# Understanding How Microbiomes Influence the Systems they Inhabit: Insight from Ecosystem Ecology

**Authors:** Hall, E.K.<sup>1\*</sup>, Bernhardt, E.S<sup>2</sup>, Bier, R.L.<sup>2</sup>, Bradford, M.A.<sup>3</sup>, Boot, C.M.<sup>1</sup>, Cotner, J.B.<sup>4</sup>, del Giorgio, P.A.<sup>5</sup>, Evans, S.E.<sup>6</sup>, Graham E. B <sup>2,7.8</sup>, Jones, S.E.<sup>9</sup>, Lennon, J.T.<sup>10</sup>, Locey, K.<sup>10</sup> Nemergut, D.<sup>2</sup>, Osborne, B.<sup>11</sup>, Rocca, J.D.<sup>1,2</sup>, Schimel J.S.<sup>12</sup>, Wallenstein, M.W.<sup>1</sup>

Article Type: Perspectives Words in Abstract: 272 Words in Main Text: 4,431 2 Figures, 0 Tables, 51 References

Running Title: How microbiomes influence ecosystems Keywords: microbiome, microbial ecology, ecosystem ecology, emergent properties, community aggregated traits

Affiliations: <sup>1</sup>Colorado State University, Fort Collins CO (ed.hall@colostate.edu, claudia.boot@colostate.edu, matt.wallenstein@colostate.edu) <sup>2</sup>Duke University, Durham, N.C. (ebernhar@duke.edu, ebgraham2@colorado.edu, diana.nemergut@duke.edu) <sup>3</sup>Yale University, New Haven, CT (mark.bradford.yale@gmail.com) <sup>4</sup>University of Minnesota, Saint Paul, MN (cotne002@umn.edu) <sup>5</sup>Université du Québec à Montréal, Montréal, CA (del\_giorgio.paul@uqam.ca) <sup>6</sup>Michigan State University, Hickory Corners, MI (evanssar@gmail.com) <sup>7</sup>Institute of Arctic and Alpine Research, University of Colorado at Boulder <sup>8</sup> Pacific Northwest National Laboratory, Richland, WA, USA <sup>9</sup>Notre Dame University, South Bend, IN (sjones20@nd.edu) <sup>10</sup>Indiana University, Bloomington, IN (lennonj@indiana.edu, ken@weecology.org) <sup>11</sup>Brown University, Providence, Rhode Island (brooke\_osborne@brown.edu) <sup>12</sup>University of California, Santa Barbara, Santa Barbara, CA (josh.schimel@lifesci.ucsb.edu)

\* corresponding author: Dr. Ed Hall, Natural Resource Ecology Laboratory, Department of Ecosystem Science and Sustainability, Campus Delivery 1499, Colorado State University, Fort Collins, CO 80523, ed.hall@colostate.edu, 970-491-2162

**Contribution:** All listed authors have contributed to the conceptualization, writing, and preparation of the current manuscript.

#### **Abstract**

1

2

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

24

The well-documented significance of microorganisms to the function of virtually all ecosystems has led to the assumption that more information on microbiomes will improve our ability to understand and predict system-level processes. Notably, the importance of the microbiome has become increasingly evident in the environmental sciences and in particular ecosystem ecology. However, translating the everincreasing wealth of information on environmental microbiomes to advance ecosystem science is proving exceptionally challenging. One reason for this challenge is that correlations between microbiomes and the ecosystem processes they influence are often reported without the underlying causal mechanisms. This limits the predictive power of each correlation to the time and place at which it was identified. In this paper, we assess the assumptions and approaches currently used to establish links between environmental microbiomes and the ecosystems they influence, propose a framework to more effectively harness our understanding of microbiomes to advance ecosystem science, and identify key challenges and solutions required to apply the proposed framework. Specifically, we suggest identifying each microbial process that contributes to the ecosystem process of interest a priori. We then suggest linking information on microbial community membership through microbial community properties (such as biomass elemental ratios) to the microbial processes that drive each ecosystem process (e.g. N mineralization). A key challenge in this framework will be identifying which microbial community properties can be determined from the constituents of the community (community aggregated traits, CATs) and which properties are unable to be predicted from a list of their constituent taxa (emergent properties, EPs). We view

- 25 this directed approach as a promising pathway to advance our understanding of how
- 26 microbiomes influence the systems they inhabit.

## Current approaches in linking microbial characteristics and ecosystem processes

27 Recently there has been a broad call, including the National Microbiome Initiative 28 led by the executive branch of the United States Federal Government, for a coordinated effort to evaluate the role of microorganisms in all environments<sup>1,2</sup>. 29 30 Coordinating efforts to explore microbiomes and their functioning across such a 31 broad range of systems is exciting and ambitious and holds the potential to transform 32 societies approach many of the most important challenges we currently face<sup>3</sup>. 33 However, advances in this direction require an assessment of our progress to date 34 and an attempt to identify the most promising paths forward. In most ecosystems, many processes are carried out primarily by microorganisms, 35 36 and virtually all processes are influenced by microorganisms. Therefore it is common 37 to apply high-resolution analytical techniques to better describe microbial 38 communities, assuming that greater resolution of the community, (including its 39 associated transcripts, proteins, and metabolic products) should lead to better 40 predictions of ecosystem processes. However, such justifications assume that 41 microbial metrics (e.g. 16S rRNA gene libraries, metagenomes, enzymatic activities) 42 will improve our ability to understand, model, and predict ecosystem processes. This 43 assumption may not necessarily be valid, because microbial information may not 44 provide additional explanatory power for understanding ecosystem process rates beyond what can be predicted by environmental factors alone<sup>4,5</sup>. In addition, when 45 46 correlations between microbial information and an ecosystem process are identified 47 the underlying causal associations may remain difficult to resolve and may not be 48 causal at all—e.g. the microbiome and the ecosystem may be responding in concert

