson's correlation coefficient. Although the return  
268 trajectory can be non-linear, the linear trend measures the gross trends of

269 returning (positive), little impact of feeding (zero) or remaining at depressed  
270 oxygen levels.

### 271 *2.5. Code Availability, and Execution*

272 The model was coded in the **R** programming language [37]. The 15,000  
273 model runs for the sensitivity analysis were run on the Odyssey Supercom-  
274 puter Cluster at Harvard University (Research Computing Group, FAS Divi-  
275 sion of Science, Cambridge, Massachusetts). Data, code for the simulations,  
276 and output of analyses are available in the Harvard Forest Data Archive  
277 ([harvardforest.fas.harvard.edu/harvard-forest-data-archive](http://harvardforest.fas.harvard.edu/harvard-forest-data-archive)).

## 278 **3. Results**

279 The equation representing decomposition and BOD resembles the Hill  
280 function in a general model of state changes with hysteresis (Eqs. 1 & 2).  
281 In general, when a Hill function is used in a basic alternative states model  
282 (e.g.,  $rf(x) > b$  in Eq. 1), the inflection point (e.g., half-saturation constant  
283  $K_w$ ) determines the threshold (Fig. 1A). Thus, modeling decomposition and  
284 BOD using a Hill function provided us with sufficient flexibility to yield a  
285 variety of state changes.

286 The simulations with the model produced dynamics observed in the em-  
287 pirical pitcher plant microecosystem. Because photosynthesis is nutrient-  
288 limited in *Sarracenia* [35], addition of prey increased modeled photosynthe-  
289 sis (Fig. 3A) relative to oligotrophic, prey-free, pitchers. In the oligotrophic

290 state, and when no prey was added, BOD remained low throughout the  
291 entire simulation (black line in Fig. 3B). After prey was added on, for ex-  
292 ample, days 4–6 ( $t = 6480$  to  $t = 9360$  minutes), the system jumped into  
293 its alternative state: BOD increased rapidly then declined slowly as prey  
294 was mineralized (grey line in Fig. 3B). The combination of the smooth, slow  
295 recovery response of photosynthesis to prey addition and the abrupt shift in  
296 BOD following prey addition (Fig. 3A & B) resulted in an abrupt shift in  
297 the system from an oxygenated state into an anoxic state and a very slow  
298 (hysteretic) recovery (Fig. 3C). The hysteresis of the system was apparent  
299 when oxygen concentration was plotted as a time-lagged phase plot (lag =  
300 1440 minutes starting at  $t=720$ ), which shows the change in oxygen following  
301 addition of prey at  $t = 6480$  and the slow return due to high BOD (Fig. 3D).  
302 These results were corroborated by observations from field and greenhouse  
303 experiments in which oxygen was observed to decline with the capture or  
304 addition of insect prey to the pitcher [15], and demonstrate the presence of  
305 both state changes and hysteresis (i.e., Fig. 3D) for at least some parameter-  
306 izations of the model.

307 The parameter fitting and sensitivity analysis revealed several key effects  
308 of the parameters that we varied. First, the LSE model-fitting procedure  
309 resulted in an estimate for  $\beta$  of  $\bar{x} = 0.00041 \pm 0.0004$  [SE] (Fig. 4, vertical  
310 lines). Second, varying  $\beta$  had a large effect both on the percent time spent  
311 in an hypoxic state and on the return rate (steeper contours with increasing  
312  $\beta$  in Fig. 4). Last, varying the amount of prey by two orders of magnitude



313 produced a sharp threshold for the effect of varying  $\beta$  on hypoxia and return  
314 rate (Fig. 4).

315 Although varying  $\beta$  has a potentially larger effect on the dynamics of  
316 the microecosystem than varying  $K_w$ , the latter played an important role in  
317 determining the return trajectory of the oxygen. For simulations with lower  
318 values of  $K_w$ , the oxygen concentration was still exponentially increasing  
319 when the simulation ended (Fig. 5A). Relative to simulations with higher  
320  $K_w$ , the return rate was faster when  $\beta$  was low enough, and there was prey  
321 (i.e.,  $w_t$ ) remaining in the pitcher at the last observed time (Fig. 5B). Thus,  
322 in this part of the parameter space, if another round of feeding were to occur  
323 at a similar level of prey input, the system would never recover, and would  
324 remain in or near an hypoxic state.

#### 325 4. Discussion

326 General theoretical work in complex systems has suggested that the defi-  
327 nition of system boundaries is arbitrary and carries the potential for system  
328 dynamics to be mechanistically connected to, but unpredictable from, lower  
329 levels (or scales) of organization [40, 41]. However, others have argued that  
330 food-web dynamics of whole ecosystems can be inferred from the compo-  
331 nents (i.e., motifs and modules) of these ecosystems [42]. Overall, our model  
332 of the *Sarracenia* microecosystem supports the latter assertion: a focus on  
333 particular pathways (e.g., photosynthesis, decomposition, etc.) reproduced  
334 the non-linear behavior of its oxygen dynamics, including state changes and

335 hysteresis. The results of the sensitivity analysis also revealed that the car-  
336 rying capacity of the bacterial community (as simulated by the effect of  $K_w$ )  
337 could contribute to observed non-linear state-changes of the *Sarracenia* mi-  
338 croecosystem.