50

51

52

53

54

55

56

57

58

59

60

61

62

63

64

65

66

67

68

69

70

71

72

to a common underlying driver; limiting the predictive ability of each result across additional systems. Two recent meta-analyses suggest that the current research approach at the intersection of ecosystem and microbial ecology has the potential to be better focused to more effectively achieve the anticipated insights into how microbiomes influence ecosystems<sup>6,7</sup>. The first meta-analysis evaluated studies that related the relative abundance of protein encoding gene copy or transcript abundance to associated biogeochemical processes<sup>6</sup>. Of 416 identified studies that attempted to address the correlation between the relative abundance of a protein-encoding gene (or transcript) and an ecosystem process, only 56 measured both genes or transcript copy number and the corresponding process. Within these 56 studies, 14% of the observations showed a significant negative correlation between gene copy number and process rate, 38% had a significant positive relationship, and 48% had no significant relationship. Thus, the effect size for the relationship between gene copy number and process rate had an approximately normal distribution with a mean near zero<sup>6</sup>. The second meta-analysis evaluated links between microbial community composition and ecosystem processes in response to an experimentally-induced disturbance in the environment (e.g. drought, warming, nutrient addition)<sup>7</sup>. Whereas 40% of published papers reported concomitant changes in microbial community structure and ecosystem function, only about a third of those cases (only 12% of total studies) attempted to identify a statistical relationship between community composition and an ecosystem process. Interestingly, many of the studies that did not measure both community composition and a corresponding ecosystem process still framed their study in the microbial structure-ecosystem function framework.

These meta-analyses illustrate that links between microbial characteristics and ecosystem processes are often assumed to be present but are rarely tested. When linkages are explicitly tested, connections between microbial structure and ecosystem processes are more often than not weak or non-existent<sup>6</sup>. These findings suggest that our current approach to linking microbial metrics to ecosystem function should be refocused with more attention paid to empirically identifying explicit linkages between microbial characteristics and the ecosystem processes that they influence.

#### Linking microbial characteristics and ecosystem processes

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

93

94

95

96

A key challenge in linking microbial information to an ecosystem process is that conceptual research frameworks often do not align directly with the available methods or the data they generate. For example, environmental factors act on the physiology of individual organisms, which alters their competitive ability, relative abundance, collective physiology, and ultimately their contribution to ecosystem processes (Figure 1a). However, designing an observational study or experiment from this conceptual framework (Figure 1a) assumes that environmental metrics can be empirically linked to measureable microbial characteristics across multiple categories of ecological organization (i.e. individuals, populations, and communities) at the appropriate temporal and spatial scales. Yet in nature, relationships between environmental variables and microbial characteristics are dynamic and non-linear<sup>8</sup>, simultaneously affected by a plethora of biotic and abiotic variables<sup>9,10</sup> and decoupled in both time and space<sup>7</sup>. Each of these aspects of microbial-environment interactions obscures the relationships among microbial characteristics collected at each level of ecological organization and the ecosystem processes they affect (Figure 2b). To address this challenge we propose a framework that explicitly

98

99

100

101

102

103

104

105

106

107

108

109

110

111

112

113

114

115

116

117

118

119

120

121

identifies the ecosystem process of interest and how it relates to microbial characteristics. The proposed framework illustrates the relationship among different categories of microbial characteristics and conceptually defines their contribution to ecosystem processes. *Identifying the Ecosystem Process* The first step to understand how microorganisms influence an ecosystem process is to define each of its sub-processes, the set of constituent reactions that combine to dominate the net flux of the ecosystem process of interest. Ecosystem processes are defined as a change in a pool size or a flux from one pool to another (e.g. NH<sub>4</sub><sup>+</sup> to NO<sub>3</sub><sup>-</sup> or dissolved organic matter mineralization to CO<sub>2</sub>). Few, if any of these processes are carried out by a single physiological pathway or a single organism. Rather, ecosystem processes are aggregate processes consisting of complementary or antagonistic sub-processes carried out by a breadth of phylogenetically diverse microorganisms<sup>11</sup>. For example, net ecosystem productivity (NEP) is the balance between C-fixation and Cmineralization. Each sub-process of NEP can be further partitioned into a series of metabolic pathways (e.g. heterotrophic fermentation and aerobic respiration or chemoautotrophic nitrification and photoautotrophic C-fixation). Partitioning each ecosystem process in this hierarchical manner can continue until the sub-process maps directly to specific microbial metabolic pathways (e.g. acetoclastic methanogenesis). Subsequently each of these metabolic pathways can be categorized as either broad or narrow<sup>12</sup>. Broad processes are phylogenetically common (i.e. widely distributed among taxa), whereas narrow processes are phylogenetically conserved (i.e. limited to a specific subset of taxa). For example, denitrification is broad, while both methanogenesis and methanotrophy are narrow (with some notable exceptions<sup>13</sup>).

123

124

125

126

127

128

129

130

131

132

133

134

135

136

137

138

139

140

141

142

143

144

145

146

The second step is to identify the controls or constraints on each constituent subprocess. For example, kinetics of a given metabolic pathway in a model organism may help understand the rate limiting steps of a narrow process, but insights from model organisms are much less likely to be useful for a broad process where phenotypic variation among phylogenetically diverse organisms should be much greater. Defining the ecosystem process, its critical sub-processes, and the known phylogenetic distribution of the metabolic pathways that drive those sub-processes in this manner creates an explicit conceptual pathway that directly links the ecosystem process to the microorganisms that influence them. Once this conceptual pathway has been identified a concerted empirical approach can be applied to investigate how the microbiome influences the ecosystem process of interest. Understanding the relationship between categories of microbial characteristics Categories of microbial characteristics At present, researchers are measuring a wide variety of characteristics of microbial communities (e.g. sequence or relative abundance of genes, transcripts or proteins, enzyme expression, and process rates). Much of that work does not clearly articulate how these measurements differ in their specificity, precision, or linkage among each other or how they inform the microbial contribution to ecosystem processes. We propose that by categorizing microbial characteristics into three distinct categories, 1) microbial processes, 2) microbial community properties, and 3) microbial community membership (Figure 2), we can frame how different metrics interact with each other and how they can elucidate the microbial contribution to an ecosystem process. Microbial Processes Microbial processes are the collective physiology of the microbial community that drives changes in pools and fluxes of important elements

148

149

150

151

152

153

154

155

156

157

158

159

160

161

162

163

164

165

166

167

168

169

170

171

or compounds<sup>14</sup> (Figure 2). They are the level of microbial information that can most readily be incorporated into ecosystem-level models because many of these processes are the sub-processes that contribute to an overall pool or flux. Examples include nitrogen fixation, denitrification, nitrification, phosphorus uptake and immobilization, primary production, respiration, and carbon use efficiency. The rates of many microbial processes can be estimated through physiological assays, and while they do not open the "black box" of the microbial community, they do directly quantify the microbial contribution to the transformation of substrates moving through the box. However, physiological assays to estimate microbial processes are often logistically challenging, require experimental manipulations that inevitably deviate from the *in situ* conditions, and often depend on the environment in which they were measured. For example, the relationship between microbial process rates and temperature vary across geographical temperature gradients<sup>15</sup> (enzyme activity) and among seasons<sup>16</sup> (phosphorus use efficiency, PUE). Thus observations of the effect of temperature on either enzyme activity or PUE are time and place dependent. Therefore, without an underlying physiological mechanism, to accurately quantify the microbial process the relationship between temperature and community physiology must be measured through a direct assay at each location and at each time. Because of these limitations, a microbial community property that can be measured in situ or collected and preserved in the field for later analysis in the laboratory has several advantages over using bioassays to directly measure microbial processes. A community property may include characteristics of community biomass such as elemental ratios (biomass C:N or C:P ratios) that estimate potential to mineralize or immobilize limiting nutrients, or the relative abundance of genes that encode for an

173

174

175

176

177

178

179

180

181

182

183

184

185

186

187

188

189

190

191

192

193

194

195

known physiology or physiological response (such as the relative abundance of cold shock proteins to estimate cold tolerance)<sup>17</sup>. We refer to these *in situ* measurements that allow estimation of microbial processes as microbial community properties (hereafter community properties, Figure 2). Community properties represent an integrated characteristic of the extant microbial community that has the potential to estimate the microbial process of interest. Microbial Community Properties Microbial community properties can be separated into two categories, emergent properties (EPs) such as biofilm thickness, which cannot be determined from the properties of their constituent populations<sup>18</sup>, and community aggregated traits (CATs) such as nitrification potential, which can be estimated from community membership or at least characteristics such as relative gene abundance (e.g. AmoA), of those members<sup>17</sup>. The potential importance of EPs to influence ecosystem processes was demonstrated in series of experiments conducted in flow-through flumes that mimicked development and metabolism of stream biofilms<sup>19</sup>. Both transient storage (i.e. an increase in residence time of the water and its solutes near the biofilm relative to the flow around it) and the biofilm community's ability to use arabinose relative to glucose increased as the microbial biofilm density increased and porosity decreased. Microbial biofilm thickness and density are both EPs that affected the important ecosystem processes of hydrological transient storage and substrate use specificity<sup>19</sup>. In this case biofilm thickness was affected by physical factors (i.e. flow) but biofilm thickness may also be influenced by other environmental characteristics such as P availability<sup>20</sup>.

197

198

199

200

201

202

203

204

205

206

207

208

209

210

211

212

213

214

215

216

217

218

219

220

While EPs are powerful metrics for understanding ecosystem processes (Figure 2, Letter F or H) they cannot, by definition, be estimated from a list of constituent taxa or characteristics of those taxa (Figure 2, Letter E) and thus must remain as an intermediary between environmental drivers such as flow or P availability (Figure 2, Letter C) and ecosystem processes. Unlike EP's CATs can be estimated from characteristics of their constituents and provide one pathway to link microbial community membership to the community properties that drive ecosystem processes<sup>17</sup> (Figure 2, Letter E and F). Microbial community biomass stoichiometry (e.g. biomass C:N or C:P) is one example of a putative CAT that has been shown to be a useful predictor of nutrient immobilization or mineralization during litter decomposition<sup>21</sup>, and can predict both respiration and N-mineralization better than microbial biomass alone<sup>22</sup>. In a study of soil microcosms amended with organic carbon and reactive N, the relationship between the resource C:N and microbial biomass C:N was able to better predict whether C would be respired or immobilized relative to microbial biomass alone<sup>22</sup>. Biomass stoichiometry has been shown to differ among phylogenetically different organisms. For example, at the broadest level microbial biomass C:N differs between fungi and bacteria<sup>24</sup>, and has been shown to be variable among a wide range of taxa grown on the same media<sup>25,26</sup>. Thus community biomass stoichiometry has the potential to be empirically deconstructed into the biomass stoichiometry of its constituent taxa<sup>27</sup>, linking community membership and a community property (e.g. biomass C:N) with the power to estimate an important microbial process (e.g. N-mineralization). Microbial Community Membership Although analysis of community membership by sequencing phylogenetic marker genes provides organism identity, the direct coupling of microbial phylogeny to its