339 Predictions based on the model are highly sensitive to changes in the pa-  
340 rameterization of decomposition (e.g.  $\beta$ ). In the initial parameterization of  
341 this model, we started with an empirical estimate of decomposition rate in  
342 which  $> 99\%$  of the average amount of prey captured could be decomposed in  
343 a single day [15, 43]. This is extremely rapid decomposition relative to a set  
344 of 58 previously published food webs [44], in which 1.27–66.2% of available  
345 detritus or organic matter is decomposed each day. When we set the de-  
346 composition parameter ( $\beta$ ) equal to 2.57E-6, the overall decomposition rate  
347 approached the mean of the published food webs ( $24.22 \pm 2.79\%$  [SE]). This  
348 value for  $\beta$  is within the parameter space that we observed experimentally,  
349 and used in our sensitivity analysis, and suggests that insights gained from  
350 the *Sarracenia* microecosystem should be scalable to larger systems.

351 Although the dynamics of the *Sarracenia* microecosystem share similar-  
352 ities with lake, stream and other large-scale ecosystems, there are several  
353 differences that should be noted. First, oxygen levels in the pitcher plant are  
354 dynamically controlled by photosynthesis of the plant that serves as a strong  
355 driver of oxygen levels. In lakes, the primary oxygen production is carried out  
356 by phytoplankton, which are immersed in the aquatic system. Second, lake  
357 food webs are “green” (i.e., plant-based); whereas pitcher plant food webs

358 are “brown” (detritus-based [17]). In lakes, the shift to a eutrophic state  
359 occurs through addition of limiting nutrients (usually N or P), accumulation  
360 of producer biomass that is uncontrolled by herbivores (see [45], and subse-  
361 quent decomposition that increases biological oxygen demand [46, 47]. The  
362 *Sarracenia* microecosystem’s “brown” food web also experiences an increase  
363 in oxygen demand and microbial activity; however, this occurs during the  
364 breakdown of detritus that is characteristic of its shift from an oligotrophic  
365 to a eutrophic state [15]. Even though the source of the nutrients in the  
366 *Sarracenia* microecosystem is “brown”, the functional shape of the pathways  
367 involved in its nutrient cycling are similar to those in lakes with “green” food  
368 webs and are likely to lead to similar qualitative dynamics of both systems.

369 The results of our model and sensitivity analyses, combined with pre-  
370 viously published empirical data [15], suggest that the *Sarracenia* microe-  
371 cosystem could be employed as a powerful system with which to develop  
372 new understanding of the dynamics of complex ecosystems. The food web of  
373 *S. purpurea* consists of species that share an evolutionary history, multiple  
374 trophic levels, and interactions that have been shaped by both environmental  
375 and co-evolutionary forces [21, 48]. Its abiotic environment and food web are  
376 comparable in complexity to large lakes [18, 49, 50]. It features similar criti-  
377 cal transitions and non-linear dynamical behavior that are of broad interest  
378 for theoretical ecologists.

379 Mesocosm studies have been critiqued for lacking any or all of these char-  
380 acteristics [10], but a recent meta-analysis of the scaling relationships of

381 the half-saturation constant ( $K_w$ ) provides evidence that uptake of nutrients  
382 such as nitrogen and phosphorus by food webs, and inter-trophic nutrient  
383 transfers, all are nearly invariant to spatial scale [32]. At the same time, the  
384 dynamics of the *Sarracenia* microecosystem play out over days, rather than  
385 years, decades, centuries, or even longer. Thus we conclude that, similar to  
386 previous work that has demonstrated the ability to scale up ecosystems pro-  
387 cesses (e.g., [51]), the pitcher-plant microecosystem provides an experimental  
388 system and computational model with which to study the linkages between  
389 “green”, and “brown” food webs [17, 52, 53] in the context of a food-web  
390 with an evolutionary history. Therefore, this system provides a powerful tool  
391 for identifying early warning signals of state changes in ecosystems that are  
392 of crucial importance for environmental management [54, 55].

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- 398 [1] M. Scheffer, S. Carpenter, J. A. Foley, C. Folke, B. Walker, Catastrophic  
399 shifts in ecosystems, *Nature* 413 (2001) 591–596.
- 400 [2] M. Scheffer, J. Bascompte, W. A. Brock, V. Brovkin, S. R. Carpenter,  
401 V. Dakos, H. Held, E. H. van Nes, M. Rietkerk, G. Sugihara, Early-  
402 warning signals for critical transitions, *Nature* 461 (2009) 53–59.  
403 URL <http://dx.doi.org/10.1038/nature08227>
- 404 [3] P. S. Petraitis, S. R. Dudgeon, Cusps and butterflies: multiple stable  
405 states in marine systems as catastrophes, *Marine and Freshwater Re-*  
406 *search* 67 (1) (2016) 37. doi:10.1071/MF14229.
- 407 [4] A. D. Barnosky, E. A. Hadly, J. Bascompte, E. L. Berlow, J. H. Brown,  
408 M. Fortelius, W. M. Getz, J. Harte, A. Hastings, P. A. Marquet, N. D.  
409 Martinez, A. Mooers, P. Roopnarine, G. Vermeij, J. W. Williams,  
410 R. Gillespie, J. Kitzes, C. Marshall, N. Matzke, D. P. Mindell, E. Re-  
411 villa, A. B. Smith, Approaching a state shift in Earth’s biosphere., *Na-*  
412 *ture* 486 (7401) (2012) 52–8. doi:10.1038/nature11018.  
413 URL <http://dx.doi.org/10.1038/nature11018>
- 414 [5] B. W. Brook, E. C. Ellis, M. P. Perring, A. W. Mackay,  
415 L. Blomqvist, Does the terrestrial biosphere have planetary tip-  
416 ping points?, *Trends in Ecology & Evolution* 28 (7) (2013) 396–401.  
417 doi:<http://dx.doi.org/10.1016/j.tree.2013.01.016>.  
418 URL <http://www.sciencedirect.com/science/article/pii/S0169534713000335>

- 419 [6] H. F. Wilson, J. E. Saiers, P. A. Raymond, W. V. Sobczak, Hydrologic  
420 drivers and seasonality of dissolved organic carbon concentration, nitro-  
421 gen content, bioavailability, and export in a forested new england stream,  
422 *Ecosystems* 16 (4) (2013) 604–616. doi:10.1007/s10021-013-9635-6.  
423 URL <http://dx.doi.org/10.1007/s10021-013-9635-6>
- 424 [7] B. T. Bestelmeyer, A. M. Ellison, W. R. Fraser, K. B. Gorman, S. J.  
425 Holbrook, C. M. Laney, M. D. Ohman, D. P. C. Peters, F. C. Pilsbury,  
426 A. Rassweiler, R. J. Schmitt, S. Sharma, Detecting and managing abrupt  
427 transitions in ecological systems, *Ecosphere* 2 (2011) 129.
- 428 [8] S. R. Carpenter, W. A. Brock, Rising variance: a leading indicator of  
429 ecological transition, *Ecology Letters* 9 (2006) 311–318.
- 430 [9] S. R. Carpenter, J. J. Cole, M. L. Pace, R. Batt, W. A. Brock, T. Cline,  
431 J. Coloso, J. R. Hodgson, J. F. Kitchell, D. A. Seekell, L. Smith, B. Wei-  
432 del, Early warnings of regime shifts: a whole-ecosystem experiment, *Sci-  
433 ence* 332 (6033) (2011) 1079–1082.  
434 URL <http://dx.doi.org/10.1126/science.1203672>
- 435 [10] S. R. Carpenter, Microcosm Experiments have Limited Relevance for  
436 Community and Ecosystem Ecology, *Ecology* 77 (3) (1996) 677–680.  
437 doi:10.2307/2265490.  
438 URL <http://doi.wiley.com/10.2307/2265490>
- 439 [11] G. G. Mittlebach, A. M. Turner, D. J. Hall, J. E. Rettig, C. W. Osenberg,

- 440 Perturbation and resilience: a long-term, whole-lake study of predator  
441 extinction and reintroduction, *Ecology* 76 (1995) 2347–2360.
- 442 [12] R. Contamin, A. M. Ellison, Indicators of regime shifts in ecological  
443 systems: what do we need to know and when do we need to know it?,  
444 *Ecological Applications* 19 (2009) 799–816.
- 445 [13] R. Biggs, S. R. Carpenter, W. A. Brock, Turning back from the brink:  
446 detecting an impending regime shift in time to avert it, *Proceedings of*  
447 *the National Academy of Sciences, USA* 106 (2009) 826–831.
- 448 [14] V. Dakos, S. R. Carpenter, W. A. Brock, A. M. Ellison, V. Guttal,  
449 A. R. Ives, S. Kéfi, V. Livina, D. A. Seekell, E. H. van Nes, M. Scheffer,  
450 Methods for detecting early warnings of critical transitions in time series  
451 illustrated using simulated ecological data., *PLOS ONE* 7 (7) (2012)  
452 e41010. doi:10.1371/journal.pone.0041010.  
453 URL [/pmcc/articles/PMC3398887/?report=abstract](https://pmcc/articles/PMC3398887/?report=abstract)
- 454 [15] J. Sirota, B. Baiser, N. J. Gotelli, A. M. Ellison, Organic-matter loading  
455 determines regime shifts and alternative states in an aquatic ecosystem,  
456 *Proceedings of the National Academy of Sciences* 110 (19) (2013) 7742–  
457 7747. doi:10.1073/pnas.1221037110.  
458 URL <http://www.pnas.org/cgi/doi/10.1073/pnas.1221037110>
- 459 [16] H. T. Odum, Scales of ecological engineering, *Ecol. Eng. Eng.* 6 (7)

460 (1996) 7–19.

461 URL <http://ecotechnics.edu/wp-content/uploads/2011/08/0dum-scales-ecol-eng.pdf>

462 [17] J. L. Butler, N. J. Gotelli, A. M. Ellison, Linking the brown and green:  
463 nutrient transformation and fate in the *Sarracenia* microecosystem,  
464 Ecology 89 (2008) 898–904.

465 [18] D. S. Srivastava, J. Kolasa, J. Bengtsson, A. Gonzalez, S. P. Lawler,  
466 T. E. Miller, P. Munguia, T. Romanuk, D. C. Schneider, M. K. Trzcinski,  
467 Are natural microcosms useful model systems for ecology?, Trends  
468 Ecol.Evol 19 (7) (2004) 379–384. doi:10.1016/j.tree.2004.04.010.  
469 URL <http://www.ncbi.nlm.nih.gov/pubmed/16701289>

470 [19] D. E. Schnell, Carnivorous plants of the United States and Canada,  
471 Vol. 2, Timber Press, Portland, Oregon, USA, 2002.

472 [20] A. Arber, On the morphology of the pitcher-leaves in *Heliamphora*, *Sar-*  
473 *racenia*, *Darlingtonia*, *Cephalotus*, and *Nepenthes*, Annals of Botany 5  
474 (1941) 563–578.

475 [21] A. M. Ellison, N. J. Gotelli, Energetics and the evolution of carnivorous  
476 plants - Darwin's "most wonderful plants in the world", Journal of  
477 Experimental Botany 60 (2009) 19–42.

478 [22] J. L. Butler, D. Z. Atwater, A. M. Ellison, Red-spotted newts: an un-  
479 usual nutrient source for northern pitcher plants, Northeastern Natural-  
480 ist 12 (2005) 1–10.