222

223

224

225

226

227

228

229

230

231

232

233

234

235

236

237

238

239

240

241

242

243

244

245

physiology and ecology is often weak<sup>28</sup> (Figure 2, Letter G). For example, most organic carbon molecules can be metabolized by a phylogenetically diverse suite of organisms and denitrification is also a phylogenetically broad process. The result is that with the exception of some specialists (e.g. nitrifiers), we can infer very little about the function of microbial communities from a list of their constituent taxa. Whereas it is clear that microbial populations are non-randomly distributed in space and time<sup>29</sup> and some microbial traits appear to be conserved at coarse taxonomic scales 30,31,32 the underlying physiological mechanisms for phylogenetic sorting across environmental gradients is often unknown. This prevents an explicit link between the organism's relative abundance and their role in the collective community physiology that influences ecosystems. The paucity of associated physiological data that accompanies phylogenetic sequence data in most studies limits the ecological insight from phylogenetic analyses and constrains our ability to attribute microbial processes to community membership even of well-defined consortia. In addition to the paucity of ecological inference provided by an analysis of community membership, community analysis using current methods has two exceptional challenges that need to be addressed in order to gain insight from community membership to drivers of ecosystem processes. First, bulk extraction of DNA from environmental samples, often the first step in analysis of microbial communities, may or may not represent the extant and active microbial community at the time of sampling. In microbial ecology, unlike in plant or animal ecology, the number, biomass, and identity of different populations cannot be assessed with confidence<sup>33,34</sup>. When DNA is extracted from the environment the presence or relative abundance of a given sequence is not necessarily proportionate to the

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

262

263

264

265

266

267

268

269

270

absolute abundance or biomass of that organism within the community. One reason for this is that the extracted nucleic acid may not have been contained within a microbial cell at the time of extraction. This idea has been well established<sup>35</sup> and the proportion of extracellular DNA is known to vary among ecosystems<sup>36</sup> (e.g. there may be more DNA in sediment than the water column). It is clear that a portion of sequences derived from any environmental sample are not from intact cells<sup>37,38,39</sup>. Even when specific phylotypes can be empirically linked to intact cells (e.g. using in situ hybridization), the viability and metabolism of that cell typically remains unknown. Because of this, the presence of a sequence within a "community" does not indicate the associated organism is participating in the microbial process of interest, and if active, it does not indicate that that the organism's contribution is proportionate to the relative abundance of its DNA sequence. While these facts are readily acknowledged, they can be very challenging to address, and therefore are often overlooked or ignored as a problematic but non-addressable constraint, something of an inconvenient truth of microbial ecology. The second major challenge in linking relative abundance of microbial populations with microbial processes is that the diversity, growth rate, and metabolic complexity of environmental microorganisms are orders of magnitude greater than all other organisms. The result is that when scaling from individuals through populations to microbial communities, the obfuscating factors described in Figure 1b introduce and ultimately accumulate more uncertainty at higher levels of ecological organization. relative to the same analyses applied to macroorganism-ecosystem linkages. This uncertainty is further confounded because community composition and processes are routinely measured at different spatial scales<sup>40</sup>. Translating microbial measurements to an ecosystem flux typically requires linking measurements from

272

273

274

275

276

277

278

279

280

281

282

283

284

285

286

287

288

289

290

291

292

293

294

295

microbial physiology (10<sup>-12</sup> m), to microbial process measurements (10<sup>-1</sup> or 10<sup>-2</sup> m), to ecosystem process measurements (m to 10<sup>3</sup> m). This enormous scale (15 orders of magnitude) over which to interpolate data, raises challenges analogous to mechanistically analyzing the global carbon cycle, i.e. linking experiments on grams of carbon in soils (e.g. g C per g dry weight) to global fluxes of Petagrams (10<sup>15</sup> g). However, unlike global C cycle research, microbial-to-ecosystem research often addresses these in a descriptive manner using correlative approaches, without the rigor and quantitative modeling approaches typically applied to global biogeochemical cycles<sup>41</sup>. Each of these challenges must be effectively addressed in order to rigorously incorporate the growing wealth of information on microbiomes to system level processes. A Way Forward at the Intersection of Microbial and Ecosystem Science The conceptual diagram presented here (Figure 2) provides a road map for organizing and linking the diverse suite of microbial characteristics that are commonly measured. Ecosystem ecology has traditionally been confined to the interactions depicted within the horizontal arrow, moving from environmental parameters to ecosystem processes (Letter A, Figure 2). As the role of the microbiome has come to the forefront of environmental sciences it is clear that microbial ecology has a great deal to contribute. However, microbial ecology has traditionally been confined to interactions depicted within the vertical arrow, moving from individuals (microbial community membership) to process rates (microbial processes) of populations or more recently communities (Figure 2). The excitement to integrate microbial metrics into ecosystem science has led to a range of novel approaches of linking characteristics on the microbiome to ecosystem processes. Direct connections between microbial community membership and ecosystem

297

298

299

300

301

302

303

304

305

306

307

308

309

310

311

312

313

314

315

316

317

318

319

processes (Figure 2, Letter J) or community properties and ecosystem processes (Figure 2, Letter I) are almost exclusively correlative in nature. Whereas many of these results are intriguing, the relationships discovered across these pathways, in the absence of the defining physiological mechanism, are often restricted to the time and place they are identified. Because of this, moving from correlative, phenomenological approaches to a causative and mechanistic understanding is a challenging but necessary step for microbiome science. Because microbial processes can be estimated by community properties (e.g. N mineralization and biomass C:N), understanding the drivers of community properties is a way to more explicitly link environmental microorganisms with the ecosystem processes they control. We propose that identifying which community properties best describe microbial processes (Figure 2, Letter F), then identifying whether or not the community properties that best describe each process are a CAT (Letter E, Figure 2) or an EP (Figure 2 Letter C) provides clear pathway to understand whether environmental drivers or microbial drivers dominate ecosystem processes. Currently, many microbial community properties are EPs (i.e. cannot be predicted from their constituent members or their characteristics). Understanding when community properties can be predicted by membership is an important and open research direction. Distinguishing which community properties represent aggregated traits (CATs) and which are actually EPs may be an essential link in advancing our ability to apply microbial information to ecosystem science. This is not a trivial task. however a suite of existing methods already provides the ability to directly pursue this challenge. Applying the proposed framework

321

322

323

324

325

326

327

328

329

330

331

332

333

334

335

336

337

338

339

340

341

342

343

344

Understanding the principal drivers of community properties will require a series of complementary approaches applied in concert, including; stable isotope probing of cultured isolates and mixed communities, single cell methods that can assay cells in the same physiological state they occur in in the environment, sorting of complex communities into subsets of populations or consortia for further investigation, and physiological assays of isolates grown in culture. Studies that use either labeled substrates or single cell techniques (or both) have been successful in linking community composition and process rates. One example, stable isotope probing (SIP, in which an isotopically labeled element (or elements) from a defined substrate can be tracked into microbial biomass) is a method used to identify which organisms may be participating in an ecosystem process of interest<sup>42</sup>. For example, a study of sulfate reduction in a Scottish peatland revealed that a single species of Desulfosporosinus was most likely responsible for the totality of sulfate reduction within the peatland even though it only comprised 0.0006% of the retrieved sequences<sup>43</sup>. In this case the *Desulfosporosinus* species represented the only known sulfate reducer within the community and thus the kinetics of this organism seemingly defined the kinetics of sulfate reduction in the entire system. Whereas this is a single example of using confirmative ecophysiology to link categories of microbial information (pathway G in Figure 2), there is a suite of culture-free techniques (such as Raman microspectroscopy (MS), NanoSIMS, or X-ray microanalysis, XRMA) that complement sequence-based microbiome analysis by reporting on the physiological and compositional characteristics of individual cells in situ<sup>27, 44, 45</sup>. All three methods (Raman MS, NanoSIMS and XRMA) can be coupled with phylogentic labels (in situ hybridization) that can visualize identification of phylotypes simultaneously with macromolecular (Raman MS), isotopic (NanoSIMS),

346

347

348

349

350

351

352

353

354

355

356

357

358

359

360

361

362

363

364

365

366

367

368

369

and elemental composition (XRMA). In addition, both Raman MS and NanoSIMS can trace substrates that are labeled with stable isotopes into cells, providing the ability to identify populations that are participating in specific metabolic pathways within complex communities. These powerful approaches applied in concert with sequence analysis have the potential to empirically link the categories of microbial information defined here (Figure 2). These methods also have the greatest potential to begin to unravel, which community properties are EPs, and which are CATs. In addition to direct visualization of individuals cells from mixed populations or consortia using single cell approaches, there are abundant examples of immunocapture (e.g. bromodeosyuridine, BrdU)<sup>46</sup> or other labeling and cell sorting approaches (eg. fluorescent in situ hybridization(FISH) coupled with flow cytometry cell sorting, FACS)<sup>47</sup> that provide powerful tools to constrain the complexity of microbial communities and link community membership to microbial characteristics that influence ecosystem processes. Labeling and sorting techniques allow the cells that incorporate a labeled substrate or can be targeted with a stain or fluorescent reporter to be separated from the broader community and then assayed for membership or for biomass composition. For example, a study of a North Atlantic bacterial community labeled the actively growing component of the community using BrdU and then separated those populations from the rest of the community using an immune capture technique<sup>46</sup>. Similarly, cells can be labeled with phylogenetic probes (e.g. FISH) that fluorescent at different wavelengths and can be separated from the general community using a flow cytometry to only select those cells that reported for the phylogenetic label<sup>47</sup>. In addition to these approaches for assaying natural communities, there is a need for broader representation of cultured taxa that better represent physiologies that are

371

372

373

374

375

376

377

378

379

380

381

382

383

384

385

386

387

388

389

390

391

392

393

394

similar to those phyla found in the environment. For example, microbial community carbon use efficiency (the amount of carbon allocated to biomass production relative to carbon respired or stored) is a central parameter in many ecosystem level carbon cycling models. Differences in CUE among phyla depend largely on the relative plasticity of growth rate, carbon storage capacity, and maintenance respiration. E. coli, the "poster bacterium" for physiological assays is unique both in the plasticity of its growth rate and its capacity to store energy as organic carbon and thus not necessarily informative to develop a better understanding of CUE<sup>48</sup>. Streptococcus Bovis has proven to as a more appropriate organism to study energy cycling in bacteria<sup>48</sup>. Physiological studies of isolates from a broader distribution of representative phyla are key to advancing our understanding of environmental microbiomes. However, it is unlikely that information about specific phenotypes estimated in isolation from pure-culture studies can be directly used to estimate community properties because of the plasticity of organismal physiology and because of the prevalence of competitive interactions when isolates are grown in coculture with even one other organism<sup>49</sup>. Therefore studies of isolates grown in culture would provide more powerful information if they reported the plasticity of a given phenotype, rather than only the phenotype under a single set of environmental conditions. For example, a recent study of 24 freshwater bacterial isolates showed differences in phenotypic plasticity in the biomass stoichiometry of the taxa studied<sup>26</sup>. Some of the taxa had highly variable biomass C:P ratios whereas others demonstrated a high degree of homeostasis in their C:P ratios. This allowed the authors to hypothesize how communities or consortia with homeostatic phenotypes would respond to environmental drivers compared to communities or consortia composed of populations with more plastic phenotypes.