- 481 [23] F. M. Jones, The most wonderful plant in the world, *Natural History* 23  
482 (1923) 589–596.
- 483 [24] J. F. Addicott, Predation and prey community structure: an experiment  
484 study of the effect of mosquito larvae on the protozoan communities of  
485 pitcher plants, *Ecology* 55 (1974) 475–492.
- 486 [25] S. B. Heard, Pitcher plant midges and mosquitoes: a processing chain  
487 commensalism, *Ecology* 75 (1994) 1647–1660.
- 488 [26] D. L. Cochran-Stafira, C. N. von Ende, Integrating bacteria into food  
489 webs: studies with *Sarracenia purpurea* inquilines, *Ecology* 79 (1998)  
490 880–898.
- 491 [27] P. Boynton, Ecological Patterns and Processes in *Sarracenia* Carnivo-  
492 rous Pitcher Plant Fungi, Ph.D. thesis, Harvard University (10 2012).  
493 URL <https://dash.harvard.edu/handle/1/9824173>
- 494 [28] C. N. Peterson, S. Day, B. E. Wolfe, A. M. Ellison, A. M. Kolter,  
495 A. Pringle, A keystone predator controls bacterial diversity in the pitcher  
496 plant (*Sarracenia purpurea*) microecosystem, *Environmental Microbiol-*  
497 *ogy* 10 (2008) 2257–2266.
- 498 [29] D. R. Gallie, S. C. Chang, Signal transduction in the carnivorous plant  
499 *Sarracenia purpurea* - Regulation of secretory hydrolase expression dur-  
500 ing development and in response to resources, *Plant Physiology* 115 (4)  
501 (1997) 1461–1471.

- 502 [30] J. L. Butler, A. M. Ellison, Nitrogen cycling dynamics in the carnivorous  
503 pitcher plant, *Sarracenia purpurea*, *Functional Ecology* 21 (2007) 835–  
504 843.
- 505 [31] W. E. Bradshaw, R. A. Creelman, Mutualism between the carnivorous  
506 purple pitcher plant *Sarracenia purpurea* and its inhabitants, *American*  
507 *Midland Naturalist* 112 (1984) 294–304.
- 508 [32] C. Mulder, A. J. Hendriks, Half-saturation constants in func-  
509 tional responses, *Global Ecology and Conservation* 2 (2014) 161–169.  
510 doi:10.1016/j.gecco.2014.09.006.  
511 URL [www.elsevier.com/locate/gecco](http://www.elsevier.com/locate/gecco)
- 512 [33] A. C. Northrop, R. K. Brooks, A. M. Ellison, N. J. Gotelli, B. A. Bal-  
513 lif, Environmental proteomics reveals taxonomic and functional changes  
514 in an enriched aquatic ecosystem, *Ecosphere* 8 (10) (2017) e01954.  
515 doi:10.1002/ecs2.1954.  
516 URL <http://doi.wiley.com/10.1002/ecs2.1954>
- 517 [34] T. J. Givnish, E. L. Burkhardt, R. E. Happel, J. D. Weintraub, Car-  
518 nivory in the bromeliad *Brocchinia reducta*, with a cost/ benefit model  
519 for the general restriction of carnivorous plants to sunny, moist nutrient-  
520 poor habitats, *American Naturalist* 124 (1984) 479–497.
- 521 [35] A. M. Ellison, Nutrient limitation and stoichiometry of carnivorous  
522 plants, *Plant Biology* 8 (2006) 740–747.

- 523 [36] C. Darwin, *Insectivorous Plants*, John Murray, London, UK, 1875.
- 524 [37] R Core Team, *R: A Language and Environment for Statistical Comput-*  
525 *ing*, Vienna, Austria (2016).  
526 URL <https://www.r-project.org/>
- 527 [38] J. G. Kingsolver, Thermal and Hydric Aspects of Environmental Het-  
528 erogeneity in the Pitcher Plant Mosquito, *Ecological Monographs* 49 (4)  
529 (1979) 357–376. doi:10.2307/1942468.  
530 URL <http://doi.wiley.com/10.2307/1942468>
- 531 [39] R. Vaquer-Sunyer, C. M. Duarte, Thresholds of hypoxia for marine  
532 biodiversity., *Proceedings of the National Academy of Sciences, USA*  
533 105 (40) (2008) 15452–7. doi:10.1073/pnas.0803833105.  
534 URL <http://www.ncbi.nlm.nih.gov/pubmed/18824689>  
535 <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=PMC2556360>
- 536 [40] R. E. Ulanowicz, *Growth and Development: Ecosystems Phenomenol-*  
537 *ogy*, Springer Science & Business Media, 2012.
- 538 [41] N. M. Levine, K. Zhang, M. Longo, A. Baccini, O. L. Phillips,  
539 S. L. Lewis, E. Alvarez-Dávila, A. C. Segalin de Andrade, R. J. W.  
540 Brienen, T. L. Erwin, T. R. Feldpausch, A. L. Monteagudo Mendoza,  
541 P. Nuñez Vargas, A. Prieto, J. E. Silva-Espejo, Y. Malhi, P. R. Moor-  
542 croft, Ecosystem heterogeneity determines the ecological resilience of  
543 the Amazon to climate change, *Proceedings of the National Academy of*