396

397

398

399

400

401

402

403

404

405

406

407

408

409

410

411

412

413

414

415

416

417

418

419

Whereas linking microbial membership to system level processes is an exceptional challenge the tools and approaches to address this challenge already exist. These existing and established microbiological methods both relatively novel (NanoSIMS) and foundational in the field (physiological assays of isolates) can be applied in concert to begin to parse the exceptional complexity of environmental microbiomes. Designing microbiome research to maximize insights into system-level processes The meta-analyses discussed above clearly illustrate that a more directed approach to microbiome research is necessary. We suggest that rather than looking for linkages among microbial community membership and system-level processes in every study, research efforts would benefit from strategically targeting the linkages and processes for which an a priori understanding of microbial physiology should allow us to improve our understanding of the ecosystem process. These cases may be identified first by noting patterns in which environmental factors explain little of the variability in an ecosystem process, or where system-level responses deviate significantly from rates that are predicted from environmental factors alone. These deviations could include spatial or temporal heterogeneity in an ecosystem response where the environmental characteristics do not have the same level of heterogeneity. For example, the discovery of novel microbial metabolic pathways (i.e. annamox) has helped explain otherwise puzzling chemical transformations, such as the oxidation of ammonium under anoxic conditions<sup>50</sup>. Such advances are most likely to be cases where the microbial process of interest can be directly linked to a phylogenetically constrained group (i.e. a narrow process) and where the systemlevel behavior of the process reflects the organismal-level physiological controls, such as N-sensitivity of methane monooxygenase<sup>51</sup>.

421

422

423

424

425

426

427

428

429

430

431

432

433

434

435

436

437

438

439

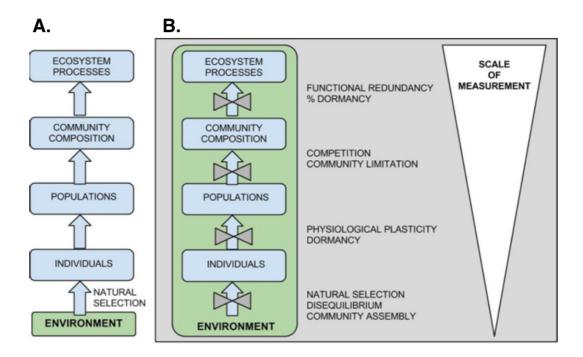
440

441

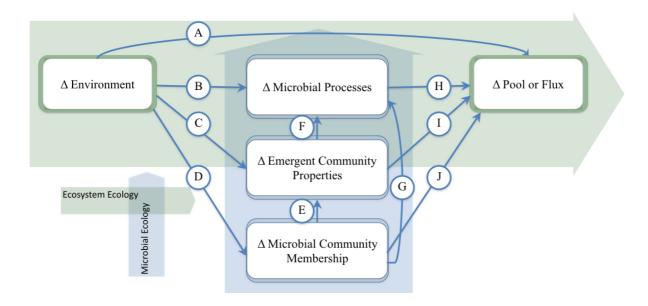
442

The framework presented here provides one approach to formalize inquiry across microbiome science and encourage empirical linkages between the presence of organisms in a system and the processes that characterize that system. Whereas we draw examples from environmental microbiomes and the ecosystems they inhabit, the framework presented here should also benefit the analysis of microbiomes associated with other systems such as host organisms and those of engineered environments. We assert that this framework provides an important and straightforward starting point as the global research community aims to undergo one of the most exciting concerted efforts in the microbial sciences to date. **Acknowledgements** This work is a product of the Next Generation of Ecosystem Indicators Working Group, supported by the USGS John Wesley Powell Center for Synthesis and Analysis. Preparation of this manuscript was supported by NSF DEB IOS #1456959 awarded to EKH. Chuck Pepe-Ranney provided valuable feedback on a previous version of this manuscript.

### **Figures**



**Figure 1** Diagram of microbial-ecosystem linkages A) how linkages are commonly conceptualized across levels of ecological organization and B) the series of environmental filters that create challenges when attempting to link metrics from one level of ecological organization to the other.



**Figure 2** Reframing how we study microbial-ecosystem linkages. Shown is the intersection between microbial (vertical) and ecosystem (horizontal) ecology with each of the three categories of microbial information (microbial processes, emergent community properties, and microbial community membership) as defined in the text. We argue for an increased focus on studies that elucidate pathways E, F and H. In addition we note that pathways G, J and I are less likely to effectively incorporate microbial information into ecosystem science. The delta symbol in each category indicates an emphasis on how changes within a category may lead to a change in a connected category.