- 544 Sciences, USA 113 (3) (2016) 793–797. doi:10.1073/pnas.1511344112.  
545 URL <http://www.pnas.org/lookup/doi/10.1073/pnas.1511344112>
- 546 [42] K. S. McCann, Food Webs, Princeton University Press, 2012.  
547 URL <http://www.jstor.org/stable/j.ctt7rr0s>
- 548 [43] B. Baiser, R. S. Ardeshiri, A. M. Ellison, Species richness and trophic  
549 diversity increase decomposition in a co-evolved food web, PLOS ONE  
550 6 (5) (2011) e20672. doi:10.1371/journal.pone.0020672.  
551 URL <http://dx.plos.org/10.1371/journal.pone.0020672>
- 552 [44] M. Lau, S. Borrett, D. Hines, P. Singh, enaR: Tools for Ecological Net-  
553 work Analysis, r package version 2.10.0 (2017).  
554 URL <https://CRAN.R-project.org/package=enaR>
- 555 [45] K. A. Wood, M. T. O’Hare, C. McDonald, K. R. Searle,  
556 F. Daunt, R. A. Stillman, Herbivore regulation of plant abundance  
557 in aquatic ecosystems, Biological Reviews 92 (2) (2017) 1128–1141.  
558 doi:10.1111/brv.12272.  
559 URL <http://dx.doi.org/10.1111/brv.12272>
- 560 [46] S. R. Carpenter, D. L. Christensen, J. J. Cole, K. L. Cottingham, X. He,  
561 J. R. Hodgson, J. F. Kitchell, S. E. Knight, M. L. Pace, . et al., Bio-  
562 logical Control of Eutrophication in Lakes, Environmental Science &  
563 Technology 29 (3) (1995) 784–786. doi:10.1021/es00003a028.  
564 URL <http://pubs.acs.org/doi/abs/10.1021/es00003a028>

- 565 [47] M. F. Chislock, E. Doster, R. A. Zitomer, A. E. Wilson, Eutrophication:  
566 Causes, Consequences, and Controls in Aquatic Ecosystems., *Nature*  
567 *Education Knowledge* 4 (4) (2013) 10.
- 568 [48] L. S. Bittleston, N. E. Pierce, A. M. Ellison, A. Pringle, Convergence in  
569 Multispecies Interactions, *Trends in Ecology & Evolution* 31 (4) (2016)  
570 269–280. doi:10.1016/j.tree.2016.01.006.
- 571 [49] R. L. Kitching, *Food webs and container habitats : the natural history*  
572 *and ecology of phytotelmata*, Cambridge University Press Cambridge,  
573 2000.
- 574 [50] B. Baiser, R. Elhesha, T. Kahveci, Motifs in the assembly of food web  
575 networks, *Oikos* 125 (4) (2016) 480–491. doi:10.1111/oik.02532.  
576 URL <http://dx.doi.org/10.1111/oik.02532>
- 577 [51] K. A. Capps, R. Rancatti, N. Tomczyk, T. B. Parr, A. J. K. Calhoun,  
578 M. Hunter, Biogeochemical hotspots in forested landscapes: The role  
579 of vernal pools in denitrification and organic matter processing, *Ecosys-*  
580 *tems* 17 (8) (2014) 1455–1468. doi:10.1007/s10021-014-9807-z.  
581 URL <http://dx.doi.org/10.1007/s10021-014-9807-z>
- 582 [52] E. M. Wolkovich, S. Allesina, K. L. Cottingham, J. C. Moore, S. A.  
583 Sandin, C. de Mazancourt, Linking the green and brown worlds: the  
584 prevalence and effect of multichannel feeding in food webs, *Ecology*

- 585 95 (12) (2014) 3376–3386. doi:10.1890/13-1721.1.  
586 URL <http://doi.wiley.com/10.1890/13-1721.1>
- 587 [53] M. I. Sitvarin, A. L. Rypstra, J. D. Harwood, Linking the green and  
588 brown worlds through nonconsumptive predator effects, *Oikos* 125 (8)  
589 (2016) 1057–1068. doi:10.1111/oik.03190.  
590 URL <http://doi.wiley.com/10.1111/oik.03190>
- 591 [54] M. J. Abbott, L. L. Battaglia, Purple pitcher plant (*sarracenia rosea*)  
592 dieback and partial community disassembly following experimental  
593 storm surge in a coastal pitcher plant bog, *PLOS ONE* 10 (4) (2015)  
594 1–12. doi:10.1371/journal.pone.0125475.  
595 URL <https://doi.org/10.1371/journal.pone.0125475>
- 596 [55] D. Hoekman, Turning up the heat: Temperature influences the relative  
597 importance of top-down and bottom-up effects, *Ecology* 91 (10) (2010)  
598 2819–2825. doi:10.1890/10-0260.1.

599 **6. Tables**

600 Table 1: Terms, units, and interpretation for the model of oxygen dy-  
601 namics in pitcher plant fluid.

Term	Units	Interpretation
$t$	minutes	Time (model iteration)
$f$	$1/t$	Constant adjusting sine wave of diurnal PAR for frequency of measurements
$x_t$	$\text{mg L}^{-1}$	$[\text{O}_2]$ : concentration of oxygen in the pitcher fluid
<b>Photosynthesis</b>		
$A_t$	$\text{mg L}^{-1}$	Production of oxygen by photosynthesis, and infused from the plant into the fluid during the day
$a_t$	$\text{mg L}^{-1}$	Photosynthetic rate augmentation by microbial nutrient mineralization
$a_{min}, a_{max}$	$\text{mg L}^{-1}$	Minimum or maximum possible photosynthetic augmentation
PAR	$\mu\text{mol} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$	Photosynthetically active radiation
<b>Respiration</b>		
$w_t$	mg	Mass of prey remaining at time $t$
$W$	mg	Maximum mass of decomposable prey (set at $75\mu\text{g}$ )
$K_w$	$\text{mg min}^{-1}$	Half-saturation constant for bacterial carrying capacity
$m$	$\text{mg L}^{-1}$	Basal metabolic oxygen used by bacteria (respiration)
<b>Nutrients</b>		
$n_t$	$\text{mg L}^{-1}$	Quantity of nutrients mineralized by decomposition; a function of $w_t$ , and $x_t$
$s$	dimensionless	Sigmoidal curve steepness relating nutrient mineralization to augmentation
$d$	mg	Inflection point of sigmoidal curve relating mineralization to augmentation
$\beta$	dimensionless	The rate of prey decomposition



## 602 7. Figure legends

Figure 1: The threshold dynamics of the Hill function are determined in part by the inflection parameter  $h$ . A) Plotted output of the Hill function for different values of  $h$  (different lines shaded darker for lower values), ranging from 0.1 to 150 with  $p = 10$ . B) Lagged ( $k = 1$  lag term) phase plot of the Hill function with  $h = 71.11$ , showing the state transition (lower-left to upper-right). A small amount of random variation was introduced to the series to reveal overlapping points within the two states.

Figure 2: A) The pitcher plant model shown as a network diagram. The nodes in the graph and their corresponding variables in the model are Prey ( $w$ ), the microbially-dominated food web (controlled by  $K_w$ ), Nitrogen ( $n$ ), Oxygen ( $x$ ) and the pitcher plant itself, which is included to show the fluxes of nitrogen and oxygen as they relate to the plant. B) Photosynthetically active radiation (PAR) from the sun or other light source modeled as a sine wave. Although negative values of PAR are set equal to zero in our model's photosynthesis function, we show here the full sine-wave to illustrate the mathematical function from which PAR is derived. C) The relationship between PAR and photosynthesis in the pitcher plant. D) The output of dissolved oxygen in the pitcher fluid as a function of pitcher-plant photosynthesis. E) The decomposition of prey over time. F) The impact of prey addition ( $t = 2160$  min) on dissolved oxygen in the pitcher-plant fluid.

Figure 3: In model simulations, addition of prey impacts both photosynthetic oxygen production via augmentation from nutrients mineralized from prey and oxygen depletion through the biological oxygen demand (BOD) of microbial metabolism. (A) shows how photosynthesis increases when prey is added (grey) on days 4–6 ( $t = 5040$  to  $t = 7920$  minutes; indicated by open circles), relative to when no prey was added (black). (B) The quantity of oxygen used via the BOD of microbial decomposition. The net impact in this parameterization was a decrease in dissolved oxygen when prey was added to the system; (C) shows oxygen present in the pitcher at mid-day. (D) A time-lagged phase plot ( $t_0 = 720$ , lag = 1440 min) showing the change in oxygen production during the prey addition simulation. Beginning and end points of the simulation are indicated by closed circles. When prey was added at  $t = 5040$ ,  $t = 6480$ , and  $t = 7920$  (open circles), it was decomposed rapidly by the microbially-dominated food web, resulting in oxygen depletion. The altered return trajectory (i.e., hysteresis) resulting from the biological oxygen demand in the system is shown by the arrows indicating the direction of the change in oxygen through time.

Figure 4: Sensitivity analysis of the pitcher-plant model revealed non-linear effects of varying the parameters  $\beta$  and  $K_w$  ( $n = 15000$  simulations). Contour plots show the percent time the system spent in an hypoxic state (top row), and the Pearson correlation coefficient for the decycled trend (bottom row). The sensitivity simulations were repeated for additions of prey corresponding to  $1 \text{ mg mL}^{-1}$  (left column), and  $100 \text{ mg mL}^{-1}$  (right column) of prey added to the microecosystem. The LSE estimate for  $\beta$  ( $\pm 1 \text{ SE}$ ) is plotted in each contour plot as vertical solid and dashed lines, respectively.

Figure 5: Oxygen dynamics in three simulations using different levels of bacterial carrying capacity ( $K_w$ ; light-grey = 0.1, dark-grey = 0.5, and black = 1) with the same rate of decomposition ( $\beta = 2.0\text{E-}6$ ) illustrating hysteresis (i.e., altered return trajectory) of the system. (A) Lower levels of  $K_w$  produce slower return rates over the course of the simulation. Prey addition (open circles) depressed mid-day oxygen curves at lower values of  $K_w$ . Closed circles indicate the first and last mid-day prey addition points. (B) A time-lagged ( $t = 1440$  minutes) phase plot for the same simulations showing that lower values of  $K_w$  led to the oxygen being at lower levels for more time following prey addition (open circles), but followed a similar return trajectory as prey was decomposed by the food web (closed circles also indicate the beginning and end of each series). Although all three simulations ran for the same amount of time, the lengths of the trajectories are different in phase space because lower values of  $K_w$  resulted in the system spending more time with the same amount of oxygen (i.e.,  $x_t = x_{t+1440}$ ).