#### References

- Alivisatos A. P., M. J. Blaser, E. L. Brodie, M. Chun, J. L. Dangl, T. J. Donohue, P. C. Dorrestein, J. A Gilbert, J. L. Green, J. K. Jansson, R. Knight, M. E. Maxon, M. J. McFall-Ngai, J. F. Miller, K. S. Pollard, E. G. Ruby, S. A. Taha, Unified Microbiome Initiative Consortium 2015: A unified initiative to harness Earth's microbiomes. Science 350, 503-504. DOI: 0.1126/science.aac84809. Angilletta, M. J. Jr. (2009). Thermal Adaptation. A Theoretical and Empirical Synthesis. Oxford: Oxford University Press, New York, New York
- 2. Dubilier N., McFall-Mgai, M. and Zhao L. (2015): Create a global microbiome effort. Nature 529 (631-634).DOI: 10.1038/526631a
- 3. Blaser, M.J., Cardon, Z. G., Cho, M.K., Dangl, J.L, Donohue, T.J., Green, J.L., Knight, R., Maxon, M.E., Northen, T.R., Pollard, K.S. and Brodie, E.L. 2016 Toward a Predictive Understanding of Earth's Microbiomes to Address 21st Century Challenges 7(3) e00714-16
- 4. Graham E. B., Wieder, W.R., Leff, J.R., Weintraub, S. W., Townsend, A.R., Cleveland, C.C., Philippot L., Nemergut D.R. 2014 Do we need to understand microbial communities to predict ecosystem function? A comparison of statistical models of nitrogen cycling processes Soil Biology and Biochemistry 68, 279–282
- Graham E. B. et al. 2016 Microbes as engines of ecosystem function: when does community structure enhance predictions of ecosystem processes? Frontiers in Microbiology 7: doi: DOI=10.3389/fmicb.2016.0021
- 6. Rocca J.D., Hall E.K., Lennon J.T., Evans S.E., Waldrop M.P., Cotner J.B., Nemergut D.R., Graham E.B., Wallenstein M.D. (2015). ISME J. 9, 1693–1699
- 7. Bier R.L., E.S. Bernhardt, CM Boot, EB Graham, EK Hall, JT Lennon, D Nemergut, B Osborne, C Ruiz-González, JP Schimel, MP Waldrop, MD. Wallenstein How are we forging conceptual, analytical, and mechanistic links between microbial community structure and ecosystem process? (in review at FEMS Microbiol.)
- 8. Felip M., Pace M.L., Cole, J.J. (1996). Regulation of planktonic bacterial growth rates: The effects of temperature and resources. Microb Ecol. 31(1):15-28.
- 9. Hochachka P.W., Somero G.N. (2002). Biochemical adaptation: Mechanism and process in physiological evolution. Oxford University Press, New York
- 10. Angilletta, M. J. Jr. (2009). Thermal Adaptation. A Theoretical and Empirical Synthesis. Oxford: Oxford University Press, New York, New York
- 11. Schimel, J.P., J. Bennett, and N. Fierer. 2005. Microbial community composition and soil N cycling: is there really a connection? In: Biological diversity and function in soils. Bardgett, R.D., D.W. Hopkins, and M.B. Usher (Eds.) Cambridge University Press. pp. 171-188
- 12. Schimel, J.P. (1995). Ecosystem consequences of microbial diversity and community structure in Arctic and Alpine Biodiversity: Patterns, Causes and Ecosystem Consequences Ecological Studies Volume 113, pp 239-254
- 13. Lenhart K., M. Bunge, S. Ratering, T. R. Neu, I. Schüttmann, M. Greule, C. Kammann, S. Schnell, C. Müller, H. Zorn & F. Keppler. (2012). Evidence for methane production by saprotrophic fungi Nature Comm. 3:1046
- 14. Wallenstein, M. D., and E. K. Hall. (2012). A trait-based framework for predicting when and where microbial adaptation to climate change will affect ecosystem functioning. Biogeochem.: 109: 1-3, 35-47
- 15. German D.P., Marcelo, K.R.B., Stone, M.M and Allison, S.D. (2012). The michaelis-menten kinetics of soil extracellular enzymes in response to temperature: a cross-latitudinal study Global Change Biology 18,1468-1479

- 16. Hall, E. K., A. R. Dzialowski, S. M. Stoxen, and J. B. Cotner. (2009). The effect of temperature on the coupling between phosphorus and growth in lacustrine bacterioplankton communities. Limnol. and Oceanogr. 54:880-889.
- 17. Fierer, N., A. Barberán, D. Laughlin. (2014). Seeing the forest for the genes: Using metagenomics to infer the aggregated traits of microbial communities. Frontiers in Microbiology. 5:614
- 18. Konopka, A. (2009) What is microbial community ecology ISME J., 3:11, 1223-12230
- 19. Battin, T., L. A. Kaplan, L. Newbold, X. Cheng, and C. Hansen. (2003). Effects of current velocity on the nascent architecture of stream microbial biofilms. Appl. and Env. Microbiol. 69:5443-5452.
- 20.Pepe-Ranney, C.P. and Hall, E.K. 2015 Carbon subsidies affect planktonic niche partitioning and recruitment of bacteria to marine biofilms http://dx.doi.org/10.3389/fmicb.2015.00703 Frontiers in Aquatic Microbiology
- 21. Manzoni, S., R. B. Jackson, J. A. Trofymow, and A. Porporato. (2008). The global stoichiometry of litter nitrogen mineralization. Science 321:684-686.
- 22. Buchkowski, R.W., Schmitz, O.J., Bradford, M.A. (2015) Microbial stoichiometry overrides biomass as a regulator of soil carbon and nitrogen cycling. Ecology, 96, 1139-1149.
- Strickland M.S., and Rousk J. 2010 Considering fungal:bacterial dominance in soils- methods, controls and ecosystem implications Soil Biology and Biochemistry 42: 1385-1396
- 24. Keiblinger, K.M.† Hall, E.K†, Szukics, U., Hämmerle, I., Ellersdorfer, G., Sterflinger, K., Wanek, W., Richter, A., Jandl R., and Zechmeister -Boltenstern, S. 2010 The effect of resource quantity and resource stoichiometry on microbial carbon use efficiency FEMS Microbiology Ecology 73(3): 430-440
- 25. Mouginot, C., Kawamura, R., Matulich, K.L., Berlemont, R., Allison, S.D., Amend, A.S. and Martiny, A.C. (2014) Elemental stoichiometry of fungi and bacteria strains from grassland leaf litter Soil Biol. & Biochem. 76, 278-285
- 26. Godwin, C.M. and J.B. Cotner. 2015 Stoichiometric Flexibility in Diverse Aquatic Heterotrophic Bacteria Is Coupled to Differences in Cellular Phosphorus Quotas. Name: Frontiers in Microbiology 6 (2015): doi:10.3389/fmicb
- 27. Hall, E.K., Maixner F., Franklin O., Daims, H., Richter, A. and Battin, T. (2011) Linking microbial and ecosystem ecology using ecological stoichiometry: A synthesis of conceptual and empirical approaches Ecosystems 14, 261-273
- 28. Newton, R. J, S. E. Jones, A. Eiler, K. D McMahon, and S. Bertilsson. (2011). A guide to the natural history of freshwater lake bacteria. Microbiol. and Molec. Biol. Rev. 75:1, doi:10.1128/MMBR.00028
- 29. Martiny, J.B.H., B. Bohannan, J. Brown, R. Colwell, J. Fuhrman, J. Green, M.C. Horner-Devine, M. Kane, J. Krumins, C. Kuske, P. Morin, S. Naeem, L. Ovreas, A.-L. Reysenbach, V. Smith, J. Staley. 2006. "Microbial biogeography: Putting microorganisms on the map." Nature Reviews Microbiology 4: 102-112
- 30. Martiny JBH, Jones SE, Lennon JT, Martiny AC (2015) Microbiomes in light of traits: a phylogenetic perspective. Science 350: aac9323 (pdf)
- 31. Lennon JT, Aanderud ZA, Lehmkuhl BK, Schoolmaster DR (2012) Mapping the niche space of soil microorganisms using taxonomy and traits. Ecology 93: 1867-1879 (pdf)
- 32. Treseder KK, Lennon JT (2015) Fungal traits that drive ecosystem dynamics. Microbiology and Molecular Biology Reviews 79: 243-262 (pdf)

- 33. Tilman D, PB Reich, J Knops, D Wedin, T Mielke, C Lehman. (2001). Diversity and productivity in a long-term grassland experiment Science, 294:5543, 843-845
- 34. Reich P. B., David Tilman, Forest Isbell, Kevin Mueller, Sarah E. Hobbie, Dan F. B. Flynn, and Nico Eisenhauer. (2012). Impacts of biodiversity loss escalate through time as redundancy rades Science: 336:6081, 589-592.
- 35. Klein, D. A., 2011 Bulk Extraction-Based Microbial Ecology: Three Critical Questions Microbe Sept
- 36. Corinaldesi C., R. Danovaro, and A. Dell'Anno (2005) Simultaneous Recovery of Extracellular and Intracellular DNA Suitable for Molecular Studies from Marine Sediments Appl. and Environ. Microbiol., 71:1, 46–50
- 37. Dell'Anno A. and Roberto Danovaro (2005) Extracellular DNA Plays a Key Role in Deep-Sea Ecosystem Functioning Science 309 (5744), 2179
- 38. Torti, A., Leverb, M.A., Jørgensen, B.B. 2015 Origin, dynamics, and implications of extracellular DNA pools in marine sediments Marine Genomics doi:10.1016/j.margen.2015.08.007
- 39. Carini P., Patrick J. Marsden, Jonathan W. Leff, Emily E. Morgan, Michael S.Strickland, Noah Fierer 2016 Relic DNA is abundant in soil and obscures estimates of soil microbial diversity http://dx.doi.org/10.1101/043372
- 40. Clark, J. S., Bell, D. M., Hersh, M. H., Kwit, M. C., Moran, E., Salk, C., Stine, A., Valle, D. and Zhu, K. (2011) Individual-scale variation, species-scale differences: inference needed to understand diversity. Ecol. Lett., 14, 1273–1287
- 41. Thornton P.E., Doney S.C., Linsay K., Moore J.K., Mahowald N., Randerson J.T., Fung I., Lamarque J.F., Feddema J.J., Lee, Y.H. (2009) Carbon-nitrogen interactions regulate climate-carbon cycle feedbacks: results from an atmosphere-ocean general circulation model Biogeosciences, 6, 2099-2120
- 42. Neufeld J.D., Vohra J., Dumont M.G., Lueders T., Manefield M., Friedrich M.W., Murrell J.C. 2007 DNA stable-isotope probing. Nat Protoc. 2:4, 860-866
- 43. Pester, M., Bittner, N., Deevong, P., Wagner, M., and Loy, A. (2010). A 'rare biosphere' microorganism drives sulfate reduction in a peatland. ISME J. 4:1591–1602.
- 44. Wagner, M. (2009). Single cell ecophysiology of microbes as revealed by raman microspectroscopy or secondary ion mass spectrometry imaging. Annu Rev Microbiol 63:411-429.
- 45. Behrens, S., Kappler, A., Obst M. (2012). Linking environmental processes to the in situ functioning of microorganisms by high-resolution secondary ion mass spectrometry (NanoSIMS) and scanning transmission X-ray microscopy (STXM) Env. Microbiol. 14:11, 2851-69
- 46. Galand P.E., L. Alonso-Sáez, S. Bertilsson, C. Lovejoy, and E. O. Casamayor (2013) Contrasting activity patterns determined by BrdU incorporation in bacterial ribotypes from the Arctic Ocean in winter Front Microbiol. 2013; 4: 118
- 47. Czechowska, K., Johnson, D.R., van der Meer, J.R. (2008) Use of flow cytometric methods for single-cell analysis in environmental microbiology Curr. Opin. In Microbiol. 11:3, 205-212
- 48. Russell, J.B. (2007). The energy spilling reactions of bacteria and other organisms. Journal of Molecular Microbiology and Biotechnology 13, 1-11
- 49. Foster, K.R. and Bell, T. 2012 Competition, not cooperation, dominates interations among culturable microbial species Curr. Biol. 22, 1845-1850

- 50. Mulder A., A. A. Van De Graaf, L. A. Robertson, and J. G. Kuenen. (1995). Anaerobic ammonium oxidation discovered in a denitrifying fluidized bed reactor, FEMS Microbiology Ecology, 16:3, 177–184,
- 51. Paul LE Bodelier, Interactions between nitrogenous fertilizers and methane cycling in wetland and upland soils, Current Opinion in Environmental Sustainability, Volume 3, Issue 5, October 2011, Pages 379-388, ISSN 1877-3435, http://dx.doi.org/10.1016/j.cosust.2011.06.002.