Figure 1

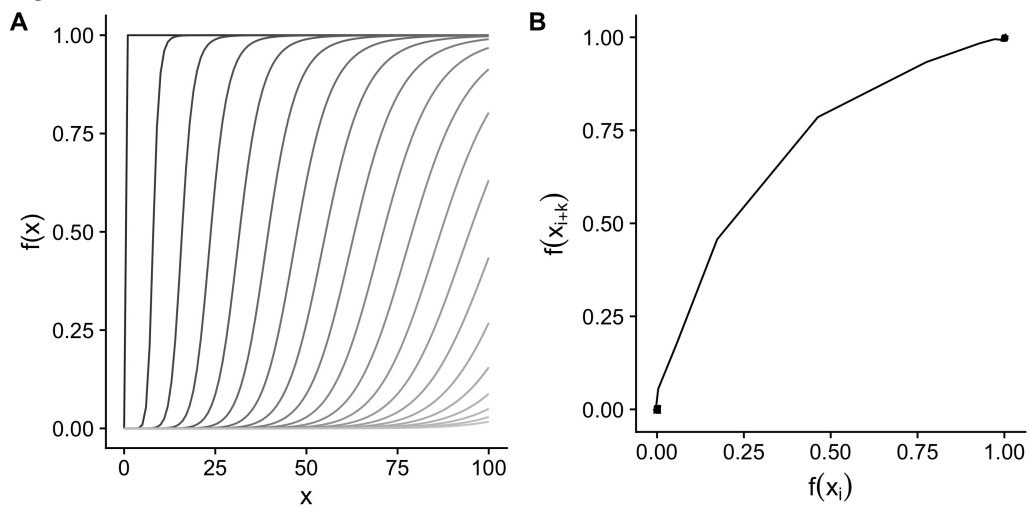


Figure 2

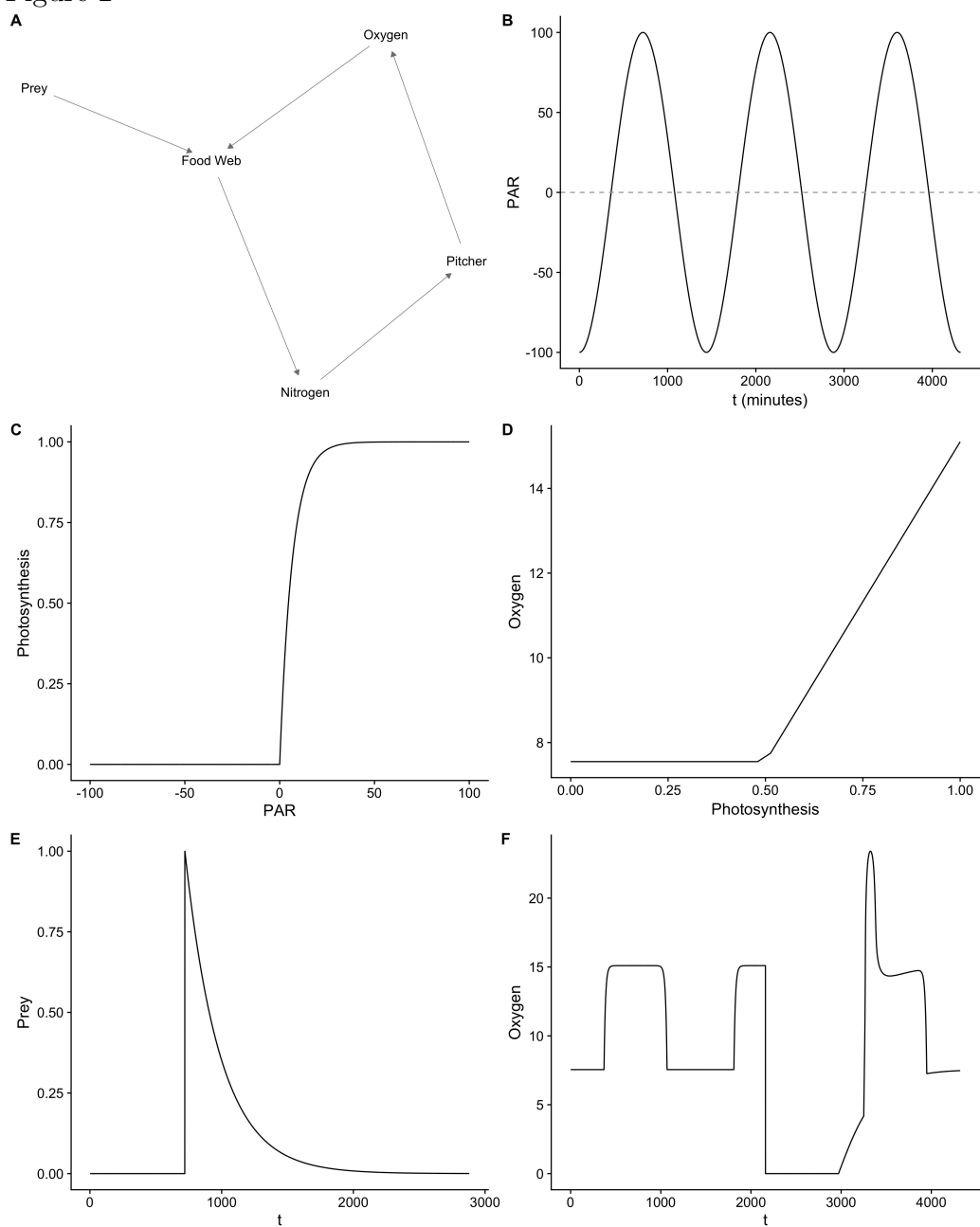


Figure 3

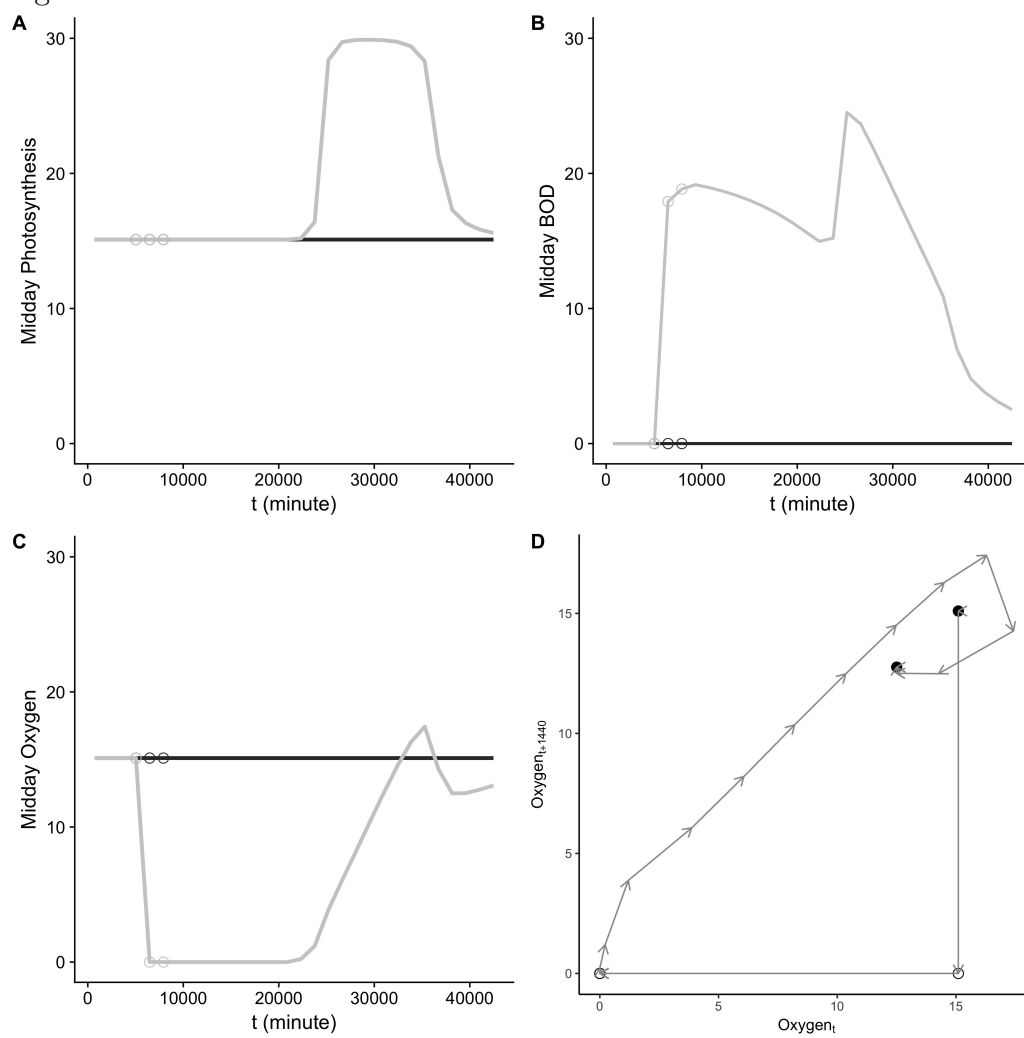


Figure 4

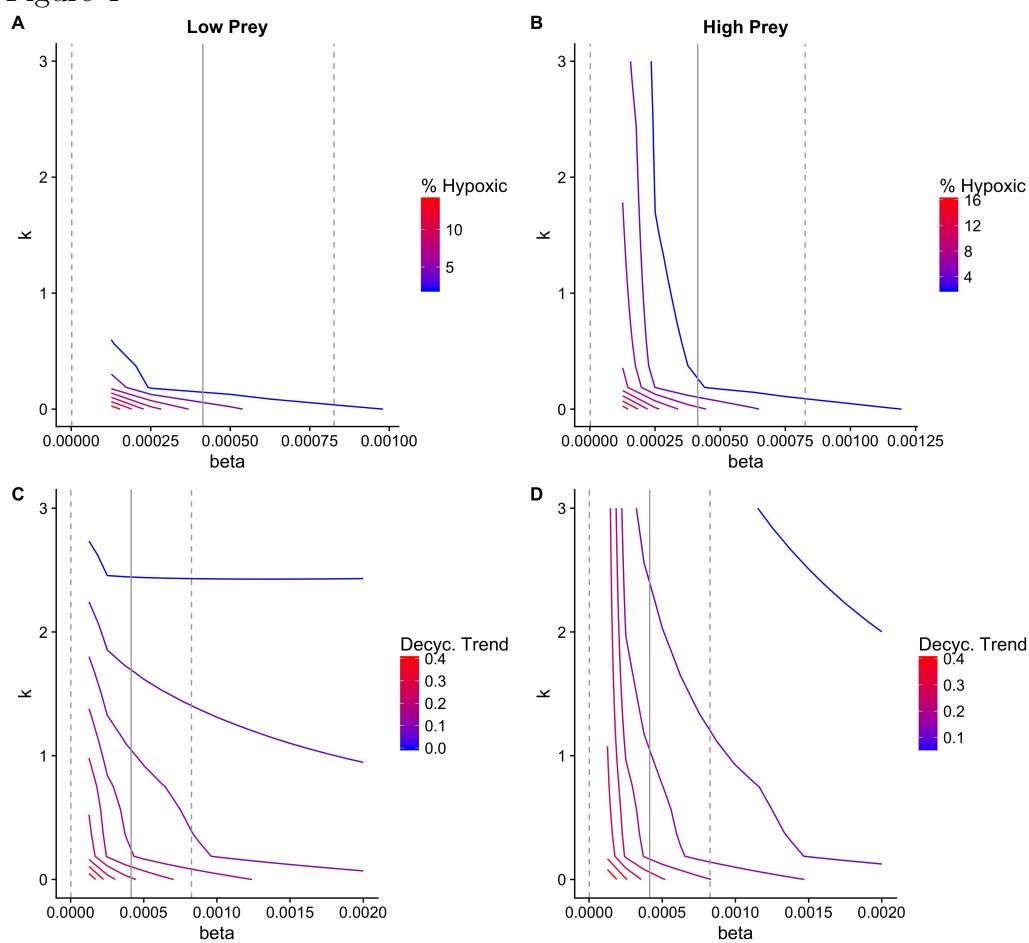


Figure 5

